1 Process-based modeling framework for sustainable irrigation

2 management at the regional scale: Integrating rice production, water

3 use, and greenhouse gas emissions

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16 Abstract

Rice cultivation faces multiple challenges of rising food demand while increasing water 17 scarcity and greenhouse gas emissions, intensifying the tension of the food-water-18 climate nexus. Process-based modeling is pivotal for developing effective measures to 19 balance these challenges. However, current models struggle to simulate their complex 20 21 relationships under different water management schemes, primarily due to inadequate representation of critical physiological effects and lack of efficient spatially explicit 22 modeling strategies. Here, we propose an advancing framework that addresses these 23 problems by integrating a process-based soil-crop model with vital physiological 24 effects, a novel method for model upscaling, and the NSGA-II multi-objective 25 optimization algorithm at a parallel computing platform. Applying the framework 26 27 accounted for 52%, 60%, 37%, and 94% of the experimentally observed variations in rice yield, irrigation water use, methane and nitrous oxide emissions in response to 28 irrigation schemes. Compared with the origin model using traditional parameter 29 upscaling methods, the advancing framework significantly reduced simulation errors 30 by 35%-85%. Moreover, it well reproduced the multivariable synergies and tradeoffs 31 observed in China's rice fields and identified additional 18% areas feasible for 32 33 irrigation optimization, along with an additional 11% and 14% reduction potentials of water use and methane emissions, without compromising production. Over 90% of the 34 potentials could be realized at the cost of 4% less yield increase and 25% higher nitrous 35 oxide emissions under multiple objectives. Overall, this study provides a valuable tool 36 for multi-objective optimization of rice irrigation schemes at a large scale. The 37 advancing framework also has implications for other process-based modelling 38 39 improvements efforts.

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41 Key points

- This study significantly improved rice yield simulations under various irrigation
 schemes by incorporating critical physiological processes into a process-based
 model.
- This study developed a novel upscaling method of model parameterization that
 well reproduced observed synergies and tradeoffs among multiple objectives (i.e.,
 rice yield, irrigation water use, methane emissions, and nitrous oxide emissions).
- This study provides a practical tool for multi-objective optimization of water
 management to deliver co-benefits of ensuring food production, saving water, and
 reducing greenhouse gas emissions of rice fields.

52 1 Introduction

Rice is the staple food for more than half of the world's population and is also the most 54 55 water-intensive cereal crop with a significant contribution to greenhouse gas emissions 56 (GHGs) (Lampayan et al., 2015; Carlson et al., 2017). Rice cultivation currently accounts for 40% of global irrigation water use (IRR), 30% of methane (CH₄), and 11% 57 of nitrous oxide (N₂O) emissions in agriculture (Yuan et al., 2021). To meet the demand 58 of the growing population, a 50-60% increase in global rice production along with a 59 15% increase in water use are required by 2050, potentially leading to higher 60 greenhouse gas emissions and intensifying the food-water-climate tensions of rice 61 62 fields (Flörke et al., 2018). Therefore, ensuring food security while conserving water resources and reducing GHGs in rice cultivation is essential for achieving multiple 63 United Nations Sustainable Development Goals. 64

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Optimizing water management is promising to address the multiple challenges. 66 However, different water management schemes can lead to a wide range of outcomes 67 in rice yield (-16.9% to 21.9%), IRR (-68.0% to -0.3%), CH₄ (-85.5% to -0.1%) and 68 N₂O (0% to 364%) across climatic zones, reflecting complex interactions between 69 70 environmental factors and management strategies (Bo et al., 2022). Process-based models are powerful tools for predicting and managing the complicated interactions in 71 responses to water management, given their strength in simulating crop growth, water 72 dynamics, and soil biogeochemical processes under diverse genotype \times environment \times 73 74 management conditions (Tian et al., 2021; Chen et al., 2022; Yan et al., 2024). Despite 75 with several relevant studies at site-scales, extrapolation of optimized water management schemes from limited sites to the broader rice growing regions is hindered 76 77 by the diverse climate, soil, crop variety, field management, etc. (Yan et al., 2024; Liang et al., 2021). Region-specific simulations of the food-water-climate nexus are thus 78 urgently needed to identify tailored solutions. Nevertheless, current models face 79 challenges in accurately predicting yield responses to various water management 80 practices and adequately reproducing the spatial heterogeneity of these responses. 81

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Despite extensive experimental research to understand critical physiological effects 83 underlying yield responses, these processes have not been fully represented in models, 84 especially the compensation mechanisms. Compared to continuous flooding, imposing 85 moderate water deficit and then rewatering the field could increase both effective leaf 86 87 area and net photosynthetic rate upon re-irrigation to enhance photosynthesis for biomass production (Yang and Zhang, 2010). In addition, harvest index could increase 88 due to enhanced remobilization of assimilates and accelerated grain filling rate (Zhang 89 et al., 2008). However, prevailing models (for example, ORYZA, DSSAT, APSIM, 90 WHCNS) primarily focus on the negative impacts of water deficit (i.e., reduced 91 photosynthesis or leaf rolling), while neglecting or indirectly simulating crop 92 adaptation processes (e.g., enhanced root growth and water uptake in deeper soil layers) 93 (Bouman et al., 2001; Li et al., 2017; Liang et al., 2021; Tsuji et al., 1998). As a 94

95 consequence, yield sensitivities to water management could be overestimated, as 96 evidenced by evaluations of the ORYZA (v3) model (Xu et al., 2018). Moreover, 97 physiological processes respond differently to water availability at different growth 98 stages, while crop models generally use constant water effect coefficient throughout the 99 rice growing season (Ishfaq et al., 2020). These imply model deficiencies in predicting 100 yield response to water management, although no assessment across large scales exists.

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Accurate model parameters are essential for reproducing spatial heterogeneity of yield, 102 IRR, and GHGs. Previous studies usually used either the same parameters at different 103 pixels, calibrated against all observations, or the spatial proximity principle to 104 extrapolate model parameters for regional simulations, as a result of lacking enough 105 observations (Zhang et al., 2024; Zhang et al., 2016). However, critical model 106 parameters varied considerably when calibrated under different environmental and 107 management conditions, reflecting important impact of these factors on underlying 108 physiological and biogeochemical processes (Tan et al., 2021). As a consequence, 109 traditional model parameterization approaches are unlikely to capture variability of 110 yield, IRR, and GHGs due to their neglect of the environmental and management-111 related impacts (Song et al., 2023; Zhang et al., 2023). Besides, previous studies only 112 evaluated simplified irrigation protocols (i.e., once drainage at midseason or alternative 113 wetting and drying with constant threshold across the growing season) or only set bi-114 objectives as optimization targets (Tian et al., 2021; Chen et al., 2022), which likely 115 underestimated the regulation potentials. Therefore, an integrated framework composed 116 of a reliable modelling platform, broader water management schemes and multi-117 118 objective optimization targets are required for sustainable water management 119 optimization.

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121 To address these challenges, this study proposed an advancing framework that integrated a process-based soil-crop model (Soil Water Heat Carbon Nitrogen Simulator, 122 WHCNS) with key physiological effects, a novel model upscaling method, and a 123 multi-objective optimization algorithm (Non-dominated Sorting Genetic Algorithm II, 124 NSGA-II) at a parallel computing platform (see Fig.1 for workflow). This study focused 125 on rice yield (Yield), irrigation water use (IRR), methane (CH₄), and nitrous oxide 126 emissions (N₂O) of irrigated rice fields. First, three physiological effects were 127 quantified and embedded into WHCNS to enhance the prediction of yield responses. 128 Regionalized model parameters were then derived by developing parameter transfer 129 130 functions for regional simulations. The model's ability to reproduce the variations in the food-water-climate nexus was extensively validated against field observations. 131 Multi-objective optimization was conducted using the NSGA-II algorithm to 132 investigate tradeoffs within the food-water-climate nexus and assess the regulation 133 potentials of water management optimization. This framework was applied to China's 134 rice cropping system as an example, considering its position as the world's largest rice 135 producer and the ongoing conflicts between production demand, water scarcity, and 136 greenhouse gas emissions. This study aims to provide a valuable framework for 137 predicting and regulating rice's food-water-climate nexus towards sustainable water 138

139 management.

