



1 **FLAML version 2.3.3 model-based assessment of gross**  
2 **primary productivity at forest, grassland, and cropland**  
3 **ecosystem sites**

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14 **Abstract**

15 Accurately estimating Gross Primary Productivity (GPP) in terrestrial ecosystems is  
16 essential for understanding the global carbon cycle. Satellite-based Light Use  
17 Efficiency (LUE) models are commonly employed for simulating GPP. However, the  
18 variables and algorithms related to environmental limiting factors differ significantly  
19 across various LUE models. In this work, we developed a series of FLAML-LUE  
20 models tailored for different ecosystems. These models utilize the Fast Lightweight  
21 Automated Machine Learning (FLAML) framework, using variables of LUE models,  
22 to investigate the potential of estimating site-scale GPP. Incorporating meteorological  
23 data, eddy covariance measurements, and remote sensing indices, we employed  
24 FLAML-LUE models to assess the impact of various variable combinations on GPP  
25 across different temporal scales, including daily, 8-day, 16-day, and monthly intervals.  
26 Cross-validation analyses indicated that the effectiveness of FLAML-LUE models for  
27 forest ecosystems varied significantly across different sites, with  $R^2$  values ranging



28 from 0.56 to 0.94. For grassland ecosystems,  $R^2$  values ranged from 0.62 to 0.87, and  
29 for cropland ecosystems,  $R^2$  values ranged from 0.78 to 0.88. Extending the time scale  
30 of input data could significantly enhance the accuracy of model simulations.  
31 Specifically, the average  $R^2$  increased from 0.82 to 0.92 for forest ecosystems, 0.79 to  
32 0.83 for grassland ecosystems, and 0.84 to 0.87 for farmland ecosystems. Additionally,  
33 the importance ranking method indicated that vegetation index and temperature were  
34 the most important variables for GPP estimation in forest, grassland, and farmland  
35 ecosystems, while the importance of the moisture index was relatively low. This study  
36 offers an approach to estimate GPP fluxes and evaluate the impact of variables on GPP  
37 estimation. It has the potential to be applied in predicting GPP for different vegetation  
38 types at a regional scale.

39 **Keywords:** Light Use Efficiency; Gross Primary Productivity; Automated Machine  
40 Learning; Fast Lightweight Automated Machine Learning

## 41 **1. Introduction**

42 The global carbon budget mainly addresses the carbon reserves in the atmosphere,  
43 oceans, and terrestrial (Barbour, 2021), with terrestrial ecosystems being vital for  
44 regulating the global carbon cycle (Gherardi and Sala, 2020; Landry and Matthews,  
45 2016). Terrestrial ecosystems primarily absorb atmospheric carbon dioxide through the  
46 process of plant photosynthesis, which is crucial for regulating climate and mitigating  
47 global warming (Sellers et al., 2018; Beer et al., 2010; Cox et al., 2000). Gross primary  
48 productivity (GPP) is a critical measure of carbon exchange between terrestrial  
49 ecosystems and the atmosphere. (Menefee et al., 2023). Accurate quantification of GPP



50 is essential for evaluating carbon balance and comprehending the response of terrestrial  
51 ecosystems to climate change (Sellers et al., 2018).

52 The primary method currently used for measuring CO<sub>2</sub> exchange between  
53 ecosystems and the atmosphere is the eddy covariance technique (Chen et al., 2020; Yu  
54 et al., 2016). This technique precisely measures Net Ecosystem Exchange (NEE), which  
55 is the difference between the carbon released by ecosystem respiration (ER) and the  
56 carbon taken up by photosynthesis (Bhattacharyya et al., 2013). While flux observation  
57 sites based on the eddy covariance technique can dynamically monitor site-scale carbon  
58 fluxes, expanding their findings to larger regional scales remains challenging, mainly  
59 due to the sparse and spatially non-uniform distribution of flux sites (Xie et al., 2023;  
60 Jung et al., 2020). Remote sensing data is widely used in ecosystem carbon cycle  
61 research as it can provide information on spatial dynamics of vegetation and climate at  
62 a larger scale (Xiao et al., 2019). By extrapolating spatially using models that  
63 incorporate remote sensing and climate data, it is possible to estimate global GPP based  
64 on observations of GPP at the site level. Therefore, remote sensing has become a crucial  
65 data resource for estimating GPP (Cai et al., 2021; Xiao et al., 2019; Wang et al., 2011).

66 Light Use Efficiency (LUE) models based on satellite observations are commonly  
67 employed to simulate GPP. (Zhang et al., 2023; Zhang et al., 2015; Jiang et al., 2014).  
68 Such models include Physiological Principles Predicting Growth using Satellite data  
69 (3-PGS, Coops and Waring, 2001), the Carnegie-Ames- Stanford Approach Model  
70 (CASA, Potter et al., 1993), the Eddy Covariance–Light Use Efficiency Model (EC-  
71 LUE, Yuan et al., 2010, 2007), the MODIS Global Terrestrial Gross and Net Primary



72 Production (MOD17, Running et al., 2004), the Vegetation Photosynthesis Model  
73 (VPM, Xiao et al., 2003), and the Vegetation Photosynthesis and Respiration Model  
74 (VPRM, Mahadevan et al., 2008). Among all the forecasting methods (Coops and  
75 Waring, 2001; Potter et al., 1993), the LUE model is widely utilized for simulating the  
76 spatio-temporal dynamics of GPP due to its simplicity and strong theoretical foundation.  
77 Over the past few decades, numerous GPP models utilizing LUE have been developed  
78 (Pei et al., 2022).

79 Despite significant advances in LUE theory for GPP estimation, uncertainties  
80 persist in GPP models utilizing LUE. Firstly, differences in environmental limiting  
81 factors among various LUE models contribute significantly to the uncertainty in GPP  
82 estimation. For example, Cai et al. (2014) found a strong positive correlation between  
83 water effectiveness and GPP estimate factors, while other studies found that the LUE  
84 model estimates of GPP were strongly correlated with the vegetation index, which  
85 affects the photosynthetic capacity of vegetation through leaf nitrogen content  
86 (Peltoniemi et al., 2012; Ercoli, 1993).

87 Recently, with the massive accumulation of satellite data and ground-based  
88 observations, more and more studies have applied machine learning (ML) methods to  
89 model ecosystem processes (Zhao et al., 2019; Alemohammad et al., 2017; Chaney et  
90 al., 2016). ML is a modeling solution that is fundamentally different from simple  
91 regression models and complex simulation models. It is very effective in handling  
92 large-scale multivariate data with complex relationships between predictors (Reichstein  
93 et al., 2019; Tramontana et al., 2016). These data-driven ML models are well-suited for



94 addressing nonlinear and complex issues across different ecosystems. They provide  
95 innovative approaches for simulating GPP by solving the nonlinear relationship. These  
96 models are less reliant on theoretical assumptions. Therefore, many researchers prefer  
97 this method in recent years. Kong et al. (2023) developed a hybrid model that combines  
98 ML and LUE model to estimate GPP. This hybrid model improves the LUE model by  
99 integrating a machine learning approach (MLP, multi-layer perceptron), and estimates  
100 GPP using the MLP-based LUE framework along with additional required inputs.  
101 Chang et al. (2023) constructed RFR-LUE models that utilize the RFR algorithm with  
102 variables of LUE models to assess the potential of site-scale GPP estimation.

103 Lately, Automated Machine Learning (AutoML) has demonstrated significant  
104 potential in constructing data-driven models automatically (C. Zhang et al., 2023;  
105 Zheng et al., 2023). Numerous sophisticated open-source AutoML frameworks have  
106 been suggested by computer scientists, including AutoWeka (Thornton et al., 2013),  
107 H2O (LeDell and Poirier, 2020), TPOT (Melanie, 2023), AutoGluo (Erickson et al.,  
108 2020), FLAML (C. Wang et al., 2021), and AutoKera (Rosebrock, 2019). These  
109 frameworks are extensively used in finance, manufacturing, healthcare, and mobile  
110 communications, among other fields (Adams et al., 2020), with FLAML being  
111 particularly favored for its efficiency in rapid prototyping and deployment in research  
112 and production settings. FLAML (Fast Lightweight Automated Machine Learning) is a  
113 powerful framework for AutoML, known for its speed in identifying top-performing  
114 models and optimal hyperparameters through parallel optimization and smart search  
115 algorithms. FLAML integrates several effective search strategies, outperforming other



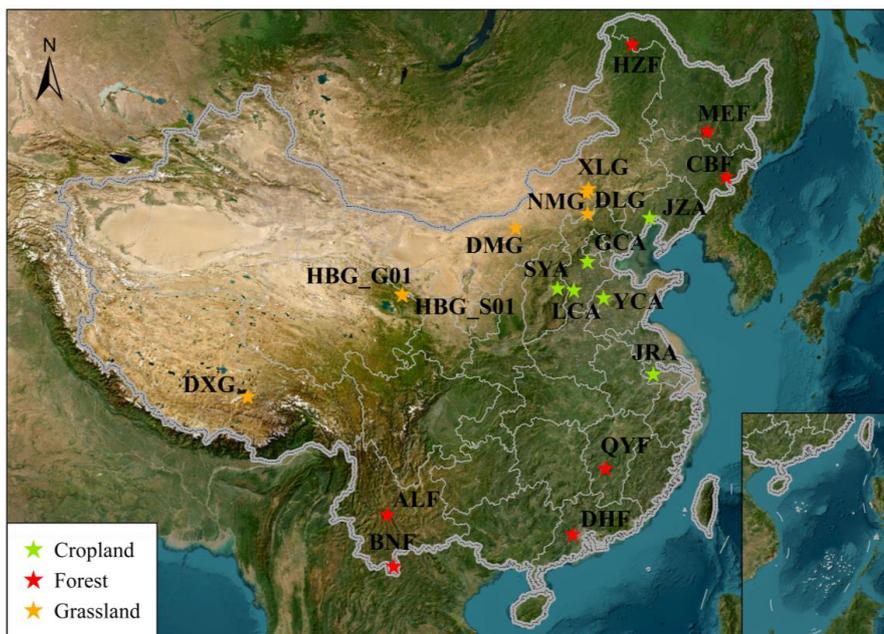
116 leading AutoML libraries on large benchmarks even with constrained budgets(C. Wang  
117 et al., 2021).

118 In this research, a new model called FLAML-LUE was created by combining  
119 FLAML model with LUE-based models, the latter provides the key variables of  
120 vegetation growth for modeling. Such knowledge-and-data-driven models aim to  
121 reduce the large uncertainty in estimating GPP. Considering the variations of the  
122 optimal moisture factor and vegetation index factor for different ecosystems (Wang et  
123 al., 2023; Wu et al., 2010), this study thus develops different models specifically for  
124 forest, grassland, and cropland ecosystems. The main goals of this study were (1) to  
125 compare the overall performance of the models simulating GPP with different input  
126 variables (moisture factor and vegetation index) and at four temporal scales; (2) to  
127 analyze monthly differences between observed and simulated values in different cover  
128 types; (3) to analyze the importance of the various input indicators for GPP modeling  
129 under different ecosystems.

## 130 **2. Material and methods**

### 131 **2.1 Site description**

132 Fig. 1 displays the geographical locations of the 20 flux sites selected for the study.  
133 These sites are situated in various climatic zones and ecosystem types including forest,  
134 grassland, and cropland. The observation data for these sites comes from the Science  
135 Data Bank (SDB, <https://www.scidb.cn/en/>). Detailed information about the sites is  
136 provided in Table 1.



137  
 138 **Fig. 1.** The location map of the flux site is based on the map approved by the National Surveying  
 139 and Mapping Bureau of China (Approval No. GS (2019)1822). The topographic map is derived  
 140 from data provided by Esri, Maxar, Earthstar Geographics, and the GIS User Community (Service  
 141 Layer Credits).

142 **Table1**

143 Basic information on the 20 flux stations.

Site	Ecosystem type	Surface cover type	Time Range	Classified
HZF	Forest	Coniferous forest	2014-2018	Needle-leaved
MEF	Forest	Deciduous broad-leaved forests	2016-2018	Deciduous Broadleaved
CBF	Forest	Broad-leaved Korean pine forests	2003-2010	Mixed
QYF	Forest	Artificial coniferous forests	2003-2010	Needle-leaved
DHF	Forest	Mixed coniferous and broad-leaved forests	2003-2010	Mixed
ALF	Forest	Evergreen Broadleaved forests	2009-2013	Evergreen Broadleaved
BNF	Forest	Tropical rainforest	2003-2015	Evergreen Broadleaved
XLG	Grassland	Mowing grasslands	2006-2014	Grassland
NMG	Grassland	Temperate steppe	2003-2010	Grassland
DLG	Grassland	Typical grasslands	2006-2015	Grassland
DMG	Grassland	Temperate desert steppe	2015-2018	Grassland
HBG_G01	Grassland	Alpine meadow	2015-2020	Alpine Meadow
HBG_S01	Grassland	Alpine shrub-meadow	2003-2013	Shrub
DXG	Grassland	Alpine meadow	2003-2010	Alpine Meadow
JZA	Cropland	Spring corn	2005-2014	Single Cropping
GCA	Cropland	Winter wheat - Summer corn	2020-2022	Double Cropping



SYA	Cropland	Spring corn	2012-2014	Single Cropping
LCA	Cropland	Winter wheat - Summer corn	2013-2017	Double Cropping
YCA	Cropland	Winter wheat - Summer corn	2003-2010	Double Cropping
JRA	Cropland	Winter wheat - Summer rice	2015-2020	Double Cropping

## 144 2.2 Data

### 145 2.2.1 Eddy covariance data

146 Eddy covariance (EC) data were collected at 20 sites, including 7 forests, 7  
147 grasslands, and 6 cropland (Table 1). Back third of long-time series data from ALF,  
148 CBF, and QYF Stations data were used for forest model validation, and in the same  
149 way, a third of DLG, DXG, and HBG Stations data were used for grassland models  
150 validation, a third of JZA and YCA Stations data were used for cropland models  
151 validation. None of the validation data were involved in the model training process.

152 Flux and meteorological data were collected every half hour from the mentioned  
153 sites. The flux and meteorological data underwent standardized quality control and  
154 corrections, ensuring high reliability and making them suitable for validating various  
155 GPP models and remote sensing observations. However, some sites have no ER data,  
156 so this study is based on the nocturnal breathing extrapolation method: Lloyd & Taylor  
157 equation (Reichstein et al., 2005; Lloyd and Taylor, 1994). The shortwave radiation  $R_g$   
158 values ( $10\text{W/m}^2$ ) determined the separation of daytime and nighttime data. In this study,  
159 the response function established by the temperature of nocturnal ER data was extended  
160 to the daytime to obtain the daytime ER.

$$161 \quad R_{eco} = R_{eco.ref} \exp \left( E_0 \left( \frac{1}{T_{ref} - T_0} - \frac{1}{T_{air} - T_0} \right) \right) \quad (1)$$

162 In the above equation,  $R_{eco}$  is the nocturnal ecosystem respiration value,  $R_{eco.ref}$  is the  
163 ER value at the reference temperature,  $T_{ref}$  is the reference temperature (298.16K),  $E_0$



164 is constant (308.56K),  $T_0$  is the minimum temperature at which respiration stops, set at  
165 227.13K, and  $T_{\text{air}}$  is the air temperature or soil temperature (K).

166 We can then estimate the total ecosystem productivity of the ecosystem during the  
167 day by subtracting the net ecosystem exchange from the total ER during the day.

$$168 \quad GPP = ER - NEE \quad (2)$$

169 In the above equation, GPP represents the carbon uptake by plants during  
170 photosynthesis. ER denotes  $\text{CO}_2$  released through ecosystem respiration from  
171 aboveground plant parts, roots, and soil, occurring both day and night. NEE reflects the  
172 net carbon gain or loss within the ecosystem.

173 The pre-processed flux data are first aggregated into daily, 8-day, 16-day, and  
174 monthly intervals. Then, daily values are further aggregated to 8-day, 16-day, and  
175 monthly resolutions applying suitable methods. A detailed flow illustrating the  
176 processing of each variable is shown in Fig. 2.

