



1 Estimation of above- and below-ground ecosystem parameters for 2 the DVM-DOS-TEM v0.7.0 model using MADS v1.7.3: a synthetic 3 case study

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13 **Abstract.**

14 The permafrost region contains a significant portion of the world's soil organic carbon, and its thawing, driven by accelerated
15 Arctic warming, could lead to the substantial release of greenhouse gases, potentially disrupting the global climate system.
16 Accurate predictions of carbon cycling in permafrost ecosystems hinge on the robust calibration of model parameters.
17 However, manually calibrating numerous parameters in complex process-based models is labor-intensive and further
18 complicated by equifinality - the presence of multiple parameter sets that can equally fit the observed data. Incorrect calibration
19 can lead to unrealistic ecological predictions. In this study, we employed the Model Analysis and Decision Support (MADS)
20 software package to automate and enhance the accuracy of parameter calibration for carbon dynamics within the coupled
21 Dynamic Vegetation Model, Dynamic Organic Soil Model, and Terrestrial Ecosystem Model (DVM-DOS-TEM), a process-
22 based ecosystem model designed for high-latitude regions. The calibration process involved adjusting rate-limiting parameters
23 to accurately replicate observed carbon and nitrogen fluxes and stocks in both soil and vegetation. Gross primary production,
24 net primary production, vegetation carbon, vegetation nitrogen, and soil carbon and nitrogen pools served as synthetic
25 observations for a black spruce boreal forest ecosystem. To validate the efficiency of this new calibration method, we utilized
26 model-generated synthetic observations. This study demonstrates the calibration workflow, offers an in-depth analysis of the
27 relationships between parameters and synthetic observations, and evaluates the accuracy of the calibrated parameter values.



28 **1 Introduction**

29 The permafrost region contains 1,440-1,600 petagrams of organic carbon in its soils, representing nearly half of the world's
30 soil organic carbon pool (Hugelius et al., 2014; Schuur et al., 2022). Accelerated warming in the Arctic leads to permafrost
31 thaw, resulting in the decomposition and potential release of a substantial portion of this stored carbon as greenhouse gases,
32 significantly impacting the global climate system (Natali et al., 2021; Schuur et al., 2022; Treharne et al., 2022). The permafrost
33 carbon-climate feedback remains one of the largest sources of model uncertainty for future climate predictions, as critical
34 ecological and biogeochemical processes are poorly represented and constrained in ecosystem models, if included at all
35 (McGuire et al., 2016, 2018; Schädel et al., 2024). To predict future permafrost evolution, models rely on various parameters
36 that contribute to a wide uncertainty range in predictions of permafrost warming (Andresen et al., 2020; Harp et al., 2016;
37 Schädel et al., 2024). Thus, the development of parameter calibration methods is an essential step toward improving prediction
38 accuracy and deepening our understanding of permafrost dynamics and future permafrost carbon-climate feedbacks.

39
40 Calibration involves estimating and adjusting model parameters and constants to enhance the agreement between model
41 outputs and observed data, with the model serving as a mathematical representation of ecological and physical processes
42 (Rykiel, 1996). These parameters are often rate or transport constants that are onerous or impractical to empirically estimate,
43 though model outputs can be highly sensitive to them. Since many model representations are grounded in physics, generalized
44 physical laws are often used to describe ecological and cryohydrological processes. Typically, model outputs are validated
45 against data from laboratory experiments, idealized mathematical models, or site-specific observations, also referred to as
46 target data. During this validation, model parameters are adjusted so that model outputs match the target data. The validated
47 model is then applied to broader geographic locations and/or different time periods, assuming that the validation data represent
48 the environment or ecosystem for which the parameters were calibrated.

49
50 Parameter calibration for complex process-based models is often constrained by the significant labor required and the limited
51 availability of sites with the necessary observations, especially in permafrost regions. Despite these challenges, process-based
52 models remain essential because they encapsulate our current understanding of ecosystem functions and structures, serving as
53 powerful tools for extrapolation. The assumption of representativeness is intrinsic to these models, as they are designed to
54 simulate processes that reflect our best understanding of ecosystem dynamics, allowing for their application beyond the
55 specific sites where they have been initially parameterized. The approach of extrapolating model parameterization for
56 ecosystems of the same type, across wider regions is standard and widely used within ecosystem modeling communities
57 (McGuire et al., 2018; Matthes et al., 2024). Additionally, the role of ecosystem diversity on the spatio-temporal patterns of
58 ecosystem carbon dynamics in the permafrost region has been characterized by numerous empirical studies (Euskirchen et al.,
59 2014; Melvin et al., 2015) and evaluated by modeling investigations (Lara et al., 2016). Therefore, a critical step in improving
60 model accuracy involves calibrating the model against data for a representative diversity of ecosystem types in the Arctic



61 where observations are available. To prepare an ecosystem model for this extensive calibration task, it is essential to develop
62 robust calibration tools and methods that can automate the process of efficiently optimizing model parameters.

63 Another well-known and significant issue in optimizing model parameters through calibration, also referred to as parameter
64 estimation or optimization, is the existence of equifinality (Jafarov et al., 2020; Nicolsky et al., 2007; Tran et al., 2017).
65 Parameterization equifinality occurs when different sets of parameter values result in the same or similar model predictions,
66 given that the model, forcing data, and observations used in calibration are the same (Beven and Freer, 2001). Model
67 equifinality can subsequently lead to different outcomes in model projections. Multiple random initial guesses are used to
68 address this challenge. If the majority of calibration tests with different initial guesses yield a good fit with observations and
69 result in optimal parameter sets that are similar or closely aligned, it increases confidence that the recovered parameter set is
70 indeed optimal. This approach mitigates the risk of converging on a local minimum and ensures a more robust and reliable
71 parameter estimation process (Hansen, 1998).

72 Various methods have been employed to improve the calibration of model parameters across multiple scientific disciplines,
73 utilizing sophisticated techniques and integrating diverse data sources such as remote sensing and field measurements, while
74 accounting for model and data uncertainty (Dietze et al., 2018; Efstratiadis and Koutsoyiannis, 2010; Luo et al., 2016).
75 Optimization-based inverse methods have been successfully used to calibrate parameters in physical models, including snow
76 properties and subsurface thermo-hydrological properties (Jafarov et al., 2014, 2020), as well as soil properties for permafrost
77 modeling (Nicolsky et al., 2007, 2009). However, inverse modeling can become computationally intractable when applied to
78 complex process-based models (Linde et al., 2015).

79 Markov Chain Monte Carlo (MCMC) and data assimilation (DA) techniques have been employed to optimize model
80 parameters by synchronizing model outputs with observed data, thereby enhancing model prediction accuracy (Brunetti et al.,
81 2023; Fer et al., 2018; Xu et al., 2017). These methods often leverage Bayesian inference to address structural uncertainties
82 within models. Nonetheless, the computational demand required for conducting MCMC simulations can outweigh the gains
83 in model accuracy, particularly when dealing with complex process-based models with slow turnover rates that necessitate
84 long simulations to reach equilibrium.

85 In recent years, DA techniques have been applied to optimize both model state variables (Fox et al., 2018; Ling et al., 2019)
86 and parameters (Bloom et al., 2016; Peylin et al., 2016; Scholze et al., 2016; Schürmann et al., 2016). However, DA also
87 encounters challenges related to unbalanced outputs and the need for extended simulations to achieve equilibrium. Persistent
88 issues include the incorrect characterization of the error covariance matrix, which can lead to inaccurate posterior parameter
89 values due to unaccounted model structural errors and observation biases (MacBean et al., 2016; Wutzler and Carvalhais,
90 2014).



91 Various surrogate-based optimization approaches have been proposed to alleviate the computational burden associated with
92 parameter calibration (Koziel et al., 2011; Queipo et al., 2005). Surrogate models, also known as reduced-order models,
93 simplify certain physical processes to approximate the underlying dynamics of the real model while being computationally
94 less demanding (Forrester et al., 2006). By simplifying specific aspects of the model, surrogate models retain essential
95 characteristics of the original system, allowing for faster and more efficient calibration without significantly compromising
96 accuracy (Razavi et al., 2012; Regis and Shoemaker, 2007). However, simplifying complex models presents significant
97 challenges. It is often unclear which assumptions can be safely made and which should be avoided, potentially leading to a
98 loss of model accuracy. Surrogate models must carefully balance the trade-off between simplification and the retention of
99 critical model characteristics to ensure reliable performance. This complexity necessitates rigorous validation to confirm that
100 the surrogate model provides an adequate approximation of the real system without introducing significant errors.

101 In recent years, machine learning-based emulators, often referred to as "models of models," have emerged as a promising
102 approach to reduce the computational burden associated with parameter calibration in complex ecosystem models (Castelletti
103 et al., 2012; Fer et al., 2018; Reichstein et al., 2019). These emulators aim to approximate the outputs of physical and process-
104 based models by learning the relationships between model inputs and outputs through multi-dimensional matrices, significantly
105 enhancing computational efficiency. Unlike traditional surrogate models, which simplify the physical processes within a
106 model, emulators strive to mimic the full complexity of the original model while requiring less computational power. For
107 instance, Dagon et al., (2020) utilized artificial neural networks to emulate the Community Land Model version 5 outputs,
108 focusing on biophysical parameter estimation and global calibration. By integrating machine learning techniques, they were
109 able to explore parameter spaces more efficiently and achieve better alignment with observed data. This method demonstrates
110 the potential of machine learning emulators in improving the accuracy and efficiency of parameter calibration in ecosystem
111 models, particularly when faced with the challenge of high computational demands.