140141 2 Data and Methods

142 2.1 WHCNS model and input data

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144 The soil Water Heat Carbon Nitrogen Simulator (WHCNS) model was improved and incorporated into the advancing framework in this study to simulate rice yield, irrigation 145 water use (IRR), methane (CH₄), and nitrous oxide (N₂O) emissions of irrigated rice 146 fields at each pixel. The WHCNS model is a process-based agroecosystem model that 147 runs at a daily time step and comprises six major components: surface ponding water 148 dynamic, soil water movements, soil heat transfer, soil N transformation and transport, 149 soil organic turnover, and crop growth. Detailed model descriptions can be found in 150 (Liang et al., 2022; Liang et al., 2023; Liang et al., 2021). This model was chosen for 151 several considerations: (i) the model directly outputs all four target variables 152 simultaneously. This avoids biogeochemical models relying on crop models for detailed 153 physiological parameters to simulate yield and calculating IRR externally to obtain all 154 four targets as previously done (Tian et al., 2021; Yan et al., 2024), (ii) the model has 155 been proven to simulate frequent dry-wet cycles effect reasonably well in China rice 156 fields, due to simulating water and nitrogen dynamics in surfacing ponding water layer 157 that is specific for rice fields (Liang et al., 2021), (iii) the model is executable at both 158 site and regional scales with high efficiency and performs well in capturing spatial 159 variation in key processes (Liang et al., 2023), (iv) the model has a very flexible 160 irrigation setup, which allows for the precise control of paddy field water surface levels 161 162 by setting the minimum and maximum irrigation thresholds. It also enables calculating water usage for paddy field irrigation under various water management scenarios (Jiang 163 et al., 2021). The model is particularly suitable for simulating the regional food-water-164 climate nexus of rice fields. 165

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This study ran the model at both site and regional scales (0.5-degree spatial resolution). 167 Model input data includes daily meteorological variables, soil properties by depth, and 168 management variables related to planting, fertilization, and irrigation (Table S1). For 169 site-scale simulations, these variables were obtained from experimental studies, if 170 unreported, were extracted from spatial datasets according to geographical locations. 171 All spatial datasets were all resampled to 0.5-degree spatial resolution for regional 172 simulations. (1) Meteorological variables, including daily mean, maximum and 173 174 minimum air temperature, wind speed, precipitation, humidity, and downward solar radiation, were obtained from the fifth generation ECMWF reanalysis (ERA5) at 0.25-175 degree resolution (Hersbach et al., 2018). (2) Soil data including bulk density, clay 176 contents, and soil hydraulic properties (i.e., saturated water content, field water capacity, 177 wilting point, saturated hydraulic conductivity) at soil depths of 5, 15, 30, 60, 100, and 178 200 cm was obtained from SoilGrids (10 km) (Han et al., 2015). (3) The planting and 179 harvest dates were obtained from the crop calendar data of Global Gridded Crop Model 180 Intercomparisons (GGCMI) Phase 3 (Jägermeyr et al., 2021). (4) Fertilization practices 181 were conducted by the auto-fertilization component of the WHCNS model, assuming 182

no nitrogen stress (Liang et al., 2023). (5) Irrigation practices are defined by three 183 variables at daily step, including upper threshold (U_{IRR}), lower threshold (L_{IRR}, with a 184 positive value representing field water level and a negative value representing soil water 185 potential at 15 cm below the soil surface) and maximum allowable field water level 186 after rainfall (H_p, also refers to as bund height). Since there is no spatially explicit 187 188 information about realistic water management schemes, daily irrigation thresholds were set following Chen et al. (2022) for regional simulations. The model simulates field 189 water level of surface ponding layer and soil water potential of stratified layers at daily 190 step. Irrigation would be triggered whenever field water level ($L_{IRR} > 0$) or soil water 191 potential at 15 cm below the soil surface ($L_{IRR} < 0$) reach the predetermined L_{IRR} . 192 Irrigation demand is then calculated as the differences between L_{IRR} and U_{IRR}. 193

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2.2 Compilation of experimental observations

198 Extensive literature reviews were conducted to collect experimental observations for 199 model improvement and parameters calibration. Relevant studies should meet the 200 following criteria: (1) only field experiments covering an entire growing season were included, while pot and laboratory experiments under controlled environmental 201 202 conditions were excluded, (2) the control and treatments only differed concerning water management with continuous flooding (CF) as control and non-continuous flooding 203 irrigation (NCF) as treatment, but not concerning other agronomic practices (e.g., 204 cropping intensity, fertilizer management, and tillage). This was to isolate water 205 206 management effects while avoiding confounding effects of other factors, (3) upper and lower irrigation thresholds were explicitly reported, and lower thresholds were 207 indicated by soil water potential measured at the soil depth of 15-20 cm. Observations 208 based on soil water potential at the other soil depth or the other soil-water indicators 209 (e.g., soil water contents) were excluded, (4) at least one of target variables were 210 observed, including rice yield (Yield), irrigation water use (IRR), methane emissions 211 (CH_4) , nitrous oxide emissions (N_2O) , leaf area index (LAI), net photosynthetic rate 212 (Pn), and harvest index (HI). For LAI and Pn, the growth stages of observations (i.e., 213 tillering, booting, heading, and ripening stage) were recorded to account for growth 214 stage-dependent effects. As a result, we collected observations of 119 experiments from 215 37 studies covering 28 sites in 6 countries (i.e., China, India, Philippines, Japan, 216 Bangladesh, and Peru) (Fig. S1). These observations were split into two datasets 217 according to target variables. The first dataset including Yield, IRR, CH4, or N2O 218 219 observations was used for calibration of model parameters. The second dataset of LAI, Pn, or HI observations was used to quantify water management effects on physiological 220 processes for model improvement (Sect. 2.3). 221

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For each paired observation under the control and treatment, the effects of noncontinuous flooding irrigation were calculated as the ratio of observations under treatment to that under control (Equation 1). This yielded 251 records for R^{Yield} , 235 for R^{IRR} , 37 for R^{CH4} , 14 for R^{N2O} , 561 for R^{LAI} (including 61 from tillering stage, 159 from booting stage, 202 from heading stage and 139 from ripening stage), 84 for R^{Pn} (including 42 from tillering stage, and 42 from filling stage), and 351 for R^{HI} .

$$R^{X} = \frac{X_{NCF}}{X_{CF}}$$
(1)

where R^X represents non-continuous flooding effects (*NCF*) on target variables *X* (including *Yield*, *IRR*, *CH*₄, *N*₂*O*, *LAI*, *Pn*, and *HI*), *X*_{NCF} and *X*_{CF} represent variable values under non-continuous flooding (*NCF*) and continuous-flooding irrigation (*CF*), respectively. Relative changes of target variables were calculated as (R^X -1)×100 for interpretation and representation (e.g., $\Delta Yield$, ΔIRR , ΔCH_4 , ΔN_2O).

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237 For each paired observation, four categories of information were also collected. First, climatic variables included mean daily air temperature (T), precipitation (P), and crop 238 evapotranspiration (PETc) during growing season. The difference between P and PETc 239 was further calculated to indicate climatological water availability (CWA). Second, soil 240 variables included sand content, bulk density (BD), soil organic carbon (SOC), pH, and 241 242 soil hydrological properties (e.g., saturated water content (SAT), field water capacity (FWC)). Third, management-related variables included nitrogen application rate and 243 244 timing, as well as lower (L_{AWD}) and upper (U_{AWD}) irrigation thresholds. Fourth, experimental parameters included geographical location (latitude, longitude), dates of 245 seeding (also transplanting date in transplanted systems), anthesis, and harvest. These 246 variables were used for running WHCNS (Sect. 2.1) and conducting correlation 247 248 analyses (Sect. 3.1).