### 177 **2.2.2 Remote sensing data**

178 In this study, remote sensing data primarily came from MODIS and ERA5-LAND.  
179 MODIS data offer a spatial resolution of 500 meters and an 8-day temporal resolution,  
180 while ERA5-LAND data have a spatial resolution of  $0.1^\circ$  and a daily temporal  
181 resolution. These datasets were sourced from the Google Earth Engine (GEE) platform  
182 (Gorelick et al., 2017). To align with the spatial and temporal scales of flux tower  
183 observations and reduce the impact of missing data (Schmid, 2002), we applied the  
184 Savitzky-Golay smoothing filter with a window size of 10 to process the vegetation  
185 indices. MODIS data from GEE were used to derive vegetation and water indices,

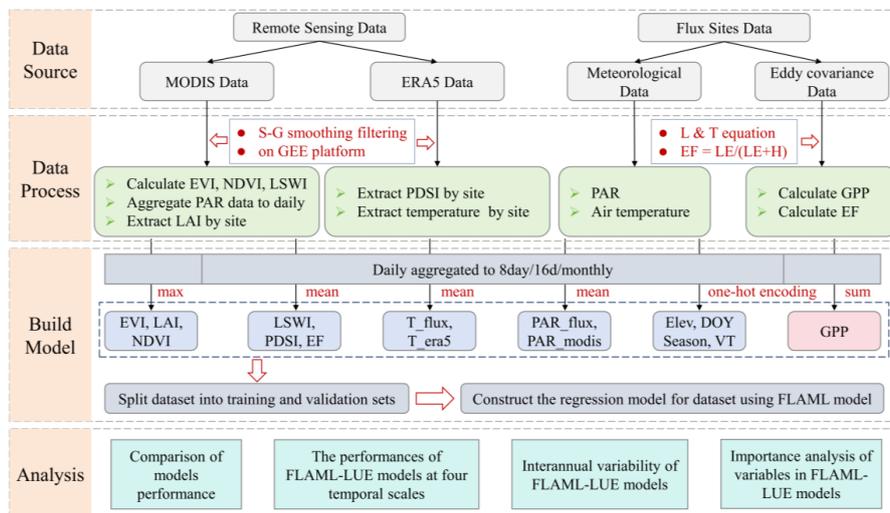


186 including EVI, NDVI, LAI, and LSWI, which were calculated using the formulas in  
 187 Table 2. Temperature and PDSI index data were obtained from the ERA5-LAND  
 188 product. The Maximum Value Composite (MVC) method was used to aggregate multi-  
 189 temporal vegetation indices (VIs), ensuring alignment with the model simulation time  
 190 steps.

### 191 2.3 Model Construction

192 Most LUE models usually have four groups of variables: PAR, VIs, temperature,  
 193 and water. In past studies, NDVI, EVI, or LAI were used as indicators of the proportion  
 194 of PAR absorbed. In addition, different moisture indices were added to the LUE model  
 195 to account for water stress, including LSWI, Palmer drought severity index (PDSI), and  
 196 evapotranspiration fraction (EF) indicators. In this study, all above-mentioned variables  
 197 were used to build the LUE model.

198 The flowchart of this study is shown in Fig. 2.



199 **Fig. 2.** Flowchart of this study. S-G smoothing filtering: Savitzky-Golay smoothing filtering method,  
 200 L & T equation: Lloyd & Taylor equation.  
 201



### 202 2.3.1 Data pre-processing

203 The primary datasets for estimating GPP with FLAML-LUE models include multi-  
 204 year continuous EC flux data, satellite-based observations, and climate data. Prior  
 205 research (Jung et al., 2011) has demonstrated notable seasonal fluctuations in GPP, we  
 206 divided the time series data into four distinct seasons. Additionally, we incorporate the  
 207 day of year (DOY) indicator into the model. Research has demonstrated that topography  
 208 significantly influences GPP modeling (Xie and Li, 2020). Therefore, we include  
 209 elevation as an additional variable. Moreover, the vegetation cover type, which varies  
 210 across different ecosystems, greatly impacts the accuracy of GPP simulation (Chang et  
 211 al., 2023). Hence, we integrate vegetation type as a factor in our model.

212 **Table 2**

213 Predictor variables for driving the FLAML models and their specifications.

	Variable	Acquired method (formula)	Original Spatial Resolution	Data Source
Vegetation indices	EVI	$2.5 \times (R_{nir} - R_{red}) / (R_{nir} + 6.0 \times R_{red} - 7.5 \times R_{blue} + 1)$	500m	MOD09GA
	NDVI	$(R_{nir} - R_{red}) / (R_{nir} + R_{red})$		
	LAI	-	500m	MCD15A3H
Water	LSWI	$(R_{nir} - R_{swir}) / (R_{nir} + R_{swir})$	500m	MOD09GA
	PDSI	-	~10km	ERA5
	EF (%)	$LE / (LE + H)$	~1km	SDB
Radiation	PAR ( $\mu \text{ mol m}^{-2} \text{ s}^{-1}$ )	-	~1km	SDB
	PAR ( $\mu \text{ mol m}^{-2} \text{ s}^{-1}$ )	-	500m	MCD18C2
Temperature	T_flux ( $^{\circ}\text{C}$ )	-	~1km	SDB
	T_era5 ( $^{\circ}\text{C}$ )	-	~10km	ERA5
Vegetation Types	EBF, DBF, CF, MF Grassland Croplands	One-hot encoding	invariant	-
Season	Spring, Summer, Autumn, Winter	One-hot encoding	invariant	-
DOY	Days of year	-	invariant	-
Terrain	Elevation	-	90m	SRTM90

### 214 2.3.2 Automated Machine Learning (AutoML)

215 Instead of applying a specific ML method like RF for building regression models,



216 we utilize the lightweight Python library “FLAML” version 2.3.3 (C. Wang et al., 2021)  
217 for the AutoML task. This library refines the search process by balancing computational  
218 cost and model error, and it iteratively selects the learner, hyperparameters, sample size,  
219 and resampling strategy (C. Wang et al., 2021). For our modeling approach, we set up  
220 the AutoML for regression tasks using the "auto" option for the estimator list, focused  
221 on optimizing the  $R^2$  metric, and used a time step of 120 seconds (2 minutes) for each  
222 AutoML run. The "auto" option includes a range of tree-based methods, such as  
223 LightGBM (Ke et al., 2017), XGBoost (Chen and Guestrin, 2016), CatBoost  
224 (Prokhorenkova et al., 2018), RF (Breiman, 2001), and Extra-Trees (Geurts et al., 2006).

### 225 **2.3.3 Model development**

226 Eighteen FLAML-LUE model variations were constructed for each site and time  
227 scale, using multiple permutations of eight input factor groups, as described in Eq. (3).  
228 Table 3 displays the model number based on different variable configurations.

$$229 \quad GPP = f(PAR, T, VI_i, W_j, VT, Season, DOY, Elevation) \quad (3)$$

230 Here, the  $VI_i$  include EVI, NDVI, and LAI;  $W_j$  denotes moisture factors including  
231 LSWI, EF, and PDSI;  $VT_i$  represents vegetation types, in which forest ecosystems  
232 include: Needle-leaved, Deciduous Broadleaved, Mixed, and Evergreen Broadleaved;  
233 Grassland ecosystems include grasslands, meadows and shrub, and farmland  
234 ecosystems include single cropping and double cropping. Season represents the  
235 season in which the original data were acquired. DOY represents the days of the year.

236 Each ecosystem has 18 indicator combinations, which are divided into two groups  
237 based on different data sources, the FLAML00-FLAML08 combination uses the



238 ground-based observations as the input data, and the FLAML10-FLAML18  
 239 combination uses remote sensing data as the main input data.

240 **Table 3**

241 Input data for different models

Group (Flux)	Input variables	Group (RS)	Input variables
FLAML00	PAR, T_flux, EVI, LSWI, Season, DOY, Elevation, Vegetation Types	FLAML10	PAR_modis, T_era5, EVI, LSWI, Season, DOY, Elevation, Vegetation Types
FLAML01	PAR, T_flux, EVI, PDSI, Season, DOY, Elevation, Vegetation Types	FLAML11	PAR_modis, T_era5, EVI, PDSI, Season, DOY, Elevation, Vegetation Types
FLAML02	PAR, T_flux, EVI, EF, Season, DOY, Elevation, Vegetation Types	FLAML12	PAR_modis, T_era5, EVI, EF, Season, DOY, Elevation, Vegetation Types
FLAML03	PAR, T_flux, NDVI, LSWI, Season, DOY, Elevation, Vegetation Types	FLAML13	PAR_modis, T_era5, NDVI, LSWI, Season, DOY, Elevation, Vegetation Types
FLAML04	PAR, T_flux, NDVI, PDSI, Season, DOY, Elevation, Vegetation Types	FLAML14	PAR_modis, T_era5, NDVI, PDSI, Season, DOY, Elevation, Vegetation Types
FLAML05	PAR, T_flux, NDVI, EF, Season, DOY, Elevation, Vegetation Types	FLAML15	PAR_modis, T_era5, NDVI, EF, Season, DOY, Elevation, Vegetation Types
FLAML06	PAR, T_flux, LAI, LSWI, Season, DOY, Elevation, Vegetation Types	FLAML16	PAR_modis, T_era5, LAI, LSWI, Season, DOY, Elevation, Vegetation Types
FLAML07	PAR, T_flux, LAI, PDSI, Season, DOY, Elevation, Vegetation Types	FLAML17	PAR_modis, T_era5, LAI, PDSI, Season, DOY, Elevation, Vegetation Types
FLAML08	PAR, T_flux, LAI, EF, Season, DOY, Elevation, Vegetation Types	FLAML18	PAR_modis, T_era5, LAI, EF, Season, DOY, Elevation, Vegetation Types

242 **2.3.4 Model performance evaluation methods**

243 Model performance in this study was assessed in two main ways. We assessed the  
 244 ability of the FLAML-LUE model to capture changes in GPP at different sites and time  
 245 scales (daily, 8-day, 16-day, monthly), as well as its representativeness of interannual  
 246 changes in GPP. We compared model-derived annual average GPP to EC-GPP  
 247 measurements at each site and scale and analyzed standard deviations to measure the  
 248 model's ability to capture the magnitude of change. Performance metrics included  
 249 coefficient of determination ( $R^2$ ), root mean square error (RMSE), mean bias, and  
 250 regression slope between simulated and observed values. Paired t-tests were used to  
 251 determine whether the differences in performance between different temporal  
 252 resolutions were statistically significant, with a significance level of 0.05. Statistical



253 analyses were performed in Python 3.9 using the following libraries: numpy, pandas,  
254 scipy, matplotlib, sklearn, and flaml. Additionally, R was used with the following  
255 libraries: ggplot2, ggpubr, and readxl.

$$256 \quad R = \frac{\frac{1}{T} \sum_{t=1}^T (f_t - \bar{f})(o_t - \bar{o})}{\sigma_f \sigma_o} \quad (4)$$

$$257 \quad nuRMSE = \frac{uRMSE}{\sigma_o} = \frac{1}{\sigma_o} \sqrt{\frac{1}{T} \sum_{t=1}^T [(f_t - \bar{f}) - (o_t - \bar{o})]^2} \quad (5)$$

$$258 \quad \sigma_f = \frac{\sigma_f}{\sigma_o} = \frac{1}{\sigma_o} \sqrt{\frac{1}{T} \sum_{t=1}^T ((f_t - \bar{f}))^2} \quad (6)$$

$$259 \quad \sigma_o = \sqrt{\frac{1}{T} \sum_{t=1}^T ((o_t - \bar{o}))^2} \quad (7)$$

260 The Taylor diagram (Taylor, 2001) visually represents the alignment between  
261 model simulations and observations by displaying the correlation coefficient (R),  
262 normalized unbiased root mean square error (nuRMSE), and normalized standard  
263 deviation (SD). The Taylor Skill Score (TSS) quantifies how closely a model's  
264 simulation aligns with observations in this diagram. It is defined as follows:

$$265 \quad TSS = \frac{4(1+R)}{\left(\hat{\sigma}_f + \frac{1}{\hat{\sigma}_f}\right)^2 (1+R_0)} \quad (8)$$

$$266 \quad \hat{\sigma}_f = \frac{\sigma_f}{\sigma_o} \quad (9)$$

267 Where  $\sigma_f$  and  $\sigma_o$  represent the standard deviations of the model simulation and  
268 observations, respectively, and  $R_0$  denotes the maximum possible correlation  
269 coefficient (in this study,  $R_0 = 1$ ). The TSS ranges from 0 to 1, with a higher TSS  
270 indicating better overall model performance relative to the observations.

### 271 2.3.5 Feature Importance Analysis

272 In Data Science, “feature importance” scores indicate how useful a feature is in  
273 predicting the target variable. These scores differ depending on the learning algorithm,



274 resulting in varying magnitudes. For instance, Extra-Trees assesses feature importance  
275 by the reduction in mean squared error, LightGBM by the frequency a feature is used  
276 in tree splits, and XGBoost by the average information gain from splits. However,  
277 model interpretability remains a complex challenge, and there is no consensus on the  
278 best technique for determining the significance of features. Shapley Additive  
279 exPlanations (SHAP, Lundberg and Lee, 2017) provide a unified approach for model  
280 interpretation. However, their assumption that ML predictions can be broken down into  
281 individual feature contributions may not apply to highly nonlinear models (Gosiewska  
282 and Biecek, 2019). Thus, we use the default feature importance metrics from the  
283 AutoML-selected algorithm, as they are widely accepted by researchers in the field.

284 Then, we introduce a “ranking score” metric to standardize feature importance  
285 comparisons across various algorithms. For each estimator, features are ranked from  
286 least to most important and assigned a score accordingly: the least important feature  
287 gets a score of 1, the next gets 2, and so on. This approach normalizes feature  
288 importance across different models, providing a unified scale for comparison, ranging  
289 from 1 (least important) to the total number of features (most important).

### 290 **3. Results**

#### 291 **3.1 Overall FLAML models performances on forest ecosystem**

##### 292 **3.1.1 Performance Evaluation of Models**

293 To examine the performance of each model in forest ecosystems and at the site  
294 level, the accuracy of the 18 FLAML-LUE models was evaluated using the site data  
295 from ALF, CBF, and QYF stations as the forest ecosystems model test set. The



296 algorithms adopted by each FLAML-LUE model under the forest ecosystems are  
 297 shown in Table S1. Table 4 shows the  $R^2$ , RMSE, and SD of the 18 FLAML-LUE  
 298 models in the forest station test set. Cross-validation analysis shows that there are few  
 299 differences between FLAML-LUE models under different combinations of input data.

300 **Table 4**

301  $R^2$ , SD, RMSE for the forest ecosystems model test set.

FLAML	$R^2$	SD	RMSE	nuRMSE	TSS
FLAML00	0.90	0.864	0.974	0.311	0.9552
FLAML01	0.88	0.832	1.056	0.338	0.9412
FLAML02	0.88	0.838	1.047	0.335	0.9431
FLAML03	0.89	0.882	1.033	0.330	0.9522
FLAML04	0.89	0.888	1.027	0.330	0.9558
FLAML05	0.88	0.875	1.049	0.335	0.9521
FLAML06	0.89	0.878	1.000	0.320	0.9550
FLAML07	0.89	0.881	1.019	0.326	0.9553
FLAML08	0.89	0.875	1.022	0.327	0.9544
FLAML10	0.89	0.896	0.997	0.319	0.9606
FLAML11	0.88	0.861	1.070	0.343	0.9491
FLAML12	0.87	0.871	1.096	0.351	0.9483
FLAML13	0.88	0.876	1.053	0.337	0.9532
FLAML14	0.88	0.885	1.093	0.351	0.9528
FLAML15	0.87	0.880	1.130	0.362	0.9476
FLAML16	0.88	0.880	1.049	0.335	0.9531
FLAML17	0.89	0.964	1.015	0.325	0.9710
FLAML18	0.87	0.898	1.099	0.352	0.9551
Flux(average)	0.89	0.868	1.025	0.328	
ERA5(average)	0.88	0.890	1.067	0.342	
Forest(average)	0.88	0.879	1.046	0.335	

302 As shown in Table 4, the cross-validation analysis showed that the average  $R^2$  for  
 303 the four temporal scales under forest ecosystems was 0.82-0.93. There was little  
 304 difference in performance between the models driven with flux data (FLAML00 -  
 305 FLAML08,  $R^2 = 0.89$ , RMSE = 1.025  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) and the models driven with ERA5



306 (FLAML10 - FLAML18,  $R^2 = 0.88$ ,  $RMSE = 1.067 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ). However, the models  
307 driven using EVI ( $R^2 = 0.89$ ,  $RMSE = 1.040 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) performed slightly better than  
308 NDVI ( $R^2 = 0.88$ ,  $RMSE = 1.064 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) and LAI ( $R^2 = 0.89$ ,  $RMSE = 1.034 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ). The model driven with LSWI ( $R^2 = 0.89$ ,  $RMSE = 1.018 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) performed  
309 slightly better than PDSI ( $R^2 = 0.89$ ,  $RMSE = 1.047 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) and EF ( $R^2 = 0.88$ ,  
310  $RMSE = 1.074 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ).