112 To facilitate the automation of the calibration process while minimizing computational demand and avoiding the
113 oversimplification of ecological processes and feedbacks, we employed a non-linear least squares approach for our calibration.
114 We utilized the Model Analysis and Decision Support (MADS) software package (Barajas-Solano et al., 2015; O'Malley and
115 Vesselinov, 2015) for parameter calibration of a terrestrial ecosystem permafrost-enabled model. MADS has been actively
116 developed since 2010, and its conversion to the Julia programming language has provided automatic differentiation capabilities
117 suitable for calibration problems, improving computational efficiency (Vesselinov V.V., 2022).

118 In this study, we developed an automated parameter calibration method for a process-based terrestrial ecosystem model
119 developed for high-latitude regions and characterized by a high level of complexity. To demonstrate its efficacy, we utilized
120 synthetic data and evaluated the capacity of the calibration method to recover the data after perturbing initial guesses (a given
121 set of parameters) using random sampling. The model was run using known parameter values, and the resulting outputs were
122 treated as observations. The primary objective was to illustrate that the parameter calibration method could recover the



123 synthetic parameter set successfully. The secondary objective was to optimize and reduce the labor and time associated with
124 manual parameter calibration. We developed and tested our calibration method for the coupled dynamic vegetation model,
125 dynamic organic soil, and terrestrial ecosystem model (DVM-DOS-TEM) and tested our approach using synthetic observations
126 at a black spruce forest site, a dominant community type in Interior Alaska.

127 2 Methods

128 2.1 Synthetic data for Black Spruce Forest site

129 The two most common forest types in interior Alaska are evergreen stands of black spruce and mixed spruce-deciduous
130 broadleaf forests. Approximately 39% of Interior Alaska is covered by evergreen forest stands and 24% by deciduous forest
131 stands (Calef et al., 2005; Jean et al., 2020). In our study, we developed model calibration for a black spruce (*Picea mariana*)
132 forest community type (CMT), using observations collected in a site located within the Tanana Valley State Forest, just outside
133 Fairbanks, Alaska (64°53'N, 148°23'W). Carbon (C) and nitrogen (N) cycling and environmental monitoring in this forest
134 stand were originally observed by Melvin et al., (2015). The Murphy Dome fire 1958, which covered 8,930 hectares, burned
135 this area and resulted in complete stand mortality.

136
137 We used Gross Primary Productivity without N limitation (GPP*), Net Primary Productivity (NPP), Vegetation C, and
138 Vegetation N stocks by compartments (i.e. roots, stems, and leaves) as synthetic observations shown in Table 1. Synthetic
139 observations are model-generated data that simulate actual measurements using known parameter values, referred to as
140 synthetic target values. To generate these target values, we used existing parameters and the setup described in Section 2.3.
141 The target values shown in Table 1 represent the state of the ecosystem where vegetation and below-ground C stocks are in a
142 steady state. Table 2 includes the below-ground target values. The model was previously manually calibrated using
143 observations from the site. The actual observations were collected and prepared from the measured data at the site and from
144 existing literature and published datasets. Data pre-processing was required before the time series data could be analyzed. Pre-
145 processing was performed to identify and resolve missing data, inconsistencies, and potential outliers. In addition, site
146 observations were aggregated to a monthly resolution to match the temporal resolution of the model outputs, and unit
147 transformations were applied when needed to standardize the units of each variable. Target values for the site were compiled
148 from various data literature sources containing information on C and N stocks, plant biomass, soil horizon depths, and
149 productivity. However, following the initial calibration, the model outputs were similar but did not exactly match the target
150 observations. As stated above, we choose synthetic targets because we know a set of parameters used to produce them and can
151 compare how closely we can recover known parameter values. Therefore, we used the actual model output as our synthetic
152 target values.

153 **Table 1:** Synthetic vegetation target values for the black spruce forest site used in the parameter calibration process



Above-ground Target Names	Notation	Units	Plant Functional Types			
			Evergreen Tree	Deciduous Shrub	Deciduous Tree	Moss
Gross Primary Productivity without nitrogen limitation	GPP^*	[gC/m ² /year]	307.17	24.53	46.53	54.23
Net Primary Productivity	NPP	[gC/m ² /year]	113.08	11.3	24.02	32.41
Vegetation Carbon Leaf	C_{leaf}	[gC/m ²]	572.36	8.35	6.14	136.54
Vegetation Carbon Stem	C_{stem}	[gC/m ²]	1894.03	98.90	477.80	
Vegetation Carbon Root	C_{root}	[gC/m ²]	474.55	33.19	7.17	
Vegetation Nitrogen Leaf	N_{leaf}	[gC/m ²]	14.79	0.38	0.57	1.15
Vegetation Nitrogen Stem	N_{stem}	[gC/m ²]	30.26	2.6	12.53	
Vegetation Nitrogen Root	N_{root}	[gC/m ²]	9.51	0.72	0.16	

154
 155
 156

Table 2: Synthetic below-ground target values for the black spruce forest site used in the parameter calibration process

Below-ground Targets Names	Notation	Unit	Value
Carbon Shallow	$C_{shallow}$	g/m ²	888.91
Carbon Deep	C_{deep}	g/m ²	3174.53
Carbon Mineral Sum	$\Sigma C_{mineral}$	g/m ²	19821.50
Available Nitrogen Sum	ΣN_{avail}	g/m ²	0.76

157

158 2.2 DVM-DOS-TEM description

159 DVM-DOS-TEM is a process-based biosphere model designed to simulate biophysical and biogeochemical processes between
 160 the soil, vegetation, and atmosphere. DVM-DOS-TEM has been applied extensively in Arctic and Boreal ecosystems in



161 permafrost and non-permafrost regions (Briones et al., 2024; Euskirchen et al., 2022; Genet et al., 2013, 2018; Jafarov et al.,
162 2013; Yi et al., 2009, 2010). This model focuses on representing C and N cycles in high-latitude ecosystems and how they are
163 affected at seasonal (i.e., monthly) to centennial scales by climate, disturbances (Genet et al., 2013, 2018; Kelly et al., 2013),
164 biophysical processes such as soil thermal and hydrological dynamics (McGuire et al., 2018; Yi et al., 2009; Zhuang et al.,
165 2002), snow cover (Euskirchen et al., 2006), and plant canopy development (Euskirchen et al., 2014). Modeled vegetation is
166 structured into multiple tiers: (1) the CMT represents the land cover class and characterizes vegetation composition and soil
167 structure at the gridcell level (spatial unit, e.g. black spruce forest, tussock tundra, bog), (2) plant functional types (groups of
168 species sharing similar ecological traits) characterize the vegetation composition within every CMT (e.g. black spruce forest
169 community would be composed of evergreen trees, deciduous shrubs and sphagnum and feather moss plant functional types),
170 and (3) plant structural compartments (leaves, stems, roots). The soil column is split into multiple horizons (fibric, humic,
171 mineral, and rock/parent material). Every horizon is split into multiple layers for which C, N, temperature, and water content
172 are simulated individually. The biophysical processes represented in DVM-DOS-TEM include radiation and water fluxes
173 between the atmosphere, vegetation canopy, snow, and soil. Soil moisture and temperature are updated at a pseudo-daily time
174 step (from linear interpolation of monthly climate forcings). A two-directional Stefan Algorithm is used to predict the positions
175 of freezing/thawing fronts in the soil. The Richards equation is used to calculate soil moisture changes in the unfrozen layers
176 of soil. Both the thermal and hydraulic properties of soil layers are affected by their water content (Yi et al., 2009, 2010;
177 Zhuang et al., 2002). The ecological processes represented in DVM-DOS-TEM include C and N dynamics for every plant
178 functional type of the vegetation community and every layer of the soil column. C and N dynamics are driven by climate,
179 atmospheric chemistry, soil and canopy environment, and wildfire occurrence and severity. C and N cycles are coupled in the
180 soil and the vegetation processes. The GPP of each plant function type is limited by N availability. When resources in N are
181 limited, GPP is downregulated for all plant functional types (PFTs) based on a comparison of N demand (N required to build
182 new tissues) and N supply in the ecosystem (Euskirchen et al., 2009). C and N from the litterfall are divided into aboveground
183 and belowground. Aboveground litterfall is assigned only to the top layer of the soil column, while belowground litterfall (root
184 mortality) is assigned to different layers of the three soil horizons based on the fractional distribution of fine roots with depth.
185

186 **2.3 Input data used for equilibrium run**

187 The driving inputs for the DVM-DOS-TEM model comprise spatial distribution of CMTs, landform, and mineral soil texture.
188 These initialization data were forced to field observations at the study site (Melvin et al., 2015). The spatiotemporal dynamics
189 of the model are driven by an annual time series of atmospheric CO₂ concentration (not spatially explicit), annual time series
190 of spatially explicit distribution of fire scars and dates, and a spatially explicit monthly time series of climate, including mean
191 air temperature, total precipitation, net incoming shortwave radiation, and vapor pressure (Genet et al., 2018). For the present
192 study, we use historical climate data from 1901 to 2015, sourced from the Climatic Research Unit time series version 4.0 (CRU
193 TS4.0; Harris et al., 2014) and downscaled at a 1-km resolution using the delta method (Pastick et al., 2017). For the



194 equilibrium run, the model was driven using the averaged climate forcings from the 1901-1930 period for the study site
195 location, repeated continuously for a sufficient period so equilibrium of vegetation and below-ground C fluxes and stocks was
196 achieved. The resulting modeled ecosystem state for each site then serves as the baseline for historical simulations, however,
197 the calibration process described here only utilized outputs from the equilibrium.