250 **2.3 Model improvement**

251 2.3.1 Incorporation of physiological effects

In the original WHCNS model, water management effects on crop growth were 253 simulated by calculating water stress factor based on the Feddes reduction function 254 (Feddes and Zaradny, 1978). Specifically, the water stress factor is calculated at daily 255 step as a function of soil water potential to reduce root water uptake, assuming 70 kpa 256 and 1500 kpa as thresholds of when root water uptake starts to decrease and approaches 257 0 (Equation 2-3). The calculated water stress factor was used to reduce the simulated 258 actual biomass production rate, which further indirectly impact produced biomass 259 allocated for leaf growth and yield formation (Equation 4-6). 260

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262 $T_{a} = \int_{L_{R}} S(h, h_{\Phi}, z) dz = T_{p} \int_{L_{R}} a_{w}(h, z) a_{s}(h_{\Phi}, z) b(z) dz$

(2)

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$$cf(w) = \frac{T_a}{T_p} = \begin{cases} \int a_w(h,z)a_s(h_{\Phi},z)b(z)dz \\ \frac{L_R}{\omega} = \frac{\omega}{\omega} = 1 \qquad \omega > \omega_c \\ \int a_w(h,z)a_s(h_{\Phi},z)b(z)dz \\ \frac{L_R}{\omega_c} = \frac{\omega}{\omega_c} < 1 \end{cases} \qquad (3)$$

$$Fgc = DL \times \frac{AMAX}{K_e} \times \ln[\frac{AMAX + CC}{AMAX + CC \times (-LAI \times K_e)}]$$
(4)

$$Fgass = Fgc \times \frac{30}{44} \times cf(w) \times cf(N)$$
(5)

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267 $GAA(org) = Fgass \times fr(org)$ (6)

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where T_a and T_p are actual and potential root water uptake (cm d⁻¹). L_R indicates root 269 length (cm). $a_w(h,z)$ and $a_s(h_{\varphi,z})$ are water and salt stress functions. b(z) is root 270 distribution function. w_c is the critical threshold of volumetric soil water content w271 272 above which root water uptake is reduced in water limited layers of the root zone, but the plant compensates by uptaking more water from other layers that have sufficient 273 available water. Fgc is daily potential dry matter production accounting for the light 274 interception, radiation use efficiency, and the CO₂ effects (kg hm⁻² d⁻¹). AMAX is the 275 maximum assimilation rate accounting for temperature effect (kg hm⁻² h⁻¹). DL, K_e , 276 and CC indicate day length (h d^{-1}), extinction coefficient (-) and actual radiation use 277 (kg hm⁻² h⁻¹). Fgass is daily actual dry matter production (kg hm⁻² d⁻¹) accounting 278 for water (cf(w)) and nitrogen stress (cf(N)). GAA indicates produced biomass 279 allocated to organs (leaf or grains) (kg hm⁻² d⁻¹) with the fraction of fr(org). 280

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To modify the WHCNS, NCF effects on leaf expansion, photosynthesis rate, and 282 assimilate partition were quantified based on experimental observations and 283 incorporated into WHCNS (Fig. S2). To do so, mean values of observed effects were 284 285 first calculated by experimental gradient of soil water potential (SWP, negative values) and growth stages (RDS, 0-1) (Table S2-S4). RDS corresponds to planting, tillering, 286 booting, heading, filling, and maturity stages was quantified as 0, 0.20, 0.40, 0.55, 0.75, 287 288 and 1. Effects at other levels of SWP and RDS were then estimated by bilinear interpolation (i.e., F^{LAI}(SWP, RDS), F^{Pn}(SWP, RDS), F^{HI}(SWP)). Three functions were 289 thus developed involving three new genetic parameters to account for differences in 290 cultivar sensitivities (P^{LAI}, P^{Pn}, P^{HI}, Equations 7-9). The three functions were added to 291 the origin crop growth module to modify simulations of leaf area index, net 292 photosynthesis rate and biomass allocated into grains (Equation 10-12, Fig. 2a). 293 294

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$$R^{LAI}(SWP, RDS) = 1 + \left[\left(F^{LAI}(SWP, RDS) \right) - 1 \right] \times P^{LAI}$$
(7)

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$$R^{P_n}(SWP, RDS) = 1 + \left[\left(F^{P_n}(SWP, RDS) \right) - 1 \right] \times P^{P_n}$$
(8)

297
$$R^{HI}(SWP) = 1 + \left[F^{HI}(SWP) - 1\right] \times P^{HI}$$
(9)

$$LAI' = GAA(leaf) \times SLA \times R^{LAI}$$
(10)

$$AMAX' = AMAX \times R^{Pn} \tag{11}$$

$$GAA(grains)' = Fgass \times fr(grains) \times R^{HI}$$
(12)

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where R^{LAI} , R^{Pn} , R^{HI} represent NCF effects on leaf area index, net photosynthetic rate 302 and harvest index, respectively. SWP represents soil water potential at 15-20 cm soil 303 depth. RDS represents relative development stages (0-1). P^{LAI} , P^{Pn} , and P^{HI} are genetic 304 parameters indicating cultivar sensitivities to irrigation regulation that were calibrated 305 based on observations (Sect. 2.4). LAI and SLA are leaf area index $(m^2 m^{-2})$ and 306 specific leaf area (m² kg⁻¹). LAI', AMAX' and GAA(grains)' denote simulations of the 307 modified model. It is worth noting that the three functions can be flexibly coupled to 308 the other process-based crop models to modify the simulation of leaf area growth, 309 biomass production, and allocation processes. The genetic parameters are needed to 310 be recalibrated against observed yield responses considering different model 311 312 structures.

2.3.2 Contribution analysis 314

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Scenario simulations were conducted to isolate contributions of the three physiological 316 317 effects on yield changes (Δ Yield) (Table S5). Four scenarios were simulated by 318 considering all the three effects (S1) and omitting one of the three effects at a time (S2-S4). For each scenario, the model was run under CF and NCF conditions respectively 319 to calculate Δ Yield. The differences in the simulated Δ Yield between S1 and S2-S4 320 represent yield changes induced by changes in leaf expansion, photosynthesis rate and 321 assimilate partition, respectively (i.e., $\Delta Yield^{Pn}$, $\Delta Yield^{HI}$). Relative 322 contribution of each process was calculated as the ratio of the absolute yield change 323 324 induced by the process to the sum of absolute yield change induced by the three 325 processes (Equation 13).

$$CON^{p} = \frac{\left|\Delta Yield^{p}\right|}{\sum_{p=1}^{3} \left|\Delta Yield^{p}\right|} \times 100$$
(13)

327 where p represents the three new physiological processes (i.e., p = 1, 2, 3), CON^p indicates relative contribution of the process p to Δ Yield, Δ Yield^p is yield changes 328

induced by the process *p*. 329

2.4 Parameters regionalization

- Spatially explicit model parameters are critical for reasonably reproducing spatial 333 variabilities of target variables. In this study, seven key model parameters were selected 334 and mapped at 0.5-degree spatial resolution due to their high influence on target 335 336 variables, including accumulated temperature for crop maturity (Cumtemp), minimum assimilation rates (AMIN), the maximum CH₄ production rate per soil weight at 30 °C 337 (MPmax), maximum portion of denitrification to N₂O production ($f_{N2O d}$) and the three 338 new genetic parameters $(P^{LAI}, P^{Pn}, P^{HI})$. These parameters were first finely calibrated at 339 site-scales (Sect. 2.4.1) and then upscaled to regional scales (Sect. 2.4.2). To capture 340 spatial variabilities of NCF effects, different parameters were used under CF and NCF 341 342 conditions, except for genetic parameters. This was consistent with a previous 343 modelling study, aiming to indicate different potentials of methane production and denitrification under different water management regimes (Song et al., 2023). 344
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346 **2.4.1 Calibration of site-scale parameters**