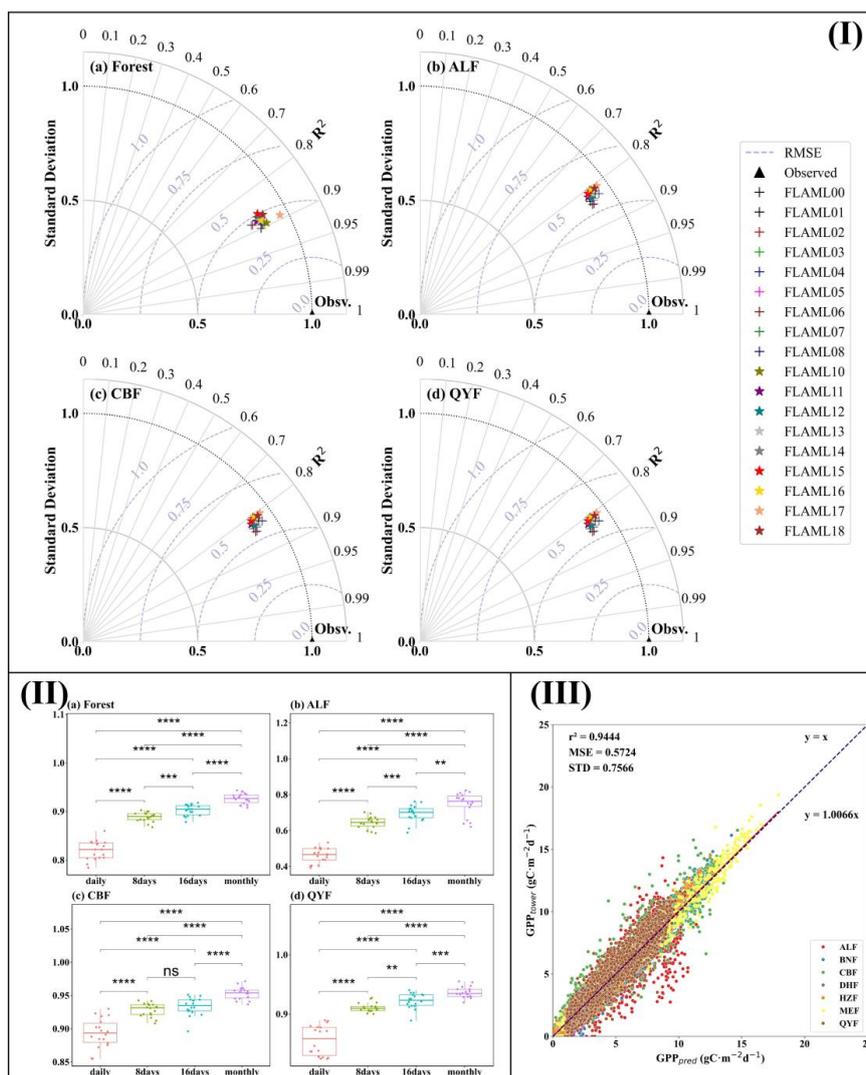
312 Fig. 3 shows the Taylor diagrams of the performance of all FLAML-LUE models  
313 in three forest sites: ALF, CBF, and QYF. The  $R^2$ , nuRMSE, and SD of different  
314 combinations of variables under forest ecosystems were slightly different, and the TSS  
315 values ranged from 0.9412 - 0.9710. The best performance was the FLAML17  
316 combination with the largest TSS of 0.9710.

317 It is worth noting that the differences in model performance are mainly between  
318 forest types rather than different combinations of input variables. For the CBS mixed  
319 forests and QYF needle-leaf, models with various input combinations show high  $R^2$  and  
320 low RMSE (Table S2, Table S3, Table S4). The average  $R^2$  of the four temporal scales  
321 of CBF broadleaf Korean pine forest was 0.92-0.94, and the average  $R^2$  of FLAML00-  
322 FLAML08 and FLAML10-FLAML18 were both 0.93 and the average RMSE was  
323  $1.153 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ,  $1.137 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ , respectively. The average  $R^2$  of the four temporal  
324 scales of the coniferous forests in QYF ranged from 0.89 to 0.93, and the average  $R^2$  of  
325 FLAML00-FLAML08 and FLAML10-FLAML18 were 0.92 and 0.90, with an average  
326 RMSE of  $0.657 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ , and  $0.719 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ , respectively. The model performed  
327 slightly better on the coniferous forest at QYF station than on the broad-leaved Korean



328 pine forest at CBF station. A significant discrepancy was observed at the ALF station,  
329 which had an average  $R^2$  for the four temporal scales ranging from 0.56 to 0.70. The  
330 average  $R^2$  of FLAML00-FLAML08 and FLAML10-FLAML18 were 0.66 and 0.61,  
331 with average RMSE values of  $1.173 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$  and  $1.261 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ , respectively. In  
332 forest ecosystems, mixed forests (CBF) and evergreen needle-leaf forests (QYF)  
333 generally show better model performance than evergreen broad-leaf forests (ALF).  
334 Mixed forests, consisting of both evergreen needle-leaved and deciduous broadleaf  
335 species, display significant seasonal variations that can be effectively captured by  
336 satellite imagery. In contrast, evergreen broad-leaf forests have minimal seasonal  
337 changes in greenness, leading to higher modeling biases in GPP estimation.

338 A best-fit line between  $\text{GPP}_{\text{tower}}$  and  $\text{GPP}_{\text{pred}}$  was determined for all sites as one  
339 dataset using linear regression (Fig. 3 (III)). The  $R^2$  for all sites differed less from the  
340 results for the sites analyzed individually. As shown in Fig. 3 (III), the slope of the fitted  
341 line was close to but slightly greater than 1, indicating that the FLAML-LUE model  
342 underestimated the GPP of forest ecosystems.



343

344 **Fig. 3.** (I) Normalized Taylor diagrams showing the overall performance of all FLAML-LUE  
 345 models in (a) forest ecosystem, (b) ALF, (c) CBF, and (d) QYF. (II) Boxplots of 18 model  
 346 performances ( $R^2$ ) at different temporal scales in forest ecosystems, ALF, CBF, and QYF. Asterisks  
 347 indicate significant differences between the  $R^2$  at the four temporal resolutions (Kruskal-Wallis test),  
 348 \*\*\*\* $p$  values  $< 0.0001$ , \*\*\* $p$  values  $< 0.001$ , \*\* $p$  values  $< 0.01$ , \* $p$  values  $\leq 0.05$ , and ns indicates  
 349 no significance ( $p > 0.05$ ). (III) Scatterplot of observed GPP vs. simulated GPP in forest ecosystems.  
 350 Different colored dots represent different site values.

351

Under forest ecosystems, for all four temporal scales, the 18 FLAML-LUE models

352

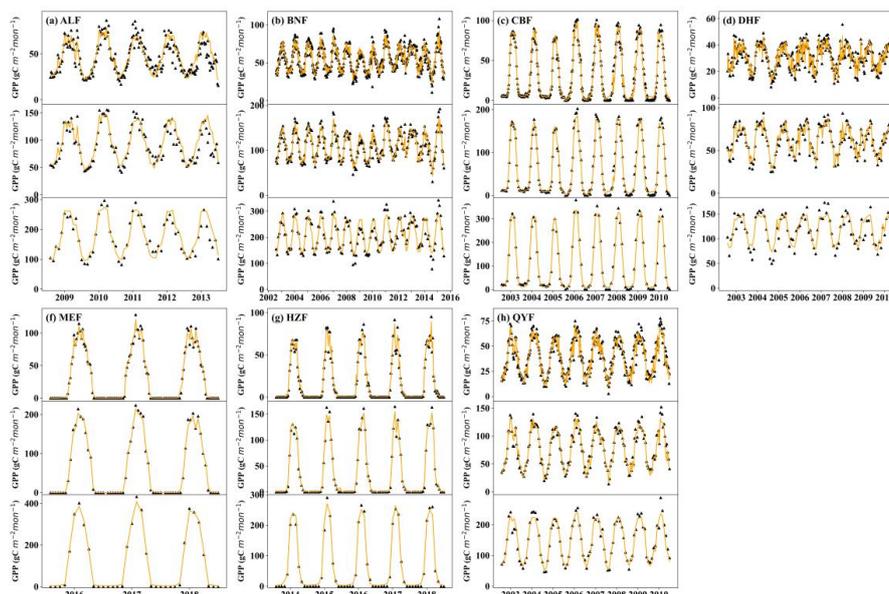
showed better accuracy as time aggregates to larger intervals. as shown by the increased



353  $R^2$  from 0.82 to 0.93. Paired t-tests revealed that the daily performance ( $R^2$ ) of the  
354 FLAML-LUE model was significantly lower than that of the other temporal scales  
355 across all sites ( $p < 0.01$ , Fig. 3(II)). In addition, the RMSE of the 8-day, 16-day, and  
356 monthly GPP (FLAML-LUE) also decreased significantly by 26.88%, 33.18%, and  
357 41.34%, respectively, when compared to the daily-scale results, suggesting that the  
358 uncertainty in these models becomes smaller at the longer temporal scale. The slopes  
359 of the linear regression relationships between the simulated and observed GPP approach  
360 1 with improving temporal resolution at ALF, CBF, and QYF sites.

### 361 3.1.2 Analysis of interannual GPP variability

362 Based on the Taylor diagram TSS skill scores, it can be seen that the forest  
363 ecosystems have the highest GPP simulation accuracy under the combination of  
364 FLAML17 indicators, as shown in Table S2.

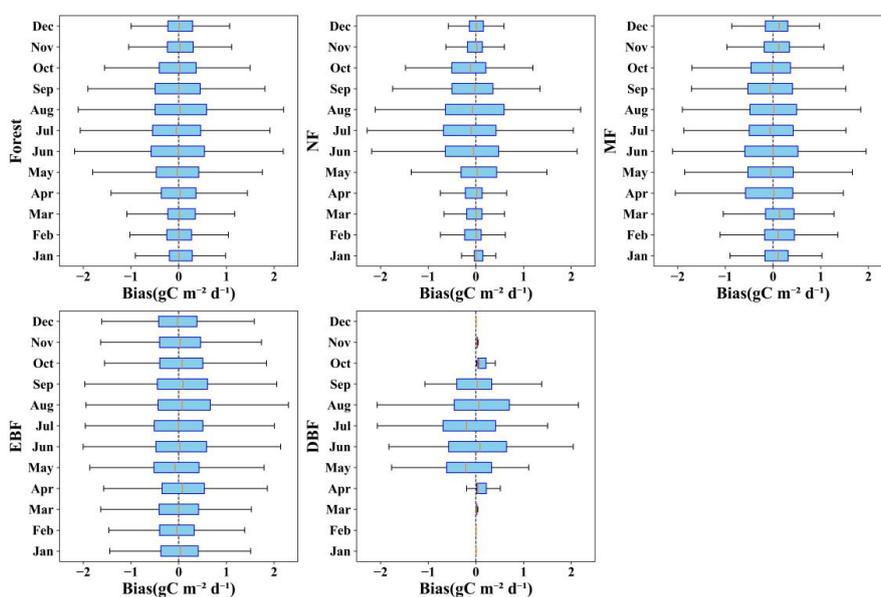


365  
366 **Fig. 4.** Plot of simulated GPP time-series variation at ALF, BNF, CBF, DHF, HZF, MEF, and QYF  
367 sites, with black triangles representing tower-based observations and orange solid lines representing



368 model-simulated data.

369 Fig. 4 shows that the simulated GPP closely aligns with the observed GPP values  
 370 in terms of seasonal patterns at the 8-day, 16-day, and monthly scales. The simulated  
 371 and observed GPP in forest ecosystems exhibit strong seasonality, with the lowest  
 372 values in spring, peaking in summer, and declining through fall and winter. Forest  
 373 ecosystems showed a peak of growth in the summer. In addition, the average annual  
 374 GPP varied greatly among sites (Table S5). Among the forest ecosystems, tropical  
 375 rainforest sites (BNF), subtropical evergreen broadleaf forests (ALF) had the highest  
 376 annual GPP, followed by subtropical planted coniferous forests (QYF), deciduous  
 377 broadleaf forests (MEF) and mixed coniferous and broadleaf forests (CBF, DHF), and  
 378 the lowest annual average GPP was found in the cold-temperate coniferous forests  
 379 (HZF). In summary, the FLAML-LUE model accurately modeled this inter-site  
 380 variation in GPP and showed seasonal variations in GPP.



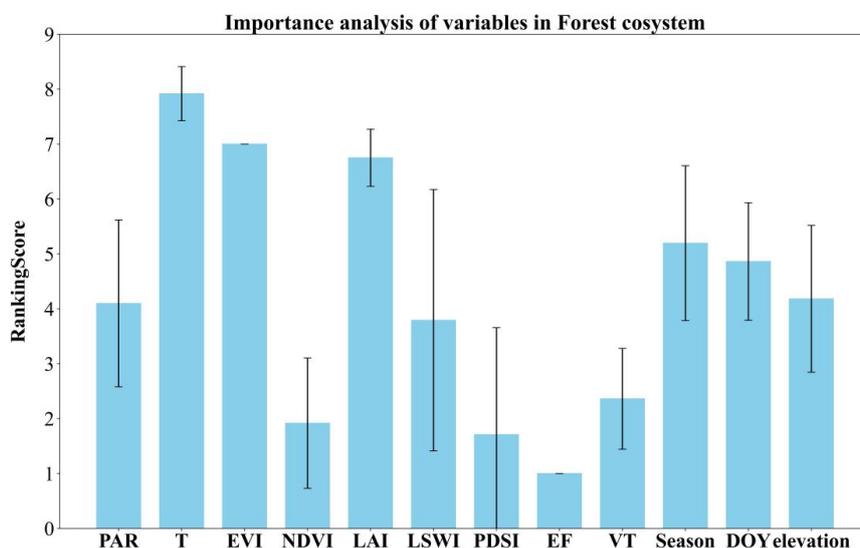
381  
 382 **Fig. 5.** The monthly bias of FLAML-LUE models among vegetation types. NF: needle-leaf forest,



383 MF: mixed forest, EBF: evergreen broad-leaf forest, DBF: deciduous broad-leaf forest.

384 We examined the monthly discrepancies between observed and simulated values  
385 across various forest types in forest ecosystems. Fig. 5 shows that the forest ecosystems  
386 model underestimated GPP in spring and summer on average, and although the forest  
387 ecosystems GPP simulation was biased in all months, it generally showed a larger bias  
388 in summer. There were significant differences in bias between forest types, with the  
389 model performing better in capturing the seasonal dynamics of coniferous and  
390 deciduous broadleaf forests.

### 391 3.1.3 Analysis of the importance of variables



392  
393 **Fig. 6.** Average variables importance of forest ecosystem in FALML-LUE models. LSWI: land  
394 surface water index, PDSI: Palmer Drought Severity Index, EF: evaporative fraction, EVI: enhanced  
395 vegetation index, NDVI: normalized difference vegetation index, LAI: leaf area index, T:  
396 temperature, PAR: photosynthetically active radiation, VT: vegetation type.

397 Fig. 6 shows the importance of each variable in the FLAML-LUE model for the  
398 forest ecosystem. The FLAML-LUE model utilizes AutoML algorithms based on  
399 different combinations of metrics to find the optimal algorithm and appropriate



400 hyperparameters. Since different ML algorithms were selected for different temporal  
401 scales and different combinations of indicators, and different methods were used to  
402 calculate the importance of each indicator, the ranking assignment method was used to  
403 assign the importance of each indicator. Based on the average importance of 4 temporal  
404 scales and 18 combinations of indicators, it can be seen that in forest ecosystems, the  
405 importance of temperature is greater than other variables in the model. The importance  
406 of EVI and LAI is much higher than that of NDVI among the three vegetation indices,  
407 which is also consistent with the results in section 3.1.1, that is, the simulation  
408 performance of the model consisting of the combination of indicators EVI and LAI is  
409 better than that of the combination of NDVI indicators. The importance of LSWI is  
410 higher than PDSI and EF among the water stress factors. Forest ecosystem GPP exhibits  
411 clear seasonal variation, with temperature and VI emerging as the most critical variables  
412 in the ML model for GPP estimation. These factors significantly impact the accuracy  
413 of predictions.