198 2.4 MADS parameter calibration

199 We employed the Model Analysis and Decision Support (MADS) software package for parameter calibration of DVM-DOS-
200 TEM, aiming to minimize the discrepancy between synthetic target and modeled data at the selected site (Barajas-Solano et
201 al., 2015; O'Malley and Vesselinov, 2015). Since its inception in 2010, MADS has undergone active development, including
202 a transition to the Julia programming language, which supports automatic differentiation suitable for calibration
203 problems(Vesselinov V.V., 2022).

204 To set up the parameter calibration using the MADS package for the DVM-DOS-TEM model (Fig. 1), several components are
205 required: the initial guess represents a set of parameter values to be passed to the DVM-DOS-TEM model; the target values;
206 and a model function that updates the parameter file and executes the DVM-DOS-TEM model using the updated values.
207 Parameter keys are used for parameter identification and tracking, and each parameter has a defined range, uniformly
208 distributed within specified limits. Parameter range limits were determined based on prior knowledge. If certain observations
209 are more critical than others, they can be weighted accordingly. For consistency of the calibration process for all parameters,
210 we did not weight parameters in our setup (weight for all parameters were set to 1.0). The experiment name is used for
211 bookkeeping purposes.

212

```
Md = Mads.create_problem(  
    initial_guess,      #the set of initial values  
    targets,           #the set of observations (targets)  
    DVMDOSTEM_run,    #function that runs the model  
    param_keys,        #list of parameter names  
    param_distributions, #the set of parameter ranges  
    observations_count, #number of observations  
    observation_weights, #the set of observation weights  
    problem_name       #the name of the experiments  
)  
Mads.calibraterandom(md, 10; tolOF=0.01, tolOFcount=4)
```

Figure 1. The example of the Julia code setup using Model Analysis and Decision Support (MADS) functions.

213

214 In Figure 1, the `calibraterandom` function initiates the calibration process by randomly distributing parameter values
215 within the specified ranges and then running the model calibration for the generated parameter sets. This function constructs



216 an objective function to minimize the difference between observed and modeled values (detailed in Section 2.5). The calculated
217 residuals are used to assess method convergence. The calibration process employs a tolerance value for the objective function,
218 denoted as tol_{OF} , as the convergence criterion. The $\text{tol}_{\text{OFcount}}$ represents the number of iterations after which calibration
219 ceases if the change in the objective function is minimal between iterations. While increasing the number of iterations could
220 enhance calibration accuracy, it would also raise computational time. More information on the MADS functions can be found
221 at the MADS website (<https://madsjulia.github.io/Mads.jl>)

222 **2.5 MADS minimization method**

223 The MADS package utilizes the Levenberg-Marquardt (LM) algorithm (Levenberg, 1944; Marquardt, 1963; Pujol, 2007) to
224 minimize the difference (the sum of squared residuals) between observations and modeled predictions. In SI1, we provide
225 more details on LM algorithm. The LM optimization method designed to solve non-linear least squares
226 optimization/minimization problems, which are common in the field of history matching, model inversion, curve fitting, and
227 parameter estimation. It combines two approaches: the first-order steepest-descent gradient method and the second-order
228 Gauss-Newton method. This steepest-descent gradient method updates parameter values in the direction opposite to the
229 gradient, thereby it is generally efficient in finding local minima. The Gauss-Newton method assumes that in a region close to
230 the solution, the solved objective function behaves quadratically.

231 The algorithm begins by selecting an initial estimate for the parameters that need to be optimized (Fig S1). This initial guess
232 is important as it sets the starting point for the optimization process. In our experiment, the initial guess is randomly generated
233 from within the provided range near `true` parameter values. Alternatively, users can provide the initial guess. However,
234 exploring a set of random initial guesses provides an efficient approach to exploring the parameter space and discrimination
235 between local and global minima. In LM, we set the damping parameter (the Marquardt lambda) to 0.01. This parameter helps
236 in adjusting the steps taken during the optimization process, balancing between the two optimization strategies (the first- and
237 the second order techniques discussed above).

238 The main advantages of the LM method are its robustness and minimal computational demand. It effectively handles ill-
239 conditioned problems where other optimization methods might fail (Lin et al., 2016; Pujol, 2007). Additionally, for problems
240 well-suited to the Gauss-Newton method, LM often converges faster than gradient descent, making it an efficient choice for
241 many non-linear least squares problems.

242 The disadvantage of the Levenberg-Marquardt (LM) method is its sensitivity to the initial parameter guesses. In addition, the
243 compute speed deteriorates with the higher number of parameters used in calibration. It requires the computation of the
244 Jacobian matrix and its pseudo-inverse, which can be computationally expensive for large-scale problems. Additionally, like
245 many optimization methods, it can be sensitive to the initial parameter guess, potentially affecting its efficiency and
246 convergence. In these cases, MADS provides alternative efficient approaches to address these computational challenges, such
247 as (1) initializing the calibration with random initial guesses, (2) multiple restarts of the LM algorithms throughout the



248 minimization process, and (3) exploration of a series of alternative values for various parameters controlling LM performance
249 (Lin et al., 2016).

250

251 **2.6 Parameters and Calibration Targets**

252 The calibration process in DVM-DOS-TEM is currently focused on the C and N annual cycles. Thus, calibrated parameters
253 are associated with and adjusted to the major C and N fluxes and stocks in the vegetation and the soil. The calibration process
254 follows a hierarchical approach (Figure 2), in which parameters to be calibrated are organized in hierarchical levels associated
255 with (1) model complexity and feedback and (2) turnover of the processes the parameters are associated with. Therefore,
256 parameters related to vegetation dynamics are calibrated first, followed by the slowest soil-related parameters.

257 The first step of the calibration relates to the simplest, fastest, first-order process in DVM-DOS-TEM, and consists of adjusting
258 the rate limiting parameter of maximum C assimilation of the vegetation (c_{max}) driving vegetation GPP. Under baseline
259 climate, the main limiting parameter of vegetation productivity in the Arctic is N availability (Chapin and Kedrowski, 1983).
260 Therefore, c_{max} is calibrated to reproduce estimates of GPP from fertilization experiments where N limitation is lifted. When
261 fertilization experiments are not available for the community/region of interest, it is estimated by applying a multiplicative
262 factor to observed GPP under natural conditions. This multiplicative factor is estimated from published fertilization
263 experiments in similar communities and computed as the ratio between GPP estimated in fertilized plots and GPP estimated
264 in control plots. Based on the literature, this fertilization factor can vary from 1.25 to 1.5 (Ruess et al., 1996; Shaver and
265 Chapin, 1995).

266 The second step of the calibration process consists of turning on the representation of N limitation on vegetation productivity
267 in the model (Euskirchen et al., 2009) and calibrating the rest of the vegetation-related parameters. In the current workflow, it
268 consists of three substeps. These substeps could follow a different order based on the preference of the user and the specifics
269 of a given site. These are rate-limiting parameters for maintenance respiration (Kr_b), maximum plant N uptake (n_{max}), C and
270 N litterfall (c_{fall} and n_{fall} respectively). These parameters are adjusted until DVM-DOS-TEM outputs match observations of
271 GPP and NPP, plant N uptake (Nup), and vegetation C and N pools, respectively). Target values of these variables are listed
272 in Table 1. It is important to note that the parameters Kr_b , c_{fall} , and n_{fall} , as well as the variables for vegetation C and N , are
273 specified per PFT and per compartment (leaf, stem, root).

274 In the third step, the rate-limiting parameters of soil heterotrophic respiration (kdc) and rate of microbial N uptake (n_{mich}^{up})
275 are calibrated as soil processes and takes longer to run in comparison to first two steps. These parameters are adjusted until
276 DVM-DOS-TEM outputs match observations of soil organic C and available N stocks. Target values of these variables are
277 listed in Table 2. In a final state, vegetation-related parameters are checked for a final adjustment after soil calibration, as soil
278 processes can feedback to vegetation dynamics.



Figure 2. Schematics of the DVM-DOS-TEM model parameters and targets participated in the calibration process.

279 **2.8 Calibrations setup and evaluation metric**

280 Table 3 shows the parameter values used to calculate synthetic target values. We established four cases by perturbing the
 281 parameters by 10%, 20%, 50%, and 90% from their original values. For each case, the MADS calibraterandom function
 282 randomly sampled ten sets of parameters within the specified ranges (see Figure 1a). These ten sets of randomly perturbed
 283 parameters were then optimized using the MADS algorithm (Figure 1b). For each set of calibrated parameters and targets, we
 284 computed the root mean square error (RMSE) and relative error (RE) metrics. RMSE is employed to measure the magnitude
 285 of varying quantities, while RE gauges the absolute difference relative to the actual values. Given that some parameters are



286 small (less than 10^{-3}), the relative error provides more informative insights. The following equations were used to compute
 287 these metrics:

$$288 \quad RMSE = \sqrt{(\bar{x} - x)^2}, \quad (1)$$

$$289 \quad RE = \left| \frac{\bar{x} - x}{x} \right| \cdot 100\%, \quad (2)$$

290 where \bar{x} is the mean of the best five out of ten computed target/parameter matches and x is a synthetic target value.

291 To ensure the selection of the best-fitting parameters, we sorted error values from the lowest to the highest. Then, we selected
 292 the top five parameter sets, calculated their mean values, and compared these averaged parameters with the synthetic target
 293 values and known parameters.

294

295 **Table 3:** Synthetic parameter values for the black spruce forest site used in the parameter calibration process.