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348 Under CF conditions, the parameter *Cumtemp* was first determined by cultivar as the minimum cumulative daily temperature higher than 10°C (base temperature for rice 349 growth) across all experiments using the cultivar. Then AMIN, MPmax and $f_{N2O d}$ were 350 calibrated to achieve the best fit of predicted target variables with observations under 351 continuous flooding conditions (i.e., experimental control). Under NCF conditions, 352 Cumtemp and AMIN were the same with that calibrated from CF conditions. The other 353 parameters (MPmax, $f_{N2O d} P^{LAI}$, P^{Pn} and P^{HI}) were then calibrated by minimizing the 354 sum of simulated squared residuals under non-continuous flooding conditions (Table 355 S6). To obtain more accurate parameter estimates, the advanced parameter estimation 356 algorithm (PEST) was used (Doherty, 2010). As a result, 51 groups of genetic 357 parameters (*Cumtemp*, AMIN, P^{LAI} , P^{Pn} and P^{HI}), 56 parameter values of MPmax (19) 358 for control and 37 for treatment) and 24 parameter values of $f_{N2O d}$ (10 for control and 359 360 14 for treatment) were calibrated.

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362 2.4.2 Parameters upscaling

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To upscale genetic parameters (AMIN, Cumtemp, P^{LAI}, P^{Pn}, P^{HI}) calibrated at site 364 scales to regional scales, the rice cultivar for each grid was first determined. Then, the 365 calibrated genetic parameters of the cultivar were used to create the grid. Since the 366 spatial distribution of rice cultivar is unknown, cultivar of each grid cell was 367 determined as follows. First, cultivars with Cumtemp lower than the effective 368 accumulative temperature requirement of the grid were identified. This ensures the 369 cultivar could reach maturity under the grid cell's temperature conditions. The grid's 370 temperature requirement was calculated as *Cumtemp* during rice growing periods 371 specified by the crop calendar data of GGCMI Phase 3 (Jägermeyr et al., 2021). 372 Subsequently, cultivars with AMIN that closely match the baseline AMIN of the grid 373 cell were selected. The baseline AMIN was estimated using PEST to achieve the best 374

fit of yield simulation with the records in county-scale statistical yearbooks of China
(downscaled to 0.5-deg spatial resolution). These procedures were designed to ensure
that yield simulations were aligned with cultivar's genetic potential and spatially
consistent with observations.

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380 To upscale parameters *MPmax* and $f_{N2O d}$, two parameter transfer functions (PTFs) were developed. Dependent variables were the ratio of site-calibrated parameters 381 under treatment to that under control (i.e., R^{MPmax} and R^{fN2O_d}) (Equation 16-17). 382 Independent variables were determined as field water capacity (FWC) for R^{MPmax} and 383 bulk density (BD) for R^{fN2O_d} , due to their higher correlations with dependent 384 variables. The function forms were determined as the form with the highest R². As a 385 result, the relationship between field water capacity and R^{MPmax} was best fitted by an 386 exponential function ($R^2 = 0.62$, p < 0.001), and the relationship between bulk density 387 and R^{fN2O_d} was best fitted by a quadratic function (R² = 0.91, p < 0.001) (Fig. S5). 388 The importance of soil properties in regulating spatial heterogeneity of denitrification 389 potentials aligns with previous studies (Tang et al., 2024). Parameters of the PTFs 390 were calibrated using the least square method (Equation 16-17). With the calibrated 391 PTFs, the ratio of parameters under NCF relative to CF (R^{MPmax} and R^{fN2O_d}) for each 392 grid could be predicted by combining spatial dataset of FWC and BD. Then gridded 393 MP_{max} and $f_{N2O d}$ for CF conditions (MP_{max}^{CF} and $f_{N2O d}^{CF}$) were estimated using PEST 394 targeting CH₄ from the EDYGA v8.0 dataset (Crippa et al., 2024) and N₂O emissions 395 estimated by Cui et al. (2024) (Fig. S4). These parameters were estimated for 2013 396 and 2015 and subsequently validated for 2014 and 2016 to assess their ability to 397 reproduce the spatial variability of target variables (Fig. S3). Finally, MP_{max} and f_{N2O} d 398 for NCF conditions were calculated by multiplying MP_{max}^{CF} and $f_{N2O d}^{CF}$ with the 399 predicted ratio (R^{MPmax} and R^{fN2O_d}). 400

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$$R^{MP_{max}} = MP_{max}^{NCF} / MP_{max}^{CF} = 986 \times e^{-26 \times FWC}$$

$$\tag{16}$$

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 $R^{f_{N2O_{-d}}} = f_{N2O_{-d}}^{NCF} / f_{N2O_{-d}}^{CF} = 268 \times BD^2 + 789 \times BD + 581$ (17)

405

406 Where R^{MPmax} and R^{fN2O_d} represent the ratio of parameter MPmax and f_{N2O_d} 407 calibrated under non-continuous flooding (treatment) to that under continuous 408 flooding (control). *FWC* and *BD* represent field water capacity (cm³ cm⁻³) and soil 409 bulk density (g cm⁻³) obtained from SoilGrids (10 km) (Han et al., 2015). 410

411 To prove the efficacy of the PTFs, two other parameter upscaling approaches were also 412 used for comparison, including the mean parameters approach and the spatial proximity 413 approach. These approaches were widely used in previous modelling studies to derive 414 regional parameters and conduct regional simulations (Zhang et al., 2024). To adopt the 415 mean parameter approach, mean value of the site-calibrated *MPmax* and f_{N2O_d} (Sect. 416 2.4.1) were calculated respectively for CF and NCF conditions, and then the two

constant mean parameters was used in regional simulations. To adopt the spatial 417 proximity approach, the nearest site of a site was first identified according to 418 geographical coordinates. Then both MPmax and $f_{N2O d}$ calibrated from the nearest site 419 were used for simulation of this site. The three approaches were compared in their 420 performance to reproduce the observed variations in ΔCH_4 and ΔN_2O (Fig. 3). 421

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2.5 Regional scenario simulations and driver identification

Scenario simulations were conducted to test whether the proposed framework could 425 reasonably predict the response sensitivity of target variables and their relations under 426 different irrigation schemes. To do so, the well-calibrated WHCNS model was run 427 under baseline and a series of non-continuous irrigation scenarios using the parallel 428 429 computing framework (Liang et al., 2023). For baseline condition, irrigation thresholds were set according to Chen et al. (2022). For non-continuous flooding irrigation 430 scenarios, a range of the lowest irrigation threshold levels were set based on 431 observations (-5, -10, -15, -20, -30, -40 and -50 kpa). The upper irrigation thresholds 432 were kept the same with baseline for consistency with experiments. NCF effects were 433 then calculated from model simulations and compared with observed effects. Observed 434 effects were obtained from two datasets. The first is the one compiled for this study 435 (Sect. 2.2) using soil water potential to distinguish irrigation schemes. The second was 436 obtained from Bo et al. (2022), who used the ratio of days with no surface water to total 437 growing days (UFR) to differentiate irrigation schemes. To facilitate comparison, the 438 UFR of each irrigation scenarios was also calculated and output by WHCNS (Fig. S9). 439

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To identify the dominant factor driving spatial patterns of NCF effects, correlation 441 analyses between simulated NCF effects and variables were performed following Cui 442 443 et al. (2021). Climatic, soil and management-related factors were selected as independent variables, including T, P, ET, Clay, BD, SOC and fertilizer rate. The 444 445 analyses were conducted respective for Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O using 3.5°-by- 3.5° moving windows. The data resolution was 0.5° by 0.5° , meaning the surrounding 446 49 pixels were used for each grid. The correlation coefficient and its significance in 447 each grid was first calculated, and the dominant driver was then defined as the factor 448 with the largest absolute correlation coefficient. To assess the robustness of the results, 449 similar analyses were done with moving windows at higher spatial resolutions (e.g., 450 2.5° by 2.5°). 451

452

2.6 Single-objective and multi-objective optimizations 453

454

Based on scenario simulations, four single-objectives and a multiple-objective were 455 designed to identify optimal irrigation schemes. The four single-objective targets are (1) 456 maxYield: maximizing rice yield, (2) minIRR: minimizing irrigation water use, (3) 457 minCH₄: minimizing CH₄ emission, and (4) minN₂O: minimizing N₂O emissions. 458 Under all targets, yield reduction compared to CF conditions was avoided. With optimal 459 solution under the four single-objective scenarios, the largest regulation potentials to 460

increase yield and reduce IRR, CH₄, and N₂O emissions were assessed. For comparison,
the scenario simulations and optimization were also conducted using the origin
WHCNS model (Fig. 5).