## 414 **3.2 Overall FLAML models performances on grassland ecosystem**

### 415 **3.2.1 Performance Evaluation of Models**

416 To examine the performance of each model in grassland ecosystems and at the site  
417 level, the accuracy of the 18 FLAML-LUE models was evaluated using the site data  
418 from DXG, DLG and HBG\_S01 Stations as the grassland ecosystem model test set.  
419 Table 5 shows the  $R^2$ , RMSE and SD of the 18 FLAML-LUE models with the grass  
420 station test set. Table S6 shows the algorithms adopted by each FLAML-LUE model  
421 under the grassland ecosystems.



422 **Table 5**  
 423  $R^2$ , SD, RMSE for the grassland ecosystems model test set.

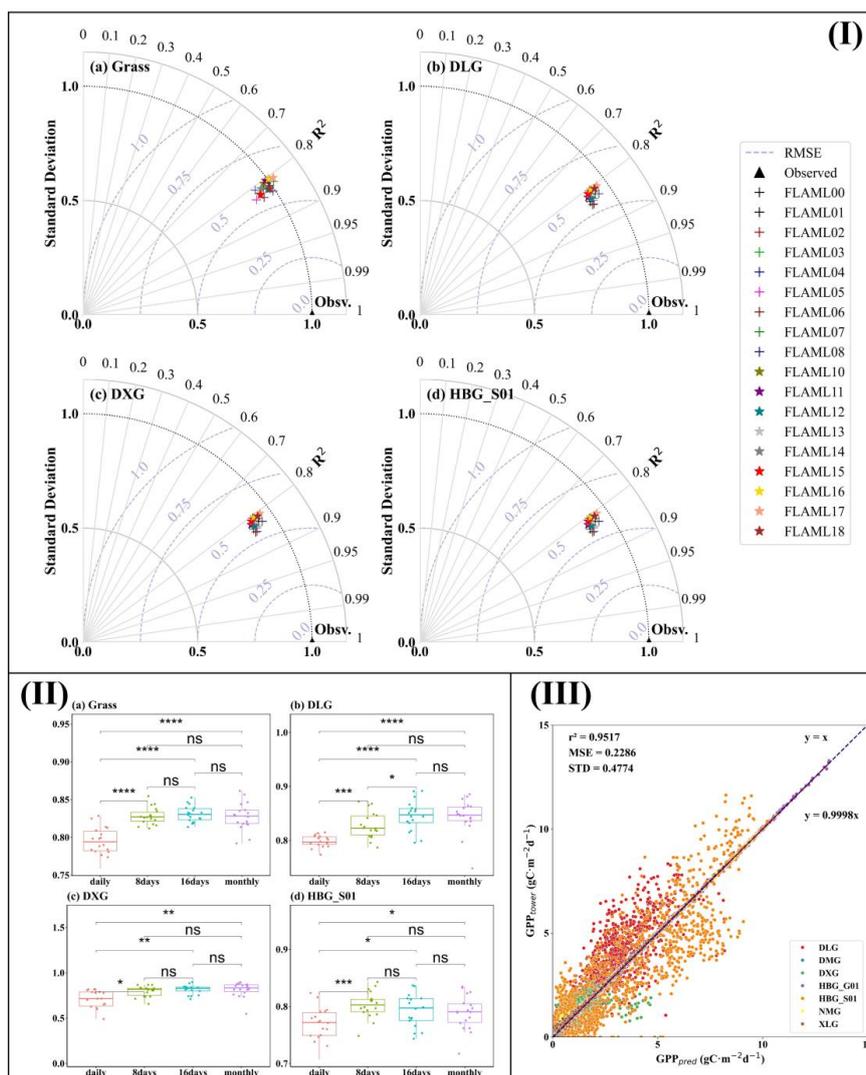
FLAML	$R^2$	SD	RMSE	nuRMSE	TSS
FLAML00	0.82	0.961	0.863	0.424	0.9525
FLAML01	0.82	0.987	0.857	0.421	0.9543
FLAML02	0.84	0.942	0.816	0.401	0.9558
FLAML03	0.82	0.935	0.858	0.422	0.9508
FLAML04	0.81	0.928	0.886	0.436	0.9466
FLAML05	0.83	0.909	0.832	0.409	0.9502
FLAML06	0.82	0.992	0.859	0.422	0.9544
FLAML07	0.82	1.015	0.865	0.425	0.9548
FLAML08	0.84	0.991	0.819	0.402	0.9585
FLAML10	0.81	0.976	0.890	0.437	0.9509
FLAML11	0.80	0.990	0.897	0.441	0.9512
FLAML12	0.83	0.976	0.845	0.415	0.9555
FLAML13	0.82	0.951	0.874	0.430	0.9508
FLAML14	0.82	0.955	0.871	0.428	0.9517
FLAML15	0.83	0.936	0.843	0.414	0.9527
FLAML16	0.81	1.004	0.895	0.440	0.9515
FLAML17	0.81	1.024	0.885	0.435	0.9528
FLAML18	0.83	0.984	0.841	0.413	0.9563
Flux(average)	0.82	0.962	0.851	0.418	
ERA5(average)	0.81	0.977	0.871	0.428	
Forest(average)	0.82	0.970	0.861	0.423	

424 As shown in Table 5, the cross-validation analysis showed that the average  $R^2$  for  
 425 the four temporal scales under grassland ecosystems was 0.80-0.84. The models driven  
 426 by the flux data performed slightly better than the one driven by the ERA5 data, with  
 427 average  $R^2$  of 0.82, 0.81, and RMSE of 0.851, 0.871  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ , respectively. In  
 428 grassland ecosystems, models driven by different vegetation indices had equal mean  $R^2$   
 429 values of 0.82 and RMSE values of 0.861  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ . The model driven with EF ( $R^2 =$   
 430 0.83, RMSE = 0.833  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) performed slightly better than LSWI ( $R^2 = 0.82$ , RMSE  
 431 = 0.873  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) and PDSI ( $R^2 = 0.81$ , RMSE = 0.877  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ).



432 Fig. 7 shows the Taylor diagrams of the performance of all FLAML-LUE models  
433 in grassland ecosystems, DXG, DL, and HBG\_S01. The  $R^2$ , nuRMSE, and SD of  
434 different combinations of variables under grassland ecosystems were slightly different,  
435 and the TSS values ranged from 0.9466 - 0.9585, among which the best performance  
436 was the FLAML08 combination with the largest TSS of 0.9585.

437 Similar to forest ecosystems, the main differences in the prediction accuracy of the  
438 FLAML-LUE model for grassland ecosystems were between grass types rather than  
439 between different combinations of indicators. It is clear that the simulation accuracy of  
440 GPP for grassland ecosystems is lower than that for forest ecosystems, and there are  
441 also significant differences between grass types. For typical grassland, the FLAML-  
442 LUE model performed best with an average  $R^2$  of 0.83 and an RMSE of  $0.779 \text{ gC}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$ ,  
443 followed by alpine scrub with an average  $R^2$  of 0.79 and an RMSE of  $0.459 \text{ gC}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$ ,  
444 and stations with alpine meadows the worst performance, with an average  $R^2$  of  
445 0.78 and an RMSE of  $0.461 \text{ gC}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$  (Table S7, S8, S9). It is worth noting that the  
446 model simulated the alpine meadows with the lowest RMSE for GPP, which is since  
447 the average daily GPP of alpine meadows is smaller than that of typical grassland and  
448 alpine scrub.



449

450 **Fig. 7.** (I) Normalized Taylor diagrams showing the overall performance of all FLAML-LUE  
 451 models in (a) grass ecosystem, (b) DXG, (c) DLG, and (d) HBG\_S01. (II) Boxplots of 18 model  
 452 performances ( $R^2$ ) at different temporal scales in grassland ecosystems, DXG, DLG, and HBG\_S01.  
 453 (III) Scatterplot of observed GPP vs. simulated GPP in grassland ecosystems.

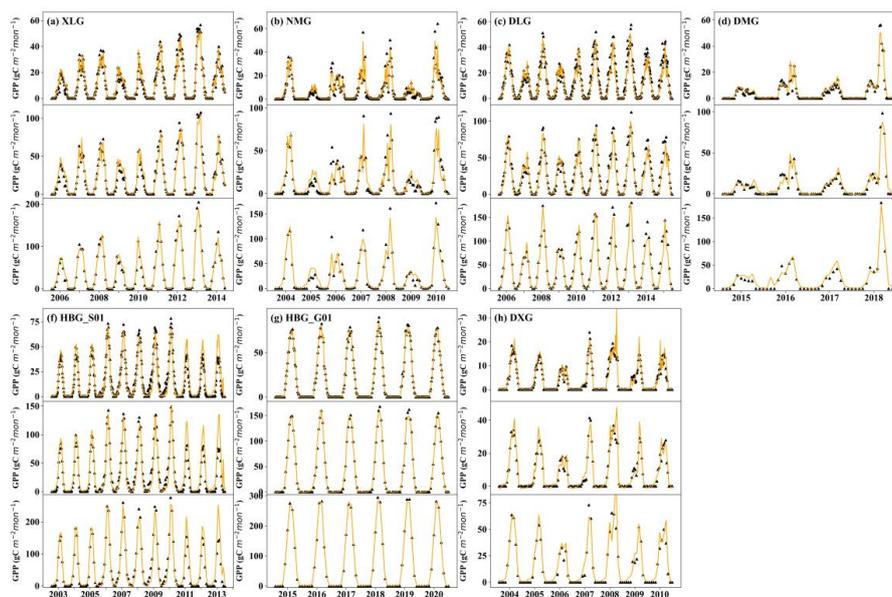
454 A best-fit line between tower-based GPP and predicted GPP was determined for all  
 455 grass ecosystem sites as one dataset using linear regression (Fig. 7 (III)). The  $R^2$  for all  
 456 sites differed less from the results for the sites analyzed individually. As shown in Fig.  
 457 10, the slope of the fitted line was close to, but slightly less than 1, indicating that the



458 FLAML-LUE model overestimated the GPP of grassland ecosystems.

459 In grassland ecosystems, for all four temporal scales, the 18 FLAML-LUE models  
460 showed higher accuracy as temporal aggregation increased from daily to monthly. The  
461 FLAML-LUE model shows a marked improvement in validation accuracy at extended  
462 time scales, with the average  $R^2$  rising from 0.80 to 0.83. Paired t-tests revealed that for  
463 grassland ecosystems and at the DXG and DLG stations, the FLAML-LUE model's  
464 performance ( $R^2$ ) was significantly lower at the daily scale compared to other time  
465 scales ( $p < 0.01$ , Fig. 7 (II)). However, at station HBG\_S01, model performance at the  
466 daily scale was only lower than the 8-day time scale, and not significantly different  
467 from other time scales. In addition, the RMSE of the 8-day, 16-day, and monthly GPP  
468 (FLAML-LUE) were also significantly lower by 12.10%, 13.36%, and 12.62%,  
469 respectively, compared to the daily-scale results, indicating that the uncertainty  
470 associated with these models diminishes at extended time scales.

### 471 **3.2.2 Analysis of interannual GPP variability**



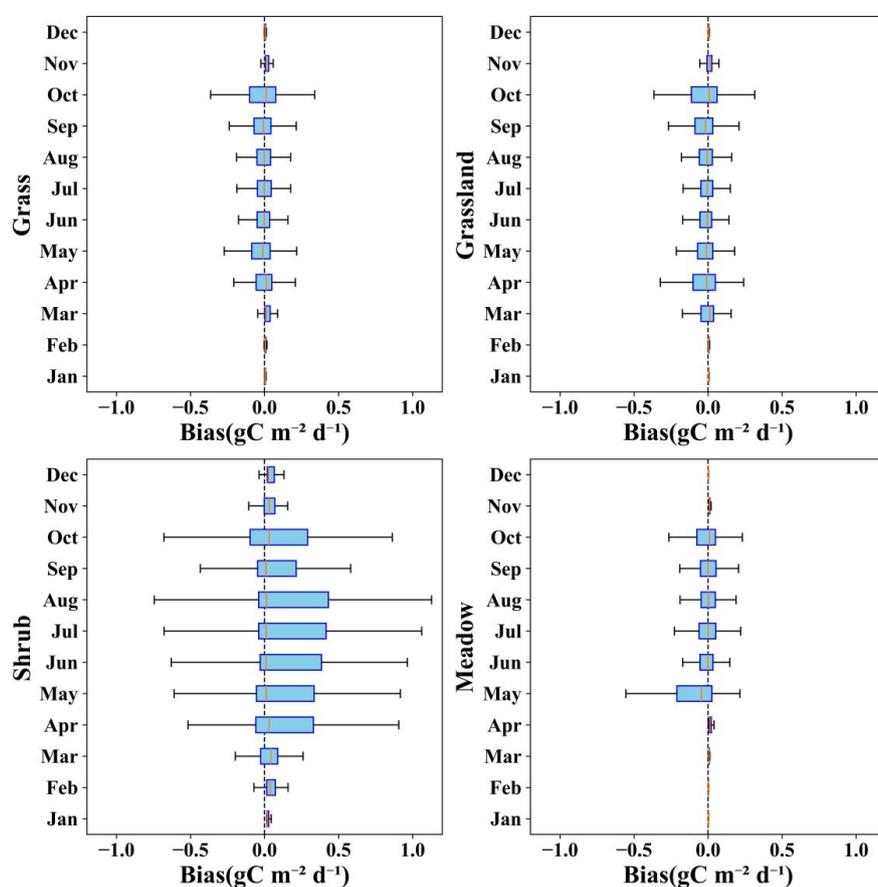
472  
473 **Fig. 8.** Plot of simulated GPP time-series variation at DLG, DMG, DXG, HBG\_G01, HBG\_S01,  
474 NMG, and XLG sites.

475 Based on the Taylor diagram TSS skill scores, it can be seen that the grassland  
476 ecosystems have the highest GPP simulation accuracy under the combination of  
477 FLAML08 indicators, as shown in Table S6.

478 Fig. 8 shows that the FLAML-LUE model can simulate seasonal dynamics similar  
479 to the observed GPP, as can be seen from their long-term evolution courses at the seven  
480 grass ecosystem sites (DLG, DMG, DXG, HBG\_G01, HBG\_S01, NMG, XLG).  
481 Although the overall trend was simulated correctly, it is clear that the FLAML-LUE  
482 model does not capture the GPP peaks in grassland ecosystems well. For the simulation  
483 of typical grassland sites (DLG, DMG, NMG, XLG), the model performance was  
484 generally poor for NMG site, and the GPP values were poorly simulated during the  
485 peak growing seasons. In addition, it is more difficult to simulate GPP at the meadow  
486 sites (DXG and HBG\_G01), especially for the summer peak simulation at DXG site,



487 which was too high compared to measured GPP. This is possibly due to the special  
488 geographic location and survival environment of alpine. In conclusion, the simulation  
489 of summer peaks of different grass types for GPP in grass ecosystems did not perform  
490 well.



491  
492 **Fig. 9.** The monthly bias of FLAML-LUE models among grass types. Grassland: typical grassland,  
493 Shrub: alpine shrub, Meadow: alpine meadow.

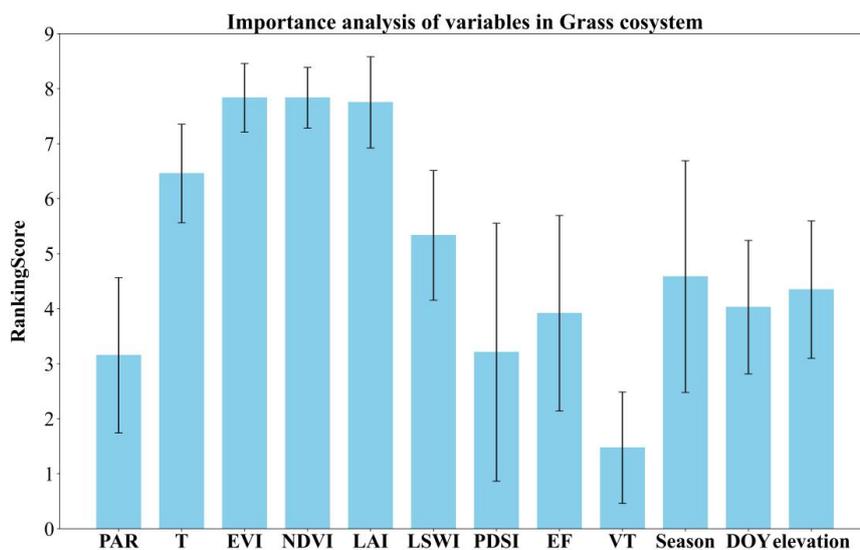
494 We examined the monthly discrepancies between observed and modeled values for  
495 different farm types in the grassland ecosystem. Fig. 9 shows that the simulated values  
496 of GPP from the grass ecosystem model for typical grassland and alpine scrub have  
497 biases in all months, and the biases were generally larger in summer and were all



498 overestimated. The gross primary productivity in spring and winter was smaller, and  
499 the corresponding deviations were smaller. Similarly, the GPP simulations for alpine  
500 meadows were underestimated and had smaller deviations, as seen in Fig. 9.

