

1 **Implementation of the ISORROPIA-lite Aerosol** 2 **Thermodynamics Model into the EMAC Chemistry Climate** 3 **Model (based on MESSy v2.55): Implications for Aerosol** 4 **Composition and Acidity**

5 Alexandros Milousis¹, Alexandra P. Tsimpidi¹, Holger Tost², Spyros N. Pandis^{3,4}, Athanasios
6 Nenes^{3,5}, Astrid Kiendler-Scharr¹⁺, and Vlassis A. Karydis¹

7 ¹Forschungszentrum Jülich GmbH, Institute for Energy and Climate Research, IEK-8 Troposphere, Jülich, Germany

8 ²Johannes Gutenberg University Mainz, Institute of Atmospheric Physics, Mainz, Germany

9 ³FORTH ICE HT, Institute of Chemical Engineering Sciences, Patras 26504, Greece

10 ⁴University of Patras, Department of Chemical Engineering, Patras 26500, Greece

11 ⁵Ecole Polytechnique Fed Lausanne, School of Architecture Civil & Environmental Engineering Lab, Atmospheric
12 Processes & Their Impacts, CH-1015 Lausanne, Switzerland

13 ⁺deceased

14
15 *Correspondence to:* Vlassis Karydis (v.karydis@fz-juelich.de)

16 **Abstract.** This study explores the differences in performance and results by various versions of the
17 ISORROPIA thermodynamic module implemented within the global atmospheric chemistry model EMAC.
18 Three different versions of the module were used, ISORROPIA II v1, ISORROPIA II v2.3, and
19 ISORROPIA-lite. First, ISORROPIA II v2.3 replaced ISORROPIA II v1 in EMAC to improve pH
20 predictions close to neutral conditions. The newly developed ISORROPIA-lite has been added to EMAC
21 alongside ISORROPIA II v2.3. ISORROPIA-lite is more computationally efficient and assumes that
22 atmospheric aerosols exist always as supersaturated aqueous (metastable) solutions while ISORROPIA II
23 includes the option to allow the formation of solid salts at low RH conditions (stable state). The predictions
24 of EMAC by employing all three aerosol thermodynamic models were compared to each other and
25 evaluated against surface measurements from three regional observational networks (IMPROVE, EMEP,
26 EANET) in the polluted Northern Hemisphere. The differences between ISORROPIA II v2.3 and
27 ISORROPIA-lite were minimal in all comparisons with the normalized mean absolute difference for the
28 concentrations of all major aerosol components being less than 11 % even when different phase state
29 assumptions were used. The most notable differences were lower aerosol concentrations predicted by
30 ISORROPIA-lite in regions with relative humidity in the range of 20% to 60% compared to the predictions
31 of ISORROPIA II v2.3 in stable mode. The comparison against observations yielded satisfactory agreement
32 especially over the US and Europe, but higher deviations over East Asia, where the overprediction of
33 EMAC for nitrate was as high as 4 $\mu\text{g m}^{-3}$ (~ 20%). The mean annual aerosol pH predicted by ISORROPIA-
34 lite was on average less than a unit lower than ISORROPIA II v2.3 in stable mode, mainly for coarse mode
35 aerosols over Middle East. The use of ISORROPIA-lite accelerated EMAC by nearly 5 % compared to the
36 use of ISORROPIA II v2.3 even if the aerosol thermodynamic calculations consume a relatively small
37 fraction of the EMAC computational time. ISORROPIA-lite can therefore be a reliable and computationally
38 efficient alternative to the previous thermodynamic module in EMAC.

39
40 **Keywords:** atmospheric aerosols, aerosol thermodynamics, nitrate, acidity, aerosol phase state.

41

42 1. Introduction

43 Aerosols in the atmosphere have a significant impact on climate and air pollution. They contribute
44 to the deterioration of air quality, especially in heavily industrialised regions, leading to increased
45 mortality rates and decreased life expectancy (H eroux et al., 2015). Particulate matter with
46 diameter less than 2.5 μm ($\text{PM}_{2.5}$) is the largest contributor to stroke, cancer, heart conditions and
47 chronic obstructive pulmonary diseases (Brook et al., 2010; Pope et al., 2011) with ambient
48 pollution causing approximately 4.2 million premature deaths in 2019 alone (WHO, 2022). Tarin-
49 Carrasco et al. (2021) predicted that mortality rates in Europe due to air pollution could increase
50 in the next thirty years in the more extreme emission scenarios (e.g., RCP8.5). In addition to the
51 direct threat aerosols pose to humans and ecosystems through their effects on air quality, they can
52 also affect other climate-related processes. For example, they can act as cloud condensation nuclei
53 and modify cloud lifetime and optical properties (Andreae et al., 2005; Klingm uller et al., 2020).
54 Aerosols also affect the energy balance of our planet by reflecting additional solar radiation back
55 to space and thus cooling the atmosphere or by absorbing solar radiation warming the atmosphere
56 (Klingm uller et al., 2019; Miinalainen et al., 2021). Some major inorganic aerosol components
57 also affect various ecosystems. For example, nitrates and sulfates can harm flora by lessening its
58 lifetime and variety (Honour et al., 2009; Manisalidis et al., 2020), and can affect wildlife by
59 causing water eutrophication (Doney et al., 2007). A critical property of atmospheric particles that
60 regulates their impacts on clouds and ecosystems is their acidity (Karydis et al., 2021). Depending
61 on its levels, acidity can affect air quality and human health (Lelieveld et al., 2015) but also the
62 aerosols' hygroscopic characteristics (Karydis et al., 2016). The aerosol pH also drives the
63 partitioning of semi-volatile inorganic components between the gas and aerosol phases (Nenes et
64 al., 2020). Finally, aerosol acidity plays a role in the activation of halogens in aerosols (Saiz-Lopez
65 and von Glasow, 2012), their toxicity (Fang et al., 2017) and also in secondary organic aerosol
66 formation (Marais et al., 2016).