464

476

The multi-objective optimization was conducted by combining the improved WHCNS 465 model and the NSGA-II algorithm (Deb et al., 2002). First, a set of 100 parental 466 populations was initialized with random solutions. Each population includes 1993 467 individuals, corresponding to 1993 grid cells of irrigated rice areas. Second, the 468 objective functions were computed with each solution by executing the WHCNS model 469 (Equation 18). Third, the performance of each population was evaluated by ranking the 470 fitness of its objective functions. Fitness is a measure of how well a solution performs 471 and is calculated based on the non-dominated sorting rank. Then, a new generation was 472 473 generated through selection, crossover, and mutation based on fitness. Finally, Pareto fronts were generated after 100 generations had been evaluated (that is 10000 474 populations). 475

477
$$f_{objective} = \begin{bmatrix} f_{max} \{\sum_{n=1}^{N} WHCNS(yield)\} \\ f_{min} \{\sum_{n=1}^{N} WHCNS(IRR)\} \\ f_{min} \{\sum_{n=1}^{N} WHCNS(GWP)\} \end{bmatrix}$$
(18)

478
$$WHCNS(GWP) = 27.2 \times WHCNS(CH_4) + 273 \times WHCNS(N_2O)$$
(19)

where *f_{objective}*(*yield*, *IRR*, *GWP*) denotes the collection of objective functions, *f_{max}* 479 denotes the objective that needs to be maximized (e.g., rice yield), and f_{min} denotes the 480 objective that needs to be minimized (e.g., IRR, GWP). GWP is the integrated global 481 warming potential of combined emissions of CH₄ and N₂O emissions and is 482 calculated based on WHCNS simulations (Equation 19) (Forster et al., 2021). It 483 should be noted that this study set equal weight for each target variable to evaluate the 484 fitness of each solution. Decision-makers can simply set the weight values of different 485 objectives according to their preferences, or adopt advanced multi-objective criteria 486 decision-making methods such as the efficiency coefficient method (Guo et al., 2021). 487 The regulation potentials of multiple-objective optimization were calculated as the 488 averaged NCF effects (Δ Yield, Δ IRR, Δ CH₄, Δ N₂O, Δ GWP) of all non-dominated 489 solutions. The potentials were further compared with that from single-objective 490 optimizations to investigate tradeoffs between target variables (Fig. 6). 491 492



Figure 1 Research framework of this study. The framework mainly combines data 494 compilation, model improvement, parameter regionalization, scenario simulations, and 495 *multi-objective optimization*. The framework can be flexibly adapted with alternative 496 irrigation scenarios, optimization objectives, and optimization algorithms in other 497 modelling studies. LAI, Pn, and HI represent leaf area index, net photosynthetic rate, 498 and harvest index. AMIN, MPmax, f_{N2O d}, P^{LAI}, P^{Pn}, and P^{HI} are model parameters 499 calibrated and mapped in this study (Sect. 2.4). CF and NCF represent continuous 500 flooding and non-continuous flooding irrigation. SWP and UFR represent soil water 501 potential and the ratio of unflooded days to total rice growing days, indicating different 502 irrigation schemes. See the Appendix for detailed descriptions of parameters and 503 504 variables.

493

506 **3 Results and discussion**

507 **3.1 Performance of model improvement**

508

The origin WHCNS model was first evaluated in reproducing variabilities of rice yield 509 and irrigation water use under various irrigation schemes. For rice yield, model 510 performance is satisfying when mixing observations under continuous flooding (CF, 511 experimental control) and non-continuous flooding (NCF, experimental treatments) 512 irrigation schemes together ($R^2 = 0.41$, normalized root mean square error *nRMSE* = 513 11%) (Fig. S6). In particular, with fine-turned crop genetic parameters (i.e., Cumtemp 514 and AMIN), the origin model performed well under CF condition ($R^2 = 0.74$, nRMSE =515 1 3%), while had worse performance under NCF condition ($R^2 = 0.22$, nRMSE = 13%) 516 (Fig. S6). As a consequence, the origin model failed to reproduce variations in observed 517 yield changes (Δ Yield) ($R^2 = 0.03$, nRMSE = 17%) (Fig. 2b). More importantly, the 518

simulations could not reproduce Δ Yield sensitivities to soil water potentials presented 519 in field experiments (Fig. 2d). In contrast to yield, model performance in simulating 520 irrigation water use responses (Δ IRR) variabilities and its sensitivities to soil water 521 potentials was acceptable (Fig. 2c and 2e). These results highlight the primary 522 modelling deficiency in simulating Δ Yield. Given the satisfying model performance in 523 simulating yield under CF and Δ IRR, the underperformance is likely due to lacking 524 critical physiological processes responsible for yield responses to NCF rather than 525 uncertainties of crop parameters. 526

527

After incorporating the three functions of NCF effects and fine calibration of genetic 528 parameters (Sect. 2.3, Fig. 2a), the model performance was substantially improved. The 529 explained variabilities of Δ Yield increased from 3% to 52% and *nRMSE* decreased from 530 17 % to 11% (Fig. 2b). The observed Δ Yield sensitivities to soil water potential (9% 531 kpa⁻¹, P < 0.001) could be reasonably reproduced by the modified model (13% kpa⁻¹, 532 P < 0.001) rather than the origin mdoel (P > 0.05) (Fig. 2d). The cultivar differences of 533 yield responses could also be simulated (R = 0.67) (Fig. S7). Across the three processes, 534 leaf area growth ($\Delta Yield^{LAI}$) was primarily responsible for yield losses, while net 535 photosynthetic rate ($\Delta Yield^{Pn}$) and biomass translocation ($\Delta Yield^{HI}$) contributed to vield 536 increases (Sect. 2.3.2, Fig. S8). The positive contributions are larger in warmer and 537 more humid areas, and in acidic soils with larger field water holding capacity and higher 538 SOC. These findings conform with empirical relationships between Δ Yield and 539 environmental factors reported by previous meta-analysis (Carrijo et al., 2017). These 540 results prove efficacy of the modified model to predict and regulate Δ Yield under 541 542 diverse irrigation schemes and environmental conditions.





Figure 2 Model improvements by incorporating water effects on physiological 545 processes. (a) Schematic of critical physiological effects in response to different 546 irrigation schemes and their representation in the WHCNS model. (b-c) Model 547 performance for simulating Δ Yield (b) and Δ IRR (c) based on the origin (blue) and 548 modified (orange) WHCNS model. Darker colored dots indicate lower soil water 549 potential (unit: kpa). (d-e) Sensitivity of Δ Yield and Δ IRR to lower irrigation threshold 550 of soil water potential. Black, blue, and orange colors show the results of observations 551 and simulations based on the origin and modified WHCNS model, respectively. Circles 552

are mean values; error bars show the 25–75% interquartile range. The lines are the linear regression lines with dashed lines indicating non-significant relationships based on two-sided t-test (P > 0.05). The shaded areas around each line represent the 95% confidence interval.