### 501 **3.2.3 Analysis of the importance of variables**

502 Fig. 10 shows the importance of the variables in the FLAML-LUE model for  
503 grassland ecosystems. It can be seen that the importance of NDVI is the highest among  
504 all the variables in the grass ecosystem model. The importance of LAI was the lowest  
505 among the three vegetation indices, while it is still higher than that of the other variables,  
506 indicating that vegetation indices are very important for modeling the GPP of grassland  
507 ecosystems. The importance score of temperature ranked just below the three  
508 vegetation indices, proving that temperature is also one of the most important indicators  
509 for estimating GPP in grassland ecosystems. In grassland ecosystems, the moisture  
510 index LSWI had a higher importance in modeling the GPP compared to PDSI and EF,  
511 and overall, the grass ecosystem showed a higher importance score for the moisture  
512 index than the forest ecosystem. Generally, forest vegetation has stronger water storage  
513 capacity and a higher ability to utilize deep soil water when compared to grasses, thus  
514 making forests more resistant to drought during meteorological droughts. Therefore,  
515 grass ecosystem simulated GPP were more sensitive to the moisture index.



516

517 **Fig. 10.** Average variables importance of grassland ecosystems in FALML-LUE models.

### 518 **3.3 Overall FLAML models performances on cropland ecosystem**

#### 519 **3.3.1 Performance Evaluation of Models**

520 To examine the performance of each model in the cropland ecosystem and at the  
521 site level, the accuracy of the 18 FLAML-LUE models was evaluated using the site  
522 data from JZA and YCA stations as the cropland ecosystems model test set. Table 6  
523 shows the  $R^2$ , RMSE, and SD of the 18 FLAML-LUE models with the cropland station  
524 test set. The algorithms adopted by each FLAML-LUE model under the cropland  
525 ecosystems are shown in Table S11.

526 As shown in Table 6, the cross-validation analysis showed that the average  $R^2$  for  
527 the four temporal scales under cropland ecosystems was 0.82-0.89. The models driven  
528 by the flux data performed slightly better than the one driven by the ERA5 data, with  
529 their average  $R^2$  of 0.88, 0.85, and RMSE of 1.908, 2.108  $\text{gC}\cdot\text{m}^{-2}\text{d}^{-1}$ , respectively.  
530 However, the models driven using EVI ( $R^2 = 0.87$ ,  $\text{RMSE} = 1.955 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ )



531 performed slightly better than NDVI ( $R^2 = 0.85$ ,  $RMSE = 2.069 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) and LAI  
 532 ( $R^2 = 0.86$ ,  $RMSE = 1.999 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ). The model driven with PDSI ( $R^2 = 0.87$ ,  $RMSE$   
 533  $= 1.952 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ) performed slightly better than EF ( $R^2 = 0.87$ ,  $RMSE = 1.991 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ )  
 534 and LSWI ( $R^2 = 0.85$ ,  $RMSE = 2.080 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$ ).

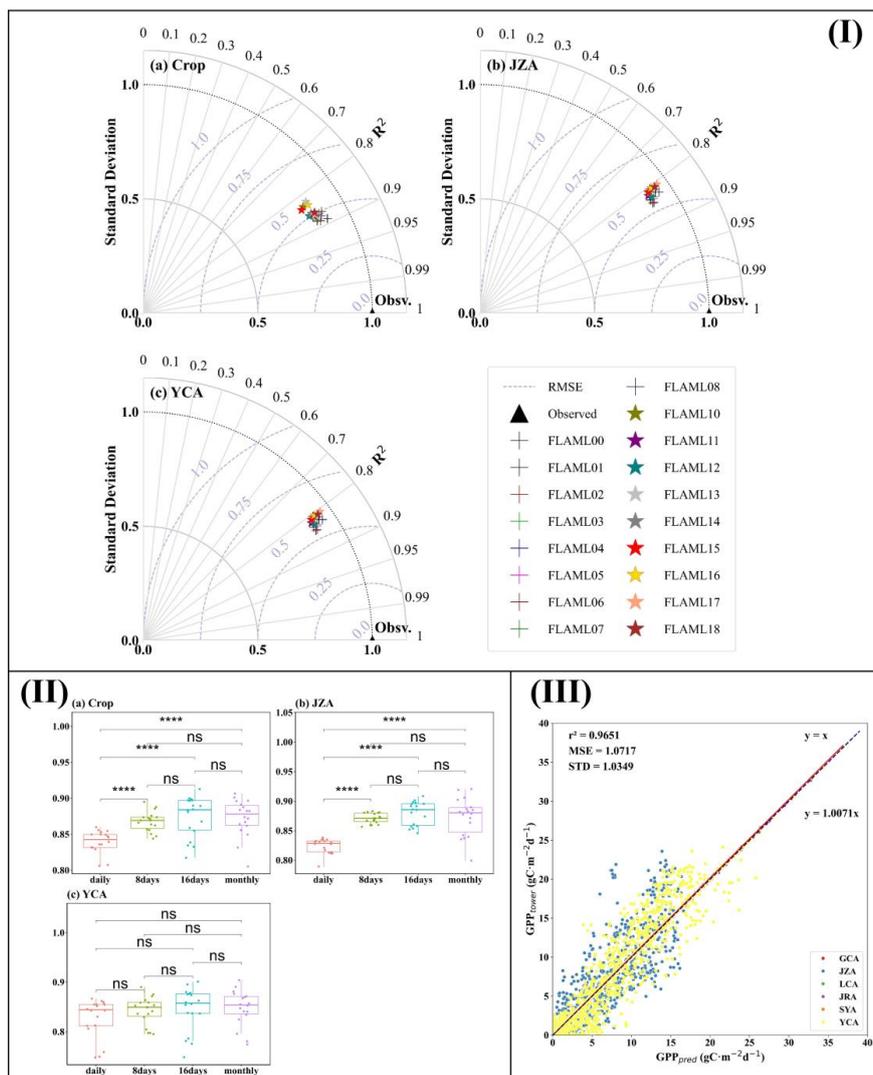
535 Fig. 11 shows the Taylor diagrams of the performance of all FLAML-LUE models  
 536 in cropland ecosystems, JZA station, and YCA station. The  $R^2$ , nuRMSE, and SD of  
 537 different combinations of variables under cropland ecosystems were slightly different,  
 538 and the TSS values ranged from 0.9253 – 0.9622, among which the best performance  
 539 was the FLAML00 combination with the largest TSS of 0.9622.

540 **Table 6**  
 541  $R^2$ , SD, RMSE for the cropland ecosystems model test set.

FLAML	$R^2$	SD	RMSE	nuRMSE	TSS
FLAML00	0.89	0.904	1.812	0.626	0.9622
FLAML01	0.88	0.859	1.858	0.611	0.9490
FLAML02	0.89	0.873	1.832	0.594	0.9542
FLAML03	0.87	0.883	1.966	0.647	0.9524
FLAML04	0.87	0.851	1.963	0.640	0.9425
FLAML05	0.87	0.872	1.967	0.592	0.9475
FLAML06	0.87	0.897	1.981	0.616	0.9532
FLAML07	0.88	0.864	1.882	0.672	0.9483
FLAML08	0.88	0.886	1.912	0.596	0.9535
FLAML10	0.83	0.838	2.230	0.621	0.9282
FLAML11	0.87	0.858	1.983	0.598	0.9430
FLAML12	0.86	0.840	2.015	0.635	0.9365
FLAML13	0.82	0.861	2.281	0.633	0.9319
FLAML14	0.86	0.853	2.042	0.644	0.9384
FLAML15	0.84	0.825	2.195	0.603	0.9253
FLAML16	0.83	0.861	2.212	0.668	0.9348
FLAML17	0.87	0.864	1.985	0.631	0.9418
FLAML18	0.86	0.868	2.025	0.629	0.9454
Flux(average)	0.88	0.877	1.908	0.622	



ERA5(average)	0.85	0.852	2.108	0.629
Forest(average)	0.86	0.864	2.008	0.625



542  
 543 **Fig. 11.** (I) Normalized Taylor diagrams showing the overall performance of all FLAML-LUE  
 544 models in (a) cropland ecosystem, (b) JZA, and (c) YCA. (II) Boxplots of 18 model performances  
 545 ( $R^2$ ) at different temporal scales in crop ecosystem, JZA, and YCA. (III) Scatterplot of observed  
 546 GPP vs. simulated GPP in crop ecosystem.

547 Unlike forest and grassland ecosystems, the performance of the FLAML-LUE  
 548 model did not differ significantly among different farm types in cropland ecosystems.



549 The average  $R^2$  was 0.86 and the average RMSE was  $1.724 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$  for the single  
550 cropping farmland station (JZA), while the average  $R^2$  was 0.84 and the average RMSE  
551 was  $2.400 \text{ gC}\cdot\text{m}^{-2}\text{d}^{-1}$  for the double cropping farmland (YCA). The simulation of the  
552 single-cropping farmland was slightly better than the double-cropping farmland (Table  
553 S12, S13).

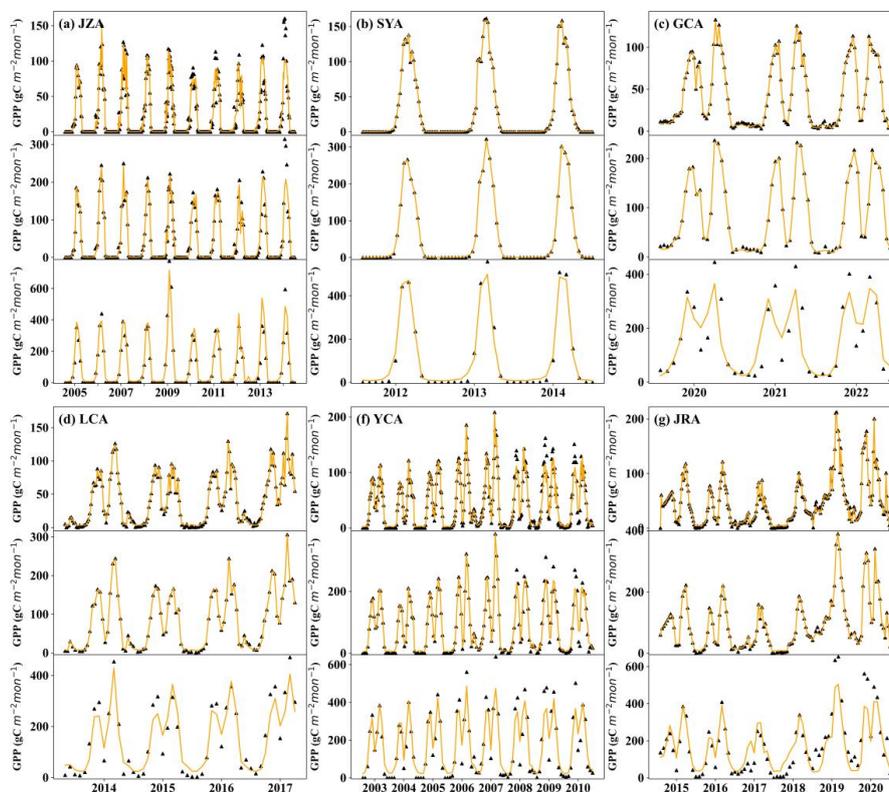
554 A best-fit line between  $\text{GPP}_{\text{tower}}$  and  $\text{GPP}_{\text{pred}}$  was determined for all cropland  
555 ecosystem sites as one dataset using linear regression (Fig. 11 (III)). The  $R^2$  for all sites  
556 differed less from the results for the sites analyzed individually. As shown in Fig. 11  
557 (III), the slope of the fitted line was close to but slightly less than 1, indicating that the  
558 FLAML-LUE model overestimated the GPP of cropland ecosystems.

559 In the cropland ecosystems, the average  $R^2$  increased from 0.84 at the daily scale  
560 to 0.87 at the 16-day scale as can be seen, and the 18 FLAML-LUE models showed  
561 higher accuracy as the temporal aggregation increased from the daily to the monthly.  
562 The model generally showed significantly lower performance ( $R^2$ ) at the daily scale  
563 than at other time scales ( $p < 0.001$ , Fig. 11(II)(a)), while there was no remarkable  
564 difference in the model performance at all four time scales for the YCA ( $p > 0.05$ , Fig.  
565 11(II) (c)). The difference in model performance between the 16-day and monthly  
566 scales was not significant at all stations ( $p > 0.05$ , Fig. 11(II)). In addition, RMSE of  
567 the GPP (FLAML-LUE) was also significantly reduced by 14.70%, 18.61%, and 19.79%  
568 for the 8-day, 16-day, and monthly GPP, respectively, when compared to the daily-scale  
569 results, suggesting that the uncertainty in these models becomes smaller at the longer  
570 temporal scale. At JZA and YCA, the slopes of the linear regression relationship



571 between the modeled GPP and the observed GPP converge to 1 as the time scale  
572 improves.

### 573 3.3.2 Analysis of interannual GPP variability



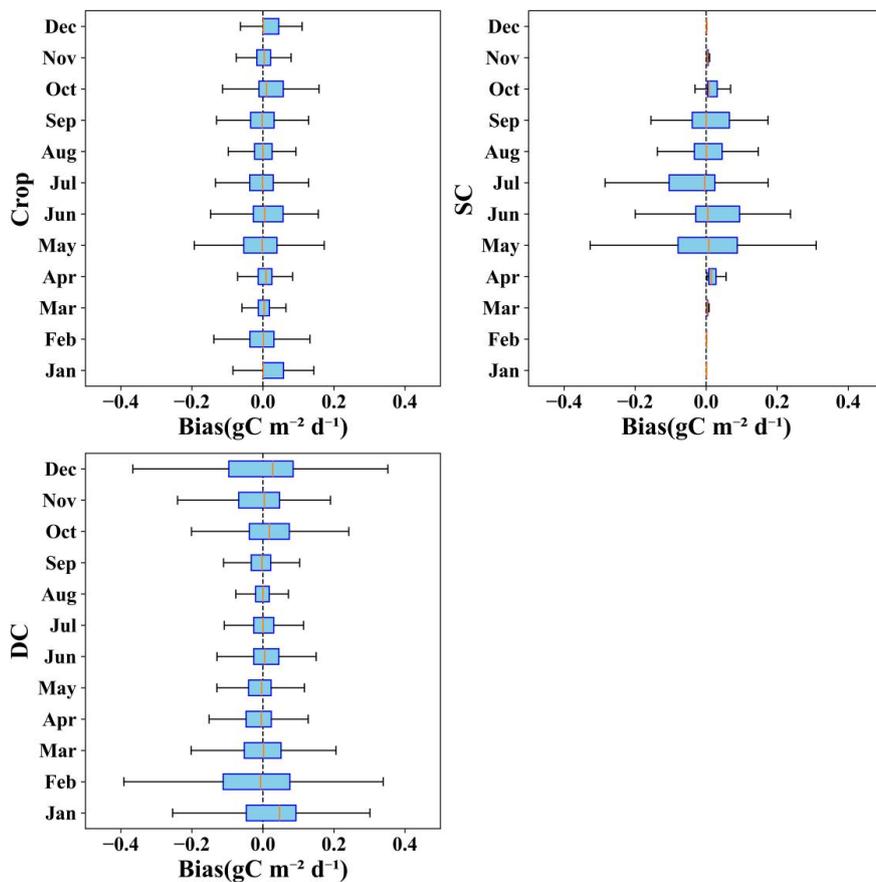
574  
575 **Fig. 12.** Plot of simulated GPP time-series variation at GCA, JRA, JZA, LCA, SYA, and YCA sites.

