FLAML version 2.3.3 model-based assessment of gross primary productivity at forest, grassland, and cropland ecosystem sites

Jie Lai [a, b], Yuan Zhang [a], Anzhi Wang [a], Wenli Fei [a], Yiwei Diao [c], Rongping Li [d], Jiabin Wu [a, *]

[a] CAS Key Laboratory of Forest Ecology and Silviculture, Institute of Applied Ecology, Chinese Academy of Sciences, Shenyang 110016, China

[b] University of Chinese Academy of Sciences, Beijing 101408, China

[c] Key Laboratory of Ecosystem Carbon Source and Sink, China Meteorological Administration (ECSS-CMA), Wuxi University, Wuxi, 214105, China

[d] Institute of Atmospheric Environment, China Meteorological Administration, Shenyang 110016, China

* Correspondence: wujb@iae.ac.cn

Abstract

Accurately estimating Gross Primary Productivity (GPP) in terrestrial ecosystems is essential for understanding the global carbon cycle. Satellite-based Light Use Efficiency (LUE) models are commonly employed for simulating GPP. However, the variables and algorithms related to environmental limiting factors differ significantly across various LUE models, leading to high uncertainty in GPP estimation. In this work, we developed a series of FLAML-LUE models with different variable combinations. These models utilize the Fast Lightweight Automated Machine Learning (FLAML) framework, using variables of LUE models, to investigate the potential of estimating site-scale GPP. Incorporating meteorological data, eddy covariance measurements, and remote sensing indices, we employed FLAML-LUE models to assess the impact of various variable combinations on GPP across different temporal scales, including daily, 8-day, 16-day, and monthly intervals. Cross-validation analyses indicated that the FLAML-LUE model performs excellently in GPP prediction, accurately simulating

both its temporal variations and magnitude, particularly in mixed forests and coniferous

forests, with average R^2 values for daily-scale simulations reaching 0.92 and 0.91,

respectively. However, the model performed less effectively in alpine shrubland and

typical grassland ecosystems, though it still outperformed both MODIS GPP and PML

GPP in terms of performance. Furthermore, the model's adaptability under extreme

climate conditions was evaluated, and the results showed that high temperatures and

high VPD lead to a slight decrease in model accuracy, though R^2 remains around 0.8.

Under drought conditions, the model's performance improved slightly in croplands and

evergreen broadleaf forests, although it declined at some sites. This study offers an

approach to estimate GPP fluxes and evaluate the impact of variables on GPP estimation.

It has the potential to be applied in predicting GPP for different vegetation types at a

regional scale.

Keywords: Light Use Efficiency; Gross Primary Productivity; Automated Machine

Learning; Fast Lightweight Automated Machine Learning

1. Introduction

The global carbon budget mainly addresses the carbon reserves in the atmosphere,

oceans, and terrestrial (Barbour, 2021), with terrestrial ecosystems being vital for

regulating the global carbon cycle (Gherardi and Sala, 2020; Landry and Matthews,

2016). Terrestrial ecosystems primarily absorb atmospheric carbon dioxide through the

process of plant photosynthesis, which is crucial for regulating climate and mitigating

global warming (Sellers et al., 2018; Beer et al., 2010; Cox et al., 2000). Gross primary

productivity (GPP) is a critical measure of carbon exchange between terrestrial

ecosystems and the atmosphere (Menefee et al., 2023). Accurate quantification of GPP

is essential for evaluating carbon balance and comprehending the response of terrestrial

ecosystems to climate change (Sellers et al., 2018).

The primary method currently used for measuring CO_2 exchange between

ecosystems and the atmosphere is the eddy covariance technique (Chen et al., 2020; Yu

et al., 2016). This technique precisely measures Net Ecosystem Exchange (NEE), which

is the difference between the carbon released by ecosystem respiration (ER) and the

carbon taken up by photosynthesis (Bhattacharyya et al., 2013). While flux observation

sites based on the eddy covariance (EC) technique can dynamically monitor site-scale

carbon fluxes, expanding their findings to larger regional scales remains challenging,

mainly due to the sparse and spatially non-uniform distribution of flux sites (Xie et al.,

2023; Jung et al., 2020). Remote sensing data is widely used in ecosystem carbon cycle

research as it can provide information on the spatial dynamics of vegetation and climate

at a larger scale (Xiao et al., 2019). By extrapolating spatially using models that

incorporate remote sensing and climate data, it is possible to estimate global GPP based

on observations of GPP at the site level. Therefore, remote sensing has become a crucial

data resource for estimating GPP (Cai et al., 2021; Xiao et al., 2019; Wang et al., 2011).

Light Use Efficiency (LUE) models based on satellite observations are commonly

employed to simulate GPP (Zhang et al., 2023; Zhang et al., 2015; Jiang et al., 2014).

Such models include Physiological Principles Predicting Growth using Satellite data

(3-PGS, Coops and Waring, 2001), the Carnegie-Ames- Stanford Approach Model

(CASA, Potter et al., 1993), the Eddy Covariance–Light Use Efficiency Model (EC-

72　LUE, Yuan et al., 2010, 2007), the MODIS Global Terrestrial Gross and Net Primary

73　Production (MOD17, Running et al., 2004), the Vegetation Photosynthesis Model

74　(VPM, Xiao et al., 2003), and the Vegetation Photosynthesis and Respiration Model

75　(VPRM, Mahadevan et al., 2008). Among all the forecasting methods (Coops and

76　Waring, 2001; Potter et al., 1993), the LUE model is widely utilized for simulating the

77　spatio-temporal dynamics of GPP due to its simplicity and strong theoretical foundation.

78　Over the past few decades, numerous GPP models utilizing LUE have been developed

79　(Pei et al., 2022).

80　　Despite significant advances in LUE theory for GPP estimation, uncertainties

81　persist in GPP models utilizing LUE. Firstly, differences in environmental limiting

82　factors among various LUE models contribute significantly to the uncertainty in GPP

83　estimation. For example, Cai et al. (2014) found a strong positive correlation between

84　water effectiveness and GPP estimate factors, while other studies found that the LUE

85　model estimates of GPP were strongly correlated with the vegetation index, which

86　affects the photosynthetic capacity of vegetation through leaf nitrogen content

87　(Peltoniemi et al., 2012; Ercoli, 1993).

88　　Recently, with the massive accumulation of satellite data and ground-based

89　observations, more and more studies have applied machine learning (ML) methods to

90　model ecosystem processes (Zhao et al., 2019; Alemohammad et al., 2017; Chaney et

91　al., 2016). ML is a modeling solution that differs from simple regression models and

92　complex simulation models in its approach. It is very effective in handling large-scale

93　multivariate data with complex relationships between predictors (Reichstein et al., 2019;

94 Tramontana et al., 2016). These data-driven models are particularly suited for capturing

95 nonlinear ecosystem dynamics but often require large training datasets and may lack

96 explicit links to real-world processes. However, their ability to uncover spatial patterns

97 without process-based constraints makes them valuable for spatial predictions.

98 Consequently, ML-based approaches have gained popularity in recent years. For

99 example, Kong et al. (2023) developed a hybrid model that combines ML and LUE

100 model to estimate GPP. This hybrid model improves the LUE model by integrating a

101 machine learning approach (MLP, multi-layer perceptron), and estimates GPP using the

102 MLP-based LUE framework along with additional required inputs. Chang et al. (2023)

103 constructed RFR-LUE models that utilize the Random Forest Regression (RFR)

104 algorithm with variables of LUE models to assess the potential of site-scale GPP

105 estimation.

106　　Lately, Automated Machine Learning (AutoML) has demonstrated significant

107 potential in constructing data-driven models automatically (Zheng et al., 2023).

108 Numerous sophisticated open-source AutoML frameworks have been suggested by

109 computer scientists, including Automated WEKA (Auto-WEKA, Thornton et al., 2013),

110 H2O AutoML (H2O, LeDell and Poirier, 2020), Tree-based Pipeline Optimization Tool

111 (TPOT, Melanie, 2023), Automated Machine Learning with Gluon (AutoGluon,

112 Erickson et al., 2020), Fast Lightweight Automated Machine Learning (FLAML, Wang

113 et al., 2021), and AutoKeras (Rosebrock, 2019). These frameworks are extensively used

114 in finance, manufacturing, healthcare, and mobile communications, among other fields

115 (Adams et al., 2020), with FLAML being particularly favored for its efficiency in rapid

prototyping and deployment in research and production settings. FLAML is a powerful framework for AutoML, known for its speed in identifying top-performing models and optimal hyperparameters through parallel optimization and smart search algorithms. FLAML integrates several effective search strategies, outperforming other leading AutoML libraries on large benchmarks even with constrained budgets (Wang et al., 2021).

In this research, a new model called FLAML-LUE was created by combining FLAML model with LUE-based models, the latter provides the key variables of vegetation growth for modeling. Such knowledge-and-data-driven models aim to reduce the large uncertainty in estimating GPP. The specific objectives of this study are: (1) to evaluate the overall performance of models using different input variables, including the fraction of photosynthetically active radiation absorbed by vegetation (fPAR) and various water stress indicators, across multiple sites and vegetation types based on eddy covariance observations; (2) to assess model performance under extreme climatic conditions, such as high temperature, elevated vapor pressure deficit (VPD), and drought.

2. Materials and methods

2.1 Site description

Figure 1 displays the geographical locations of the 20 flux sites selected for the study. These sites are situated in various climatic zones and ecosystem types including forest, grassland, and cropland. The observation data for these sites comes from the Science Data Bank (SDB, https://www.scidb.cn/en/). Detailed information about the

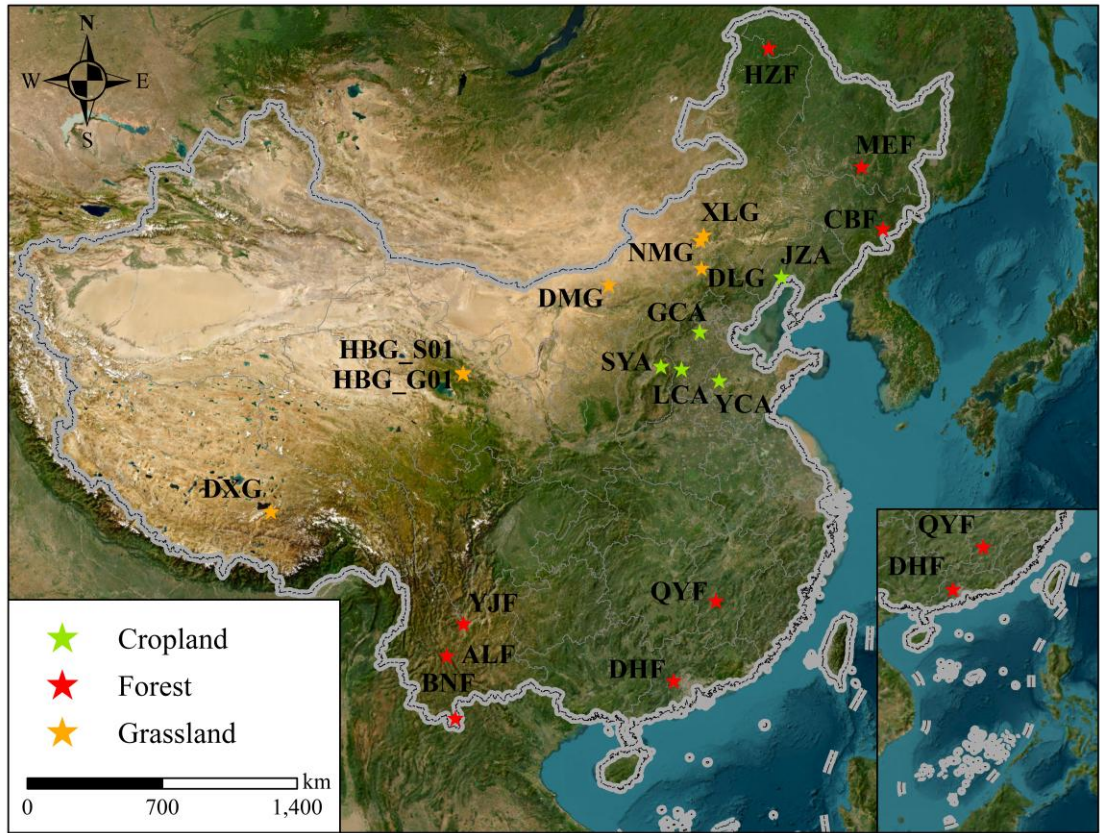138   sites is provided in **Table 1**.

140   **Figure 1.** The location map of the flux site is based on the map approved by the National Surveying
141   and Mapping Bureau of China (Approval No. GS (2019)1822). The topographic map is derived
142   from data provided by Esri, Maxar, Earthstar Geographics, and the GIS User Community (Service
143   Layer Credits).