557

558 Besides being coupled to WHCNS as an integrated system, the new functions also contribute to advancing related modelling studies by directly involving positive 559 physiological effects and considering stage-dependent response sensitivities (Li et al., 560 2017). By contrast, most prevailing crop models only account for negative effects of 561 soil drying and reduced transpiration, while does not incorporating direct compensation 562 effects (such as increased photosynthesis rate upon rewatering). Moreover, constant 563 564 stress sensitivity parameters were generally used for all growth stages (such as ORYZA and DSSAT) (Bouman et al., 2001; Tsuji et al., 1998). These models could flexibly 565 incorporate the three new functions and recalibrate the genetic parameters (i.e., P^{LAI} , 566 P^{Pn} , and P^{HI}) following the procedures of this study to improve their performance in 567 predicting yield responses. 568

569

570 **3.2 Performance of regionalized parameters**

571

572 To simulate regional NCF effects, the model was first run respectively for CF (baseline) and NCF conditions using the parallel computing framework at a spatial resolution of 573 0.5-deg. NCF effects were then calculated using model simulations following Equation 574 1 (Fig.1 and Sect. 2.4). Using the PEST-calibrated gridded model parameters for CF 575 576 (Sect. 2.4.1), the *nRMSE* between model simulations and their spatial datasets were 20% 577 to 29% for yield, ~7% for IRR, ~4% for CH4, and 4% to 6% for N2O during the validation period (year 2014 and 2016) (Fig.S2). It was noted that the *nRMSE* of rice 578 yield was relatively larger than that of other target variables, despite being within an 579 acceptable range (<30% for the validation periods). This could be caused by interannual 580 cultivar changes, which was difficult to consider in large-scale simulations due to the 581 of spatial distribution of rice cultivars. Overall, these results reveal a satisfying 582 lack model calibration to simulate baseline values and spatial variabilities of target variables. 583 584

To reproduce observed variabilities of NCF effects on target variables, NCF effects on 585 key model parameters (*MPmax* and $f_{N2O d}$) were incorporated for constraining model 586 simulations. To do so, NCF effects on model parameters were first quantified from site-587 scale calibrations and extrapolated to regional scale (Sect. 2.4). Three approaches of 588 589 parameter extrapolation were tested and compared, including developing parameter transfer functions (PTFs), using mean site-calibrated parameters (mean), and using 590 spatially nearest calibrated parameters (spatial) (Sect. 2.4.3). Results showed that 591 developing PTFs performed the best to reproduce observed variabilities of ΔCH_4 and 592 ΔN_2O (Fig. 3). Model simulations using parameters estimated by PTFs explained 37% 593 and 94% of variations in ΔCH_4 and ΔN_2O , with *nRMSE* being 25% for ΔCH_4 and 10% 594 for ΔN_2O (Fig. 3a-b). By contrast, simulations based on the other two approaches could 595 hardly reproduce observed variabilities of ΔCH_4 and ΔN_2O , with *nRMSE* achieving 66% 596

to 72% for Δ CH₄ and 29% to 73% for Δ N₂O (Fig. 3c-f). These results prove the efficacy of the developed PTFs and suggest soil variables as good predictors for spatial extrapolation of site-calibrated parameters to simulate CH₄ and N₂O. The PTFs could also be referred by other biogeochemical models for regional simulations of CH₄ and N₂O (such as the Denitrification-Decomposition model and the Dynamic Land Ecosystem Model) (Zhang et al., 2016).

603



604

Figure 3 Comparison of model parameter upscaling approaches. Model 605 performance in simulating methane and nitrous oxide emissions changes based on 606 parameters derived from (a-b) parameter transfer functions (PTFs), (c-d) mean site-607 calibrated parameters, and (e-f) spatially nearest parameters. The color of the dots 608 indicates lower irrigation thresholds of soil water potential under non-continuous 609 flooding irrigation (unit: kpa). The solid lines are regression lines with dashed lines 610 indicating non-significant relationships (P > 0.05). Blue shading around each line 611 represents the 95% confidence interval. 612

613

614 Considering scarce observations of NCF effects across space, it was impractical to 615 directly evaluate the regionalized parameters in reproducing spatial variability of NCF

effects. Therefore, the proposed framework was evaluated in terms of the response 616 sensitivity of target variables and their relationships under different irrigation schemes 617 (Sect. 2.5). Scenario simulations broadly conformed with observations regarding the 618 magnitude of NCF effects and response sensitivity across soil water potential gradients 619 (Fig. S9). With decreased soil water potential threshold, Δ Yield decreased quasi-linearly, 620 ΔCH_4 and ΔIRR decreased at a decelerating rate, while ΔN_2O showed slight variabilities 621 (Fig. S9a). The decelerating decrease in ΔCH_4 was also observed in experiments, 622 suggesting the model ability to simulate maximum potentials of CH₄ mitigation 623 (Balaine et al., 2019). The response sensitivity was further validated using an alternative 624 observation dataset (Fig. S9b). Besides, the observed synergy or tradeoffs of the yield-625 IRR-GHGs nexus were broadly covered by scenario simulations using the modified 626 627 model rather than using the origin model (Fig. S9c). Such bias could further impact 628 assessment of regulation potentials of the food-water-climate nexus.

629

630 **3.3 Assessment of regional regulation potentials**

631

632 Scenario simulations revealed large spatial variabilities of NCF effects on all target variables (Fig. 4). Applying the same irrigation scheme (e.g., lower irrigation 633 threshold of -15 or -30 kpa) could induce larger yield increase in the southwest 634 single-rice region (XNS: 2.4% to 3.4%), while larger yield losses in northern regions 635 (HHH: -3.2%) (Fig. 4a and b). The HHH region also showed larger yield sensitivity 636 with decreased lower irrigation threshold $(-0.24\% \text{ kpa}^{-1})$ (Fig. 4c). For IRR, 637 relatively larger water saving benefits occurred in south regions, whereas response 638 sensitivity was larger in northeast regions $(-1.7\% \text{ kpa}^{-1})$. For CH₄, north rice growing 639 regions showed relatively higher reductions (NES: 64% to 82%, HHH: 77% to 88%) 640 and higher response sensitivity to decreased soil water potential threshold. The 641 findings about larger water saving benefits in south China and larger CH₄ mitigation 642 in north China were consistent with previous assessments (Tian et al., 2021). 643 However, N₂O emissions showed widespread increase regardless of lower irrigation 644 threshold, except for northeast regions, indicating low opportunities to reduce N₂O by 645 only optimizing water management. 646

647

To further understand the drivers shaping the spatial variations in NCF effects, 648 correlation analyses were conducted for each target variable across varying lower 649 irrigation threshold. Overall, climatic and edaphic variables were the most important 650 drivers, while management-related variables were less important (Fig. 5). Exceptions 651 occurred in the south double rice region (HND) for Δ Yield and the southwest single 652 rice region (XNS) for $\Delta N_2 O$, where higher fertilizer application rate was associated 653 with larger yield increase but decreased N₂O reduction potentials (Fig. S10 and S11). 654 For both Δ Yield and Δ IRR, clay content was the most important driver at higher 655 irrigation thresholds, while climate factors showed increasing importance with 656 decreased irrigation thresholds (Fig. 5a and b). By contrast, reduction potentials for 657 CH₄ and N₂O emissions were dominated by edaphic factors regardless of irrigation 658 threshold (i.e., clay for CH₄ and bulk density for N₂O) (Fig. 5c and d). These findings 659

- 660 highlight the complex interplay of factors influencing regulation potentials of rice
- 661 production, irrigation water use and greenhouse gas emissions through NCF adoption.
- 662