576 Fig. 12 shows that simulated GPP aligns closely with the observed GPP values,  
577 showing comparable seasonal patterns at the 8-day, 16-day, and monthly scales. In  
578 farmland ecosystems, simulated GPP values from different farm types show different  
579 seasonal dynamics. Farmland with spring maize (JZA, SYA), a single-crop system,  
580 shows a single GPP peak during the harvest season. In comparison, double-cropping  
581 systems, with cycles of winter wheat and summer corn, display GPP peaks in both May  
582 and August. In addition, the average annual GPP of farmlands with different crop



583 rotation schemes varied greatly (Table S14). The annual mean GPP of double-cropping  
584 farmland was higher than that of single-cropping farmland. In conclusion, the FLAML-  
585 LUE model accurately modeled the differences in GPP among farmland types and  
586 showed seasonal variations in GPP among farmland types.

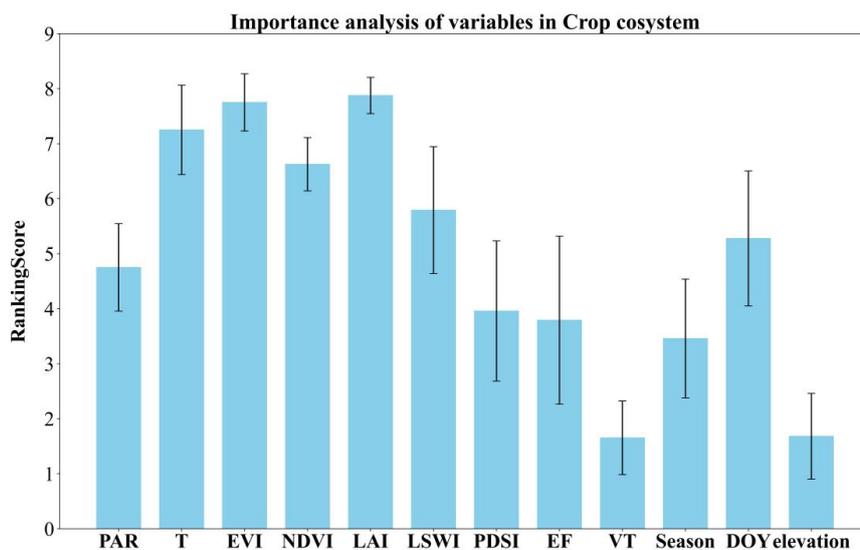
587 We examined the monthly discrepancies between observed and modeled values for  
588 different farm types in the agroecosystem. Fig. 13 shows that the agroecosystem model  
589 averagely overestimated GPP values in spring and fall (positive bias), while slightly  
590 underestimated it in summer. Although the agroecosystem GPP simulations were biased  
591 in all months, the biases were generally larger in spring and fall. There were significant  
592 differences in bias between farmland types. The model over double cropping farmland  
593 showed small biases in simulated GPP for all months of the year, while it overestimated  
594 GPP in the spring and fall, and underestimated GPP in the summer over the single  
595 cropping farmland. This suggests that the model performance for the single cropping  
596 farmland still need to be improved.



597

598 **Fig. 13.** The monthly bias of FLAML-LUE models among cropland types. SC: single cropping, DC:  
599 double cropping.

600 **3.3.3 Analysis of the importance of variables**



601  
602 **Fig. 14.** Average variables importance of farm ecosystem in FALML-LUE models.

603 Fig. 14 shows the importance of the variables in the cropland ecosystem FLAML-  
604 LUE model. It can be seen that the importance of LAI is the highest among all the  
605 variables in the farm ecosystem model. The importance of NDVI was the lowest among  
606 the three vegetation indices, while it is still higher than that of the other variables,  
607 indicating that vegetation indices are very important for modeling the GPP of cropland  
608 ecosystems. The importance score of temperature was second only to LAI and EVI, and  
609 similar to forest and grassland ecosystems, temperature is also one of the important  
610 indexes for modeling GPP in farm ecosystem. In addition, the moisture stress factor  
611 was also important, and unlike forest and grass ecosystems, the most important  
612 moisture factor for constructing the GPP simulation model in cropland ecosystems was  
613 LSWI, followed by PDSI, and EF was the lowest.

#### 614 **4. Discussion**

615 Model performance is highly influenced by the algorithms used, the underlying



616 processes, and how GPP responds to varying environmental conditions (Chang et al.,  
617 2023). A detailed comparison of the FLAML-LUE models across different ecosystems  
618 showed that performance varied depending on the input variables, vegetation types, and  
619 time scales (Chang et al., 2023; Harris et al., 2021).

#### 620 **4.1 Performance comparison of FLAML-LUE models for different** 621 **ecosystems**

622 In this study, FLAML-LUE models were constructed for different ecosystems,  
623 different combinations of variables and different time scales based on AutoML  
624 algorithms. On the whole, the modeled GPP values agree well with the GPP estimated  
625 based on the EC tower, and the FLAML-LUE models performed better in capturing the  
626 magnitude and seasonal dynamics of the GPP, which indicated that it was feasible to  
627 estimate the GPP using AutoML algorithms. Further, all three ecosystems showed good  
628 model performance driven by observational data. Comparisons across various  
629 ecosystems indicate that the model exhibited superior performance over forest  
630 ecosystems compared to grassland and agricultural ecosystems, as evidenced by the  
631 average  $R^2$  values.

632 Additionally, the models constructed for each ecosystem showed different  
633 performances under different combinations of indicators, while the differences were not  
634 significant, and the main differences in prediction accuracy were ascribed to site  
635 differences rather than model differences. FLAML-LUE had the best simulation  
636 performance for mixed forests in CBF, and planted coniferous forests in QYF, with  $R^2$   
637 of 0.93 and 0.90, respectively; followed by single cropping farmland in the Jinzhou site,



638 double cropping farmland in the Yucheng site and typical grassland DLG. Over the  
639 alpine meadow and alpine ecosystem, the model performance was poorer, with an  $R^2$   
640 of 0.79; and even worse at the MEF site, with an average  $R^2$  of 0.63. Mixed forests  
641 display clear seasonal variations that satellite imagery can effectively capture. However,  
642 evergreen broadleaf forests (ALF) have slight seasonal variations in vegetation cover  
643 or greenness, making it difficult for the model to predict. For non-forest ecosystems,  
644 the highest  $R^2$  was found in agricultural fields and typical grasslands, followed by  
645 alpine meadows and alpine scrub. In addition, the differences in model performance  
646 were also reflected in different temporal scales. In general, the model simulation  
647 performance at 16-day and monthly scales was better than that at daily scale, and the  
648 performances of different temporal scales for forest, grassland, and cropland  
649 ecosystems were consistent with previous studies.

650       Discrepancies in the comparison between observed and simulated values varied  
651 across ecosystems, with models for grassland and forest ecosystems generally  
652 underestimating GPP (exhibiting a negative bias) in spring and summer, while  
653 displaying satisfactory performance in other seasons. The GPP during spring and winter  
654 remains relatively low, and hence correspondingly smaller deviations of modeling  
655 values. Overall, the FLAML-LUE model performed well in capturing interannual  
656 variability in GPP, while it encounters challenges in accurately capturing the dynamic  
657 fluctuations of GPP throughout the growing season.

658       In addition, our results indicate that forest and agricultural fields have greater  
659 carbon sequestration capacity and higher annual fluxes than grasslands (Table S5, S10,



660 S14), aligning with previous research outcomes (Y. Wang et al., 2021; Zhang et al.,  
661 2007). However, due to the annual harvest of crops, approximately 76% of the on-farm  
662 biomass is removed, resulting in limited long-term carbon storage capacity (Zhang et  
663 al., 2007). With the exception of tropical rainforests (i.e., BNF), the annual carbon  
664 production of planted forests (i.e., QYF) is higher than that of natural forests (i.e., CBF,  
665 DHF), which implies that planted forests possess significant potential for carbon  
666 assimilation, functioning as robust carbon sinks.

#### 667 **4.2 Impact of variables on GPP estimation**

668 The estimated GPP in this research closely matched the GPP measured by the EC  
669 tower. However, the important characterizing factors affecting the models varied across  
670 different ecosystems. For forest ecosystem, temperature was the most primary variable  
671 for model construction, while the vegetation index was the most important factor for  
672 characterizing grass ecosystem and agroecosystem GPP. Our study is consistent with  
673 previous studies, indicating that, in addition to temperature data, vegetation index are  
674 the crucial drivers for accurately predicting GPP. High variability in greenness existed  
675 in grassland and scrub over the phenological cycle, as well as in agricultural land under  
676 anthropogenic management patterns, while models were less effective in predicting  
677 evergreen broadleaf forests, with lower variability in greenness. A common problem is  
678 the high uncertainty in predicting evergreen forest GPP with many satellite-driven GPP  
679 models. This study found that the FLAML-LUE model using EVI slightly  
680 outperformed the one using NDVI, highlighting EVI's superior role in GPP estimation.  
681 EVI offers better atmospheric correction and is less affected by green radiation

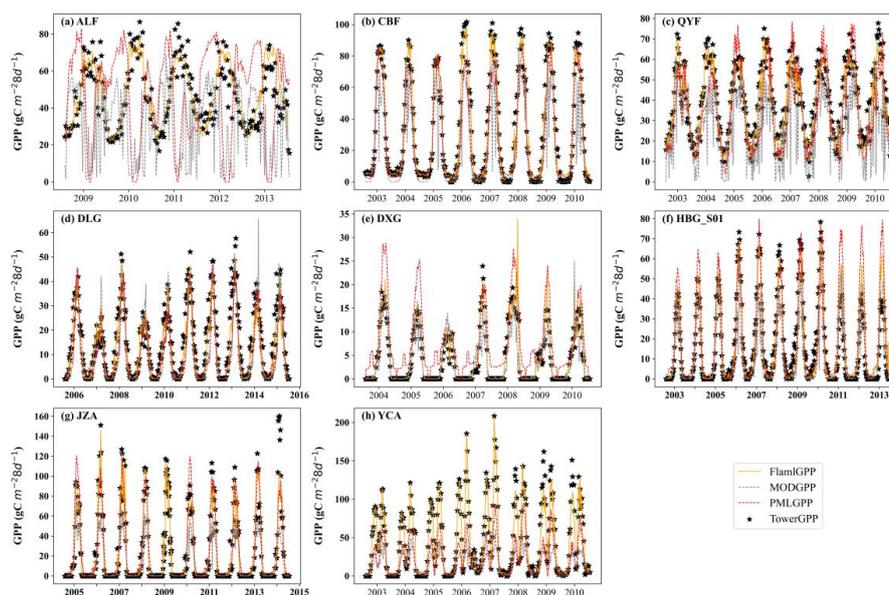


682 saturation compared to NDVI. Recent research indicates that satellite observations of  
683 solar-induced chlorophyll fluorescence (SIF) provide a more accurate picture of the  
684 dynamics of plant photosynthesis. It is a more effective indicator for modeling  
685 subtropical evergreen vegetation. Future studies should consider incorporating SIF into  
686 models to assess its potential for improving performance in evergreen forests.

687 Compared to temperature and radiation, moisture plays a more crucial role in  
688 regulating GPP. Recent research suggests that water stress is the primary source of  
689 uncertainty in GPP estimations (Zhang and Ye, 2022). At the same site, the FLAML-  
690 LUE model's performance driven by the three moisture indices was highly consistent  
691 across the six sites (QYZ, CBS, DLG, HBG\_S01, JZA, YCA). However, for the DXG  
692 and ALF stations, the performance of the model varied with the moisture index. When  
693 PDSI was used for DXG and ALF, the  $R^2$  values of these models were low at 0.75 and  
694 0.60, respectively. Our results showed low importance for all moisture indices at all  
695 sites. However, moisture indices were more important in non-forest than in forest,  
696 suggesting that forests are less sensitive to water stress. This finding is consistent with  
697 the results of previous studies (Zhang et al., 2015; Sims et al., 2014; Xie et al., 2014),  
698 which may be due to that forest vegetation has strong water storage capacity and the  
699 ability to utilize the deep soil water (Bi et al., 2015). Water variables were more crucial  
700 for grasslands compared to other ecosystems, indicating that grasslands with shallow  
701 root vegetation are less drought tolerant. In this context, future grassland management  
702 should prioritize scientific conservation planning and improved water management  
703 strategies.



704 **4.3 Comparison with other studies**



705  
706 **Fig. 15.** Comparing 8-day GPP from FLAML-LUE, PML, MOD17 models, and EC observations.

707 This study attempted to predict the GPP of different sites using the FLAML model  
708 based on the LUE model variables. The results showed that the AutoML algorithm is a  
709 promising GPP estimation method, which explains on average 63%-93% of the GPP  
710 variation.

711 Compared to two GPP products (MODIS GPP, PML GPP), the GPP from this study  
712 showed the highest precision (Table 7) and better consistency with flux tower-based  
713 GPP under different ecosystems. Overall, the FLAML-LUE model used in this study  
714 had the best simulation performance. These findings highlight the potential of the  
715 FLAML algorithm for accurately estimating GPP. The FLAML-LUE model is a data-  
716 driven ML approach that builds relationships based on dependent and explanatory  
717 variables. This enables it to effectively simulate the complex nonlinear interactions  
718 across diverse ecosystems (Tramontana et al., 2016). This advantage is even more



719 prominent at the global scale considering that more flux tower data are available for  
720 model construction.

721 **Table 7**

722  $R^2$  of 8-day GPP simulated by FLAML-LUE, PML and MOD17 at different ecosystems validation  
723 sites.

Ecosystem	Station	FLAML_R <sup>2</sup>	MOD_R <sup>2</sup>	PML_R <sup>2</sup>
Forest	ALF	0.79	0.24	0.33
	CBF	0.98	0.78	0.93
	QYF	0.96	0.54	0.74
Grass	DLG	0.93	0.76	0.77
	DXG	0.89	0.74	0.82
	HBG_S01	0.92	0.81	0.83
Crop	JZA	0.94	0.84	0.85
	YCA	0.96	0.71	0.78

724 However further work is needed to evaluate the FLAML-LUE model's suitability  
725 and accuracy considering its limitations. In particular, it tends to underestimate high  
726 GPP and overestimate low GPP. In addition, the model performance in GPP estimation  
727 is highly dependent on ecosystem type. Our findings indicated that mixed forests,  
728 deciduous broadleaf forests, and agricultural lands had higher prediction accuracies.  
729 While grass sites such as alpine scrub and alpine meadows were predicted with large  
730 uncertainties, consistent with results from other studies (Y. Wang et al., 2021; Yuan et  
731 al., 2014). This is still a big challenge in accurately estimating GPP.

732 In general, satellite imagery accurately captures the seasonal leaf phenology of  
733 DBF and MF canopies (e.g., spring leaf unfolding and fall senescence). Additionally,  
734 the key environmental factors influencing vegetation production during different  
735 phenological phases are well-defined (Yuan et al., 2014), making them well-suited for  
736 FLAML-LUE modeling. In contrast, the ambiguous seasonal leaf area changes in EBF



737 and the low variability of GPP in NMG ecosystems result in poorer model performance,  
738 and empirical methods struggle to estimate GPP variability in these areas (Tramontana  
739 et al., 2016).

740 Model performance is heavily influenced by the quality of the driver data and the  
741 typicality of the flux towers. In this study, meteorological indices are obtained directly  
742 from spatially explicit reanalysis products. Remotely sensed variables (e.g., NDVI and  
743 EVI, LAI, LSWI) serve as proxies for vegetation growth and seasonal changes and are  
744 crucial for scaling simulations from site to regional levels. These gridded indices are  
745 directly derived from satellite reflectance bands. Large-area EFs can be obtained using  
746 LE and Hs calculations from ERA5 reanalysis data or can be derived using NDVI  
747 temperature triangulation (Venturini et al., 2004). PDSI can be obtained from ERA5  
748 reanalysis data. Thus, the model can be extended from the site scale to the regional and  
749 even global scale.