144   **Table 1**

145   Basic information on the 20 flux stations.

Site	Longitude (°E)	Latitude (°N)	Ecosystem type	Time Range	Classified
HZF	123.018	51.781	Forest	2014-2018	NF
MEF	127.668	45.417	Forest	2016-2018	DBF
CBF	128.096	42.403	Forest	2003-2010	MF
QYF	115.058	26.741	Forest	2003-2010	NF
ALF	101.028	24.541	Forest	2009-2013	EBF
DHF	112.534	23.173	Forest	2003-2010	MF
BNF	101.577	21.614	Forest	2003-2015	EBF
YJF	101.827	26.080	Forest	2013-2015	SAV
XLG	116.671	43.554	Grassland	2006-2014	GRA
NMG	116.404	43.326	Grassland	2003-2010	Grassland
DLG	116.284	42.047	Grassland	2006-2015	Grassland
DMG	110.328	41.644	Grassland	2015-2018	Grassland

HBG_G01	101.313	37.613	Grassland	2015-2020	MEA
HBG_S01	101.331	37.665	Grassland	2003-2013	SHR
DXG	91.066	30.497	Grassland	2003-2010	MEA
JZA	121.202	41.148	Cropland	2005-2014	SC
GCA	115.735	39.149	Cropland	2020-2022	DC
SYA	113.200	37.750	Cropland	2012-2014	SC
LCA	114.413	37.531	Cropland	2013-2017	DC
YCA	116.570	36.829	Cropland	2003-2010	DC

Note: Vegetation types in the table are classified based on the land cover characteristics of each flux site and are used in subsequent model simulations. NF: Needle-leaved Forest; DBF: Deciduous Broadleaved Forest; MF: Mixed Forest; EBF: Evergreen Broadleaved Forest; SAV: Savannas; GRA: Typical Grassland; MEA: Alpine Meadow; SHR: Shrubs; SC: Single Cropping; DC: Double Cropping.

2.2 Data

2.2.1 Eddy covariance data

EC data were collected at 20 sites, including 8 forests sites, 7 grasslands sites, and 5 cropland sites (**Table 1**). Flux and meteorological data were collected every half hour from the mentioned sites. The flux and meteorological data underwent standardized quality control and corrections, ensuring high reliability and making them suitable for validating various GPP models and remote sensing observations. However, ER data were missing at some sites (DLG, LCA, XLG). To address this, the Lloyd & Taylor equation (Reichstein et al., 2005; Lloyd and Taylor, 1994) was applied to estimate ER based on nocturnal respiration data. Daytime and nighttime periods were distinguished using shortwave radiation (Rg), with a threshold of 10 W/m². The temperature–response relationship derived from nighttime ER was extrapolated to estimate daytime ER. This is a commonly used method for processing flux data at flux tower sites.

$$R_{eco} = R_{eco.ref} \exp \left(E_0 \left(\frac{1}{T_{ref} - T_0} - \frac{1}{T_{air} - T_0} \right) \right) \tag{1}$$

$$GPP = ER - NEE \tag{2}$$

164 In equation (1), R_{eco} is the nocturnal ecosystem respiration value, $R_{eco.ref}$ is the ER value

165 at the reference temperature, T_{ref} is the reference temperature (298.16K), E_0 is constant

166 (308.56K), T_0 is the minimum temperature at which respiration stops, set at 227.13K,

167 and T_{air} is the air temperature or soil temperature (K). Daytime GPP was then estimated

168 by subtracting NEE from the total daytime ER.

169 **2.2.2 MODIS data**

170 In this study, remote sensing data were primarily obtained from the Moderate

171 Resolution Imaging Spectroradiometer (MODIS). MODIS data offer a spatial

172 resolution of 500 meters and an 8-day temporal resolution. These datasets were sourced

173 from the Google Earth Engine (GEE) platform (Gorelick et al., 2017). To align with the

174 spatial and temporal scales of flux tower observations and reduce the impact of missing

175 data (Schmid, 2002), we applied the Savitzky-Golay smoothing filter with a window

176 size of 10 to process the vegetation indices. Vegetation and water indices derived from

177 MODIS data included the enhanced vegetation index (EVI), normalized difference

178 vegetation index (NDVI), and land surface water index (LSWI), which were calculated

179 using the formulas presented in **Table 2**.

180 **2.2.3 ERA5-LAND**

181 ERA5-Land (Hersbach et al., 2020) is a global high-resolution reanalysis dataset

182 produced by the European Centre for Medium-Range Weather Forecasts (ECMWF)

183 under the Copernicus Climate Change Service (C3S). It provides hourly land surface

184 variables at a spatial resolution of 0.1°, generated using a dedicated land surface model

185 driven by the ERA5 climate reanalysis. The dataset integrates advanced land surface

186   modeling and data assimilation techniques, offering a wide range of variables such as

187   air temperature, soil moisture, precipitation, and snow depth. In this study, site-specific

188   variables including air temperature (T), soil water content (SW), precipitation (Pre),

189   and leaf area index (LAI) were extracted from ERA5-Land. In addition,

190   photosynthetically active radiation (PAR), evapotranspiration fraction (EF), VPD and

191   relative humidity (RH) were calculated and derived from available ERA5-Land

192   variables using GEE.

193   **2.2.4 SPEI Database, Version 2.10**

194   The SPEI Database, Version 2.10 (Vicente-Serrano et al., 2010) provides global

195   data of the Standardized Precipitation-Evapotranspiration Index (SPEI) across temporal

196   scales from 1 to 48 months. Developed by the Climatic Research Unit (CRU), this

197   dataset combines precipitation and potential evapotranspiration (PET) to assess drought

198   conditions. Negative SPEI values indicate drought, while positive values signify wet

199   periods. In this study, SPEI values less than -1.5 were used to identify drought months

200   at each flux station, highlighting significant moisture deficits that affect vegetation

201   growth and ecosystem productivity (Qian et al., 2024).

202   **2.3 Model Construction**

203   Most LUE models typically incorporate four main groups of variables: PAR, fPAR,

204   temperature, and water-related stress indicators. In previous studies, vegetation indices

205   such as EVI, NDVI, or LAI have been widely used as proxies for fPAR, representing

206   the fraction of PAR absorbed by the plant canopy (Chang et al., 2023; Qian et al., 2024).

207   In this study, we selected six water-related indicators based on their ecological

208 relevance: plant-based indicators (LSWI and EF), soil-based indicators (SW), and

209 atmospheric indicators (VPD, precipitation, and relative humidity). Previous research

210 has shown that plant-based indicators like LSWI and EF effectively capture canopy-

211 level drought stress (Anderson et al., 2007; Xiao et al., 2004). Soil moisture regulates

212 water availability at the root level, which strongly influences photosynthetic activity,

213 particularly under water-limited conditions (Vicca et al., 2014; Reichstein et al., 2007).

214 Meanwhile, atmospheric indicators such as VPD, precipitation, and RH influence

215 stomatal conductance and transpiration by altering the vapor pressure gradient between

216 the leaf surface and the surrounding air (Wang et al., 2018; Novick et al., 2016). To

217 assess the relative importance of these different types of water stress indicators in

218 estimating GPP, we developed machine learning models using each group individually.

219 This allowed us to identify the most effective type of water-related variable for

220 simulating GPP across diverse ecosystems within the LUE modeling framework.
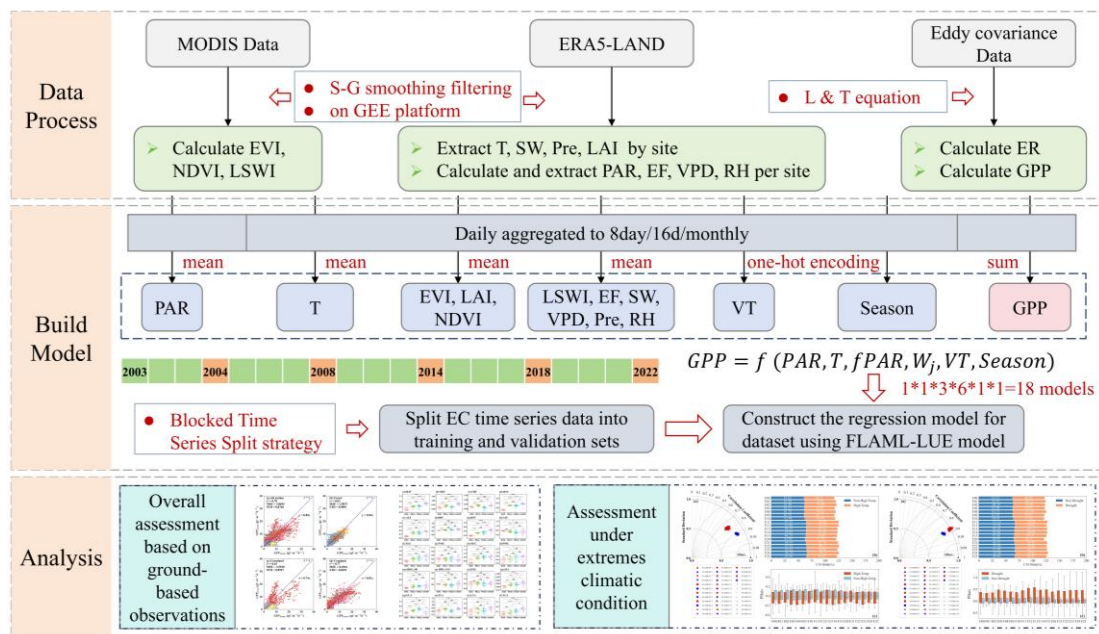
221   The flowchart of this study is shown in **Figure 2**.

222

2.3.1 Data pre-processing and splitting strategy

The primary datasets for estimating GPP with FLAML-LUE models include multi-year continuous EC flux data, satellite-based observations, and ERA5-Land climate reanalysis data. Prior research (Jung et al., 2011) has demonstrated notable seasonal fluctuations in GPP, we divided the time series data into four distinct seasons. Moreover, the vegetation cover type, which varies across different ecosystems, greatly impacts the accuracy of GPP simulation (Chang et al., 2023). Hence, we integrate vegetation type as a factor in our model.

The pre-processed dataset was divided into training and testing sets using the Blocked Time Series Split strategy. Given the temporal dependency of the data, standard cross-validation is not suitable for time series analysis (Reichstein et al., 2019). Instead, a block-based and non-continuous split is applied to preserve the temporal structure. In this approach, the time series is partitioned into several non-overlapping continuous training blocks (e.g., 2003-2005, 2007-2009, 2011-2013, 2015-2017, 2019-2021), with independent years reserved as the validation set following each training block (e.g., 2006, 2010, 2014, 2018, 2022). This strategy ensures that the temporal order is maintained, preventing future data from leaking into the training process and thus avoiding invalid predictions. Additionally, the method incorporates validation over multiple periods, enabling the assessment of model generalization across different climate conditions, which is crucial for evaluating the model's robustness under varying environmental scenarios.

Table 2

Predictor variables for driving the FLAML models and their specifications.

	Variable	Acquired method (formula)	Original Spatial Resolution	Data Source
fPAR	EVI	$2.5 \times (R_{nir} - R_{red})/ (R_{nir} + 6.0 \times R_{red} - 7.5 \times R_{blue} + 1)$	500m	MOD09GA
	NDVI	$(R_{nir} - R_{red})/ (R_{nir} + R_{red})$		
	LAI	-	~10km	ERA5-Land
Water	LSWI	$(R_{nir} - R_{swir})/ (R_{nir} + R_{swir})$	500m	MOD09GA
	EF (%)	$EF = LE/(LE+H)$	~10km	ERA5-Land
	SW (m³/ m³)	-	~10km	ERA5-Land
	VPD	$VPD = e_s - e$ $e = 6.112 \times \exp((17.67 \times T_d) + (243.5 + T_d))$ $e_s = 6.112 \times \exp((17.67 \times T) + (243.5 + T))$	~10km	ERA5-Land
	Pre (mm)	-	~10km	ERA5-Land
	RH (%)	$RH = (e/e_s) \times 100$	~10km	ERA5-Land
Radiation	PAR(μ mol m⁻² s⁻¹)	-	~10km	ERA5-Land
Temperature	T（℃）	-	~10km	ERA5-Land
VT	EBF, DBF, NF, MF, GRA, MEA, SHR, SC, DC	One-hot encoding	invariant	-
Season	Spring, Summer, Autumn, Winter	One-hot encoding	invariant	-

Note: EVI: Enhanced Vegetation Index, NDVI: Normalized Difference Vegetation Index, LAI: Leaf Area Index, LSWI: Land Surface Water Index, EF: Evaporative Fraction, SW: Surface Soil Moisture, VPD: Vapor Pressure Deficit, Pre: Precipitation, RH: Relative Humidity, PAR: Photosynthetically Active Radiation, and T: Air Temperature. NF: Needle-leaved Forest; DBF: Deciduous Broadleaved Forest; MF: Mixed Forest; EBF: Evergreen Broadleaved Forest; SAV: Savannas; GRA: Typical Grassland; MEA: Alpine Meadow; SHR: Shrubs; SC: Single Cropping; DC: Double Cropping. In the formulas for EVI and NDVI, R_{nir}, R_{red}, R_{blue}, R_{swir} represent the surface reflectance in the near-infrared (NIR), red, and blue spectral bands, respectively. In the EF calculation formula, LE refers to latent heat flux, while H represents sensible heat flux. In the RH formula, e is the actual vapor pressure, e_s is the saturation vapor pressure, T_d is the dew point temperature, and T is the air temperature.

2.3.2 Automated Machine Learning (AutoML)

Instead of applying a specific ML method like RF for building regression models, we utilize the lightweight Python library "FLAML" version 2.3.3 (Wang et al., 2021) for the AutoML task (Metin and Bilgin, 2024). FLAML optimizes the search process by balancing computational cost and model error, iteratively selecting the learner, hyperparameters, sample size, and resampling strategy (Wang et al., 2021).