Name	Parameters	Units	Plant Functional Types			
			Evergreen Tree	Deciduous Shrub	Deciduous Tree	Moss
Maximum rate of atmospheric CO ₂ assimilation	c_{max}	gC/m ² /month	381.19	113.93	210.48	93.31
Maximum rate of plant N uptake	n_{max}	gN/m ² /month	3.38	1.55	1.0	3.55
rate limiting factor for C litterfall for leaf	c_{fall}^{leaf}	month ⁻¹	0.0011	0.05	0.025	0.02
... for stem	c_{fall}^{stem}	month ⁻¹	0.0034	0.0048	0.0036	
... for root	c_{fall}^{root}	month ⁻¹	0.0052	0.0012	0.026	
Rate limiting factor for N litterfall for leaf	n_{fall}^{leaf}	month ⁻¹	0.0102	0.045	0.018	0.007
... for stem	n_{fall}^{stem}	month ⁻¹	0.001	0.001	0.005	
... for root	n_{fall}^{root}	month ⁻¹	0.003	0.007	0.008	
Rate limiting factor for maintenance respiration for leaf	Kr_b^{leaf}	month ⁻¹	-6.0	-3.45	-2.95	-4.65
... for stem	Kr_b^{stem}	month ⁻¹	-4.88	-5.15	-6.65	
... for root	Kr_b^{root}	month ⁻¹	-8.2	-6.2	-3.2	



296

297 **Table 4:** Synthetic below-ground target values for the black spruce forest site used participated in the parameter calibration
298 process

Name	Parameters	Unit	Value
Rate of microbial N uptake	n_{micb}^{up}	gg^{-1}	0.4495
Rate limiting factor of litter decomposition	kdc_{rawc}	$month^{-1}$	0.634
Rate limiting factor of active pool decomposition	kdc_{soma}	$month^{-1}$	0.54
Rate limiting factor of physically resistant pool decomposition	kdc_{sompr}	$month^{-1}$	0.002
Rate limiting factor of chemically resistant pool decomposition	kdc_{somcr}	$month^{-1}$	0.00007

299

300 3 Results

301 3.1 Vegetation Targets

302 Depending on the range of parameter variance, our analysis revealed varying levels of accuracy between known synthetic
303 parameters a those determined using the MADS search approach. In general, the variance between calibrated and synthetic
304 values grew higher with a higher degree of variance (Figure S2-S5). The averaged RMSE values for all four PFTs showed
305 similar increases (Figure 3) with an exception for $C_{stem}(c_{fall})$ deciduous shrubs, which made the RMSE score for the 10%
306 variance higher than the 20% variance (Figure 3a and 3b). That is why we introduced the RE metric, which shows that the
307 departure between synthetic and calibrated parameters increases with increasing perturbation and is the smallest for the 10%
308 variance (Figure 4a).

309 3.2 Vegetation Parameters

310 The RMSE for parameters was highest for Kr_b^{root} in the evergreen tree PFT (Figure 4). Overall, Kr_b s and n_{max} parameters
311 exhibited the worst recovery compared to other parameters based on the RMSE metric. Conversely, REs were highest for c_{fall}
312 deciduous shrubs and less for Kr_b s. The RE indicated that smaller parameter values, such as n_{fall} , deviated more significantly



313 from their synthetic values. Interestingly the RE score showed the same error range for 10% and 20% variance ranges, whereas
314 RMSE showed that 10% variance has the smallest error.

315 **3.3 Soil parameters**

316 In general, the RMSE values for the sub-surface target parameters were relatively small but increased with a higher variance
317 range (Figure 5). Notably, C_{deep} and $\sum C_{mineral}$ exhibited high RMSE values of 3.34 and 9.12, respectively, for the 10%
318 variance range (Figure 5a). Despite this, the soil parameters for 10% variance showed the best match, with RMSE values less
319 than 0.01. The RE for targets revealed increasing deviations from the synthetic parameter values for $\sum N_{avail}$. The RE for
320 parameters indicated that n_{mich}^{up} , kdc_{rawc} and kdc_{soma} had higher deviations from their respective synthetic values for the
321 50% and 90% variance range, respectively.



Figure 3. a), b), c), and d) are root mean square error (RMSE) metric and e), f), g), and h) are relative error (RE) metric for 10%, 20%, 50%, and 90% variance in the parameter range, correspondingly. Targets shown on y-axis, and plant functional types are on x-axis. The colorbar represents the RMSE and RE scores.

322

323

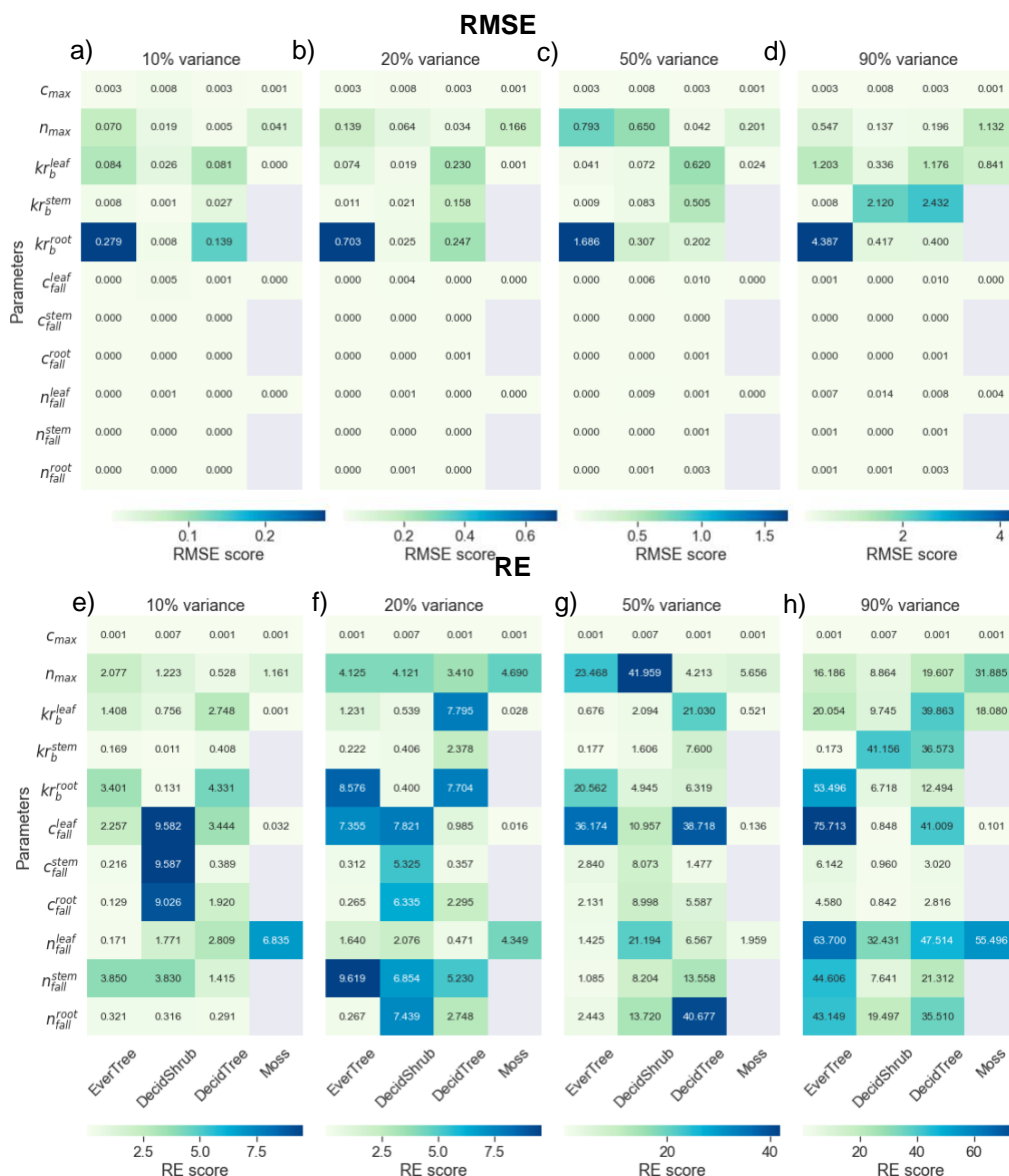


Figure 4. a), b), c), and d) are root mean square error (RMSE) metric and e), f), g), and h) are relative error (RE) metric for 10%, 20%, 50%, and 90% variance in the parameter range, correspondingly. DVM-DOS-TEM parameters shown on y-axis, and plant functional types are on x-axis. The colorbar represents the RMSE and RE scores



325

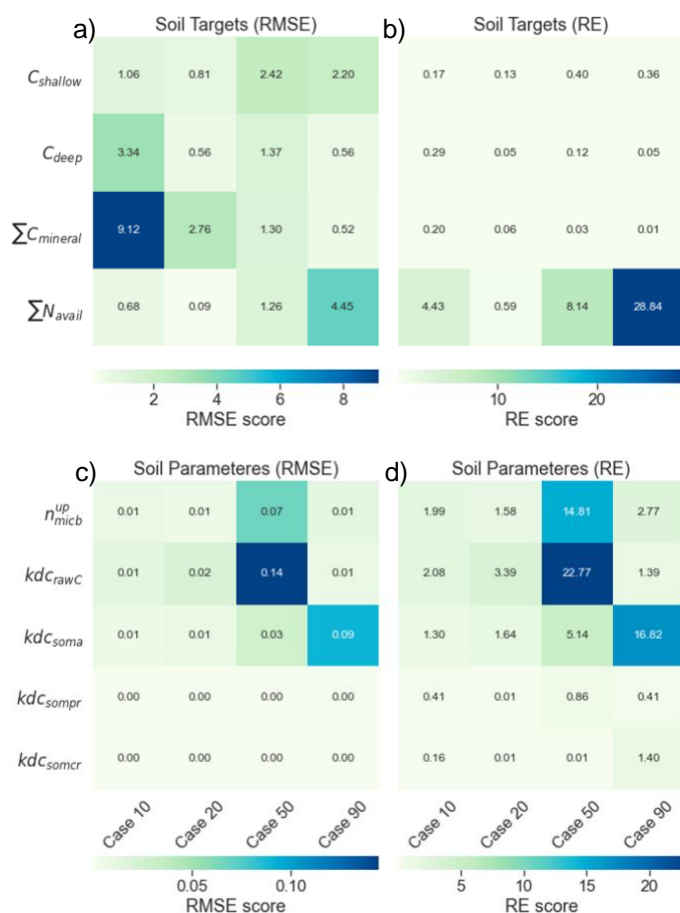


Figure 5. Comparison between calibrated and synthetic sub-surface target values (a) root mean square error (RMSE) and (b) relative error (RE) scores. Comparison between calibrated and synthetic sub-surface parameter values (a) root mean square error (RMSE) and (b) relative error (RE) scores for all range variances. The colorbar represents the RMSE and RE score.