67 Sulfate is the most important component of $\text{PM}_{2.5}$ inorganic aerosol, since it contributes
68 the most in terms of global mass burden (Szopa et al., 2021) and aerosol optical depth (AOD)
69 (Myhre et al., 2013). Nitrate contribution to the $\text{PM}_{2.5}$ aerosol composition is also important in
70 several areas (e.g., Europe, North America, East Asia) and seasons (He et al., 2001; Silva et al.,
71 2007; Weagle et al., 2018; Tang et al., 2021). The quantification of nitrate partitioning between
72 the gas and particulate phases is challenging partly because it is affected by meteorology
73 (temperature, relative humidity) and all ionic aerosol components, but also due to the lack of
74 observations to constrain the composition of the gas-phase components and the size-distribution
75 of the particulate phase. Nitrate in the form of ammonium nitrate is mainly found in the fine mode
76 (e.g., $\text{PM}_{2.5}$) (Putaud et al., 2010). This is especially the case over polluted regions where there is
77 enough ammonia remaining after the neutralization of sulfate (Karydis et al., 2011; Karydis et al.,
78 2016). In coastal and desert areas, nitrate is formed mainly by reactions of HNO_3 with sea salt and
79 dust particles (Savoie and Prospero, 1982; Wolff, 1984; Karydis et al., 2016) and therefore is found
80 mainly in the coarse particles. The importance of nitrate in the troposphere is expected to increase
81 in the following decades because SO_2 emissions are anticipated to drop while NH_3 emissions to
82 increase (Fu et al., 2017; Chen et al., 2019; Xu et al., 2020). With decreased SO_2 concentrations,

83 less ammonia is required to neutralize the sulfates and therefore more is available for ammonium
84 nitrate formation (Tsimpidi et al., 2007).

85 There have been several thermodynamic models developed in the last decades to calculate
86 the inorganic aerosol concentrations and composition in the atmosphere. Two of the first were
87 EQUIL and KEQUIL developed by Bassett and Seinfeld (1983). Then the MARS model was
88 developed by Saxena et al. (1986) with the aim of reducing the computational time required in
89 order to be incorporated into larger scale chemical transport models. MARS was the first model to
90 divide the composition domain into smaller sub-domains aiming to reduce the number of equations
91 needed to be solved. Then the SEQUILIB model by Pilinis and Seinfeld (1987) was the first to
92 incorporate sodium and chloride and the corresponding salts in the simulated aerosol system.
93 Further developments included EQUISOLV by Jacobson et al. (1996) as well as SCAPE by Kim
94 et al. (1993), which simulated temperature dependent deliquescence following Wexler and
95 Seinfeld (1991) and predicted the presence of liquid phase aerosols even at low relative humidity
96 (RH). E-AIM is another benchmark thermodynamic model which instead of solving algebraic
97 equations for equilibrium, uses the minimization of the Gibbs Free Energy approach (Wexler and
98 Clegg, 2002). Later versions of E-AIM also include selected organic aerosol components (Clegg
99 et al., 2003). Furthermore, AIOMFAC is a model that utilizes organic-inorganic interactions in
100 aqueous solutions in order to calculate activity coefficients up to high ionic strengths (Zuend et
101 al., 2008) and is based on the LIFAC model by Yan et al. (1999). Further developments in
102 AIOMFAC include a wider variety of organic compounds (Zuend et al., 2011). The EQSAM
103 thermodynamic model was developed by Metzger et al. (2002) with the basic concept that aerosol
104 activities in equilibrium are controlled by RH, and solute activity is a function of RH. The model
105 uses a domain structure based on sulphate availability to increase computational efficiency by
106 solving fewer thermodynamic equations, similar to Nenes et al. (1998). EQSAM and ISORROPIA
107 are the two available options for aerosol thermodynamics in the EMAC model.

108 Nenes et al. (1998) developed the ISORROPIA model in an effort to increase
109 computational efficiency while maintaining the accuracy of the calculations. The system simulated
110 by ISORROPIA included NH_4^+ , Na^+ , Cl^- , NO_3^- , SO_4^{2-} and H_2O . ISORROPIA also contains the
111 temperature dependent equations for deliquescence by Wexler and Seinfeld (1991) and is
112 computationally efficient so that it can be incorporated in 3D atmospheric models. In ISORROPIA,
113 the aerosol state is predicted as a weighted mean value of the dry and wet states. The weighting
114 factors depend on ambient RH, the mutual deliquescence relative humidity (MDRH) and the
115 deliquescence relative humidity (DRH) of the most hygroscopic salt in the mixture. An improved
116 version of ISORROPIA including the mineral ions K^+ , Ca^{2+} , and Mg^+ , called ISORROPIA II, was
117 developed by Fountoukis and Nenes (2007). The addition of the above crustal ions resulted in the
118 inclusion of 10 more salts and 3 more ions in the solid and aqueous phases respectively. The model
119 gained in computational efficiency by performing different calculations for different atmospheric
120 chemical composition regimes, which are determined by the abundance of each aerosol precursor
121 as well as the ambient temperature and relative humidity. Depending on the values of the so-called
122 ‘sulfate ratio’, the ‘crustal species and sodium ratio’ and the ‘crustal species’ ratio, five aerosol
123 composition regimes are determined in order to calculate the necessary equilibrium equations for
124 the species present in each regime. Furthermore, the use of pre-calculated look-up tables for the
125 activity coefficients (see Section 2.2), including their temperature dependence, is another factor