265   For our regression tasks, AutoML was configured with the "auto" option for the

266   estimator list, focusing on optimizing the R^2 metric and using a time budget of 120

267   seconds per run. Under this "auto" setting, FLAML explores a variety of built-in

268   regression estimators, including:

269   1. LightGBM (Ke et al., 2017): a histogram-based gradient boosting method designed

270   for speed and scalability;

271   2. XGBoost (Chen and Guestrin, 2016): a regularized gradient boosting framework

272   known for its robustness and accuracy;

273   3. CatBoost (Prokhorenkova et al., 2018): efficiently handles categorical features and

274   reduces overfitting via ordered boosting;

275   4. Random Forest (Breiman, 2001): an ensemble method utilizing bootstrap

276   aggregation of decision trees;

277   5. Extra Trees (Geurts et al., 2006): enhances randomness in split point selection for

278   tree construction;

279   6. Histogram-based Gradient Boosting (Brownlee, 2020), accelerate training through

280   feature binning;

281   7. K-Nearest Neighbors (Cover and Hart, 1967): a non-parametric distance-based

282   algorithm relying on local data density;

283   8. Transformer models (Vaswani et al., 2023), deep learning architectures leveraging

284   self-attention mechanisms, adapted here for structured data regression.

285   Collectively, these estimators span a broad algorithmic spectrum, including

286   ensemble learning, distance-based methods, and neural networks, enabling FLAML to

287   automatically identify the optimal model architecture for the dataset and objective.

2.3.3 Model development

289   Eighteen FLAML-LUE model variations were constructed for all sites by

290   combining different permutations of six input factor groups, as described in Eq. (3) and

291   detailed in **Table 3**. Technically, the term "FLAML-LUE" does not refer to a direct

292   implementation of a mechanistic LUE model. Instead, it reflects a hybrid modeling

293   strategy, through which we incorporate key explanatory variables that originate from

294   LUE theory—such as fPAR, light-use efficiency modifiers, and environmental stress

295   indicators (e.g., VPD, temperature, and water stress indices)—into an automated

296   machine learning framework (FLAML). These variables capture the main drivers of

297   vegetation productivity in traditional LUE models. Their integration enables FLAML

298   to build models that are both ecologically grounded and predictive, effectively

299   balancing model interpretability and accuracy.

$$GPP = f\ (PAR, T, fPAR, W_j, VT, Season) \tag{3}$$

300   where, the $fPAR$ include EVI, NDVI, and LAI; W_j denotes moisture factors including

301   LSWI, EF, SW, PDSI, Pre, RH; VT represents vegetation types, in which forest

302   ecosystems include: EBF, DBF, NF, MF, and SAV; grassland ecosystems include

303   GRA, MEA, and SHR, and farmland ecosystems include SC and DC; $Season$

304   represents the season in which the original data were acquired.

305 **Table 3**
306 Input variable combinations of fPAR and water stress indicators.

Group	Input variables	Group	Input variables	Group	Input variables
FLAML00	NDVI, LSWI	FLAML10	EVI, LSWI	FLAML20	LAI, LSWI
FLAML01	NDVI, EF	FLAML11	EVI, EF	FLAML21	LAI, EF
FLAML02	NDVI, SW	FLAML12	EVI, SW	FLAML22	LAI, SW

FLAML03	NDVI, VPD	FLAML13	EVI, VPD	FLAML23	LAI, VPD
FLAML04	NDVI, Pre	FLAML14	EVI, Pre	FLAML24	LAI, Pre
FLAML05	NDVI, RH	FLAML15	EVI, RH	FLAML25	LAI, RH

Note: EVI: Enhanced Vegetation Index, NDVI: Normalized Difference Vegetation Index, LAI: Leaf Area Index, LSWI: Land Surface Water Index, EF: Evaporative Fraction, SW: Surface Soil Moisture, VPD: Vapor Pressure Deficit, Pre: Precipitation, RH: Relative Humidity.

2.3.4 Model performance evaluation methods

To evaluate the simulation accuracy of the FLAML-LUE model in estimating GPP, we employed a suite of widely used statistical metrics to quantify the agreement between modeled and observed values (Qian et al., 2024; Chang et al., 2023; Tramontana et al., 2016). Specifically, we calculated the coefficient of determination (R^2), Pearson correlation coefficient (R), normalized unbiased root mean square error (nuRMSE), and normalized standard deviation (NSD, $\hat{\sigma}_f$), based on GPP observations from flux towers and model simulations. The Taylor diagram (Taylor, 2001) was utilized to provide a visual summary of the model's performance, incorporating R, nuRMSE, and NSD.

$$R^2 = \frac{\left[\sum_{t=1}^{T}(f_t - \bar{f})(o_t - \bar{o})\right]^2}{\sum_{t=1}^{T}(f_t - \bar{f})^2 \sum_{t=1}^{T}(o_t - \bar{o})^2} \tag{4}$$

$$R = \frac{\frac{1}{T}\sum_{t=1}^{T}(f_t - \bar{f})(o_t - \bar{o})}{\sigma_f \sigma_o} \tag{5}$$

$$nuRMSE = \frac{uRMSE}{\sigma_o} = \frac{1}{\sigma_o}\sqrt{\frac{1}{T}\sum_{t=1}^{T}\left[(f_t - \bar{f}) - (o_t - \bar{o})\right]^2} \tag{6}$$

$$\hat{\sigma}_f = \frac{\sigma_f}{\sigma_o} = \frac{1}{\sigma_o}\sqrt{\frac{1}{T}\sum_{t=1}^{T}\left((f_t - \bar{f})\right)^2} \tag{7}$$

$$\sigma_o = \sqrt{\frac{1}{T}\sum_{t=1}^{T}\left((o_t - \bar{o})\right)^2} \tag{8}$$

where, o_t represents the observed GPP from the flux tower, f_t denotes the simulated GPP from FLAML-LUE model, $\bar{o}$ represents the average of observed GPP from the flux tower, $\bar{f}$ represents the average of estimated GPP from the GPP product, t represents the corresponding ID for the GPP data, and n represents the total count of GPP data for the site. σ_o represent the standard deviations of the observed GPP. A higher R^2 value indicates better consistency between the estimated GPP and the flux GPP.

In addition, the Taylor Skill Score (TSS) was computed to quantitatively assess the overall agreement between simulations and observations, with higher values indicating better performance.

$$TSS = \frac{4(1 + R)}{\left(\hat{\sigma}_f + \frac{1}{\hat{\sigma}_f}\right)^2 (1 + R_0)} \tag{9}$$

where σ_f represent the standard deviations of the model simulation, and R_0 denotes the maximum possible correlation coefficient (in this study, $R_0 = 1$). The TSS ranges from 0 to 1, with a higher TSS indicating better overall model performance relative to the observations.

To further investigate model bias across sites, the percent bias (PBias) was introduced (Qian et al., 2024). Positive PBias values indicate overestimation by the model, while negative values suggest underestimation. The closer the PBias is to zero, the more accurate the model's estimations. The calculation formula is as follows:

$$PBias = \frac{\sum_{t=1}^{T}(f_t - o_t)}{\sum_{t=1}^{T} o_t} \times 100\% \qquad (10)$$

To evaluate the model's ability to capture GPP dynamics under extreme climate conditions, we identified heatwaves and high VPD events using the 95th percentile of historical meteorological records (Stefanon et al., 2012; Anderson and Bell, 2010). Drought events were defined as months with SPEI less than -1.5 (Ayantobo et al., 2019; Gumus, 2023) . These definitions enabled us to evaluate model performance under extreme environmental stresses (Qian et al., 2024, 2023).

$$CV_{Rmse} = \frac{\sqrt{\frac{1}{T}\sum_{t=1}^{T}(f_t - o_t)^2}}{\bar{o}} \times 100\% \qquad (11)$$

To determine whether model performance differed significantly across temporal resolutions (daily, 8-day, 16-day, and monthly), we conducted paired t-tests at a 0.05 significance level. All statistical analyses were performed in Python 3.9 using libraries including numpy, pandas, scipy, matplotlib, sklearn, and flaml. Complementary visualizations were produced in R using ggplot2, ggpubr, and readxl.

3. Results

3.1 Overall Model Evaluation Based on Ground-Based Observations

To evaluate the model performance at the site level, the accuracy of the 18 FLAML-LUE models was assessed using test datasets from individual flux tower sites. The algorithms selected by each FLAML-LUE model are listed in **Table S1**. Notably, the Extra-Trees algorithm was most frequently chosen as the best-performing model. Extra Trees is an ensemble method that constructs multiple unpruned decision trees and introduces high randomness in both feature and threshold selection, which enhances

357 generalization and reduces overfitting, particularly in noisy or high-dimensional

358 datasets. The consistent selection of Extra Trees suggests that FLAML tends to favor

359 models with higher stochasticity and ensemble structures under the given data and

360 computational constraints.

361 **Figure 3** presents the R, nuRMSE, and NSD values for the 18 models. As shown

362 in **Figure 3u**, the model performance shows relatively small differences across different

363 combinations of input indicators. Specifically (**Table 4**), the overall R^2 of the different

364 FLAML-LUE models ranged from 0.78 to 0.82, while nuRMSE values ranged from

365 0.4240 to 0.4670.

366 Among the fPAR-related indices, the model driven by EVI performed slightly

367 better ($R^2 = 0.82$, nuRMSE = 0.4265) than those driven by NDVI ($R^2 = 0.80$, nuRMSE

368 = 0.4524) and LAI ($R^2 = 0.79$, nuRMSE = 0.4561). Regarding moisture stress indicators,

369 the model using LSWI as input achieved the best performance ($R^2 = 0.82$, nuRMSE =

370 0.4298), followed by those using VPD ($R^2 = 0.80$, nuRMSE = 0.4455) and RH ($R^2 =$

371 0.80, nuRMSE = 0.4450). Models driven by EF ($R^2 = 0.80$, nuRMSE = 0.4487), SW

372 ($R^2 = 0.80$, nuRMSE = 0.4505), and Pre ($R^2 = 0.80$, nuRMSE = 0.4503) performed

373 slightly worse, though the differences were minimal.

374 As shown in **Table 5**, the performance of the FLAML-LUE model varies

375 considerably across different sites, with the average R^2 ranging from 0.17 at DXG to

376 0.92 at CBF and HBG_G01. Notably, this variation was primarily attributed to site-

377 level differences rather than the combinations of input indicators (**Figure 3**),

378 highlighting the influence of land cover type and climatic conditions on model

19

379 performance.

380 The best model performance was observed at the HZF, MEF, CBF and HBG_G01

381 sites ($R^2 > 0.85$, TSS > 0.9), followed by QYF, DLG, JZA, and SYA ($R^2 > 0.75$, TSS $>$

382 0.88). Within forest ecosystems, the model performed better in MF, NF, and DBF than

383 in EBF (ALF, BNF) and savannas (YJF). MF, which include both evergreen conifers

384 and deciduous broadleaf species, exhibit distinct seasonal variations that can be

385 effectively captured by satellite imagery. In contrast, EBF show minimal seasonal

386 greenness variation, leading to larger modeling bias in GPP estimation.

387 In grassland ecosystems, the model performed better for shrublands and typical

388 steppe than for alpine meadows (**Tables S4 and S5**). Alpine meadows, characterized

389 by short growing seasons and harsh high-altitude climates, often experience strong

390 environmental disturbances and large GPP fluctuations, making them more difficult to

391 model accurately. In contrast, typical steppe and alpine shrublands display clearer

392 phenological rhythms and stronger photosynthetic activity, making their GPP dynamics

393 easier to capture.

394 In cropland ecosystems, all sites demonstrated relatively strong model performance

395 ($R^2 > 0.6$, TSS > 0.80). Compared to natural grasslands or alpine meadows, croplands

396 are usually monocultures with stable phenology and simpler canopy structures, which

397 aid in more accurate GPP modeling.

398 Notably, at the DXG site, the model achieved a high TSS (0.8326) but a relatively

399 low R^2 (0.17), primarily due to the large performance variation among different index

400 combinations. As shown in **Table S4**, all six NDVI-driven models (FLAML10-

20

401   FLMAL15) have negative R² values, significantly reducing the overall model accuracy

402   at this site.

403   **Table 4**

404   Summary of evaluation metrics for FLAML-LUE model performance across all validation sites.

FLAML	R^2	R	NSD	nuRMSE	TSS
FLAML00	**0.82**	0.91	0.8806	<u>0.4240</u>	**0.9378**
FLAML01	0.82	0.90	0.8717	0.4301	0.9340
FLAML02	0.82	0.90	0.8810	0.4299	0.9365
FLAML03	0.82	0.91	0.8748	0.4250	0.9360
FLAML04	0.82	0.91	0.8763	0.4254	0.9363
FLAML05	0.82	0.91	0.8691	0.4244	0.9346
FLAML10	0.82	0.90	0.8638	0.4277	0.9323
FLAML11	0.79	0.89	0.8641	0.4620	0.9237
FLAML12	0.79	0.89	0.8686	0.4597	0.9256
FLAML13	0.79	0.89	0.8592	0.4539	0.9244
FLAML14	0.79	0.89	0.8629	0.4585	0.9243
FLAML15	0.80	0.89	0.8671	0.4525	0.9271
FLAML20	0.81	0.90	0.8610	0.4376	0.9291
FLAML21	0.79	0.89	0.8551	0.4542	0.9230
FLAML22	0.79	0.89	0.8597	0.4618	0.9225
FLAML23	0.79	0.89	0.8562	0.4577	0.9225
FLAML24	<u>0.78</u>	0.88	0.8543	**0.4670**	<u>0.9194</u>
FLAML25	0.79	0.89	0.8590	0.4582	0.9232
Statistics					
EVI	**0.82**	0.90	0.8756	0.4265	**0.9359**
NDVI	0.80	0.89	0.8643	0.4524	0.9262
LAI	<u>0.79</u>	0.89	0.8576	0.4561	<u>0.9233</u>
LSWI	**0.82**	0.90	0.8685	0.4298	**0.9330**
EF	<u>0.80</u>	0.89	0.8636	0.4487	0.9269
SW	0.80	0.89	0.8698	0.4505	0.9282
VPD	0.80	0.90	0.8634	0.4455	0.9276
Pre	0.80	0.89	0.8645	0.4503	<u>0.9267</u>

RH	0.80	0.90	0.8650	0.4450	0.9283

Note: The statistics represent the mean values of R², R, NSD, nuRMSE, and TSS across all combinations in which the respective variable was involved. Bold numbers indicate the highest values, while underlined numbers represent the lowest values.