326 4 Discussion

327 4.1 Importance of the initial guess

328 The importance of initial values, or the so-called initial guess, did not significantly impact our synthetic experiment because
 329 the perturbed parameter values were close enough to the true state. However, for non-synthetic calibrations, the initial state is
 330 crucial (Euskirchen et al., 2022; Yi et al., 2009). Applying calibration with parameter values far from the true values could



331 lead to non-convergence of the method and increased computation time. Therefore, when using real observations, starting with
332 a good initial guess is essential. We developed parameter sensitivity methods to provide better estimates for the initial guess
333 (Briones et al., 2024). The current experiment proves that MADS can accurately recover true parameter states well enough
334 when the initial guess is well set.

335

336 **4.2 Analysis of the recovery metrics**

337 The mean parameter values calculated from the five best-matched MADS value predictions align closely with the synthetic
338 parameter values, demonstrating the method's efficacy. The calculated REs for parameters indicate that the relative distance
339 between the calibrated and the synthetic values increases with a higher parameter variance range, except RE for soil targets
340 (Figure 5b). In many cases the RMSE for calibrated target values showed a higher distance for 10% variance range than for
341 20% variance range (Figure 3a and 3b). In addition, the RMSEs for 10% variance range for the soil targets were higher than
342 any other variance range (Figure 5a). The mixed results between 10% and 20% variance range as well as soil target high RMSE
343 for 10% variance, could be attributed to the limited number of cases participated in each variance case. The total number of
344 randomly perturbed initial guesses within the given variance range was 10. It is possible that increasing the total number of
345 searches would yield a more consistent pattern of decreasing accuracy with increasing variance.

346

347 **4.3 Parameter-target relationship and small parameter values**

348 The method demonstrated robust recovery of c_{max} values, indicating that it performs best when there is a linear relationship
349 between parameters and target values (Eq. S1). For parameters, which do not exhibit a linear relationship with their target
350 values (e.g. Kr_b , Eq. S4), the calibrated parameters showed wider variance. Additionally, small parameter values, such as
351 n_{fall} , corresponded to small range values, leading to insensitivity between n_{fall} and vegetation N . To address this, we applied
352 a logarithmic transformation to these and to some other small values for soil C rates.

353

354 **4.4 The impact of n_{max} on N uptake and NPP**

355 Sensitivity between model parameters and targets is crucial for effective parameter calibration. We observed that the sensitivity
356 between n_{max} and NPP was not strong (Eq. S2, Eq. S5), which led us to combine its calibration with the Kr_b parameter. Based
357 on (Eq. S2), n_{max} directly influences N_{uptake} . An increase in n_{max} enhances N_{uptake} , thereby increasing the total N supply.
358 Since NPP is proportional to N_{supply} and inversely proportional to $N_{required}$, a higher N supply can lead to a higher NPP ,
359 provided that other factors remain constant. Therefore, despite the initial observation of weak sensitivity, n_{max} could have a
360 considerable impact on NPP due to its role in N_{uptake} and the overall N_{supply} . However, our target values for plant N uptake
361 are poorly constrained due to a lack of sufficient observations. This underestimation of plant N uptake could account for the
362 observed lack of sensitivity of NPP to n_{max} . This issue requires further investigation and currently underscores the importance
363 of accurately calibrating n_{max} to ensure better simulation of ecosystem productivity.



364

365 **4.5 The Calibration Workflow**

366 The setup of the calibration workflow is important for achieving accurate parameter estimation in terrestrial ecosystem models,
367 considering the model organization by hierarchical levels associated with feedback and turnover of the ecological processes
368 represented. Our findings indicate that calibrating one or two parameter sets at a time, while keeping other parameters constant,
369 is more effective than calibrating all parameters simultaneously. For example in the current workflow, we combined n_{max} and
370 Kr_b (Figure 2 Step a), which was based on the low sensitivity of n_{max} to NPP . Combining multiple variables in one calibration
371 step increases the compute time and could result in low match accuracy. On the other hand, sequential parameter calibration
372 carries the risk of losing accuracy for parameters calibrated in previous steps. To mitigate this risk, we include targets from
373 previous calibration steps in the current calibration step. For example, when optimizing for n_{fall} , we include targets for NPP ,
374 vegetation C , and vegetation N .

375

376 Calibrating one parameter at a time is advantageous not only computationally but also in preventing the occurrence of an
377 underdetermined problem, which arise when the number of parameters exceeds the number of targets. Underdetermined problems
378 exhibit a lower rate of convergence due to the correlation between parameters and the sensitivity of multiple parameters to one
379 or a few similar target values. The study by Jafarov et al., (2020) showed that overdetermined problems, i.e. higher and diverse
380 number of target values, are more effective in recovering accurate parameter values.

381

382 **4.6 Sensitivity of the Kr_b parameter to NPP and vegetation C**

383 The Kr_b parameter exhibited higher sensitivity to both NPP and vegetation C compared to other parameters. Despite the
384 overall good model fitness, the deviation from the synthetic values for Kr_b was higher. This was primarily due to Kr_b^{root}
385 parameter for evergreen tree (Figure S2) persistently showed higher discrepancy. Its sensitivity can be explained by examining
386 its role in the equations governing maintenance respiration (R_m Eq. S3). The relationship between biomass and maintenance
387 respiration is non-linear; R_m increases as biomass increases, where Kr_b controls the intercept of this relationship (Tian et al.,
388 1999). Since NPP is computed as a resultant of GPP and autotrophic respiration, including R_m , any alteration in Kr_b impacts
389 NPP directly (Eq. S9). This sensitivity underscores the importance of accurately calibrating Kr_b to ensure the correct
390 simulation of ecosystem productivity and C dynamics in the DVM-DOS-TEM.

391

392 **4.7 Vegetation and Below-Ground C stocks equilibrium time**

393 Since vegetation C and N is characterized by faster turnover than soil carbon dynamic, vegetation C and N stocks and fluxes
394 equilibrate faster than soil C and N stocks and fluxes, we used a two-phase equilibration approach: 200 years for the vegetation
395 and 2000 years for the soil. However, the C stocks achieved after 200 years of equilibration for vegetation might shift when
396 the model is run for an additional 1800 years to equilibrate soil. To mitigate this issue, we developed equilibrium checks to



397 ensure that the vegetation stocks remain stable and close to their equilibrium values throughout the extended simulation period
398 required for soil stocks equilibration. These equilibrium checks help identify significant departures from the initial equilibrium
399 values of vegetation C while allowing the model to run for a longer duration to achieve below-ground C equilibrium. This
400 approach ensures the accuracy and stability of both vegetation and below-ground C stocks in long-term model simulations.

401 **5. Conclusion**

402 In this study, we showed that the developed MADS parameter calibration method for the DVM-DOS-TEM can effectively
403 recover the synthetic parameter set, optimizing labor and time, and enhancing reproducibility of the calibration process. By
404 implementing a structured workflow that calibrates one or two parameters at a time and including equilibrium checks the
405 method ensured accurate parameter estimation even for high variance parameter range. The primary advantage of the semi-
406 automated MADS calibration approach is its significant enhancement of repeatability and clear quantification of calibration
407 performance. In contrast, manual calibration processes are often difficult to reproduce as it is impractical if not impossible, to
408 record users continuous adjustments to parameters values until improved results were achieved. Additionally, appreciation of
409 model improvement by the user is often subjective as running a statistical evaluation at each parameter adjustment would be
410 too time consuming. In the approach demonstrated in this study, we introduced a calibration metric that provides a quantifiable
411 measure of the overall quality of the calibration. This metric enhances reproducibility by allowing future users working on the
412 same site to follow the established workflow and reliably reproduce the calibrated parameter and target values. The RMSE
413 quantifies the average differences between calibrated and observed (synthetic) values, while the RE metric indicates deviations
414 from the synthetic values.

415
416 In all calibration experiments, we utilized only ten randomly perturbed initial parameter sets within a specified variance range.
417 Our results indicated that perturbation ranges of 10%-20% were equally effective in achieving optimal target/parameter
418 calibration. However, increasing the number of random perturbations could potentially shift the statistics, favoring a 10%
419 variance range. Based on our findings, we recommend maintaining a small parameter variance interval, as this approach is
420 likely to provide a robust match with target values and ensure effective parameter calibration.

421
422 While the choice of the initial guess is crucial, its impact was mitigated in our study due to the design involving variance
423 around synthetic parameter values. The developed method significantly reduces the labor and time required for calibrating
424 DVM-DOS-TEM model parameters. However, it does not entirely replace the need for human intervention. Users still need
425 to understand the specifics of the model and the relationship between parameters and targets, as well as conduct post-processing
426 assessments of the fit. In future work, we will apply this method to data processed at multiple study sites to validate further
427 and refine the calibration approach.