126 for the gain in computational efficiency. Like E-AIM, ISORROPIA II can solve the
127 thermodynamic equilibrium problem under stable or metastable conditions. In the second case
128 aerosols are assumed to exist only as supersaturated aqueous solutions even at low RH, while in
129 the first the aerosols are able to form solid salts. A very slightly updated version, called
130 ISORROPIA II v2.3 was introduced to improve aerosol pH predictions close to neutral conditions
131 (Song et al., 2018). More specifically, in some subcases of the ISORROPIA II regime, NH₃
132 evaporation was not taken into account in the aerosol pH calculations, leading to unrealistic
133 estimates close to neutrality (pH~7). This error had a minimal effect on the predicted gas phase
134 NH₃ levels and consequently on the inorganic aerosol concentrations. Moreover, it only affected a
135 few subcases and only when the stable mode was used. More details on these differences can be
136 found in Song et al. (2018). The newest development of ISORROPIA II, called ISORROPIA-lite,
137 was designed to be even more computationally efficient than its predecessor and to also include
138 the effects that organic aerosol components have on particle water and the semi volatile inorganic
139 aerosol species partitioning (Kakavas et al., 2022).

140 This study aims to evaluate the newly developed ISORROPIA-lite thermodynamic module
141 within the EMAC global climate and chemistry model and to explore any discrepancies on a global
142 scale, by utilizing different aerosol phase states. For this reason, our analysis explores the
143 differences in the results between ISORROPIA-lite and ISORROPIA II over diverse conditions
144 and environments. In Section 2 the model configuration and the treatment of inorganic aerosols
145 thermodynamics is presented. In Sections 3 and 4 the results and comparisons between the
146 simulations are analyzed and in Section 5 the major conclusions are presented.

147 **2. Model Configuration**

148 **2.1 EMAC model setup**

149 The EMAC (ECHAM5/MESSy) model is a global atmospheric chemistry and climate model
150 (Jockel et al., 2006). It includes a series of submodels and links them via the Modular Earth
151 Submodel System (Jöckel et al., 2005) to the base model (core) that is the 5th generation European
152 Center Hamburg general circulation model (Roeckner et al., 2006). Gas-phase chemistry is
153 simulated by MECCA (Sander et al., 2019) with a simplified scheme similar to the one used in
154 CCMI (Chemistry-Climate Model Initiative) like in Jockel et al. (2016). Aerosol microphysics
155 along with gas/aerosol partitioning are treated by GMXe in which the aerosols are differentiated
156 between soluble and insoluble modes with a total of seven lognormal modes (Pringle et al., 2010).
157 The soluble mode contains the nucleation, Aitken, accumulation, and coarse size ranges while the
158 insoluble mode lacks only the nucleation size range. Transfer of material between the insoluble
159 and soluble modes is calculated in two processes. After coagulation, when a hydrophobic and a
160 hydrophilic particle coagulate, the resulting mass is assumed to reside in the hydrophilic mode and
161 also when soluble material condenses onto a hydrophobic particle (after gas/aerosol partitioning)
162 it is again transferred to the hydrophilic mode (Pringle et al., 2010). Wet deposition of gases and
163 aerosols is described by SCAV (Tost et al., 2006; 2007), dry deposition via DRYDEP (Kerkweg
164 et al., 2006) and gravitational sedimentation of aerosols by SEDI (Kerkweg et al., 2006). Cloud

165 properties and microphysics are calculated by the CLOUD submodel (Roeckner et al., 2006)
166 utilizing the detailed two-moment liquid and ice-cloud microphysical scheme of Lohmann and
167 Ferrachat (2010) and considering a physically based treatment of the processes of liquid (Karydis
168 et al., 2017) and ice crystals (Bacer et al., 2018) activation. The organic aerosol composition and
169 evolution in the atmosphere is calculated by the ORACLE submodel (Tsimpidi et al., 2014; 2018).

170 The model simulations in this work were nudged towards actual meteorology using ERA05
171 data (Hersbach et al., 2020). For the purposes of this study the spectral resolution applied within
172 EMAC was the T63L31 which corresponds to a grid resolution of $1.875^\circ \times 1.875^\circ$, covering
173 vertical altitudes up to 25 km with a total of 31 layers. The simulations were all done for the period
174 2009-2010, with 2009 representing the model spin-up period.