Table 5

Mean evaluation metrics for different combinations of fPAR and water stress indicators at each site.

Station Name	R^2	R	NSD	nuRMSE	TSS
HZF	0.85	0.93	0.9839	0.3685	0.9650
MEF	0.91	0.96	0.8989	0.2918	0.9679
CBF	**0.92**	0.97	0.8687	0.2716	0.9644
QYF	0.75	0.89	0.8171	0.4677	0.9057
ALF	0.64	0.83	0.6250	0.5950	0.7387
DHF	0.55	0.75	0.7831	0.6671	0.8224
BNF	0.37	0.67	0.8119	0.7540	0.8003
YJF	0.43	0.68	0.6702	0.7348	0.7130
XLG	0.49	0.75	0.9877	0.6980	0.8736
NMG	0.40	0.64	0.6334	0.7685	<u>0.6673</u>
DLG	0.78	0.89	0.9509	0.4543	0.9425
DMG	0.59	0.78	0.6941	0.6204	0.7742
HBG_G01	**0.92**	0.96	0.9040	0.2750	**0.9715**
HBG_S01	0.53	0.82	1.1390	0.6556	0.8945
DXG	<u>0.17</u>	0.83	1.3421	0.7631	0.8326
JZA	0.80	0.91	0.7697	0.4373	0.8916
GCA	0.62	0.82	0.9519	0.5950	0.9014
SYA	0.81	0.92	0.7606	0.4294	0.8854
LCA	0.64	0.80	0.7830	0.5898	0.8488
YCA	0.64	0.80	0.7117	0.5991	0.8043
All	0.80	0.90	0.8658	0.4450	0.9285
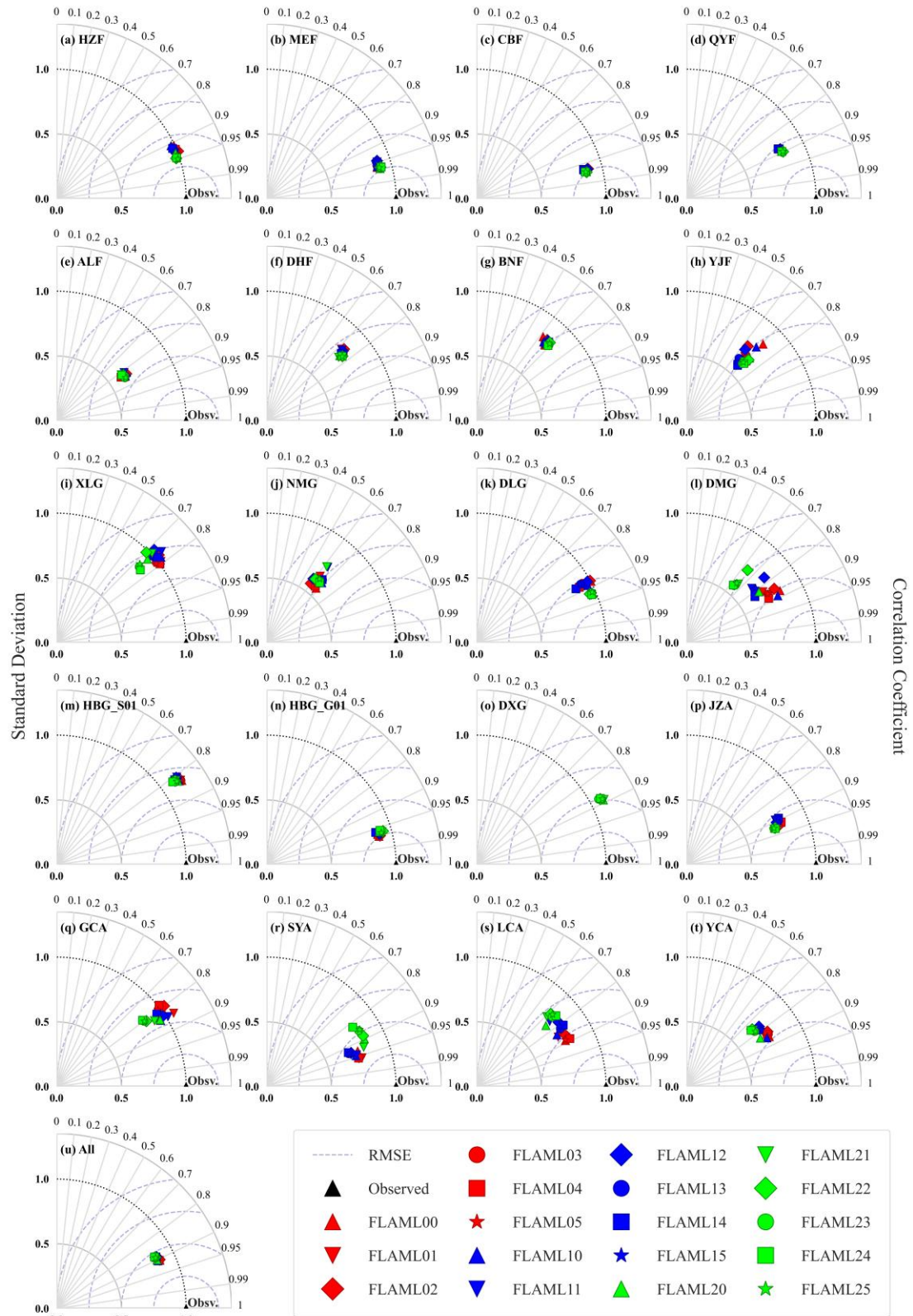
Figure 3. Normalized Taylor diagrams showing the performance of the FLAML-LUE model at various sites based on observed GPP data. Each point represents a specific combination of fPAR and water stress factor used in the model simulation. Different colors denote different fPAR products: red for EVI, blue for NDVI, and green for LAI. Marker shapes indicate the type of water stress

factor: "+" for LSWI, "×" for EF, diamond for SW, circle for VPD, square for Pre, and star for RH. Points closer to the reference point (R = 1, NSD = 1) indicate better agreement between simulated and observed GPP. Panels (a)–(h) correspond to eight forest sites, (i)–(o) to seven grassland sites, (p)–(t) to five cropland sites, and (u) presents an overall model evaluation on the validation dataset across all sites.

From an ecosystem perspective, **Table 7** indicate that the FLAML-LUE model achieves the highest fitting accuracy in forest ecosystems (R^2 = 0.83, nuRMSE = 0.4162), followed by cropland ecosystems (R^2 = 0.72, nuRMSE = 0.5258), and the lowest in grassland ecosystems (R^2 = 0.71, nuRMSE = 0.5407). The slope of the fitted line in **Figure 7** is less than 1 for all ecosystem types, indicating that the FLAML-LUE model tends to underestimate GPP, particularly in croplands and grasslands.

Tables S2, S3, and **Table 6** collectively demonstrate that the model's performance varies across ecosystem types depending on the choice of fPAR-related variables. In forest ecosystems, the model is relatively insensitive to different fPAR and water-related inputs, with the LAI-driven model achieving the best performance. This can be attributed to LAI's ability to capture forest canopy structure, thereby improving fPAR estimates. In contrast, the model's performance is more sensitive to the choice of input variables in cropland and grassland ecosystems. In croplands, the EVI-driven model performs best, followed by LAI and then NDVI, although the performance differences are moderate. In grasslands, however, the NDVI-driven model performs worst, especially at the DXG site, likely due to NDVI's sensitivity to soil background and saturation in sparse and heterogeneous vegetation. EVI, with reduced saturation and higher sensitivity to biomass, shows better performance in structured cropland areas. Overall, the EVI and LSWI driven model (FLAML00) exhibits the best performance across all ecosystem types.

24

440   To further investigate model accuracy across different land cover types, **Figure 5**

441   presents the R² values of five forest types, three grassland types, and two cropland types

442   under different models. In general, model performance varies little within the same land

443   cover type but differs substantially across types. Specifically, DBF, NF, MF, and SC

444   exhibit higher simulation accuracy, followed by GRA, SHR, and DC, while EBF, SAV,

445   and MEA perform the worst. These results are consistent with the Taylor diagram in

446   **Figure 3**.

447
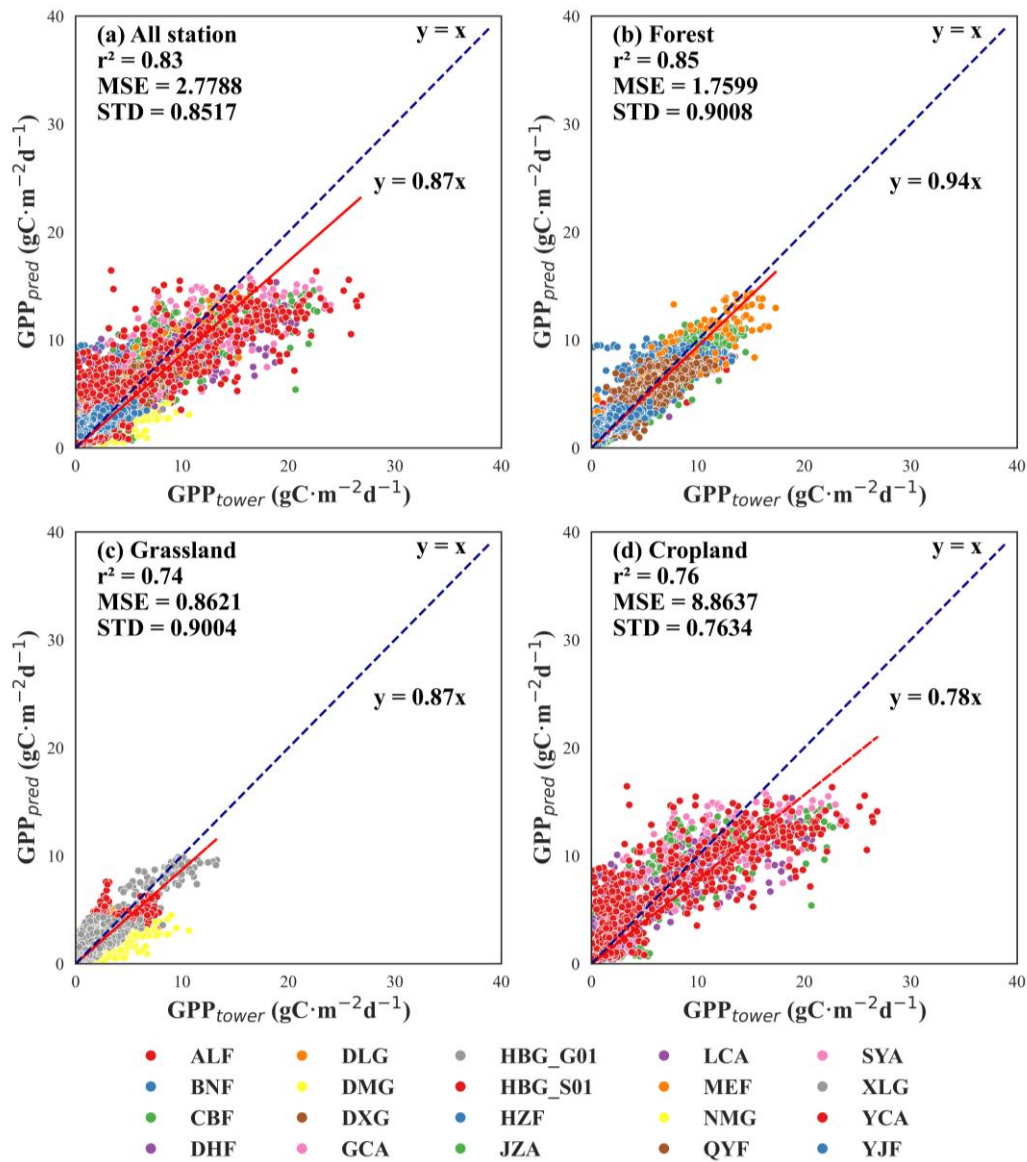
448   **Figure 4.** Scatterplot of observed GPP vs. simulated GPP. Different colored dots represent different

449   sites. Note: The simulated GPP values represent the mean of FLAML00 to FLAML25.

450

451

Table 6

Summary of evaluation metrics for FLAML-LUE model performance across all validation sites.