428



429 **6. Data and model availability**

430 The version of the model used in these simulations, along with the calibration scripts, auxiliary files (including plots presented
431 in the paper), and corresponding output files, are available at the following link: <https://doi.org/10.5281/zenodo.13772987>.

433 **7. Author contributions**

434 EEJ designed and executed the experiment. HG supervised in experiment design. VV supervised with MADS model. RR, TC,
435 and DT provided technical support on the DVM-DOS-TEM model. VB, AK, ALM, BM, C-CC, and JC tested calibration
436 approach. TS technical support on scientific computing. All authors participated in manuscript writing and editing. SMN and
437 BMR provided overall supervision and research funding.

439 **8. Competing interests**

440 The contact author has declared that none of the authors has any competing interests.

442 **9. Disclaimer**

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449 **11. References**

450 Andresen, C. G., Lawrence, D. M., Wilson, C. J., McGuire, A. D., Koven, C., Schaefer, K., Jafarov, E., Peng, S., Chen, X.,
451 Gouttevin, I., Burke, E., Chadburn, S., Ji, D., Chen, G., Hayes, D., and Zhang, W.: Soil moisture and hydrology projections of
452 the permafrost region – a model intercomparison, *The Cryosphere*, 14, 445–459, <https://doi.org/10.5194/tc-14-445-2020>, 2020.

453 Barajas-Solano, D. A., Wohlberg, B. E., Vesselinov, V. V., and Tartakovsky, D. M.: Linear functional minimization for inverse
454 modeling, *Water Resources Research*, 51, 4516–4531, <https://doi.org/10.1002/2014WR016179>, 2015.

455 Beven, K. and Freer, J.: Equifinality, data assimilation, and uncertainty estimation in mechanistic modelling of complex
456 environmental systems using the GLUE methodology, *Journal of Hydrology*, 249, 11–29, [https://doi.org/10.1016/S0022-1694\(01\)00421-8](https://doi.org/10.1016/S0022-1694(01)00421-8), 2001.

458 Bloom, A. A., Exbrayat, J.-F., Van Der Velde, I. R., Feng, L., and Williams, M.: The decadal state of the terrestrial carbon
459 cycle: Global retrievals of terrestrial carbon allocation, pools, and residence times, *Proc. Natl. Acad. Sci. U.S.A.*, 113, 1285–
460 1290, <https://doi.org/10.1073/pnas.1515160113>, 2016.



- 461 Briones, V., Jafarov, E. E., Genet, H., Rogers, B. M., Rutter, R. M., Carman, T. B., Clein, J., Euschkirchen, E. S., Schuur, E.
462 A., Watts, J. D., and Natali, S. M.: Exploring the interplay between soil thermal and hydrological changes and their impact on
463 carbon fluxes in permafrost ecosystems, *Environ. Res. Lett.*, 19, 074003, <https://doi.org/10.1088/1748-9326/ad50ed>, 2024.
- 464 Brunetti, G., Šimunek, J., Wöhling, T., and Stumpp, C.: An in-depth analysis of Markov-Chain Monte Carlo ensemble samplers
465 for inverse vadose zone modeling, *Journal of Hydrology*, 624, 129822, <https://doi.org/10.1016/j.jhydrol.2023.129822>, 2023.
- 466 Calef, M. P., David McGuire, A., Epstein, H. E., Scott Rupp, T., and Shugart, H. H.: Analysis of vegetation distribution in
467 Interior Alaska and sensitivity to climate change using a logistic regression approach, *Journal of Biogeography*, 32, 863–878,
468 <https://doi.org/10.1111/j.1365-2699.2004.01185.x>, 2005.
- 469 Castelletti, A., Galelli, S., Ratto, M., Soncini-Sessa, R., and Young, P. C.: A general framework for Dynamic Emulation
470 Modelling in environmental problems, *Environmental Modelling & Software*, 34, 5–18,
471 <https://doi.org/10.1016/j.envsoft.2012.01.002>, 2012.
- 472 Chapin, F. S. and Kedrowski, R. A.: Seasonal Changes in Nitrogen and Phosphorus Fractions and Autumn Retranslocation in
473 Evergreen and Deciduous Taiga Trees, *Ecology*, 64, 376–391, <https://doi.org/10.2307/1937083>, 1983.
- 474 Dagon, K., Sanderson, B. M., Fisher, R. A., and Lawrence, D. M.: A machine learning approach to emulation and biophysical
475 parameter estimation with the Community Land Model, version 5, *Adv. Stat. Clim. Meteorol. Oceanogr.*, 6, 223–244,
476 <https://doi.org/10.5194/ascmo-6-223-2020>, 2020.
- 477 Dietze, M. C., Fox, A., Beck-Johnson, L. M., Betancourt, J. L., Hooten, M. B., Jarnevich, C. S., Keitt, T. H., Kenney, M. A.,
478 Laney, C. M., Larsen, L. G., Loeschner, H. W., Lunch, C. K., Pijanowski, B. C., Randerson, J. T., Read, E. K., Tredennick, A.
479 T., Vargas, R., Weathers, K. C., and White, E. P.: Iterative near-term ecological forecasting: Needs, opportunities, and
480 challenges, *Proc. Natl. Acad. Sci. U.S.A.*, 115, 1424–1432, <https://doi.org/10.1073/pnas.1710231115>, 2018.
- 481 Efstratiadis, A. and Koutsoyiannis, D.: One decade of multi-objective calibration approaches in hydrological modelling: a
482 review, *Hydrological Sciences Journal*, 55, 58–78, <https://doi.org/10.1080/02626660903526292>, 2010.
- 483 Euschkirchen, E. S., McGUIRE, A. D., Kicklighter, D. W., Zhuang, Q., Clein, J. S., Dargaville, R. J., Dye, D. G., Kimball, J. S.,
484 McDONALD, K. C., Melillo, J. M., Romanovsky, V. E., and Smith, N. V.: Importance of recent shifts in soil thermal dynamics
485 on growing season length, productivity, and carbon sequestration in terrestrial high-latitude ecosystems: HIGH-LATITUDE
486 CLIMATE CHANGE INDICATORS, *Global Change Biology*, 12, 731–750, <https://doi.org/10.1111/j.1365-2486.2006.01113.x>, 2006.
- 488 Euschkirchen, E. S., McGuire, A. D., Chapin, F. S., Yi, S., and Thompson, C. C.: Changes in vegetation in northern Alaska under
489 scenarios of climate change, 2003–2100: implications for climate feedbacks, *Ecological Applications*, 19, 1022–1043,
490 <https://doi.org/10.1890/08-0806.1>, 2009.
- 491 Euschkirchen, E. S., Edgar, C. W., Turetsky, M. R., Waldrop, M. P., and Harden, J. W.: Differential response of carbon fluxes
492 to climate in three peatland ecosystems that vary in the presence and stability of permafrost: Carbon fluxes and permafrost
493 thaw, *J. Geophys. Res. Biogeosci.*, 119, 1576–1595, <https://doi.org/10.1002/2014JG002683>, 2014.
- 494 Euschkirchen, E. S., Serbin, S. P., Carman, T. B., Fraterrigo, J. M., Genet, H., Iversen, C. M., Salmon, V., and McGuire, A. D.:
495 Assessing dynamic vegetation model parameter uncertainty across Alaskan arctic tundra plant communities, *Ecological
496 Applications*, 32, <https://doi.org/10.1002/eap.2499>, 2022.