175 Anthropogenic emissions of aerosols and aerosol precursors were based on the
176 EDGARv4.3.2 inventory (Crippa et al., 2018). Open biomass burning emissions were derived by
177 the GFEDv3.1 database (van der Werf et al., 2010), and natural emissions of NH_3 (volatilization
178 from soils and oceans) were based on the GEIA database (Bouwman et al., 1997). SO_2 emissions
179 by volcanic eruptions are based on the AEROCOM dataset (Dentener et al., 2006), as are emissions
180 of sea spray aerosols using the chemical composition proposed by Seinfeld and Pandis (2016).
181 Biogenic emissions of NO from soils are calculated online according to the algorithm of Yienger
182 and Levy (1995) while NO_x produced by lightning is also calculated online based on the
183 parameterization of Grewe et al. (2001). Oceanic emissions of DMS are calculated online by the
184 AIRSEA submodel (Pozzer et al., 2006). Finally, the dust emission fluxes are calculated online
185 according to Astitha et al. (2012), by taking into account the meteorological information for each
186 grid cell (i.e., temperature and relative humidity) as well as the different thresholds of friction
187 velocities above which suspension of dust particles takes place. The emissions of crustal ions
188 (Ca^{2+} , Mg^+ , K^+ and Na^+) are estimated as a fraction of the total dust flux based on the soil chemical
189 composition of each individual grid cell (Karydis et al., 2016; Klingmüller et al., 2018). These ions
190 are emitted in the insoluble accumulation and coarse size modes and are subsequently transferred
191 to the soluble aerosols by the processes described above.

192 **2.2 Inorganic aerosol thermodynamics treatment**

193 In this study, the ISORROPIA-lite aerosol thermodynamic model has been implemented into the
194 EMAC as part of the GMXe submodel, not as a replacement but as an alternative to the previous
195 version, in order to efficiently calculate the equilibrium partitioning of the inorganic species
196 between gas and aerosol phases. Furthermore, ISORROPIA II v2.3 is used to replace ISORROPIA
197 II v1 in the model.

198 Kinetic limitations in the partitioning need to be taken into consideration because only fine
199 aerosols are able to achieve equilibrium within the time frame of one model time step, which in
200 this study equals to 10 minutes. Therefore, the partitioning calculation is done in two stages
201 according to Pringle et al. (2010). First the amount of the gas-phase species that is able to
202 kinetically condense onto the aerosol phase within the model time step is calculated by assuming
203 diffusion limited condensation (Vignati et al., 2004). Then in the second stage, the partitioning
204 between this gas phase material and the aerosol phase is performed. The partitioning calculation

205 is performed for all seven size modes, i.e. in each model timestep ISORROPIA is called separately
206 for each of them.

207 According to Kakavas et al. (2022), ISORROPIA-lite features two main modifications in
208 its code, with regard to ISORROPIA II v2.3 (Song et al., 2018) and ISORROPIA II v1 (Fountoukis
209 and Nenes, 2007). First, the routines related to the stable case have been removed, since only the
210 metastable case is considered and all salts formed are deliquesced. However, CaSO_4 is the only
211 solid salt allowed to form, as it is considered insoluble for most atmospherically-relevant RH
212 values and precipitates spontaneously. Furthermore, for the calculation of binary activity
213 coefficients, ISORROPIA-lite uses the tabulated binary activity coefficient data for each salt from
214 Kusik-Meissner (Kusik and Meissner, 1978) instead of calculating them online, and includes their
215 temperature dependence according to Meissner and Peppas (1973). This is done by combining the
216 Kusik and Meissner (1978) model for specific ionic pairs with the Bromley (1973) activity
217 coefficient mixing rule for multicomponent mixtures. More information on this procedure, can be
218 found in Fountoukis and Nenes (2007). This second modification is the major contributor to the
219 computational speed-up provided by ISORROPIA-lite, which in an offline estimation was reported
220 to be around 35% (Kakavas et al., 2022). Furthermore, this feature could explain differences in
221 inorganic aerosol estimates with the previous version of ISORROPIA using the same aerosol state
222 assumption (metastable case). Another important modification is that the effect of organic aerosol
223 water on the inorganic semi volatile aerosol components is included. This consideration slightly
224 increases the aerosol pH but more significantly drives the phase partitioning towards the aerosol
225 phase in order to satisfy equilibrium conditions (Kakavas et al., 2022). However, this feature of
226 ISORROPIA-lite was not used in the present study, as the water uptake by organics is treated by
227 other parts of the GMXe aerosol microphysics submodel in the EMAC global model. The effects
228 of the secondary organic aerosol on aerosol water and nitrate partitioning are discussed by Kakavas
229 et al. (2023).

230 In the updated version of the GMXe submodel, the users have the option to select between
231 ISORROPIA-lite and ISORROPIA II v2.3 to perform EMAC simulations depending on the
232 application and the desired phase state assumption. While ISORROPIA-lite utilizes the metastable
233 approach exclusively, ISORROPIA II v2.3 utilizes both and has the stable approach as default.

234 **3. Evaluation of New Aerosol Thermodynamic Modules within EMAC**

235 For reasons of clarity, from this point forward both in the main text as well as in any figure
236 captions, whenever different aerosol sizes are mentioned, total suspended particles (TSP) refer to
237 the sum of the 4 lognormal size modes of the aerosol microphysics submodel (i.e. nucleation,
238 Aitken, accumulation and coarse mode), fine aerosols refer to the sum of the 3 smaller size modes
239 (nucleation, Aitken and accumulation mode) and coarse aerosols refer to the largest size mode of
240 the model exclusively.