FLAML	R^2			TSS		
	Forest	Grass	Crop	Forest	Grass	Crop
FLAML00	0.83	0.73	0.77	0.9476	0.9241	0.9004
FLAML01	0.84	0.71	0.75	0.9472	0.9187	0.8946
FLAML02	0.84	0.70	0.75	0.9516	0.9167	0.8966
FLAML03	0.84	0.71	0.76	0.9485	0.9169	0.8971
FLAML04	0.84	0.72	0.76	0.9475	0.9157	0.8991
FLAML05	0.84	0.72	0.76	0.9487	0.9171	0.8927
FLAML10	0.83	0.72	0.76	0.9463	0.9213	0.8861
FLAML11	0.83	0.68	0.70	0.9464	0.9124	0.8696
FLAML12	0.84	0.67	0.70	0.9487	0.9091	0.8717
FLAML13	0.83	0.69	0.71	0.9459	0.9083	0.8696
FLAML14	0.83	0.69	0.70	0.9450	0.9060	0.8713
FLAML15	0.84	0.69	0.71	0.9486	0.9096	0.8746
FLAML20	0.85	0.73	0.73	0.9525	0.9219	0.8718
FLAML21	0.85	0.71	0.70	0.9531	0.9186	0.8575
FLAML22	0.86	0.70	0.68	0.9549	0.9150	0.8545
FLAML23	0.86	0.71	0.69	0.9539	0.9153	0.8535
FLAML24	0.85	0.72	0.67	0.9532	0.9145	0.8465
FLAML25	0.86	0.71	0.68	0.9542	0.9163	0.8561
Statistics						
EVI	0.84	**0.72**	**0.76**	0.9485	0.9182	0.8968
NDVI	<u>0.83</u>	<u>0.69</u>	0.72	0.9468	0.9111	0.8738
LAI	**0.85**	0.71	<u>0.69</u>	0.9536	0.9169	0.8566
LSWI	0.84	**0.73**	**0.75**	0.9488	0.9224	0.8861
EF	0.84	0.70	0.72	0.9489	0.9166	0.8739
SW	0.84	<u>0.69</u>	<u>0.71</u>	0.9517	0.9136	0.8743
VPD	0.84	0.70	0.72	0.9495	0.9135	0.8734
Pre	0.84	0.71	0.71	0.9486	0.9121	0.8723
RH	0.84	0.70	0.72	0.9505	0.9143	0.8745

455 **Table 7**

456 Mean evaluation metrics for different combinations of fPAR and water stress indicators across

457 various ecosystems.

Ecosystem	R^2	R	$\hat{\sigma}_f$	nuRMSE	TSS
ALL	0.80	0.90	0.8658	0.4450	0.9285
Forest	**0.83**	0.91	0.8958	0.4162	**0.9431**
Grassland	<u>0.71</u>	0.84	0.9187	0.5407	0.9154
Croplands	0.72	0.85	0.7893	0.5258	<u>0.8757</u>

458 Note: The evaluation metrics for all sites and different ecosystem types were calculated based on

459 the average of 18 simulation results.

460   Regarding CV_{RMSE}, SHR shows the largest error, followed by MEA, GRA, SC, and

461 DC, while the five forest types show the smallest errors. This may be attributed to the

462 greater GPP variability in grassland and cropland ecosystems, which are more strongly

463 influenced by climatic variability and anthropogenic activities, leading to higher model

464 uncertainty. In contrast, forest ecosystems have more stable structures and continuous

465 carbon exchange processes, resulting in more robust model performance. Although

466 alpine meadow is classified as grassland ecosystems, their extreme climatic conditions,

467 short growing season, and high sensitivity to temperature and precipitation further

468 increase the uncertainty of GPP simulation, leading to higher errors.

469   In terms of PBias, SHR consistently shows a pronounced overestimation across all

470 models. Similarly, SAV and MEA are also generally overestimated in all models,

471 though to a lesser extent than SHR. EBF exhibits a slight overestimation as well. Other

472 vegetation types display only minor underestimation or overestimation. Overall, the

473 models perform best for DBF, NF, and MF, followed by EBF, MEA, SC, and DC, while

474 the simulation accuracy is relatively poor for SAV, SC, and especially SHR.

475   Biases also differ among grassland ecosystems, especially for typical grasslands,

476   alpine meadows, and shrublands. Typical grasslands tend to be underestimated, while

477   alpine meadows and shrublands are often overestimated. These biases may result from

478   the model's limited ability to capture seasonal changes in water availability and its

479   interaction with temperature. Typical grasslands usually show high productivity when

480   water is sufficient, especially in spring and summer. If the model fails to reflect these

481   seasonal patterns, it can lead to underestimation. In contrast, productivity in alpine

482   meadows is mainly limited by low temperatures and a short growing season. If the

483   model does not fully consider these constraints, it may overestimate photosynthesis and

484   thus GPP. For shrublands, overestimation may be due to high spatial heterogeneity,

485   including a mix of shrubs, grasses, and bare soil. This complexity is difficult to capture

486   in remote sensing data (e.g., fPAR) and model inputs, leading to possible overestimation

487   of productivity.
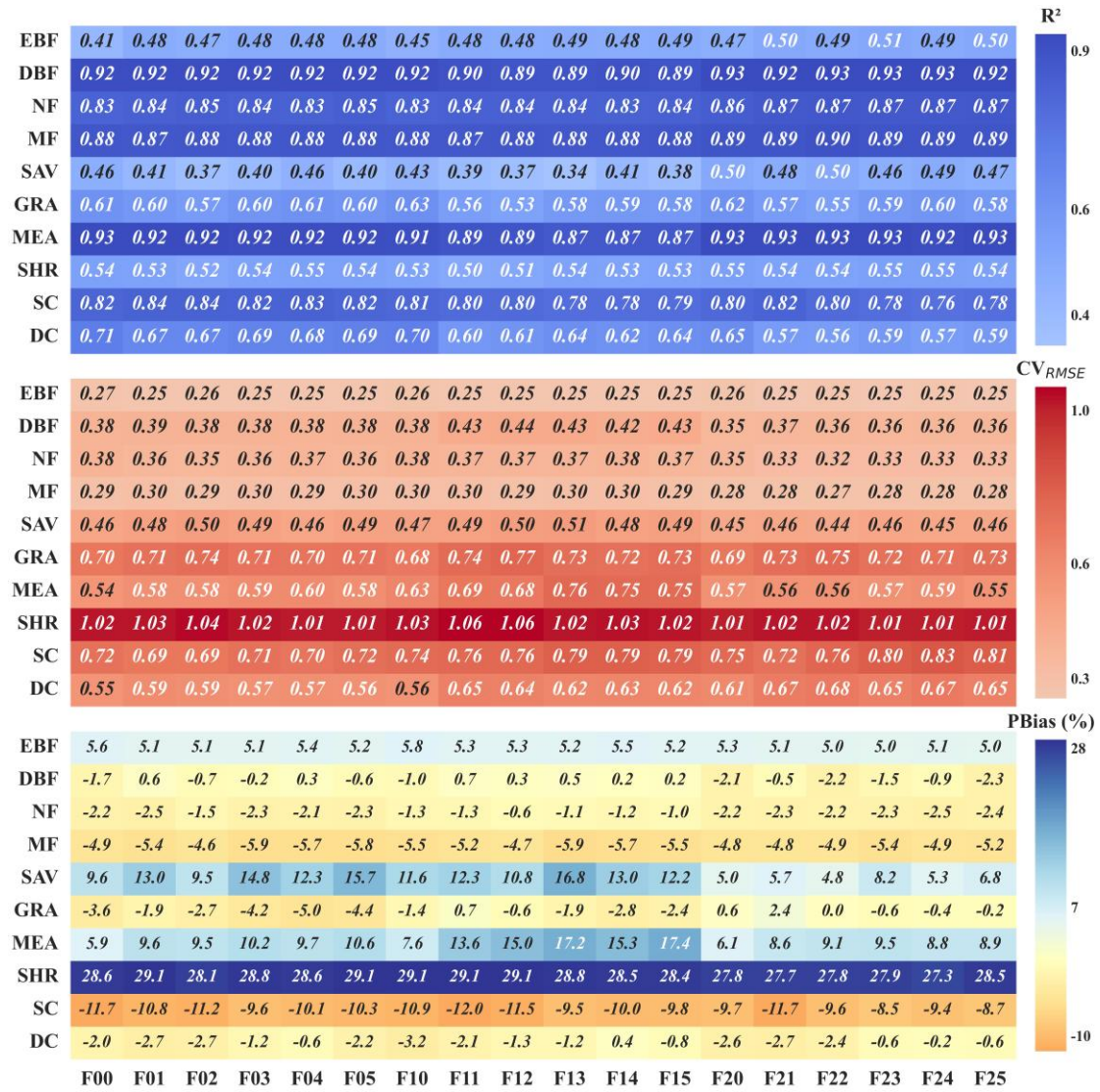
488
489 **Figure 5.** Comparison of R², CV$_{RMSE}$, and PBias of GPP estimates from different FLAML-LUE
490 models across various land cover types. Note: F00 represents FLAML00, and so on.

491   Across the four temporal scales, the performance of the 18 FLAML-LUE models

492 improves as the temporal resolution becomes coarser. The average R² across 20 sites

493 increases from 0.64 at the daily scale to 0.74 at the monthly scale (**Table S8**), while the

494 average nuRMSE decreases from 0.5518 to 0.4088. Paired t-tests show that, except for

495 YJF, NMG, DMG, DXG, and YCA, the FLAML-LUE model exhibits significantly

496 lower R² at the daily scale than at longer temporal scales ($p < 0.05$, **Figure 6**). For these

497 five sites, model performance remains relatively stable across different temporal scales.

498   Furthermore, compared to the daily scale, the nuRMSE decreases by 12.97%,

499　16.52%, and 25.92% at the 8-day, 16-day, and monthly scales, respectively, indicating

500　that the uncertainty of the FLAML-LUE model is significantly reduced at coarser

501　temporal resolutions.

502

Figure 6. Asterisks indicate significant differences between the R^2 at the four temporal resolutions (Kruskal-Wallis test), ****p values < 0.0001, ***p values < 0.001, **p values < 0.01, *p values ≤ 0.05, and ns indicates no significance (p > 0.05).

506　　　Overall, the accuracy of FLAML-LUE models constructed using different

507　combinations of fPAR and water stress indicators showed limited variation, with the

508　FLAML00 model (fPAR = EVI, water = LSWI) demonstrating the best performance.

509　However, the model exhibited considerable differences in performance across

ecosystem types, with the highest accuracy observed in forest ecosystems, followed by

croplands and then grasslands. Further analysis by specific vegetation cover types

revealed that the model performed best for DBF, NF, and MF, followed by GRA, MEA

SC, and DC, while its performance was relatively poor for EBF, SAV, and particularly

SHR (PBias > 27%, CVrmse > 1, R^2 < 0.6). In addition, evaluation across different

temporal scales indicated that model uncertainty decreased with increasing time

intervals, suggesting that the FLAML-LUE model exhibits greater robustness and

reliability at coarser temporal resolutions.

3.2 Model Evaluation Under Extreme Climatic Conditions

Numerous studies have demonstrated that climate extremes such as heatwaves,

droughts, and high atmospheric VPD can substantially alter ecosystem dynamics and

reduce carbon uptake capacity (Frank et al., 2015; Reichstein et al., 2013). These

extreme events can suppress photosynthesis, increase respiration, and disrupt the

balance of carbon exchange between vegetation and the atmosphere. In order to

evaluate the robustness and reliability of the FLAML-LUE models under such stress

conditions, this study further investigates model performance in simulating GPP under

three types of climate extremes: high temperature, high VPD, and drought. By

analyzing the response of model accuracy and bias under these scenarios, we aim to

assess its applicability and limitations in extreme environmental conditions.

3.2.1 Performance Under High Temperature Events

Figure 7 shows the performance of 18 FLAML-LUE models under high-

temperature and non-high-temperature conditions. The results indicate a significant

decline in model accuracy under high-temperature conditions. As shown in **Figure 7a**, the models perform well under non-high-temperature conditions, with the R values of all 18 FLAML-LUE models exceeding 0.9. However, under high-temperature conditions, the Taylor diagram reveals a significant decrease in model performance, with correlation coefficients dropping and a substantial increase in nuRMSE, indicating a reduced ability to capture GPP dynamics.

Interestingly, as shown in **Figure 7b**, the CV_{RMSE} values under non-high-temperature conditions are generally higher than under high-temperature conditions. This may be due to higher observed GPP values under high temperatures, resulting in a larger denominator for CV_{RMSE}, which can reduce the CV_{RMSE} despite larger prediction errors. Overall, the difference in prediction bias between high-temperature and non-high-temperature conditions is minimal.

Figure 7c shows that, under high-temperature conditions, the PBias fluctuates more significantly, with more stations showing severe overestimation or underestimation. Specifically, some models (e.g., FLAML00, FLAML01, FLAML11, FLAML15, FLAML21) overestimate GPP at certain sites under high-temperature conditions, while all models show more severe underestimation at other sites. Models driven by LAI (FLAML20 – FLAML25) exhibit smaller bias variations under non-high-temperature conditions, with PBias mainly ranging from -0.3 to 0.3.

In conclusion, high-temperature conditions increase model uncertainty, with all models exhibiting varying degrees of overestimation or underestimation across sites. Models incorporating VPD, precipitation, and relative humidity as water stress factors

554   perform better overall, indicating greater robustness under high-temperature stress.