- 497 Fer, I., Kelly, R., Moorcroft, P. R., Richardson, A. D., Cowdery, E. M., and Dietze, M. C.: Linking big models to big data:
498 efficient ecosystem model calibration through Bayesian model emulation, *Biogeosciences*, 15, 5801–5830,
499 <https://doi.org/10.5194/bg-15-5801-2018>, 2018.
- 500 Forrester, A. I. J., Bressloff, N. W., and Keane, A. J.: Optimization using surrogate models and partially converged
501 computational fluid dynamics simulations, *Proc. R. Soc. A.*, 462, 2177–2204, <https://doi.org/10.1098/rspa.2006.1679>, 2006.
- 502 Fox, A. M., Hoar, T. J., Anderson, J. L., Arellano, A. F., Smith, W. K., Litvak, M. E., MacBean, N., Schimel, D. S., and Moore,
503 D. J. P.: Evaluation of a Data Assimilation System for Land Surface Models Using CLM4.5, *J Adv Model Earth Syst*, 10,
504 2471–2494, <https://doi.org/10.1029/2018MS001362>, 2018.
- 505 Genet, H., McGuire, A. D., Barrett, K., Breen, A., Euskirchen, E. S., Johnstone, J. F., Kasischke, E. S., Melvin, A. M., Bennett,
506 A., Mack, M. C., Rupp, T. S., Schuur, A. E. G., Turetsky, M. R., and Yuan, F.: Modeling the effects of fire severity and climate
507 warming on active layer thickness and soil carbon storage of black spruce forests across the landscape in interior Alaska,
508 *Environ. Res. Lett.*, 8, 045016, <https://doi.org/10.1088/1748-9326/8/4/045016>, 2013.
- 509 Genet, H., He, Y., Lyu, Z., McGuire, A. D., Zhuang, Q., Clein, J., D’Amore, D., Bennett, A., Breen, A., Biles, F., Euskirchen,
510 E. S., Johnson, K., Kurkowski, T., (Kushch) Schroder, S., Pastick, N., Rupp, T. S., Wylie, B., Zhang, Y., Zhou, X., and Zhu,
511 Z.: The role of driving factors in historical and projected carbon dynamics of upland ecosystems in Alaska, *Ecol Appl*, 28, 5–
512 27, <https://doi.org/10.1002/eap.1641>, 2018.
- 513 Hansen, P. C.: Rank-Deficient and Discrete Ill-Posed Problems: Numerical Aspects of Linear Inversion, *Society for Industrial
514 and Applied Mathematics*, <https://doi.org/10.1137/1.9780898719697>, 1998.
- 515 Harp, D. R., Atchley, A. L., Painter, S. L., Coon, E. T., Wilson, C. J., Romanovsky, V. E., and Rowland, J. C.: Effect of soil
516 property uncertainties on permafrost thaw projections: a calibration-constrained analysis, *The Cryosphere*, 10, 341–358,
517 <https://doi.org/10.5194/tc-10-341-2016>, 2016.
- 518 Harris, I., Jones, P. D., Osborn, T. J., and Lister, D. H.: Updated high-resolution grids of monthly climatic observations - the
519 CRU TS3.10 Dataset: UPDATED HIGH-RESOLUTION GRIDS OF MONTHLY CLIMATIC OBSERVATIONS, *Int. J.
520 Climatol.*, 34, 623–642, <https://doi.org/10.1002/joc.3711>, 2014.
- 521 Hugelius, G., Strauss, J., Zubrzycki, S., Harden, J. W., Schuur, E. A. G., Ping, C.-L., Schirmer, L., Grosse, G., Michaelson,
522 G. J., Koven, C. D., O’Donnell, J. A., Elberling, B., Mishra, U., Camill, P., Yu, Z., Palmtag, J., and Kuhry, P.: Estimated stocks
523 of circumpolar permafrost carbon with quantified uncertainty ranges and identified data gaps, *Biogeosciences*, 11, 6573–6593,
524 <https://doi.org/10.5194/bg-11-6573-2014>, 2014.
- 525 Jafarov, E. E., Romanovsky, V. E., Genet, H., McGuire, A. D., and Marchenko, S. S.: The effects of fire on the thermal stability
526 of permafrost in lowland and upland black spruce forests of interior Alaska in a changing climate, *Environ. Res. Lett.*, 8,
527 035030, <https://doi.org/10.1088/1748-9326/8/3/035030>, 2013.
- 528 Jafarov, E. E., Nicolsky, D. J., Romanovsky, V. E., Walsh, J. E., Panda, S. K., and Serreze, M. C.: The effect of snow: How
529 to better model ground surface temperatures, *Cold Regions Science and Technology*, 102, 63–77,
530 <https://doi.org/10.1016/j.coldregions.2014.02.007>, 2014.
- 531 Jafarov, E. E., Harp, D. R., Coon, E. T., Dafflon, B., Tran, A. P., Atchley, A. L., Lin, Y., and Wilson, C. J.: Estimation of
532 subsurface porosities and thermal conductivities of polygonal tundra by coupled inversion of electrical resistivity, temperature,
533 and moisture content data, *The Cryosphere*, 14, 77–91, <https://doi.org/10.5194/tc-14-77-2020>, 2020.



- 534 Jean, M., Melvin, A. M., Mack, M. C., and Johnstone, J. F.: Broadleaf Litter Controls Feather Moss Growth in Black Spruce
535 and Birch Forests of Interior Alaska, *Ecosystems*, 23, 18–33, <https://doi.org/10.1007/s10021-019-00384-8>, 2020.
- 536 Kelly, R., Chipman, M. L., Higuera, P. E., Stefanova, I., Brubaker, L. B., and Hu, F. S.: Recent burning of boreal forests
537 exceeds fire regime limits of the past 10,000 years, *Proc. Natl. Acad. Sci. U.S.A.*, 110, 13055–13060,
538 <https://doi.org/10.1073/pnas.1305069110>, 2013.
- 539 Koziel, S., Ciaurri, D. E., and Leifsson, L.: Surrogate-Based Methods, in: *Computational Optimization, Methods and*
540 *Algorithms*, vol. 356, edited by: Koziel, S. and Yang, X.-S., Springer Berlin Heidelberg, Berlin, Heidelberg, 33–59,
541 https://doi.org/10.1007/978-3-642-20859-1_3, 2011.
- 542 Lara, M. J., Genet, H., McGuire, A. D., Euskirchen, E. S., Zhang, Y., Brown, D. R. N., Jorgenson, M. T., Romanovsky, V.,
543 Breen, A., and Bolton, W. R.: Thermokarst rates intensify due to climate change and forest fragmentation in an Alaskan boreal
544 forest lowland, *Global Change Biology*, 22, 816–829, <https://doi.org/10.1111/gcb.13124>, 2016.
- 545 Levenberg, K.: A METHOD FOR THE SOLUTION OF CERTAIN NON-LINEAR PROBLEMS IN LEAST SQUARES,
546 *Quarterly of Applied Mathematics*, 2, 164–168, 1944.
- 547 Lin, Y., O'Malley, D., and Vesselinov, V. V.: A computationally efficient parallel Levenberg-Marquardt algorithm for highly
548 parameterized inverse model analyses, *Water Resources Research*, 52, 6948–6977, <https://doi.org/10.1002/2016WR019028>,
549 2016.
- 550 Linde, N., Renard, P., Mukerji, T., and Caers, J.: Geological realism in hydrogeological and geophysical inverse modeling: A
551 review, *Advances in Water Resources*, 86, 86–101, <https://doi.org/10.1016/j.advwatres.2015.09.019>, 2015.
- 552 Ling, X.-L., Fu, C.-B., Yang, Z.-L., and Guo, W.-D.: Comparison of different sequential assimilation algorithms for satellite-
553 derived leaf area index using the Data Assimilation Research Testbed (version Lanai), *Geosci. Model Dev.*, 12, 3119–3133,
554 <https://doi.org/10.5194/gmd-12-3119-2019>, 2019.
- 555 Luo, Y., Ahlström, A., Allison, S. D., Batjes, N. H., Brovkin, V., Carvalhais, N., Chappell, A., Ciais, P., Davidson, E. A.,
556 Finzi, A., Georgiou, K., Guenet, B., Hararuk, O., Harden, J. W., He, Y., Hopkins, F., Jiang, L., Koven, C., Jackson, R. B.,
557 Jones, C. D., Lara, M. J., Liang, J., McGuire, A. D., Parton, W., Peng, C., Randerson, J. T., Salazar, A., Sierra, C. A., Smith,
558 M. J., Tian, H., Todd-Brown, K. E. O., Torn, M., Van Groenigen, K. J., Wang, Y. P., West, T. O., Wei, Y., Wieder, W. R.,
559 Xia, J., Xu, X., Xu, X., and Zhou, T.: Toward more realistic projections of soil carbon dynamics by Earth system models,
560 *Global Biogeochemical Cycles*, 30, 40–56, <https://doi.org/10.1002/2015GB005239>, 2016.
- 561 MacBean, N., Peylin, P., Chevallier, F., Scholze, M., and Schürmann, G.: Consistent assimilation of multiple data streams in
562 a carbon cycle dataassimilation system, *Geosci. Model Dev.*, 9, 3569–3588, <https://doi.org/10.5194/gmd-9-3569-2016>, 2016.
- 563 Marquardt, D. W.: An Algorithm for Least-Squares Estimation of Nonlinear Parameters, *Journal of the Society for Industrial*
564 *and Applied Mathematics*, 11, 431–441, 1963.
- 565 McGuire, A. D., Koven, C., Lawrence, D. M., Clein, J. S., Xia, J., Beer, C., Burke, E., Chen, G., Chen, X., Delire, C., Jafarov,
566 E., MacDougall, A. H., Marchenko, S., Nicolsky, D., Peng, S., Rinke, A., Saito, K., Zhang, W., Alkama, R., Bohn, T. J., Ciais,
567 P., Decharme, B., Ekici, A., Gouttevin, I., Hajima, T., Hayes, D. J., Ji, D., Krinner, G., Lettenmaier, D. P., Luo, Y., Miller, P.
568 A., Moore, J. C., Romanovsky, V., Schädel, C., Schaefer, K., Schuur, E. A. G., Smith, B., Sueyoshi, T., and Zhuang, Q.:
569 Variability in the sensitivity among model simulations of permafrost and carbon dynamics in the permafrost region between
570 1960 and 2009: MODELING PERMAFROST CARBON DYNAMICS, *Global Biogeochem. Cycles*, 30, 1015–1037,
571 <https://doi.org/10.1002/2016GB005405>, 2016.