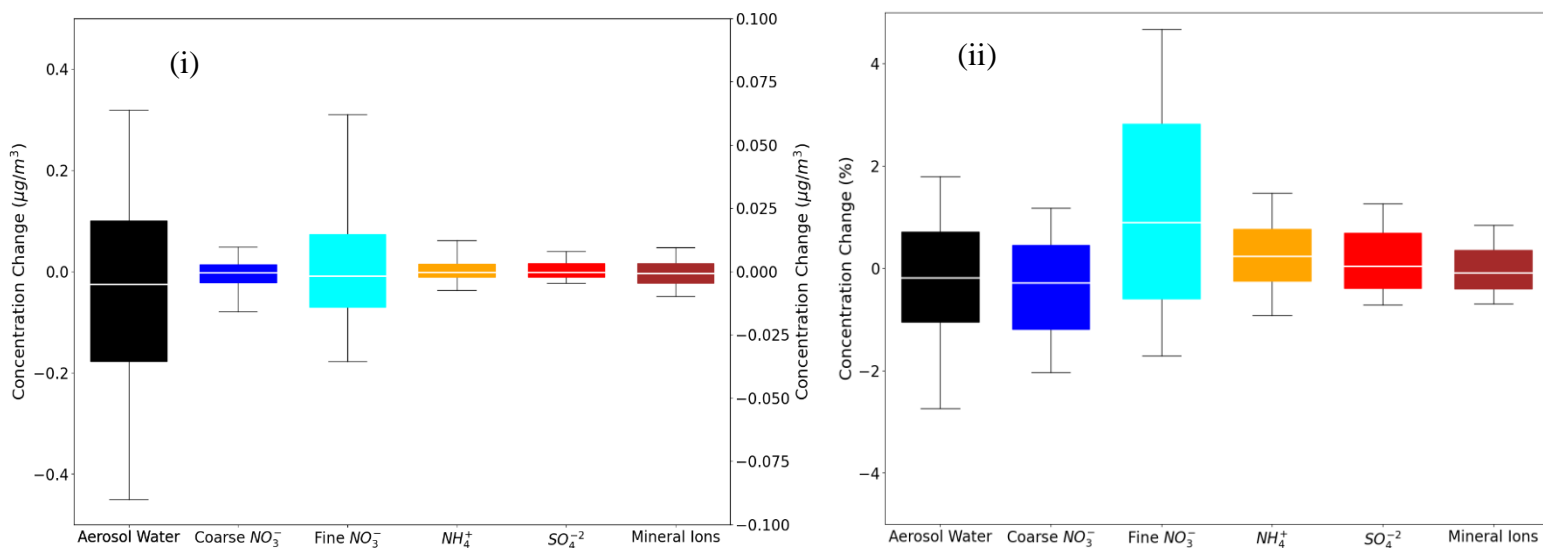
241 **3.1 Comparison of ISORROPIA II v1 against ISORROPIA II v2.3 in stable**
 242 **mode**

243 The first comparison aims to examine how ISORROPIA II v2.3 fares against ISORROPIA II v1
 244 when considering solely the stable assumption, after the latter's replacement in the newer version
 245 of the EMAC model.

246 The differences in global daily mean surface concentrations of NH_4^+ , SO_4^{2-} , mineral ions (sum
 247 of Ca^{2+} , K^+ , Mg^{+2}), aerosol water in TSP, as well as fine and coarse aerosol NO_3^- as predicted by
 248 the two versions can be seen in Figure 1. The 25th and 75th percentiles of concentration differences
 249 between the two versions for the aerosol water are below $0.2 \mu\text{g m}^{-3}$ and for the remaining species
 250 they are an order of magnitude less, which translates to differences mostly below 1 % for all
 251 species. Therefore, the predictions of inorganic aerosol composition of the two versions agree
 252 exceptionally well.

253 In order to investigate potential differences arising in specific areas, regions affected by high
 254 nitrate concentrations were selected, i.e., Europe, the Tibetan Plateau, Eastern Asia, North America
 255 and the Middle East. The differences in daily mean coarse and fine NO_3^- over these regions are
 256 shown in Figure S1. The comparison showed that the differences regarding the 25th and 75th
 257 percentiles are less than $0.05 \mu\text{g m}^{-3}$ (or less than 2.5 %) between the results of the two
 258 ISORROPIA II versions for both size modes. A statistical analysis of the results reveals that all
 259 differences between the aforementioned species are on average below 3% (Table 1). Therefore,
 260 the replacement of ISORROPIA II v1 by v2.3 in the EMAC model yields only trivial differences
 261 in the predicted aerosol ionic composition and water. The following sections focus on the
 262 comparison between the results of ISORROPIA-lite against ISORROPIA II v2.3 (called
 263 ISORROPIA II hereafter for simplicity), both in stable and metastable states.