555
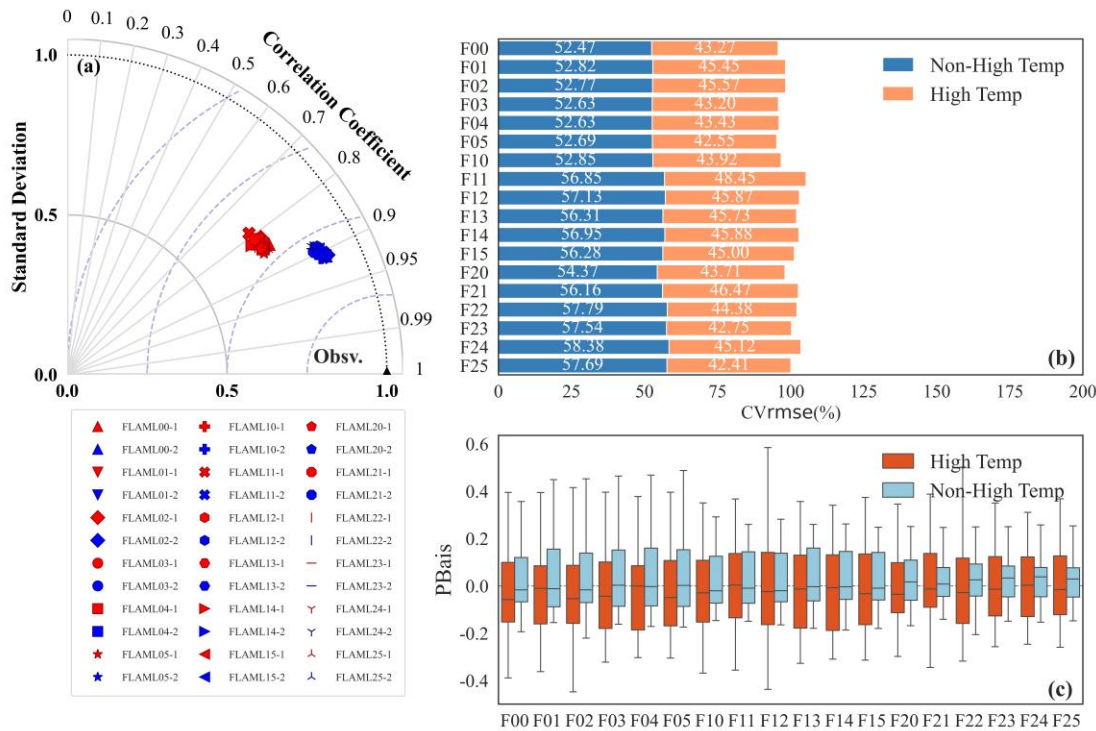
Figure 7. The comparison of GPP products performance under high temperature and non-high
temperature (In the Taylor diagram, 1 represents high temperature, 2 represents non-high
temperature).

559   Differences in model performance under high-temperature and non-high-

560   temperature conditions are pronounced across various land cover types. **Figure 8**

561   compares the estimation accuracy of different land cover types under both conditions.

562   Overall, model accuracy in simulating GPP is significantly lower under high-

563   temperature conditions, with R^2 values showing a notable decline. Specifically, for the

564   NF type, the R^2 under high temperatures approaches a negative value, indicating very

565   low explanatory power, whereas under non-high-temperature conditions, R^2 ranges

566   from 0.83 to 0.87. Notably, the FLAML13 model for Savannas shows a drastic decrease

567   in R^2 from 0.38 under non-high-temperature conditions to -1.46 under high-temperature

568   conditions, performing even worse than the mean of the data during high temperatures.

569   Corresponding to **Figure 7**, CV_{RMSE} is generally lower under high-temperature

570   conditions than under non-high-temperature conditions. The SHR type exhibits a higher

571   coefficient of variation, while PBias shows more pronounced fluctuations. For SHR

572   and EBF, the models tend to overestimate GPP under both temperature conditions, with

573   overestimation more pronounced under high temperatures. In contrast, MEA shows

574   underestimation under high-temperature conditions but overestimation under non-high-

575   temperature conditions. Overall, most land cover types exhibit a greater degree of

576   underestimation under high-temperature conditions. Nevertheless, the MF type

577   maintains relatively high simulation accuracy. In contrast, the NBF, NF, and SC types

578   are more strongly affected by high temperatures, with NF showing negative simulation

579   accuracy under high-temperature conditions and SC exhibiting marked variations in

580   PBias.
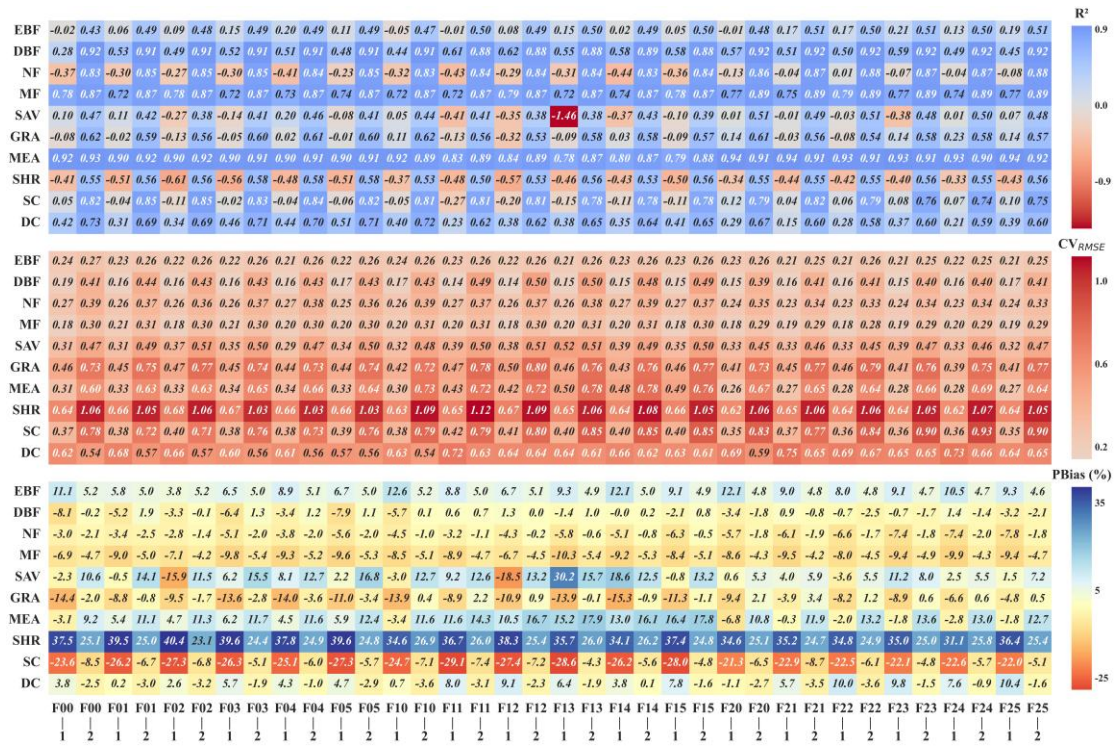
Figure 8. Comparison of statistical indicators (R^2, CV_{RMSE}, PBias) of FLAML-LUE model under high temperature conditions and non-high temperature conditions for different land cover types (1 represents high temperature, 2 represents non-high temperature).

3.2.2 Performance Under High VPD

Figure 9 shows the performance of the 18 FLAML-LUE models under high and non-high VPD conditions. Unlike the high-temperature scenario, the statistical metrics of all models exhibit only a slight decline under high VPD, indicating a less pronounced impact on model performance. As shown in **Figure 9a**, the variability in model performance increases under high VPD conditions. However, **Figure 9b** reveals that CVRMSE values are generally higher under non-high VPD conditions, a trend consistent with the results observed under high-temperature conditions.

Under high VPD, PBias exhibits significant fluctuations compared to non-high VPD conditions (**Figure 9c**). Specifically, the average PBias across sites is higher under high VPD, whereas it is lower under non-high VPD. In high VPD conditions, models

35

596   driven by EVI show smaller differences in PBias across sites, with values primarily

597   ranging from -0.4 to 0.5. In contrast, FLAML05 shows larger differences in PBias

598   between sites under non-high VPD, with overestimations at some sites. Overall, model

599   performance under high VPD shows greater uncertainty, with both overestimations and

600   underestimations occurring across different sites. In general, EVI-driven models

601   perform more consistently under both high and non-high VPD conditions.

602   Model performance also differs across land cover types under high and non-high

603   VPD conditions. **Figure 10** compares the estimation accuracy for various land cover

604   types under both conditions. Overall, GPP simulation accuracy for certain cover types

605   (e.g., DBF, MF, MEA, SC, DC) shows little difference between high and non-high VPD

606   conditions. Although R^2 values for some land cover types are significantly lower under

607   high VPD than under non-high VPD, the impact of high VPD on model performance is

608   smaller compared to high temperature. The most notable example is the FLAML13

609   model for Savannas, where R^2 drops significantly from -1.46 under non-high VPD to -

610   0.39 under high VPD, performing worse than the mean data value under high VPD.

611   Similar to high-temperature conditions, CV_{RMSE} under high VPD is generally lower

612   than under non-high VPD. MEA shows a larger coefficient of variation, and PBias

613   exhibits more noticeable fluctuations. For the EBF and SHR type, models tend to

614   overestimate GPP in both high and non-high VPD conditions, with the overestimation

615   being more pronounced under high VPD. SC and GRA models show significant

616   underestimation under high VPD. DBF, NF, and MF perform relatively well under high

617   VPD, while SC underestimates GPP under both conditions, and DC overestimates GPP

under high VPD but underestimates it under non-high VPD. Overall, compared to high-

temperature conditions, the effect of high VPD on estimation errors is smaller across
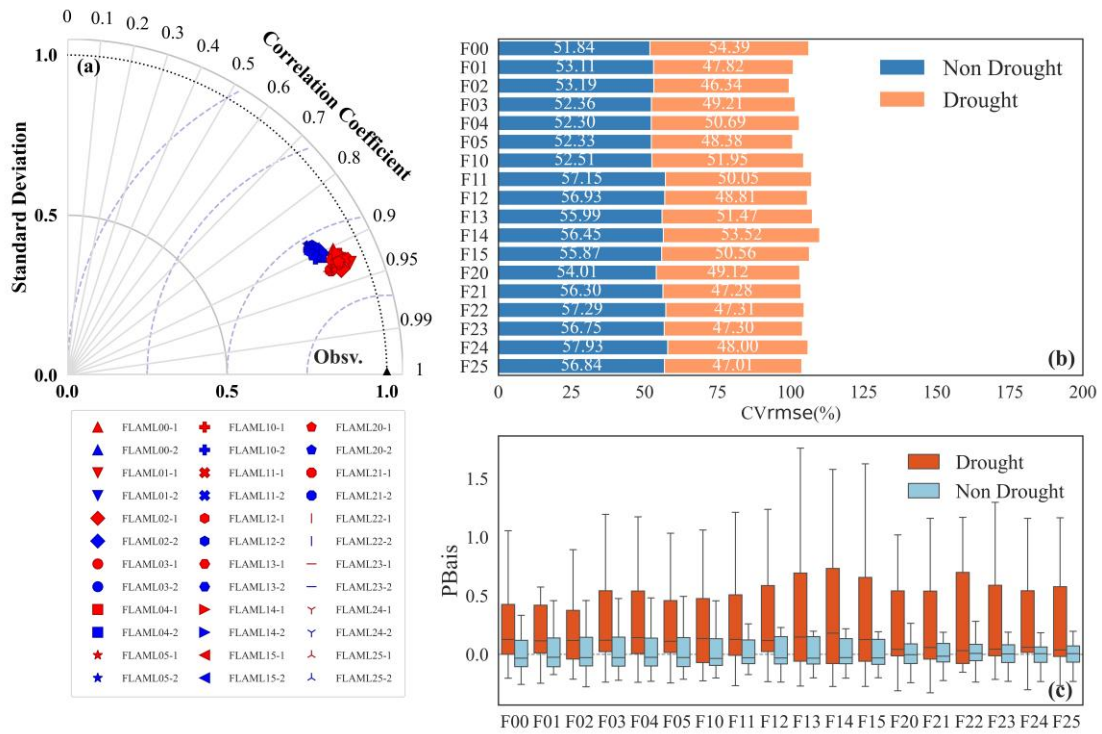
different land cover types.
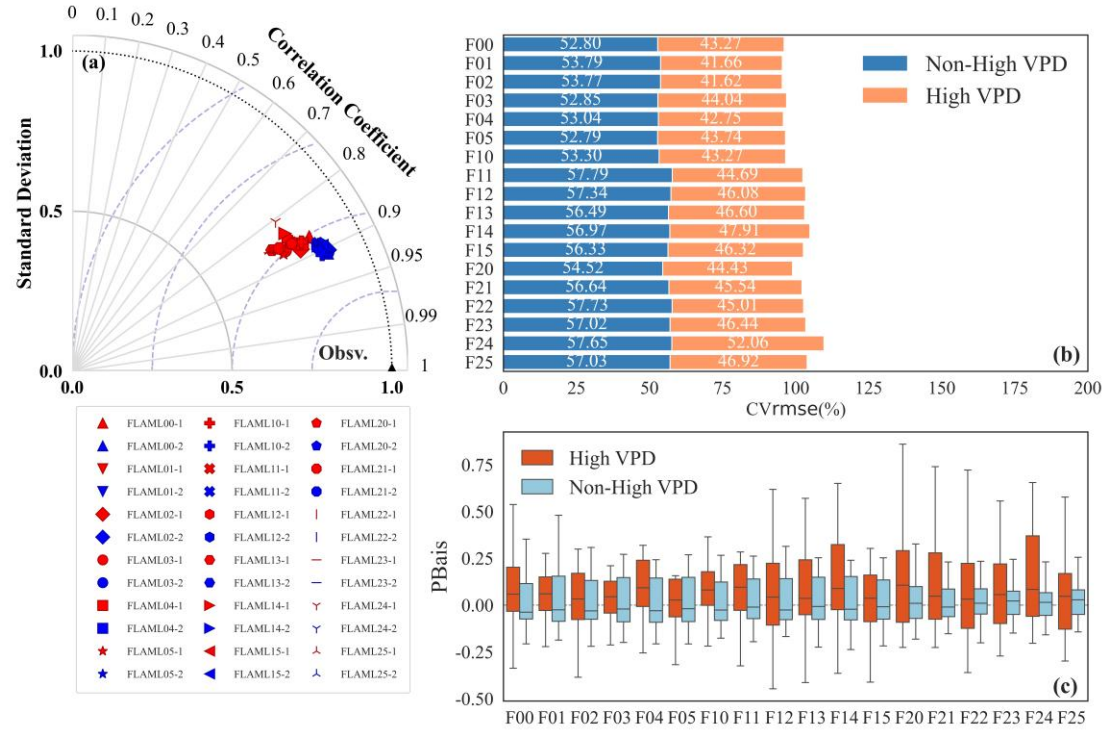
Figure 9. The comparison of GPP products performance under high VPD and non-high VPD (In
the Taylor diagram, 1 represents high VPD, 2 represents non-high VPD).
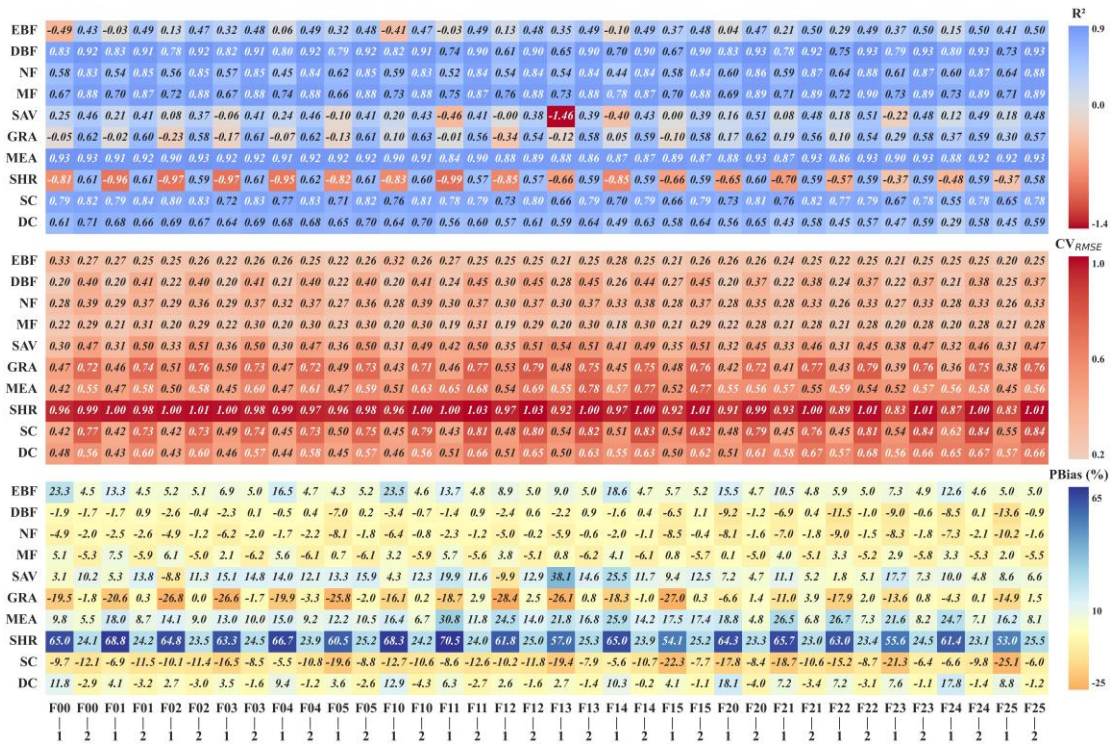
Figure 10. Comparison of statistical indicators (R^2, CV_{RMSE}, PBias) of FLAML-LUE model under high VPD conditions and non-high VPD conditions for different land cover types (1 represents high VPD, 2 represents non-high VPD).