## 750 **5. Conclusion**

751 In this study, the FLAML-LUE model was applied to estimate GPP at four different  
752 time scales across 20 sites in China. Our findings indicate that the FLAML-LUE model  
753 excels at predicting GPP, capturing both its temporal variations and magnitude. It  
754 performs particularly well in mixed and evergreen coniferous forests, with mean  $R^2$   
755 values of 0.93 and 0.91, respectively. In addition, extending the time scale of input data  
756 can further enhance model accuracy. Specifically, the mean  $R^2$  of forest ecosystems  
757 increased from 0.89 to 0.93, that of grassland ecosystems from 0.79 to 0.83, and that of  
758 farmland ecosystems from 0.84 to 0.87. Analysis of the importance of the variables by



759 using the importance ranking method showed that vegetation index and temperature  
760 were the most important variables for GPP estimation in forest, grassland and farmland  
761 ecosystems, while the importance of moisture index was relatively low. Of which,  
762 temperature were the primary variables in the construction of FLAML-LUE models for  
763 forest, grassland and farmland ecosystems. The GPP time-series plots for each site  
764 indicated that the FLAML model was able to simulate seasonal dynamics more  
765 accurately at most of the sites but generally underestimated the GPP peaks. These  
766 results suggest that the FLAML-LUE model is highly capable of predicting GPP and  
767 has significant potential for scaling up GPP from flux footprints to larger areas,  
768 enhancing our understanding of carbon dynamics. However, it is important to note that  
769 the FLAML-LUE model demonstrates limited performance in alpine meadows,  
770 highlighting the need for further research to improve GPP modeling in these ecosystems  
771 in the future.

#### 772 **CRedit authorship contribution statement**

773 J.L., Y.Z. and J.W. conceived the study. J.L. collected and processed the data. J.L.  
774 and Y.Z. drafted the manuscript. A.W., Y.Z., R.L and W.D. funded the study, J.L., Y.Z.,  
775 A.W, W.F. and J.W. checked the negatives and touched up. All authors have read and  
776 agreed to the embellished manuscript.

#### 777 **Data availability**

778 A Fast Library for Automated Machine Learning & Tuning (FLAML) is a Python  
779 library and we can find detailed documentation about FLAML  
780 at <https://microsoft.github.io/FLAML/>. The observation data for these sites can be



781 downloaded from <https://www.scidb.cn/en/> (Science Data Bank). The flux observation  
782 data and the Python source code of the FLAML-LUE used in this paper are also  
783 archived on Zenodo (<https://doi.org/10.5281/zenodo.14542880>, Laijie, 2024).

#### 784 **Declaration of competing interest**

785 The authors declare that they have no known competing financial interests or  
786 personal relationships that could have appeared to influence the work reported in this  
787 paper.

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#### 798 **References**

- 799 Adams, M.D., Massey, F., Chastko, K., Cupini, C., 2020. Spatial modelling of particulate matter air  
800 pollution sensor measurements collected by community scientists while cycling, land use  
801 regression with spatial cross-validation, and applications of machine learning for data  
802 correction. *Atmos. Environ.* 230, 117479. <https://doi.org/10.1016/j.atmosenv.2020.117479>  
803 Alemohammad, S.H., Fang, B., Konings, A.G., Aires, F., Green, J.K., Kolassa, J., Miralles, D.,  
804 Prigent, C., Gentile, P., 2017. Water, Energy, and Carbon with Artificial Neural Networks  
805 (WECANN): a statistically based estimate of global surface turbulent fluxes and gross  
806 primary productivity using solar-induced fluorescence. *Biogeosciences* 14, 4101–4124.



- 807 <https://doi.org/10.5194/bg-14-4101-2017>
- 808 Barbour, M.T., 2021. Estimating Organic Carbon Burial in Freshwater Impoundments with a Rapid-  
809 Assessment Model and Geospatial Analysis (M.S.).
- 810 Beer, C., Reichstein, M., Tomelleri, E., Ciais, P., Jung, M., Carvalhais, N., Rödenbeck, C., Arain,  
811 M.A., Baldocchi, D., Bonan, G.B., Bondeau, A., Cescatti, A., Lasslop, G., Lindroth, A.,  
812 Lomas, M., Luysaert, S., Margolis, H., Oleson, K.W., Roupsard, O., Veenendaal, E., Viovy,  
813 N., Williams, C., Woodward, F.I., Papale, D., 2010. Terrestrial Gross Carbon Dioxide  
814 Uptake: Global Distribution and Covariation with Climate. *Science* 329, 834–838.  
815 <https://doi.org/10.1126/science.1184984>
- 816 Bhattacharyya, P., Neogi, S., Singha Roy, K., Rao, K.S., 2013. Gross primary production, ecosystem  
817 respiration and net ecosystem exchange in Asian rice paddy: An eddy covariance-based  
818 approach. *Curr. Sci.* 104, 67–75.
- 819 Bi, J., Knyazikhin, Y., Choi, S., Park, T., Barichivich, J., Ciais, P., Fu, R., Ganguly, S., Hall, F.,  
820 Hilker, T., Huete, A., Jones, M., Kimball, J., Lyapustin, A.I., Möttus, M., Nemani, R.R.,  
821 Piao, S., Poulter, B., Saleska, S.R., Saatchi, S.S., Xu, L., Zhou, L., Myneni, R.B., 2015.  
822 Sunlight mediated seasonality in canopy structure and photosynthetic activity of  
823 Amazonian rainforests. *Environ. Res. Lett.* 10, 064014. <https://doi.org/10.1088/1748-9326/10/6/064014>
- 824
- 825 Breiman, L., 2001. Random Forests. *Mach. Learn.* 45, 5–32.  
826 <https://doi.org/10.1023/A:1010933404324>
- 827 Cai, W., Ullah, S., Yan, L., Lin, Y., 2021. Remote Sensing of Ecosystem Water Use Efficiency: A  
828 Review of Direct and Indirect Estimation Methods. *Remote Sens.* 13, 2393.  
829 <https://doi.org/10.3390/rs13122393>
- 830 Cai, W., Yuan, W., Liang, S., Liu, S., Dong, W., Chen, Y., Liu, D., Zhang, H., 2014. Large  
831 Differences in Terrestrial Vegetation Production Derived from Satellite-Based Light Use  
832 Efficiency Models. *Remote Sens.* 6, 8945–8965. <https://doi.org/10.3390/rs6098945>
- 833 Chaney, N.W., Herman, J.D., Ek, M.B., Wood, E.F., 2016. Deriving global parameter estimates for  
834 the Noah land surface model using FLUXNET and machine learning. *J. Geophys. Res.*  
835 *Atmospheres* 121, 13,218–13,235. <https://doi.org/10.1002/2016JD024821>
- 836 Chang, X., Xing, Y., Gong, W., Yang, C., Guo, Z., Wang, D., Wang, J., Yang, H., Xue, G., Yang, S.,  
837 2023. Evaluating gross primary productivity over 9 ChinaFlux sites based on random forest  
838 regression models, remote sensing, and eddy covariance data. *Sci. Total Environ.* 875,  
839 162601. <https://doi.org/10.1016/j.scitotenv.2023.162601>
- 840 Chen, S.-P., Cui-Hai, Y.O.U., Zhong-Min, H.U., Zhi, C., Lei-Ming, Z., Qiu-Feng, W., 2020. Eddy  
841 covariance technique and its applications in flux observations of terrestrial ecosystems.  
842 *Chin. J. Plant Ecol.* 44, 291. <https://doi.org/10.17521/cjpe.2019.0351>
- 843 Chen, T., Guestrin, C., 2016. XGBoost: A Scalable Tree Boosting System, in: Proceedings of the  
844 22nd ACM SIGKDD International Conference on Knowledge Discovery and Data Mining,  
845 KDD '16. Association for Computing Machinery, New York, NY, USA, pp. 785–794.  
846 <https://doi.org/10.1145/2939672.2939785>
- 847 Coops, N.C., Waring, R.H., 2001. The use of multiscale remote sensing imagery to derive regional  
848 estimates of forest growth capacity using 3-PGS. *Remote Sens. Environ.* 75, 324–334.  
849 [https://doi.org/10.1016/S0034-4257\(00\)00176-0](https://doi.org/10.1016/S0034-4257(00)00176-0)
- 850 Cox, P.M., Betts, R.A., Jones, C.D., Spall, S.A., Totterdell, I.J., 2000. Erratum: Acceleration of



- 851 global warming due to carbon-cycle feedbacks in a coupled climate model. *Nature* 408,  
852 750–750. <https://doi.org/10.1038/35047138>
- 853 Ercoli, L., 1993. Relationship between nitrogen and chlorophyll content and spectral properties in  
854 maize leaves. *Eur. J. Agron.*
- 855 Erickson, N., Mueller, J., Shirkov, A., Zhang, H., Larroy, P., Li, M., Smola, A., 2020. AutoGluon-  
856 Tabular: Robust and Accurate AutoML for Structured Data.  
857 <https://doi.org/10.48550/arXiv.2003.06505>
- 858 Geurts, P., Ernst, D., Wehenkel, L., 2006. Extremely randomized trees. *Mach. Learn.* 63, 3–42.  
859 <https://doi.org/10.1007/s10994-006-6226-1>
- 860 Gherardi, L.A., Sala, O.E., 2020. Global patterns and climatic controls of belowground net carbon  
861 fixation. *Proc. Natl. Acad. Sci.* 117, 20038–20043.  
862 <https://doi.org/10.1073/pnas.2006715117>
- 863 Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., Moore, R., 2017. Google Earth  
864 Engine: Planetary-scale geospatial analysis for everyone. *Remote Sens. Environ., Big  
865 Remotely Sensed Data: tools, applications and experiences* 202, 18–27.  
866 <https://doi.org/10.1016/j.rse.2017.06.031>
- 867 Gosiewska, A., Biecek, P., 2019. auditor: an R Package for Model-Agnostic Visual Validation and  
868 Diagnostics. *R J.* 11, 85–98.
- 869 Harris, N.L., Gibbs, D.A., Baccini, A., Birdsey, R.A., de Bruin, S., Farina, M., Fatoyinbo, L., Hansen,  
870 M.C., Herold, M., Houghton, R.A., Potapov, P.V., Suarez, D.R., Roman-Cuesta, R.M.,  
871 Saatchi, S.S., Slay, C.M., Turubanova, S.A., Tyukavina, A., 2021. Global maps of twenty-  
872 first century forest carbon fluxes. *Nat. Clim. Change* 11, 234–240.  
873 <https://doi.org/10.1038/s41558-020-00976-6>
- 874 He, H., Zhang, Liyun, Gao, Y., Ren, X., Zhang, Li, Yu, G., Wang, S., 2015. Regional  
875 representativeness assessment and improvement of eddy flux observations in China. *Sci.  
876 Total Environ.* 502, 688–698. <https://doi.org/10.1016/j.scitotenv.2014.09.073>
- 877 Jiang, G., Sun, R., Zhang, L., Liu, S., Xu, Z., Qiao, C., 2014. Analysis of light use efficiency and  
878 gross primary productivity based on remote sensing data over a phragmites-dominated  
879 wetland in Zhangye, China, in: *Land Surface Remote Sensing II. Presented at the Land  
880 Surface Remote Sensing II, SPIE*, pp. 571–578. <https://doi.org/10.1117/12.2068840>
- 881 Jung, M., Reichstein, M., Margolis, H.A., Cescatti, A., Richardson, A.D., Arain, M.A., Arneth, A.,  
882 Bernhofer, C., Bonal, D., Chen, J., Gianelle, D., Gobron, N., Kiely, G., Kutsch, W., Lasslop,  
883 G., Law, B.E., Lindroth, A., Merbold, L., Montagnani, L., Moors, E.J., Papale, D.,  
884 Sottocornola, M., Vaccari, F., Williams, C., 2011. Global patterns of land-atmosphere fluxes  
885 of carbon dioxide, latent heat, and sensible heat derived from eddy covariance, satellite,  
886 and meteorological observations. *J. Geophys. Res. Biogeosciences* 116.  
887 <https://doi.org/10.1029/2010JG001566>
- 888 Jung, M., Schwalm, C., Migliavacca, M., Walther, S., Camps-Valls, G., Koirala, S., Anthoni, P.,  
889 Besnard, S., Bodesheim, P., Carvalhais, N., Chevallier, F., Gans, F., Goll, D.S., Haverd, V.,  
890 Köhler, P., Ichii, K., Jain, A.K., Liu, J., Lombardozi, D., Nabel, J.E.M.S., Nelson, J.A.,  
891 O’Sullivan, M., Pallandt, M., Papale, D., Peters, W., Pongratz, J., Rödenbeck, C., Sitch, S.,  
892 Tramontana, G., Walker, A., Weber, U., Reichstein, M., 2020. Scaling carbon fluxes from  
893 eddy covariance sites to globe: synthesis and evaluation of the FLUXCOM approach.  
894 *Biogeosciences* 17, 1343–1365. <https://doi.org/10.5194/bg-17-1343-2020>



- 895 Ke, G., Meng, Q., Finley, T., Wang, T., Chen, W., Ma, W., Ye, Q., Liu, T.-Y., 2017. LightGBM: A  
896 Highly Efficient Gradient Boosting Decision Tree, in: Guyon, I., Luxburg, U.V., Bengio,  
897 S., Wallach, H., Fergus, R., Vishwanathan, S., Garnett, R. (Eds.), ADVANCES IN  
898 NEURAL INFORMATION PROCESSING SYSTEMS 30 (NIPS 2017), Advances in  
899 Neural Information Processing Systems. Presented at the 31st Annual Conference on  
900 Neural Information Processing Systems (NIPS), Neural Information Processing Systems  
901 (nips), La Jolla.
- 902 Kong, D., Yuan, D., Li, H., Zhang, J., Yang, S., Li, Y., Bai, Y., Zhang, S., 2023. Improving the  
903 Estimation of Gross Primary Productivity across Global Biomes by Modeling Light Use  
904 Efficiency through Machine Learning. *Remote Sens.* 15, 2086.  
905 <https://doi.org/10.3390/rs15082086>
- 906 Landry, J.-S., Matthews, H.D., 2016. Non-deforestation fire vs. fossil fuel combustion: The source  
907 of CO<sub>2</sub> emissions affects the global carbon cycle and climate responses. *Biogeosciences*  
908 13, 2137–2149. <https://doi.org/10.5194/bg-13-2137-2016>
- 909 LeDell, E., Poirier, S., 2020. H2O AutoML: Scalable Automatic Machine Learning.
- 910 Lloyd, J., Taylor, J.A., 1994. On the Temperature Dependence of Soil Respiration. *Funct. Ecol.* 8,  
911 315–323. <https://doi.org/10.2307/2389824>
- 912 Lundberg, S.M., Lee, S.-I., 2017. A Unified Approach to Interpreting Model Predictions, in: Guyon,  
913 I., Luxburg, U.V., Bengio, S., Wallach, H., Fergus, R., Vishwanathan, S., Garnett, R. (Eds.),  
914 ADVANCES IN NEURAL INFORMATION PROCESSING SYSTEMS 30 (NIPS 2017),  
915 Advances in Neural Information Processing Systems. Presented at the 31st Annual  
916 Conference on Neural Information Processing Systems (NIPS), Neural Information  
917 Processing Systems (nips), La Jolla.
- 918 Mahadevan, P., Wofsy, S.C., Matross, D.M., Xiao, X., Dunn, A.L., Lin, J.C., Gerbig, C., Munger,  
919 J.W., Chow, V.Y., Gottlieb, E.W., 2008. A satellite-based biosphere parameterization for net  
920 ecosystem CO<sub>2</sub> exchange: Vegetation Photosynthesis and Respiration Model (VPRM).  
921 *Glob. Biogeochem. Cycles* 22. <https://doi.org/10.1029/2006GB002735>
- 922 Melanie, 2023. TPOT: All about this Machine Learning Python library. *Data Sci. Courses*  
923 *DataScientest*. URL [https://datascientest.com/en/tpot-all-about-this-machine-learning-](https://datascientest.com/en/tpot-all-about-this-machine-learning-python-library)  
924 [python-library](https://datascientest.com/en/tpot-all-about-this-machine-learning-python-library) (accessed 6.2.24).
- 925 Menefee, D., Lee, T.O., Flynn, K.C., Chen, J., Abraha, M., Baker, J., Suyker, A., 2023. Machine  
926 learning algorithms improve MODIS GPP estimates in United States croplands. *Front.*  
927 *Remote Sens.* 4. <https://doi.org/10.3389/frsen.2023.1240895>
- 928 Pei, Y., Dong, J., Zhang, Y., Yuan, W., Doughty, R., Yang, J., Zhou, D., Zhang, L., Xiao, X., 2022.  
929 Evolution of light use efficiency models: Improvement, uncertainties, and implications.  
930 *Agric. For. Meteorol.* 317, 108905. <https://doi.org/10.1016/j.agrformet.2022.108905>
- 931 Peltoniemi, M., Pulkkinen, M., Kolari, P., Duursma, R.A., Montagnani, L., Wharton, S., Lagergren,  
932 F., Takagi, K., Verbeeck, H., Christensen, T., Vesala, T., Falk, M., Loustau, D., Mäkelä, A.,  
933 2012. Does canopy mean nitrogen concentration explain variation in canopy light use  
934 efficiency across 14 contrasting forest sites? *Tree Physiol.* 32, 200–218.  
935 <https://doi.org/10.1093/treephys/tpr140>
- 936 Potter, C.S., Randerson, J.T., Field, C.B., Matson, P.A., Vitousek, P.M., Mooney, H.A., Klooster,  
937 S.A., 1993. Terrestrial ecosystem production: A process model based on global satellite and  
938 surface data. *Glob. Biogeochem. Cycles* 7, 811–841. <https://doi.org/10.1029/93GB02725>