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284 **Figure 1:** Bar chart plots depicting the 25th, 50th and 75th percentiles (box) of the i) difference and ii)
 285 fractional difference in global daily mean surface concentrations of aerosol water (left y-axis), mineral ions,
 286 NH₄⁺ and SO₄²⁻ in TSP as well as coarse and fine aerosol NO₃⁻ (right y-axis), as predicted by EMAC using
 287 ISORROPIA II v1 and ISORROPIA II v2.3. The 10th and 90th percentiles (whiskers) for each aerosol
 288 component are also shown. Both models assume that the aerosol is at its stable state at low RH and a positive
 289 change corresponds to higher concentrations by ISORROPIA II v1.

291

292 **Table 1:** Statistical analysis of EMAC-simulated mean daily surface concentrations by
 293 employing ISORROPIA II v1 versus ISORROPIA II v2.3, both in **stable mode**.
 294 Deviations are given as ISORROPIA II v1 – ISORROPIA II v2.3.

	Mean Difference ($\mu\text{g}/\text{m}^3$)	Normalized Mean Absolute Difference (%)
Coarse NO ₃ ⁻	-8×10^{-4}	1.8
Fine NO ₃ ⁻	-0.011	2.6
HNO ₃	-3.1×10^{-4}	0.7
NH ₄ ⁺	-1.6×10^{-4}	2.0
SO ₄ ²⁻	-0.009	1.2
Na ⁺	0.007	1.6
Ca ²⁺	1.7×10^{-4}	0.4
K ⁺	1.1×10^{-4}	0.4
Mg ⁺	1.5×10^{-4}	0.4
Cl ⁻	0.040	2.3
H ₂ O	0.046	1.1
H ⁺	-2.9×10^{-5}	1.5

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297 **3.2 Comparison of ISORROPIA-lite against ISORROPIA II in metastable** 298 **mode**

299 The model results using ISORROPIA-lite are compared first against those using ISORROPIA II
 300 in metastable mode in order to determine whether the ISORROPIA-lite version can produce
 301 similar results with the more detailed module in EMAC, under same conditions. Figure 2 depicts
 302 the differences of the global daily mean surface concentrations of the same species that were
 303 examined before. The comparison yields differences for the 25th and 75th percentiles that are less
 304 than $0.5 \mu\text{g m}^{-3}$ for the aerosol water and mostly less than $0.05 \mu\text{g m}^{-3}$ for the remaining inorganic

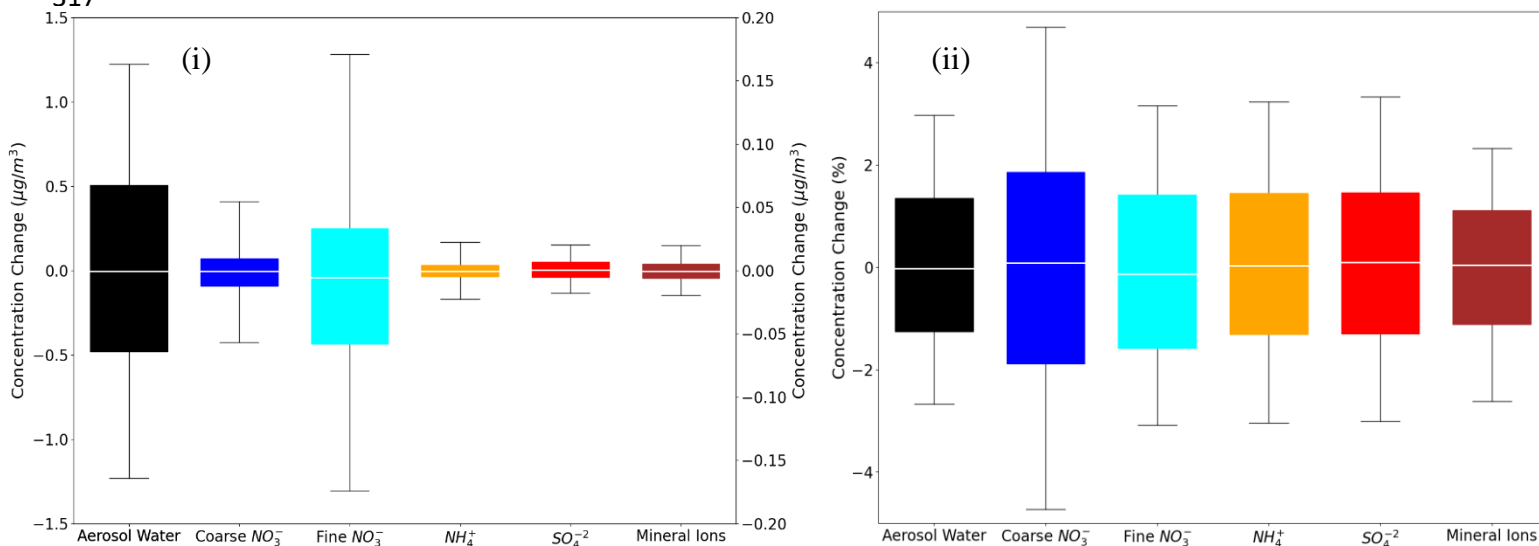
305 aerosol components, which translates into differences of less than 2% for all species most of the
306 time.

307 Figure S2 shows the comparison between predicted global daily mean coarse and fine
308 aerosol nitrate concentrations, focusing on the regions with the higher simulated mean annual
309 concentrations. Across all regions, the concentration differences for both size modes are typically
310 lower than $0.1 \mu\text{g m}^{-3}$ (or less than 3 %) and are mostly found over the Himalayan and East Asian
311 regions.