3.2.3 Performance Under Drought Conditions

Figure 11 presents the simulation performance of the 18 FLAML-LUE models under drought and non-drought conditions. Unlike the decline in performance under high temperature and high VPD conditions, the model shows similar or even slightly better accuracy under drought compared to non-drought conditions. This may be attributed to an overall reduction in GPP and its variability during drought periods, which potentially makes it easier for the models to capture the general trend and thereby improves simulation accuracy.

Compared to the boxplots under non-drought conditions, drought notably increases the variability in PBias across sites for all models, particularly due to substantial overestimation at certain sites. In contrast, the degree of underestimation remains

similar to that under non-drought conditions. Among the models, those driven by EVI

exhibit the best overall performance, followed by those using LAI as the vegetation

indicator.

Figure 12 shows that drought substantially affects GPP estimation accuracy across

most land cover types. For certain types, such as savannas and deciduous broadleaf

forests, no data were available during drought months, making performance evaluation

under drought impossible. For other land cover types, the impact of drought varies

significantly. Specifically, EBF, MEA, and DC show higher R^2 values under drought,

while NF, MF, GRA, SHR, and SC perform better under non-drought conditions.

Among them, MF and SHR have the lowest simulation accuracy under drought but

perform relatively well during non-drought periods.

Regarding CV_{RMSE}, all land cover types except MEA and NF exhibit lower values

under drought conditions, consistent with the results in **Figure 11a**. MEA shows the

largest coefficient of variation, indicating greater variability in model performance

under drought. In terms of PBias, NF, MEA, and SHR exhibit the highest errors. On

average, model errors increase under drought across most land cover types. Except for

EBF and GRA, most types show severe overestimation or underestimation during

drought periods.

658

Figure 11. The comparison of GPP products performance under drought and non-drought (In the
Taylor diagram, 1 represents drought, 2 represents non drought).

661
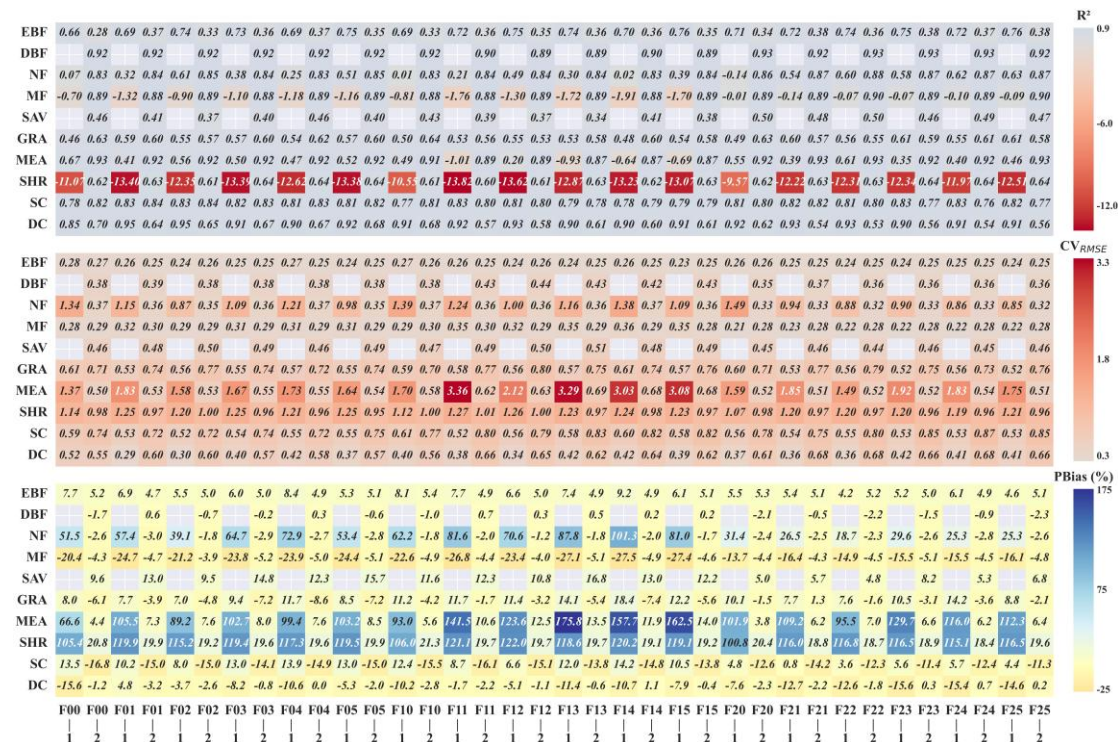
Figure 12. Comparison of statistical indicators (R^2, CV$_{RMSE}$, PBias) of FLAML-LUE model under
drought conditions and non-drought conditions for different land cover types (1 represents
drought, 2 represents non-drought).

4. Discussion

Model performance is highly influenced by the algorithms used, the underlying processes, and how GPP responds to varying environmental conditions (Chang et al., 2023). A detailed comparison of the FLAML-LUE models across different ecosystems showed that performance varied depending on the input variables, vegetation types, and time scales (Chang et al., 2023; Harris et al., 2021).

4.1 Performance comparison of FLAML-LUE models for different ecosystems

In this study, FLAML-LUE models were constructed for different combinations of variables and different time scales based on AutoML algorithms. On the whole, the modeled GPP values agree well with the GPP estimated based on the EC tower, and the FLAML-LUE models performed better in capturing the magnitude and seasonal dynamics of the GPP, which indicated that it was feasible to estimate the GPP using AutoML algorithms. Further, all three ecosystems showed good model performance driven by observational data. Comparisons across various ecosystems indicate that the model exhibited superior performance over forest ecosystems compared to grassland and agricultural ecosystems, as evidenced by the average R^2 values.

Although model performance differences across indicator combinations were minimal, EVI-driven FLAML-LUE models slightly outperformed those driven by NDVI. This highlights the key role of EVI in GPP estimation, as it offers more comprehensive atmospheric correction and is less susceptible to saturation from green reflectance compared to NDVI. Additionally, model performance varied significantly

687   across sites.

688   Based on the evaluation metrics, the optimal model selected was FLAML00 (EVI

689   + LSWI). Under this combination of indicators, the FLAML-LUE model demonstrated

690   the best performance in mixed forests at CBF, deciduous broadleaf forests at MEF, and

691   alpine meadows at HBG_G01, with R^2 values of 0.92, 0.92, and 0.93, respectively. The

692   next best performances were observed in coniferous forests at QYF and HZF, single-

693   cropping farmland at JZA and SYA stations, double-cropping farmland at YCA, and

694   typical grasslands at DLG and DMG sites. In contrast, the model performed poorly in

695   alpine shrub and alpine ecosystems, with an R^2 of 0.54, and the worst performance was

696   observed at the BNF site, with an average R^2 of only 0.28. Mixed forests exhibit distinct

697   seasonal variations that satellite imagery can effectively capture, while evergreen

698   broadleaf forests (ALF and BNF) show minimal seasonal changes in vegetation cover

699   or greenness, making accurate predictions challenging. Alpine shrublands have more

700   complex vegetation structures and less distinct seasonal variations in vegetation cover,

701   which makes it harder for the model to capture the dynamics accurately. In contrast,

702   alpine meadows exhibit more pronounced seasonal variations in vegetation cover,

703   which makes the model more effective in capturing GPP dynamics. For non-forest

704   ecosystems, the highest R^2 values were observed in agricultural fields and typical

705   grasslands, followed by alpine meadows and alpine shrublands.

706   Mixed forests display clear seasonal variations that satellite imagery can effectively

707   capture. However, evergreen broadleaf forests (ALF) have slight seasonal variations in

708   vegetation cover or greenness, making it difficult for the model to predict. For non-

forest ecosystems, the highest R^2 was found in agricultural fields and typical grasslands, followed by alpine meadows and alpine scrub. In addition, the differences in model performance were also reflected in different temporal scales. In general, the model simulation performance at the 16-day and monthly scales was better than that at the daily scale, and the performances of different temporal scales for forest, grassland, and cropland ecosystems were consistent with previous studies.

This study did not distinguish between rainfed and irrigated agricultural systems, considering only the crop rotation types. Specifically, JZA and SYA represent rainfed systems, whereas GCA, LCA, and YCA are irrigated. Future research could incorporate this distinction to improve the accuracy of carbon flux estimates in cropland ecosystems. This distinction is important for interpreting model results under water-limited conditions.

In addition, our results indicate that forest and agricultural fields have greater carbon sequestration capacity and higher annual fluxes than grasslands (Table S9, S10, S11), aligning with previous research outcomes (Y. Wang et al., 2021; Zhang et al., 2007). However, due to the annual harvest of crops, approximately 76% of the on-farm biomass is removed, resulting in limited long-term carbon storage capacity (Zhang et al., 2007). With the exception of tropical rainforests (i.e., BNF), the annual carbon production of planted forests (i.e., QYF) is higher than that of natural forests (i.e., CBF, DHF), which implies that planted forests possess significant potential for carbon assimilation, functioning as robust carbon sinks.

4.2 Model Performance Variations Under Extreme Conditions

In the context of global warming and the increasing frequency of extreme climate events, the adaptability and stability of GPP estimation models in extreme environments have become crucial. This study systematically evaluated the performance of the FLAML-LUE model under high-temperature, high-VPD, and drought scenarios by grouping the validation set. The results showed a general decline in the model's accuracy across all three extreme climate conditions, with varying performance depending on the scenario, highlighting the complexity of vegetation carbon absorption responses to climate stress.

In high-temperature conditions, the model generally underestimated GPP. This could be due to the suppression of photosynthesis caused by high temperatures. High temperatures increase transpiration stress, causing stomatal closure to reduce water loss, which limits CO_2 input and lowers photosynthetic rates (Qu et al., 2020; Reichstein et al., 2013). Additionally, high temperatures can cause leaf damage and senescence, reducing LAI and overall photosynthetic potential (A. Chen et al., 2021; Y. Chen et al., 2021). Although the FLAML-LUE model accounts for fPAR and water stress factors, it may not fully capture rapid responses such as leaf damage or sudden declines in LAI, which likely contribute to the reduced accuracy under high-temperature conditions. Moreover, the model does not explicitly account for the lag effect of leaf senescence, which may further worsen estimation bias (Frank et al., 2015).