- 572 McGuire, A. D., Lawrence, D. M., Koven, C., Klein, J. S., Burke, E., Chen, G., Jafarov, E., MacDougall, A. H., Marchenko,
573 S., Nicolsky, D., Peng, S., Rinke, A., Ciais, P., Gouttevin, I., Hayes, D. J., Ji, D., Krinner, G., Moore, J. C., Romanovsky, V.,
574 Schädel, C., Schaefer, K., Schuur, E. A. G., and Zhuang, Q.: Dependence of the evolution of carbon dynamics in the northern
575 permafrost region on the trajectory of climate change, *Proc Natl Acad Sci USA*, 115, 3882–3887,
576 <https://doi.org/10.1073/pnas.1719903115>, 2018.
- 577 Melvin, A. M., Mack, M. C., Johnstone, J. F., David McGuire, A., Genet, H., and Schuur, E. A. G.: Differences in Ecosystem
578 Carbon Distribution and Nutrient Cycling Linked to Forest Tree Species Composition in a Mid-Successional Boreal Forest,
579 *Ecosystems*, 18, 1472–1488, <https://doi.org/10.1007/s10021-015-9912-7>, 2015.
- 580 Natali, S. M., Holdren, J. P., Rogers, B. M., Treharne, R., Duffy, P. B., Pomerance, R., and MacDonald, E.: Permafrost carbon
581 feedbacks threaten global climate goals, *Proc. Natl. Acad. Sci. U.S.A.*, 118, e2100163118,
582 <https://doi.org/10.1073/pnas.2100163118>, 2021.
- 583 Nicolsky, D. J., Romanovsky, V. E., and Tzipenko, G. S.: Using in-situ temperature measurements to estimate saturated soil
584 thermal properties by solving a sequence of optimization problems, *The Cryosphere*, 1, 41–58, <https://doi.org/10.5194/tc-1-41-2007>, 2007.
- 586 Nicolsky, D. J., Romanovsky, V. E., and Panteleev, G. G.: Estimation of soil thermal properties using in-situ temperature
587 measurements in the active layer and permafrost, *Cold Regions Science and Technology*, 55, 120–129,
588 <https://doi.org/10.1016/j.coldregions.2008.03.003>, 2009.
- 589 O’Malley, D. and Vesselinov, V. V.: Bayesian-information-gap decision theory with an application to CO₂ sequestration:
590 BAYESIAN-INFORMATION-GAP DECISION THEORY, *Water Resour. Res.*, 51, 7080–7089,
591 <https://doi.org/10.1002/2015WR017413>, 2015.
- 592 Pastick, N. J., Duffy, P., Genet, H., Rupp, T. S., Wylie, B. K., Johnson, K. D., Jorgenson, M. T., Bliss, N., McGuire, A. D.,
593 Jafarov, E. E., and Knight, J. F.: Historical and projected trends in landscape drivers affecting carbon dynamics in Alaska,
594 *Ecological Applications*, 27, 1383–1402, <https://doi.org/10.1002/eap.1538>, 2017.
- 595 Peylin, P., Bacour, C., MacBean, N., Leonard, S., Rayner, P., Kuppel, S., Koffi, E., Kane, A., Maignan, F., Chevallier, F.,
596 Ciais, P., and Prunet, P.: A new stepwise carbon cycle data assimilation system using multiple data streams to constrain the
597 simulated land surface carbon cycle, *Geosci. Model Dev.*, 9, 3321–3346, <https://doi.org/10.5194/gmd-9-3321-2016>, 2016.
- 598 Pujol, J.: The solution of nonlinear inverse problems and the Levenberg-Marquardt method, *GEOPHYSICS*, 72, W1–W16,
599 <https://doi.org/10.1190/1.2732552>, 2007.
- 600 Queipo, N. V., Haftka, R. T., Shyy, W., Goel, T., Vaidyanathan, R., and Kevin Tucker, P.: Surrogate-based analysis and
601 optimization, *Progress in Aerospace Sciences*, 41, 1–28, <https://doi.org/10.1016/j.paerosci.2005.02.001>, 2005.
- 602 Razavi, S., Tolson, B. A., and Burn, D. H.: Review of surrogate modeling in water resources, *Water Resources Research*, 48,
603 2011WR011527, <https://doi.org/10.1029/2011WR011527>, 2012.
- 604 Regis, R. G. and Shoemaker, C. A.: A Stochastic Radial Basis Function Method for the Global Optimization of Expensive
605 Functions, *INFORMS Journal on Computing*, 19, 497–509, <https://doi.org/10.1287/ijoc.1060.0182>, 2007.
- 606 Reichstein, M., Camps-Valls, G., Stevens, B., Jung, M., Denzler, J., Carvalhais, N., and Prabhat: Deep learning and process
607 understanding for data-driven Earth system science, *Nature*, 566, 195–204, <https://doi.org/10.1038/s41586-019-0912-1>, 2019.



- 608 Ruess, R. W., Cleve, K. V., Yarie, J., and Viereck, L. A.: Contributions of fine root production and turnover to the carbon and
609 nitrogen cycling in taiga forests of the Alaskan interior, *Can. J. For. Res.*, 26, 1326–1336, <https://doi.org/10.1139/x26-148>,
610 1996.
- 611 Rykiel, E. J.: Testing ecological models: the meaning of validation, *Ecological Modelling*, 90, 229–244,
612 [https://doi.org/10.1016/0304-3800\(95\)00152-2](https://doi.org/10.1016/0304-3800(95)00152-2), 1996.
- 613 Schädel, C., Rogers, B. M., Lawrence, D. M., Koven, C. D., Brovkin, V., Burke, E. J., Genet, H., Huntzinger, D. N., Jafarov,
614 E., McGuire, A. D., Riley, W. J., and Natali, S. M.: Earth system models must include permafrost carbon processes, *Nat. Clim.*
615 *Chang.*, <https://doi.org/10.1038/s41558-023-01909-9>, 2024.
- 616 Scholze, M., Kaminski, T., Knorr, W., Blessing, S., Vossbeck, M., Grant, J. P., and Scipal, K.: Simultaneous assimilation of
617 SMOS soil moisture and atmospheric CO₂ in-situ observations to constrain the global terrestrial carbon cycle, *Remote Sensing*
618 *of Environment*, 180, 334–345, <https://doi.org/10.1016/j.rse.2016.02.058>, 2016.
- 619 Schürmann, G. J., Kaminski, T., Köstler, C., Carvalhais, N., Voßbeck, M., Kattge, J., Giering, R., Rödenbeck, C., Heimann,
620 M., and Zaehle, S.: Constraining a land-surface model with multiple observations by application of the MPI-Carbon Cycle
621 Data Assimilation System V1.0, *Geosci. Model Dev.*, 9, 2999–3026, <https://doi.org/10.5194/gmd-9-2999-2016>, 2016.
- 622 Schuur, E. A. G., Abbott, B. W., Commane, R., Ernakovich, J., Euskirchen, E., Hugelius, G., Grosse, G., Jones, M., Koven,
623 C., Leshyk, V., Lawrence, D., Lorant, M. M., Mauritz, M., Olefeldt, D., Natali, S., Rodenhizer, H., Salmon, V., Schädel, C.,
624 Strauss, J., Treat, C., and Turetsky, M.: Permafrost and Climate Change: Carbon Cycle Feedbacks From the Warming Arctic,
625 *Annu. Rev. Environ. Resour.*, 47, 343–371, <https://doi.org/10.1146/annurev-environ-012220-011847>, 2022.
- 626 Shaver, G. R. and Chapin, F. S.: Long-term responses to factorial, NPK fertilizer treatment by Alaskan wet and moist tundra
627 sedge species, *Ecography*, 18, 259–275, <https://doi.org/10.1111/j.1600-0587.1995.tb00129.x>, 1995.
- 628 Tian, H., Melillo, J. M., Kicklighter, D. W., McGuire, A. D., and Helfrich, J.: The sensitivity of terrestrial carbon storage to
629 historical climate variability and atmospheric CO₂ in the United States, *Tellus B: Chemical and Physical Meteorology*, 51,
630 414, <https://doi.org/10.3402/tellusb.v51i2.16318>, 1999.
- 631 Tran, A. P., Dafflon, B., and Hubbard, S. S.: Coupled land surface–subsurface hydrogeophysical inverse modeling to estimate
632 soil organic carbon content and explore associated hydrological and thermal dynamics in the Arctic tundra, *The Cryosphere*,
633 11, 2089–2109, <https://doi.org/10.5194/tc-11-2089-2017>, 2017.
- 634 Treharne, R., Rogers, B. M., Gasser, T., MacDonald, E., and Natali, S.: Identifying Barriers to Estimating Carbon Release
635 From Interacting Feedbacks in a Warming Arctic, *Front. Clim.*, 3, 716464, <https://doi.org/10.3389/fclim.2021.716464>, 2022.
- 636 Vesselinov V.V.: MADS: Model Analysis and Decision Support in Julia, <https://github.com/madsjulia/Mads.jl>, 2022.
- 637 Wutzler, T. and Carvalhais, N.: Balancing multiple constraints in model-data integration: Weights and the parameter block
638 approach, *JGR Biogeosciences*, 119, 2112–2129, <https://doi.org/10.1002/2014JG002650>, 2014.
- 639 Xu, T., Valocchi, A. J., Ye, M., Liang, F., and Lin, Y.: Bayesian calibration of groundwater models with input data uncertainty,
640 *Water Resources Research*, 53, 3224–3245, <https://doi.org/10.1002/2016WR019512>, 2017.
- 641 Yi, S., McGuire, A. D., Harden, J., Kasischke, E., Manies, K., Hinzman, L., Liljedahl, A., Randerson, J., Liu, H., Romanovsky,
642 V., Marchenko, S., and Kim, Y.: Interactions between soil thermal and hydrological dynamics in the response of Alaska
643 ecosystems to fire disturbance: SOIL THERMAL AND HYDROLOGICAL DYNAMICS, *J. Geophys. Res.*, 114, n/a-n/a,
644 <https://doi.org/10.1029/2008JG000841>, 2009.



645 Yi, S., McGuire, A. D., Kasischke, E., Harden, J., Manies, K., Mack, M., and Turetsky, M.: A dynamic organic soil
646 biogeochemical model for simulating the effects of wildfire on soil environmental conditions and carbon dynamics of black
647 spruce forests, *J. Geophys. Res.*, 115, G04015, <https://doi.org/10.1029/2010JG001302>, 2010.

648 Zhuang, Q., McGuire, A. D., O'Neill, K. P., Harden, J. W., Romanovsky, V. E., and Yarie, J.: Modeling soil thermal and
649 carbon dynamics of a fire chronosequence in interior Alaska, *J. Geophys. Res.*, 108, 8147,
650 <https://doi.org/10.1029/2001JD001244>, 2002.

651