312 In Table 2, the statistics of the results for the global surface concentrations for all examined
313 aerosol components, reveal differences that are on average less than 7%. Therefore, ISORROPIA-
314 lite does provide quite similar predictions with ISORROPIA II in the EMAC model, for
315 simulations using the metastable state assumption.

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328 **Figure 2 :** Bar chart plots depicting the 25th, 50th and 75th percentiles (box) of the i) difference and ii)
329 fractional difference in global daily mean surface concentrations of aerosol water (left y-axis), mineral ions,
330 NH_4^+ and SO_4^{2-} in TSP as well as coarse and fine aerosol NO_3^- (right y-axis) , as predicted by EMAC using
331 ISORROPIA-lite and ISORROPIA II. The 10th and 90th percentiles (whiskers) for each aerosol component
332 are also shown. Both models assume that the aerosol is at its metastable state at low RH and a positive
333 change corresponds to higher concentrations by ISORROPIA-lite.

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339 **Table 2:** Statistical analysis of EMAC-simulated mean daily surface concentrations by
 340 employing ISORROPIA-lite versus ISORROPIA II, both in **metastable mode**.
 341 Bias is given as ISORROPIA-lite – ISORROPIA II.

	Mean Difference ($\mu\text{g}/\text{m}^3$)	Normalized Mean Absolute Difference (%)
Coarse NO_3^-	-6.2×10^{-4}	3.5
Fine NO_3^-	-3.1×10^{-4}	3.9
HNO_3	-2.7×10^{-4}	2.0
NH_4^+	-1.4×10^{-5}	3.8
SO_4^{2-}	2.5×10^{-3}	4.0
Na^+	0.011	6.7
Ca^{2+}	2.9×10^{-4}	1.9
K^+	1.8×10^{-4}	2.4
Mg^+	5.8×10^{-4}	3.5
Cl^-	0.017	7.0
H_2O	0.035	1.8
H^+	-8.3×10^{-4}	4.6

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343 **3.3 Evaluation of inorganic aerosol predictions**

344 EMAC predictions using both ISORROPIA-lite and ISORROPIA II in stable mode for $\text{PM}_{2.5}$
 345 ammonium, sulfate and nitrate were compared against measurements from three observational
 346 networks. The networks cover some of the most polluted areas in the Northern Hemisphere. The
 347 EPA CASTNET network (U.S. Environmental Protection Agency Clean Air Status and Trends
 348 Network) and the IMPROVE network (Interagency Monitoring of Protected Visual Environments)
 349 with 152 stations for nitrate and sulfate and 143 stations for ammonium cover the USA, with
 350 IMPROVE concerning mostly rural and/or remote areas. The EMEP network (EMEP Programme
 351 Air Pollutant Monitoring Data) includes 9 stations for nitrate and sulfate and 7 for ammonium
 352 covering the European region. Finally, the EANET network (The Acid Deposition Monitoring
 353 Network in East Asia) with 33 stations measuring all three major aerosol components covers parts
 354 of East Asia. The number of stations refers to the year 2010 which is simulated in this work.

355 Figure 3 depicts the differences between the model-predicted and the observed mean
 356 annual concentration values for SO_4^{2-} , NH_4^+ and NO_3^- aerosols, while Tables 3, 4 and 5 contain the
 357 overall statistics for the same comparisons. Here, the mean bias (MB), mean absolute gross error
 358 (MAGE), normalized mean bias (NMB), normalized mean error (NME), and the root-mean-square
 359 error (RMSE) are calculated to assess the model performance. Starting with SO_4^{2-} , the model tends

360 to underpredict the observations but with mean bias (MB) less than $-0.5 \mu\text{g m}^{-3}$ for Europe and less
361 than $-1 \mu\text{g m}^{-3}$ for USA, capturing both the higher values of the Eastern US and the lower values
362 of the Western US. Its normalized mean error (NME) ranges from 40 to 60% being highest for the
363 East Asia region, which also has the highest MB of $-1.65 \mu\text{g m}^{-3}$ (Table 3). Seasonally, the largest
364 biases are found during summertime over Europe and the USA and during wintertime over East
365 Asia (Table S4), while the same is true also for the predictions of ISORROPIA II in stable mode
366 exhibiting quite similar metrics (Table S1). NH_4^+ is much better simulated by the model over the
367 three regions, where the agreement with observations is high with MB values less than $0.4 \mu\text{g m}^{-3}$
368 but with slightly higher NME values (Table 4). Over Eastern Asia, the only important disparity is
369 a slight underprediction of about $2 \mu\text{g m}^{-3}$ around Hong Kong following the underprediction of
370 SO_4^{2-} over the same area (Fig. 3). Seasonally, spring is the worst period for the predictions of both
371 versions, while there doesn't seem to be a consistent pattern of behavior for all three regions which
372 perform best over different periods (Table S5 & S2). Finally, the model tends to overpredict NO_3^-
373 concentrations over the three regions with MB values less than $1 \mu\text{g m}^{-3}$ albeit with high NME
374 values (Table 5). Over East Asia, with the exception of Hong Kong, the model overestimates the
375 NO_3^- concentrations by about $3 \mu\text{g m}^{-3}$, especially in the Wuhan and Guangzhou areas and also
376 around Beijing (Fig. 3). In general, besides Hong Kong, the model overpredicts the concentrations
377 of all three aerosol components examined here in the East Asian region. For all regions, the best
378 seasonal agreement between the predictions of both versions in terms of MB values is found during
379 the summer period, while the worst agreement occurs around the winter/spring period (Tables S6
380 & S3). The NME values are lowest in the summer for the USA and, surprisingly, in the winter for
381 Europe and East Asia, even though this is the period with the worst MB values for these regions.
382 Potential explanations include the coarse grid resolution used in this work as well as issues related
383 to emissions (Zakoura and Pandis, 2018). It should be noted that even though the two versions
384 perform similarly, better performance on certain statistical metrics should not be taken as an
385 indication that one state assumption is more scientifically valid than the other. While a stable state
386 could be considered more accurate under very low humidity conditions (e.g., over remote deserts;
387 Karydis et al., 2016), in regions, such as those with intermediate RH and low nitrate concentration
388 (e.g., Northeastern US), particles are mostly in metastable state (Guo et al., 2016). However, the
389 two state assumptions produce very similar results in most cases, as shown here.