Figure 4 Spatial pattern of relative changes in target variables under different 664 irrigation schemes. The four columns correspond to the four target variables Δ Yield, 665 Δ IRR, Δ CH₄, and Δ N₂O, respectively. (a) relative changes of target variables under a 666 lower irrigation potential of -15 kpa, (b) relative changes of target variables under a 667 lower irrigation potential of -30 kpa, (c) differences between (b) and (a), (d) results for 668 different rice growing regions. NES, HHH, CJS, CJD, HND, and XNS indicate six rice 669 growing areas of China, namely, Northeast Single rice, HuangHuaiHai single rice, 670 Yangtze River single rice, Yangtze River double Rice, South China Double rice, and 671 Southwest China Single rice, respectively. 672



674

Figure 5 Drivers regulating spatial variations in relative changes in yield (a), IRR

(b), CH₄ (c) and N₂O (d). The numbers and colors indicate correlation coefficients, 676 with gray indicating non-significant correlations (N.S., P > 0.05). The pie plots 677 represent the proportion of irrigated rice areas (%) for which relative changes 678 variation is regulated by the dominant drivers. The dominant driver is defined as the 679 factor with the largest absolute correlation coefficient in each grid cell, identified from 680 3.5°-by-3.5° moving windows. The numbers in blue, orange and green around the pie 681 plots denote the area proportions dominated by climate (i.e., T + P + ET), soil (i.e., 682 Clay + BD + SOC) and management-related (i.e., Fertilizer rate) factors under 683 corresponding lower irrigation threshold. Spatial distributions of dominant drivers are 684 shown in Fig. S10 and S11. 685

686

To identify the largest regulation potentials from NCF adoption, four single objective targets were designed, including maximizing rice yield, minimizing IRR, CH₄ emissions, or N₂O emissions (denoted as *maxYield*, *minIRR*, *minCH*₄, *min N₂O*, Sect.

- 690 2.6). Results indicated that the largest regulation potentials of Δ Yield, Δ IRR, Δ CH₄
- and ΔN_2O were 4.6%, -61.0%, -64.2% and -10.9%, respectively (Fig. 6a). These
- potentials could be achieved respectively over 91%, 91%, 88% and 26% of national
- 693 rice areas (Fig. 6b). Spatially, larger yield increase potential occurred in south (HND:
- 694 7.7%) and southwest regions (XNS: 6.8%) (Fig. S12A). The reduction potential of

- IRR and CH₄ showed relatively slight spatial variabilities. In contrast, reduction 695 potential of N₂O primarily concentrated in northern regions (NES: -30%) due to 696 increased N₂O in southern regions (Fig. 5a and S12A). N₂O increase in southern 697 regions is associated with higher nitrogen application rates, providing substrate for 698 nitrification and denitrification processes to facilitate N₂O emissions (Jiang et al., 699 700 2019). The results conform to previous studies in that irrigation and nitrogen should be co-regulated for these areas to avoid unintended N₂O emissions from water 701 management (Jiang et al., 2019; Kritee et al., 2018). 702
- 703
- The largest regulation potentials of Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O are not likely to 704 be achieved at the same time, as evidenced by different optimized irrigation strategies 705 between single-objective targets (Fig. 6 and S13). For example, the lower irrigation 706 707 threshold should be higher than -20kpa for most areas (84%) under maxYield, while lower than -20kpa over half areas under minIRR and minCH₄. This suggests tradeoffs 708 between yield increase and IRR/CH₄ mitigation (Bo et al., 2022). To compare, using 709 the origin model could overlook nearly 20% feasible areas for applying optimized 710 711 irrigation schemes (Fig. 6). As a consequence, regulation potentials of Δ Yield, Δ IRR, Δ CH₄ and Δ N₂O could be underestimated by 4%, 11%, 14%, and 2%, especially for 712 the southwest regions (XNS) (Fig. 6a). Moreover, optimal NCF strategies also 713 714 differed from that identified by the improved model, particular under maxYield targets (Fig. 6b). These results showed important implications of the improved framework for 715 prompting sustainable water management. 716 717







- 724 Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O for China and six rice growing regions. Blue and
- orange indicate results from the origin and modified model, respectively. (b)
- 726 Proportions of rice areas with corresponding optimized lower irrigation thresholds

(*L_{IRR}*) to total irrigated rice areas under the four single objective targets. NES, HHH,
CJS, CJD, HND, and XNS indicate six rice growing areas of China, namely,
Northeast Single rice, HuangHuaiHai single rice, Yangtze River single rice, Yangtze
River double Rice, South China Double rice, and Southwest China Single rice,
respectively.

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733 **3.4 Tradeoffs between food, water, and greenhouse gas emissions**

The NSGA-II algorithm was conducted to investigate synergies or tradeoffs of the food-735 water-climate nexus (Fig. 7 and Sect. 2.6). There were evident tradeoffs between 736 reducing CH₄ (or IRR) and N₂O (Fig. 7a). In contrast, synergies were noted between 737 738 reducing IRR and CH₄, as well as between inhibiting N₂O emissions and increasing rice yield. The relationships between yield increase and CH4 (or IRR) reductions were more 739 complicated due to the impacts of varying irrigation timing and no-flooded days (Yan 740 et al., 2024). Adopting non-dominated solutions from multi-objective optimization 741 could realize over 90% of the largest reduction potentials of IRR and CH₄, while at the 742 cost of 4% less yield increase (4.6% versus 0.5%) and 25% higher nitrous dioxide 743 744 emissions (-11% versus 14%). The N₂O increase is because this study used integrated warming potentials of CH₄ and N₂O emissions (GWP) to indicate greenhouse gas 745 746 emissions so that CH₄ outweighed N₂O due to large emission quantities (Sect. 2.6).

747

Spatially, over 90% of the reduction potentials for IRR and CH₄ could be achieved 748 across 53% and 60% of the national rice areas, primarily in southern regions (Fig. 7 and 749 750 S14). In these areas, N₂O increase was inevitable, but yield increase could be expected. By contrast, stronger tradeoffs occurred in the northern regions, where the reduction 751 potentials of IRR and CH4 were limited even with decreased yield and increased N2O 752 753 emissions. Therefore, NCF adoption should be prioritized in southern regions (e.g. XND, CJD, CJS) to achieve a national optimum balance among rice production, water 754 use, and greenhouse gas emissions mitigation. Noted that other objective functions 755 could also be designed for multi-objective optimization, such as applying other 756 indicators (e.g., water productivity, yield-scaled GWP), setting distinguished weights 757 for each indicator or grid cell. 758



Figure 7 Regulation potentials of Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O under single-761 objective and multi-objective targets. (a) Synergies or tradeoffs between target 762 variables with different solutions of multi-objective optimization. Dots color indicates 763 probability density distributions of variable changes from all non-dominated solutions 764 (N = 10000) of the NSGA II optimization. The vertical and horizontal dashed lines 765 show national regulation potentials of the target variable under single-objective 766 targets, with corresponding spatial distributions presented in panel (b). Note that the 767 results of ΔN_2O potentials (-11%) were not shown in the third, fifth, and sixth 768 subplots for a clearer view. (b) Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O under single-objective 769 targets of maximizing rice yield (maxYield), minimizing irrigation water use 770 (minIRR), minimizing CH₄ emissions (minCH₄), and minimizing N₂O emissions 771 $(maxN_2O)$. These results indicate the maximum benefits of each target variable from 772 adopting non-continuous irrigation, which could not be necessarily realized 773 simultaneously. (c) Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O under multi-objective 774 optimization. These figures show mean benefits from all non-dominated solutions of 775 776 the NSGA II optimization (N = 10000).