- 939 Prokhorenkova, L., Gusev, G., Vorobev, A., Dorogush, A.V., Gulin, A., 2018. CatBoost: unbiased  
940 boosting with categorical features, in: Bengio, S., Wallach, H., Larochelle, H., Grauman,  
941 K., CesaBianchi, N., Garnett, R. (Eds.), ADVANCES IN NEURAL INFORMATION  
942 PROCESSING SYSTEMS 31 (NIPS 2018), Advances in Neural Information Processing  
943 Systems. Presented at the 32nd Conference on Neural Information Processing Systems  
944 (NIPS), Neural Information Processing Systems (nips), La Jolla.
- 945 Reichstein, M., Camps-Valls, G., Stevens, B., Jung, M., Denzler, J., Carvalhais, N., Prabhat, 2019.  
946 Deep learning and process understanding for data-driven Earth system science. *Nature* 566,  
947 195–204. <https://doi.org/10.1038/s41586-019-0912-1>
- 948 Reichstein, M., Falge, E., Baldocchi, D., Papale, D., Aubinet, M., Berbigier, P., Bernhofer, C.,  
949 Buchmann, N., Gilmanov, T., Granier, A., Grünwald, T., Havránková, K., Ilvesniemi, H.,  
950 Janous, D., Knohl, A., Laurila, T., Lohila, A., Loustau, D., Matteucci, G., Meyers, T.,  
951 Miglietta, F., Ourcival, J.-M., Pumpanen, J., Rambal, S., Rotenberg, E., Sanz, M., Tenhunen,  
952 J., Seufert, G., Vaccari, F., Vesala, T., Yakir, D., Valentini, R., 2005. On the separation of  
953 net ecosystem exchange into assimilation and ecosystem respiration: review and improved  
954 algorithm. *Glob. Change Biol.* 11, 1424–1439. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2005.001002.x)  
955 [2486.2005.001002.x](https://doi.org/10.1111/j.1365-2486.2005.001002.x)
- 956 Rosebrock, A., 2019. Auto-Keras and AutoML: A Getting Started Guide. PyImageSearch. URL  
957 <https://pyimagesearch.com/2019/01/07/auto-keras-and-automl-a-getting-started-guide/>  
958 (accessed 6.2.24).
- 959 Running, S.W., Nemani, R.R., Heinsch, F.A., Zhao, M., Reeves, M., Hashimoto, H., 2004. A  
960 Continuous Satellite-Derived Measure of Global Terrestrial Primary Production.  
961 *BioScience* 54, 547–560. [https://doi.org/10.1641/0006-](https://doi.org/10.1641/0006-3568(2004)054[0547:ACSMOG]2.0.CO;2)  
962 [3568\(2004\)054\[0547:ACSMOG\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2004)054[0547:ACSMOG]2.0.CO;2)
- 963 Schmid, H.P., 2002. Footprint modeling for vegetation atmosphere exchange studies: a review and  
964 perspective. *Agric. For. Meteorol., FLUXNET 2000 Synthesis* 113, 159–183.  
965 [https://doi.org/10.1016/S0168-1923\(02\)00107-7](https://doi.org/10.1016/S0168-1923(02)00107-7)
- 966 Sellers, P.J., Schimel, D.S., Moore, B., Liu, J., Eldering, A., 2018. Observing carbon cycle–climate  
967 feedbacks from space. *Proc. Natl. Acad. Sci.* 115, 7860–7868.  
968 <https://doi.org/10.1073/pnas.1716613115>
- 969 Sims, D.A., Brzostek, E.R., Rahman, A.F., Dragoni, D., Phillips, R.P., 2014. An improved approach  
970 for remotely sensing water stress impacts on forest C uptake. *Glob. Change Biol.* 20, 2856–  
971 2866. <https://doi.org/10.1111/gcb.12537>
- 972 Taylor, K.E., 2001. Summarizing multiple aspects of model performance in a single diagram. *J.*  
973 *Geophys. Res. Atmospheres* 106, 7183–7192. <https://doi.org/10.1029/2000JD900719>
- 974 Thornton, C., Hutter, F., Hoos, H.H., Leyton-Brown, K., 2013. Auto-WEKA: Combined Selection  
975 and Hyperparameter Optimization of Classification Algorithms.  
976 <https://doi.org/10.48550/arXiv.1208.3719>
- 977 Tramontana, G., Jung, M., Schwalm, C.R., Ichii, K., Camps-Valls, G., Ráduly, B., Reichstein, M.,  
978 Arain, M.A., Cescatti, A., Kiely, G., Merbold, L., Serrano-Ortiz, P., Sickert, S., Wolf, S.,  
979 Papale, D., 2016. Predicting carbon dioxide and energy fluxes across global FLUXNET  
980 sites with regression algorithms. *Biogeosciences* 13, 4291–4313.  
981 <https://doi.org/10.5194/bg-13-4291-2016>
- 982 Venturini, V., Bisht, G., Islam, S., Jiang, L., 2004. Comparison of evaporative fractions estimated



- 983 from AVHRR and MODIS sensors over South Florida. *Remote Sens. Environ.* 93, 77–86.  
984 <https://doi.org/10.1016/j.rse.2004.06.020>
- 985 Wang, C., Wu, Q., Weimer, M., Zhu, E., 2021. FLAML: A Fast and Lightweight AutoML Library.  
986 <https://doi.org/10.48550/arXiv.1911.04706>
- 987 Wang, H., Shao, W., Hu, Y., Cao, W., Zhang, Y., 2023. Assessment of Six Machine Learning  
988 Methods for Predicting Gross Primary Productivity in Grassland. *Remote Sens.* 15, 3475.  
989 <https://doi.org/10.3390/rs15143475>
- 990 Wang, J., Liu, J., Cao, M., Liu, Y., Yu, G., Li, G., Qi, S., Li, K., 2011. Modelling carbon fluxes of  
991 different forests by coupling a remote-sensing model with an ecosystem process model. *Int.*  
992 *J. Remote Sens.* 32, 6539–6567. <https://doi.org/10.1080/01431161.2010.512933>
- 993 Wang, Y., Li, R., Hu, J., Fu, Y., Duan, J., Cheng, Y., 2021. Daily estimation of gross primary  
994 production under all sky using a light use efficiency model coupled with satellite passive  
995 microwave measurements. *Remote Sens. Environ.* 267, 112721.  
996 <https://doi.org/10.1016/j.rse.2021.112721>
- 997 Wei, S., Yi, C., Fang, W., Hendrey, G., 2017. A global study of GPP focusing on light-use efficiency  
998 in a random forest regression model. *Ecosphere* 8, e01724.  
999 <https://doi.org/10.1002/ecs2.1724>
- 1000 Wu, C., Munger, J.W., Niu, Z., Kuang, D., 2010. Comparison of multiple models for estimating  
1001 gross primary production using MODIS and eddy covariance data in Harvard Forest.  
1002 *Remote Sens. Environ.* 114, 2925–2939. <https://doi.org/10.1016/j.rse.2010.07.012>
- 1003 Xiao, J., Chevallier, F., Gomez, C., Guanter, L., Hicke, J.A., Huete, A.R., Ichii, K., Ni, W., Pang, Y.,  
1004 Rahman, A.F., Sun, G., Yuan, W., Zhang, L., Zhang, X., 2019. Remote sensing of the  
1005 terrestrial carbon cycle: A review of advances over 50 years. *Remote Sens. Environ.* 233,  
1006 111383. <https://doi.org/10.1016/j.rse.2019.111383>
- 1007 Xiao, X., Braswell, B., Zhang, Q., Boles, S., Frohking, S., Moore, B., 2003. Sensitivity of vegetation  
1008 indices to atmospheric aerosols: continental-scale observations in Northern Asia. *Remote*  
1009 *Sens. Environ.* 84, 385–392. [https://doi.org/10.1016/S0034-4257\(02\)00129-3](https://doi.org/10.1016/S0034-4257(02)00129-3)
- 1010 Xie, J., Sun, G., Chu, H.-S., Liu, J., McNulty, S.G., Noormets, A., John, R., Ouyang, Z., Zha, T., Li,  
1011 H., Guan, W., Chen, J., 2014. Long-term variability in the water budget and its controls in  
1012 an oak-dominated temperate forest. *Hydrol. Process.* 28, 6054–6066.  
1013 <https://doi.org/10.1002/hyp.10079>
- 1014 Xie, M., Ma, X., Wang, Y., Li, C., Shi, H., Yuan, X., Hellwich, O., Chen, C., Zhang, W., Zhang, C.,  
1015 Ling, Q., Gao, R., Zhang, Y., Ochege, F.U., Frankl, A., De Maeyer, P., Buchmann, N.,  
1016 Feigenwinter, I., Olesen, J.E., Juszczak, R., Jacotot, A., Korrensalo, A., Pitacco, A.,  
1017 Varlagin, A., Shekhar, A., Lohila, A., Carrara, A., Brut, A., Kruijt, B., Loubet, B., Heinesch,  
1018 B., Chojnicki, B., Helfter, C., Vincke, C., Shao, C., Bernhofer, C., Brümmer, C., Wille, C.,  
1019 Tuittila, E.-S., Nemitz, E., Meggio, F., Dong, G., Lanigan, G., Niedrist, G., Wohlfahrt, G.,  
1020 Zhou, G., Goded, I., Gruenwald, T., Olejnik, J., Jansen, J., Neiryneck, J., Tuovinen, J.-P.,  
1021 Zhang, J., Klumpp, K., Pilegaard, K., Šigut, L., Klemedtsson, L., Tezza, L., Hörtnagl, L.,  
1022 Urbaniak, M., Roland, M., Schmidt, M., Sutton, M.A., Hehn, M., Saunders, M., Mauder,  
1023 M., Aurela, M., Korkiakoski, M., Du, M., Vendrame, N., Kowalska, N., Leahy, P.G.,  
1024 Alekseychik, P., Shi, P., Weslien, P., Chen, S., Fares, S., Friborg, T., Tallec, T., Kato, T.,  
1025 Sachs, T., Maximov, T., di Cella, U.M., Moderow, U., Li, Y., He, Y., Kosugi, Y., Luo, G.,  
1026 2023. Monitoring of carbon-water fluxes at Eurasian meteorological stations using random



- 1027 forest and remote sensing. *Sci. Data* 10, 587. <https://doi.org/10.1038/s41597-023-02473-9>
- 1028 Xie, X., Li, A., 2020. An Adjusted Two-Leaf Light Use Efficiency Model for Improving GPP  
1029 Simulations Over Mountainous Areas. *J. Geophys. Res. Atmospheres* 125, e2019JD031702.  
1030 <https://doi.org/10.1029/2019JD031702>
- 1031 Yu, G., Ren, W., Chen, Z., Zhang, Leiming, Wang, Q., Wen, X., He, N., Zhang, Li, Fang, H., Zhu,  
1032 X., Gao, Y., Sun, X., 2016. Construction and progress of Chinese terrestrial ecosystem  
1033 carbon, nitrogen and water fluxes coordinated observation. *J. Geogr. Sci.* 26, 803–826.  
1034 <https://doi.org/10.1007/s11442-016-1300-5>
- 1035 Yuan, W., Cai, W., Xia, J., Chen, J., Liu, S., Dong, W., Merbold, L., Law, B., Arain, A., Beringer, J.,  
1036 Bernhofer, C., Black, A., Blanken, P.D., Cescatti, A., Chen, Y., Francois, L., Gianelle, D.,  
1037 Janssens, I.A., Jung, M., Kato, T., Kiely, G., Liu, D., Marcolla, B., Montagnani, L., Raschi,  
1038 A., Rouspard, O., Varlagin, A., Wohlfahrt, G., 2014. Global comparison of light use  
1039 efficiency models for simulating terrestrial vegetation gross primary production based on  
1040 the LaThuile database. *Agric. For. Meteorol.* 192–193, 108–120.  
1041 <https://doi.org/10.1016/j.agrformet.2014.03.007>
- 1042 Yuan, W., Liu, S., Yu, G., Bonnefond, J.-M., Chen, J., Davis, K., Desai, A.R., Goldstein, A.H.,  
1043 Gianelle, D., Rossi, F., Suyker, A.E., Verma, S.B., 2010. Global estimates of  
1044 evapotranspiration and gross primary production based on MODIS and global meteorology  
1045 data. *Remote Sens. Environ.* 114, 1416–1431. <https://doi.org/10.1016/j.rse.2010.01.022>
- 1046 Yuan, W., Liu, S., Zhou, Guangsheng, Zhou, Guoyi, Tieszen, L.L., Baldocchi, D., Bernhofer, C.,  
1047 Gholz, H., Goldstein, A.H., Goulden, M.L., Hollinger, D.Y., Hu, Y., Law, B.E., Stoy, P.C.,  
1048 Vesala, T., Wofsy, S.C., 2007. Deriving a light use efficiency model from eddy covariance  
1049 flux data for predicting daily gross primary production across biomes. *Agric. For. Meteorol.*  
1050 143, 189–207. <https://doi.org/10.1016/j.agrformet.2006.12.001>
- 1051 Zhang, C., Tian, X., Zhao, Y., Lu, J., 2023. Automated machine learning-based building energy load  
1052 prediction method. *J. Build. Eng.* 80, 108071. <https://doi.org/10.1016/j.job.2023.108071>
- 1053 Zhang, W.L., Chen, S.P., Chen, J., Wei, L., Han, X.G., Lin, G.H., 2007. Biophysical regulations of  
1054 carbon fluxes of a steppe and a cultivated cropland in semiarid Inner Mongolia. *Agric. For.*  
1055 *Meteorol.* 146, 216–229. <https://doi.org/10.1016/j.agrformet.2007.06.002>
- 1056 Zhang, Y., Song, C., Sun, G., Band, L.E., Noormets, A., Zhang, Q., 2015. Understanding moisture  
1057 stress on light use efficiency across terrestrial ecosystems based on global flux and remote-  
1058 sensing data. *J. Geophys. Res. Biogeosciences* 120, 2053–2066.  
1059 <https://doi.org/10.1002/2015JG003023>
- 1060 Zhang, Y., Ye, A., 2022. Uncertainty analysis of multiple terrestrial gross primary productivity  
1061 products. *Glob. Ecol. Biogeogr.* 31, 2204–2218. <https://doi.org/10.1111/geb.13578>
- 1062 Zhang, Z., Guo, J., Jin, S., Han, S., 2023. Improving the ability of PRI in light use efficiency  
1063 estimation by distinguishing sunlit and shaded leaves in rice canopy. *Int. J. Remote Sens.*  
1064 44, 5755–5767. <https://doi.org/10.1080/01431161.2023.2252165>
- 1065 Zhao, W.L., Gentile, P., Reichstein, M., Zhang, Y., Zhou, S., Wen, Y., Lin, C., Li, X., Qiu, G.Y.,  
1066 2019. Physics-Constrained Machine Learning of Evapotranspiration. *Geophys. Res. Lett.*  
1067 46, 14496–14507. <https://doi.org/10.1029/2019GL085291>
- 1068 Zheng, Z., Fiore, A.M., Westervelt, D.M., Milly, G.P., Goldsmith, J., Karambelas, A., Curci, G.,  
1069 Randles, C.A., Paiva, A.R., Wang, C., Wu, Q., Dey, S., 2023. Automated Machine Learning  
1070 to Evaluate the Information Content of Tropospheric Trace Gas Columns for Fine Particle



1071            Estimates Over India: A Modeling Testbed. J. Adv. Model. Earth Syst. 15, e2022MS003099.  
1072            <https://doi.org/10.1029/2022MS003099>  
1073