Under high VPD conditions, the model showed significant uncertainty, with some areas overestimating GPP and others underestimating it. This inconsistency likely arises

753   from the diverse water stress mechanisms induced by high VPD. Guo et al. (2015) noted

754   that high VPD does not always reflect the true level of water stress in plants, leading to

755   the potential overestimation of GPP. Conversely, in extreme VPD scenarios, where

756   stomata close to reduce carbon absorption, the model may underestimate GPP if it fails

757   to recognize this regulatory behavior (Li et al., 2016). Additionally, the FLAML-LUE

758   model does not explicitly consider leaf energy load or light inhibition, which may

759   contribute to the model's higher errors under high VPD conditions (Rigden et al., 2020).

760   Although the model's performance decreased at some sites under drought

761   conditions, its overall accuracy improved under these scenarios. This improvement may

762   be due to the stronger limiting effect of drought on vegetation growth, allowing the

763   model to more accurately capture the suppressive impact of water stress on

764   photosynthesis. In drought conditions, water scarcity limits carbon absorption, leading

765   to a substantial reduction in GPP (McDowell et al., 2008). As a result, the model's

766   estimates are more likely to align with the actual limitation of carbon absorption. Thus,

767   under drought conditions, the model may underestimate GPP, which can be more

768   accurate, while in wetter environments, where water stress is less pronounced, the

769   model may overestimate GPP, reducing its accuracy. Additionally, under drought, the

770   model is likely better at capturing the direct effects of water shortage on plant

771   physiology, reducing interference from other environmental variables and improving

772   prediction accuracy (Zhou et al., 2019).

773   Although the FLAML-LUE model demonstrates strong predictive capabilities

774   under normal climate conditions, there is still room for improvement under extreme

scenarios. One potential limitation is the insufficient representation of rapid plant response mechanisms (e.g., leaf damage and sudden declines in LAI) in the current input features (Frank et al., 2015; Reichstein et al., 2013). Future research could incorporate high-temporal-resolution vegetation indices, such as solar-induced chlorophyll fluorescence (SIF), to better capture dynamic changes in plant metabolic activity and stress responses under extreme conditions (Yi et al., 2024; Pagán et al., 2019). Including lag variables or cumulative stress indices could also enhance the model's ability to handle delayed physiological responses after stress events (Frank et al., 2015). Furthermore, future studies should expand the scope to include a broader range of climate events that affect GPP, such as floods and low temperatures, in addition to high temperature, high VPD, and drought (Wang et al., 2023). Vegetation in different regions responds differently to these events, with low temperatures and frost being especially important for high-latitude ecosystems.

4.3 Advantages of FLAML-LUE framework

In this study, FLAML (Wang et al., 2021) selected the Extra Trees algorithm as the best-performing model for GPP simulation in China. Extra Trees is an ensemble learning method that builds multiple unpruned decision trees and incorporates randomization in features selection and split thresholds determination. Compared to traditional decision tree ensembles such as Random Forests, Extra Trees typically achieves minimal variance while maintaining low bias, which makes it particularly well-suited for complex, high-dimensional datasets (Geurts et al., 2006).

The adoption of FLAML provides several significant advantages. First, it

automates the model selection and hyperparameter tuning process, eliminating the need

for extensive manual trial-and-error and reducing reliance on domain expertise (Nakano

and Liu, 2025; Wang et al., 2022). Instead of manually evaluating various algorithms

and their configurations, FLAML efficiently explores a broad search space and

identifies the most appropriate model for the dataset.

Moreover, FLAML employ a cost-aware hyperparameter optimization strategy,

enabling it to find high-performing models with relatively low computational cost

(Zhang et al., 2023; Wang et al., 2021). This feature is particularly advantageous in

scenarios with limited computational resources or the need for rapid prototyping.

Compared to conventional machine learning workflows, FLAML significantly

reduces human bias in model selection, improves reproducibility, and lowers the barrier

to applying advanced modeling techniques (He et al., 2021). Overall, the use of FLAML

in this study not only improved model performance but also streamlined the modeling

process, supporting its broader applicability in ecological and climate-related research.

4.4 Comparison with other products
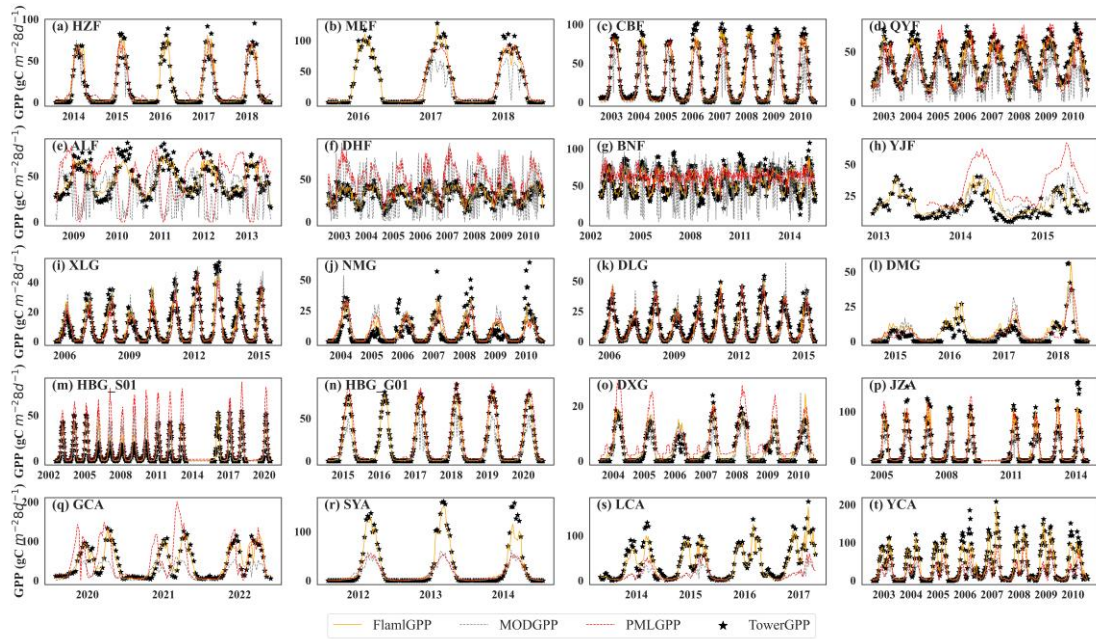
Figure 15. Comparing 8-day GPP from FLAML-LUE, PML, MOD17 models, and EC observations.

This study attempted to predict the GPP of different sites using the FLAML model based on the LUE model variables. The results showed that the AutoML algorithm is a promising GPP estimation method, which explains on average 75%-98% of the GPP variation.

Compared to two GPP products (MODIS GPP, and PML GPP), the GPP from this study showed the highest precision (**Table 7**) and better consistency with flux tower-based GPP under different ecosystems. Overall, the FLAML-LUE model used in this study had the best simulation performance. These findings highlight the potential of the FLAML algorithm for accurately estimating GPP. The FLAML-LUE model is a data-driven ML approach that builds relationships based on dependent and explanatory variables. This enables it to effectively simulate the complex nonlinear interactions across diverse ecosystems (Tramontana et al., 2016). This advantage is even more prominent at the global scale, considering that more flux tower data are available for

827   model construction.

Ecosystem	Station	R2			TSS		
		FLAML	MOD	PML	FLAML	MOD	PML
ALL	ALL	0.93	0.71	0.78	0.9657	0.2677	0.5675
Forest	HZF	0.95	**0.88**	0.91	0.9843	**0.9672**	0.9569
	MEF	**0.98**	0.78	**0.95**	0.9868	0.7664	**0.9571**
	CBF	0.98	0.78	0.93	**0.9903**	0.8860	0.9567
	QYF	0.95	0.54	0.74	0.9833	0.8634	0.9231
	ALF	0.87	0.24	0.34	0.9054	<u>0.2455</u>	<u>0.1812</u>
	DHF	0.83	0.27	0.45	0.9527	0.3030	0.5851
	BNF	0.81	<u>0.05</u>	<u>0.02</u>	<u>0.9025</u>	0.3370	0.3337
	YJF	<u>0.75</u>	0.31	0.42	0.9334	0.7759	0.5820
Grass	XLG	0.92	0.76	0.79	0.9651	0.9343	0.9008
	NMG	0.67	0.48	0.41	<u>0.8288</u>	0.8340	0.7436
	DLG	0.92	0.76	0.77	0.9787	**0.9349**	0.9320
	DMG	0.82	0.68	0.57	0.9537	0.9080	0.8611
	HBG_S01	0.89	0.78	0.81	0.9718	0.9284	<u>0.7175</u>
	HBG_G01	0.99	0.91	0.97	**0.9947**	<u>0.7546</u>	**0.9911**
	DXG	0.90	0.75	0.82	0.9737	0.9134	0.9105
Crop	JZA	0.95	0.84	0.85	**0.9786**	**0.6009**	**0.9582**
	GCA	<u>0.89</u>	<u>0.33</u>	<u>0.19</u>	0.9708	0.4889	0.6748
	SYA	**0.96**	**0.92**	**0.92**	0.9666	0.3708	<u>0.3948</u>
	LCA	0.94	0.57	0.48	0.9731	<u>0.2433</u>	0.3959
	YCA	0.93	0.71	0.78	<u>0.9657</u>	0.2677	0.5675

831   Note: Bold numbers indicate the highest values, while underlined numbers represent the lowest

832   values.

833   However, further work is needed to evaluate the FLAML-LUE model's suitability

834   and accuracy, considering its limitations. In particular, it tends to underestimate high

835   GPP and overestimate low GPP. In addition, the model performance in GPP estimation

836   is highly dependent on ecosystem type. Our findings indicated that mixed forests,

837   deciduous broadleaf forests, and agricultural lands had higher prediction accuracies.

While grass sites such as alpine scrub and alpine meadows were predicted with large

uncertainties, consistent with results from other studies (Wang et al., 2021; Yuan et al.,

2014). This is still a big challenge in accurately estimating GPP.

In general, satellite imagery accurately captures the seasonal leaf phenology of

DBF and MF canopies (e.g., spring leaf unfolding and fall senescence). Additionally,

the key environmental factors influencing vegetation production during different

phenological phases are well-defined (Yuan et al., 2014), making them well-suited for

FLAML-LUE modeling. In contrast, the ambiguous seasonal leaf area changes in EBF

and the low variability of GPP in NMG ecosystems result in poorer model performance,

and empirical methods struggle to estimate GPP variability in these areas (Tramontana

et al., 2016).

Model performance is heavily influenced by the quality of the driver data and the

typicality of the flux towers. In this study, meteorological indices are obtained directly

from spatially explicit reanalysis products. Remotely sensed variables (e.g., NDVI and

EVI, LSWI) serve as proxies for vegetation growth and seasonal changes and are crucial

for scaling simulations from site to regional levels. These gridded indices are directly

derived from satellite reflectance bands. Large-area EFs can be obtained using LE and

Hs calculations from ERA5 reanalysis data or can be derived using NDVI temperature

triangulation (Venturini et al., 2004). LAI, VPD, Pre, and RH can be obtained from

ERA5 reanalysis data. Thus, the model can be extended from the site scale to the

regional and even global scale. Building on this foundation, we will develop a long-

term gridded GPP dataset for China using the FLAML-LUE framework to analyze its

860 spatiotemporal variations over multiple years. This dataset will allow us to investigate

861 long-term GPP trends across different climate zones and vegetation types, as well as

862 their responses to key environmental drivers. By comparing GPP estimates across

863 regions and years, we will also assess model uncertainties and identify potential areas

864 for improvement.

5. Conclusion

866 In this study, the FLAML-LUE model was developed based on data from 20 flux

867 observation sites across China, integrating the FLAML algorithm with key variables

868 from the LUE model. The results demonstrate that the FLAML-LUE model performs

869 excellently in GPP prediction, accurately simulating both its temporal variations and

870 magnitude, particularly in mixed forests and coniferous forests. The average R^2 for

871 daily-scale simulations reached 0.92 and 0.91, respectively. Further analysis showed

872 that extending the temporal scale of input data significantly improves model accuracy.

873 In a comparison of models with different variable combinations, it was found that the

874 model driven by EVI outperformed those driven by NDVI and LAI. The model using

875 LSWI as the driving variable performed better than those with EF, SW, CPD, Pre, and

876 RH as primary variables, with the EVI+LSWI combination yielding the best

877 performance. Additionally, the model's prediction accuracy decreased under high

878 temperature and high VPD conditions. However, under drought conditions, the overall

879 prediction accuracy increased, although it decreased at some sites.

880 In summary, the FLAML-LUE model demonstrates strong applicability and

881 potential for wider application in GPP estimation. It holds promise for scaling from site-

level to regional or even global levels, contributing to a deeper understanding of carbon cycling processes. However, the model's applicability in unique ecosystems, such as alpine shrublands, remains limited, and its ability to adapt to extreme climate events requires further enhancement. Future work should focus on optimizing the model structure and parameter settings to improve its robustness and generalization across diverse ecological environments.

CRediT authorship contribution statement

J.L., Y.Z. and J.W. conceived the study. J.L. collected and processed the data. J.L. and Y.Z. drafted the manuscript. A.W., Y.Z., R.L and W.D. funded the study. J.L., Y.Z., A.W, W.F. and J.W. checked the negatives and touched up. All authors have read and agreed to the embellished manuscript.

Data availability

A Fast Library for Automated Machine Learning & Tuning (FLAML) is a Python library, and detailed documentation about FLAML can be found on GitHub. We have uploaded the related source code and documentation to Zenodo (https://doi.org/10.5281/zenodo.14874754, Laijie, 2025). The flux observation data and the Python source code of the FLAML-LUE used in this paper are also archived on Zenodo (https://doi.org/10.5281/zenodo.15477703, Laijie, 2024).

Declaration of competing interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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