777

778 **3.5 Uncertainties and future direction**

779

This framework is subject to several uncertainties, mainly sourced from observational gaps and management-related input data. First, the absence of field observations for baseline CH_4 and N_2O emissions across regional scales forced us to use estimates from inventory or data-driven approaches as a proxy for deriving gridded model parameters

of this study (Cui et al., 2021; Crippa et al., 2024). Despite uncertainties in predicting 784 absolute values, these parameters could reasonably reproduce the spatial patterns and 785 could be further refined given increased field observations. Second, the limited 786 experimental observations of CH₄ (n = 37) and N₂O (n = 14) under various irrigation 787 schemes have contributed to uncertainties in developing and applying parameter 788 789 transfer functions (PTFs). The values of PTFs predictors (bulk density and field water capacity) in the observation dataset $(1.34 \sim 1.48 \text{ g cm}^{-3} \text{ and } 0.25 \sim 0.30 \text{ cm}^{-3})$ did not 790 encompass the full range across national rice areas $(1.24 \sim 1.48 \text{ g cm}^{-3} \text{and } 0.22 \sim 0.32$ 791 cm³ cm⁻³), indicating potential extrapolation in parameters regionalization (Fig. S1). 792 Despite these uncertainties, the PTFs significantly improved over previous approaches 793 (constant parameters or spatial proximity approach). Lastly, current irrigation practices 794 795 across large scales remain largely unknown, so that irrigation thresholds were set following previous recommendations. However, actual farmer practices are influenced 796 by various factors and may not align with these recommendations. This discrepancy 797 could lead to an overestimation or underestimation of target variables and further 798 introduce uncertainties to the assessment of regulation potentials. 799

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801 These uncertainties provide insights to enlighten future research efforts, including conducting extensive observations and experiments and developing high-resolution 802 803 input data. On the one hand, intensive GHGs monitoring networks are essential to reduce uncertainties associated with parametrization (Arenas-Calle et al., 2024). To 804 better constrain the PTFs and reduce extrapolation uncertainty, field experiments 805 combined with incubation experiments across a broader range of climate conditions 806 807 (e.g., colder and more humid areas) and soil properties (e.g., areas with higher SOC or lower bulk density) should be conducted (Fig. S1). In addition, extensive field 808 experiments with simultaneous measurements of yield, IRR, CH₄, and N₂O emissions 809 across diverse environments are required to validate the framework further. On the other 810 hand, developing a high-resolution dataset of current irrigation schemes is crucial for 811 more accurate model parameter calibration and realistic assessment of regulation 812 potentials. This could be achieved by integrating remote sensing technologies with 813 extensive field investigations (Novick et al., 2022). 814

815

816 **4 Conclusion**

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This study introduced an advancing framework for process-based modelling of the 818 819 complex food-water-climate nexus in rice fields under various water management 820 schemes. By integrating the Soil Water Heat Carbon Nitrogen Simulator (WHCNS) with key physiological effects, a novel model upscaling method, and the NSGA-II 821 multi-objective optimization algorithm at a parallel computing platform, the framework 822 provides a comprehensive approach to optimize irrigation strategies. Applying this 823 framework to China's rice cropping system, we assessed the largest regulation 824 potentials of Δ Yield, Δ IRR, Δ CH₄, and Δ N₂O as 4.6%, -61.0%, -64.2%, and -10.9% 825 from 91%, 91%, 88%, and 26% of national rice areas. However, these regulation 826 potentials could not be simultaneously realized due to complicated tradeoffs among 827

food-water-GHGs. Based on NSGA II multi-objective optimization targeting food-828 water-GHGs co-benefits, over 90% of the reduction potentials in water use and methane 829 emissions could be realized, while at the cost of 4% less yield increase and 25% higher 830 nitrous dioxide emissions. The proposed framework is a valuable tool for irrigation 831 optimization in rice cultivation and also offers a transferable paradigm for incorporating 832 833 other management effects into process-based models, thus supporting comprehensive assessments of sustainable management measures. 834

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Appendix A: Abbreviation table

Туре	Abbreviation	Description
Target variables	Yield	Rice yield (kg ha ⁻¹)
	IRR	Irrigation water use (mm)
	CH_4	Methane emissions (kg ha ⁻¹)
	N_2O	Nitrous oxide emissions (kg ha ⁻¹)
	GWP	Integrated global warming potential of CH ₄ and
		N ₂ O at 100-yearr scale, calculated as 27.2×CH ₄
		$+273 \times N_2O \ (kg \ ha^{-1})$
	LAI	Leaf area index $(m^2 m^{-2})$
	Pn	Net photosynthetic rate (kg ha ⁻¹)
	HI	Harvest index (-)
Effect sizes	\mathbf{p} Yield \mathbf{p} IRR	Effect size of non-continuous flooding irrigation
	к,к, рСН4 рN20	(NCF) on target variables, calculated as the ratio
	R^{LAI}, R^{Pn}, R^{HI}	of observations under NCF to that under
		continuous flooding (CF) (-)
Relative	$\Delta Yield, \Delta IRR,$	Relative changes of target variables under NCF
changes	$\Delta CH_4, \Delta N_2O$	compared to CF, calculated as (R-1)×100 (%)
	Cumtemp	accumulated temperature for crop maturity (°C)
	AMIN	Minimum assimilation rates (kg $hm^{-2} h^{-1}$)
	P^{LAI}, P^{Pn}, P^{HI}	Genetic parameters accounting for cultivar
Model parameters		sensitivities to NCF effects on leaf area index,
		net photosynthetic rate, and harvest index
	MPmax	Maximum CH4 production rate per soil weight
		at 30 °C (g C $g^{-1} d^{-1}$)
	f _{N2O_d}	Maximum portion of denitrification to N2O
		production (-)
Environmental variables	Т	Mean daily air temperature during rice growing
		season (°C)
	Р	Total precipitation during rice growing season
		(mm)
	РЕТс	Total crop evapotranspiration during rice
		growing season (mm)
	CWA	Climatological water availability, calculated as

		the difference between <i>P</i> and <i>PETc</i> (<i>P-PETc</i> ,
		mm)
Soil variables	BD	Bulk density (g cm ⁻³)
	Sand	Sand content (%)
	Clay	Clay content (%)
	SOC	Soil organic carbon (%)
	SAT	Saturated water content (cm ³ cm ⁻³)
	FWC	Field water capacity (cm ³ cm ⁻³)
	L_{AWD}	Lower irrigation threshold, indicated by SWP
		(kpa)
Management	U_{AWD}	Upper irrigation threshold (cm)
variables	SWP	Soil water potential (kpa)
	UFR	Ratio of unflooded days to total growing days
		(%)
	maxYield	Maximizing rice yield
Optimization	minIRR	Minimizing irrigation water use
objectives	minCH ₄	Minimizing CH ₄ emission
	minN ₂ O	Minimizing N ₂ O emissions

839 Code and data availability

- 840 The origin code of WHCNS model and required model input files are available at
- 841 <u>https://figshare.com/s/139f3ad8a70faa99724d</u>. Spatial dataset of harvested area of
- 842 irrigated rice is available from <u>https://doi.org/10.7910/DVN/KAGRFI</u>. Origin climate
- 843 data is available from https://cds.climate.copernicus.eu/datasets/reanalysis-era5-
- 844 <u>single-levels?tab=download</u>. Origin soil data is available from
- https://doi.org/10.1002/2013MS000293. Processed climate and soil data for model
- running are included in the figshare repository (see Readme for detailed explanationsof each file). Crop calendar data are available from
- https://zenodo.org/record/5062513. All other data that support the findings of this
- study are available in the main text or the Supplementary Information.
- 850

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855 Author contributions

- 856 F.Z. designed the study. Y.B. and H.L. performed all computational analyses. Y.B.,
- H.L. and F.Z. drafted the paper. Y.B., H.L., T. L. and F.Z. reviewed and commented onthe manuscript.
- 858 the 859

860 **Competing interests**

- 861 The authors declare no conflicts of interest.
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863 **References**

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