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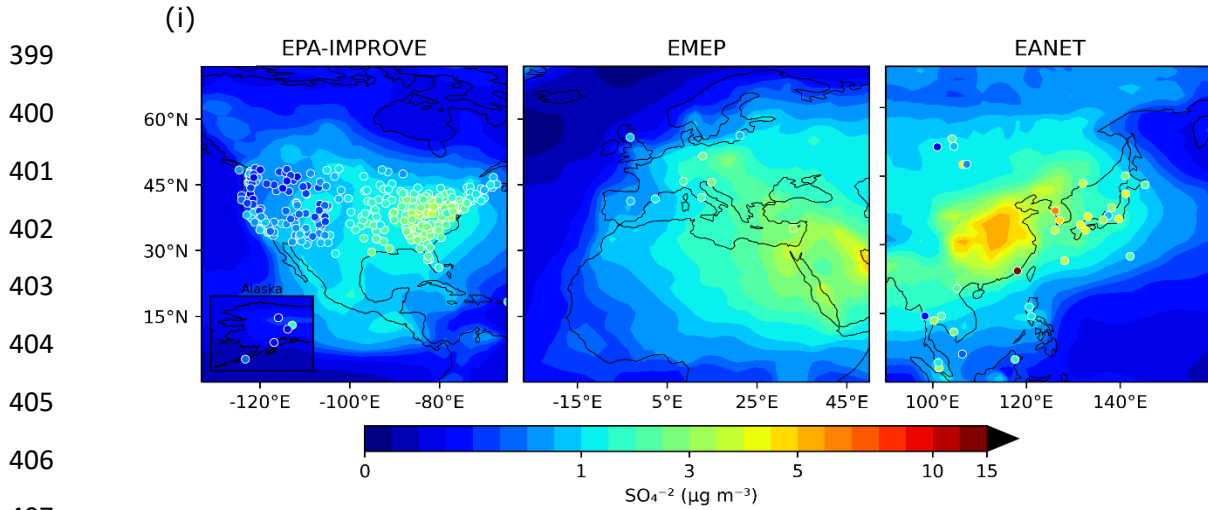
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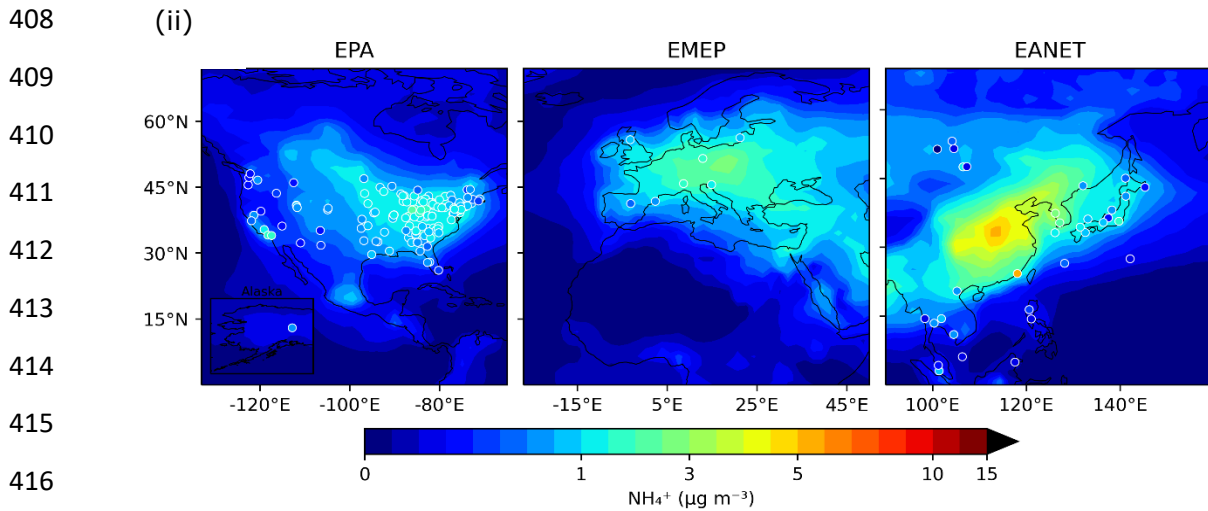
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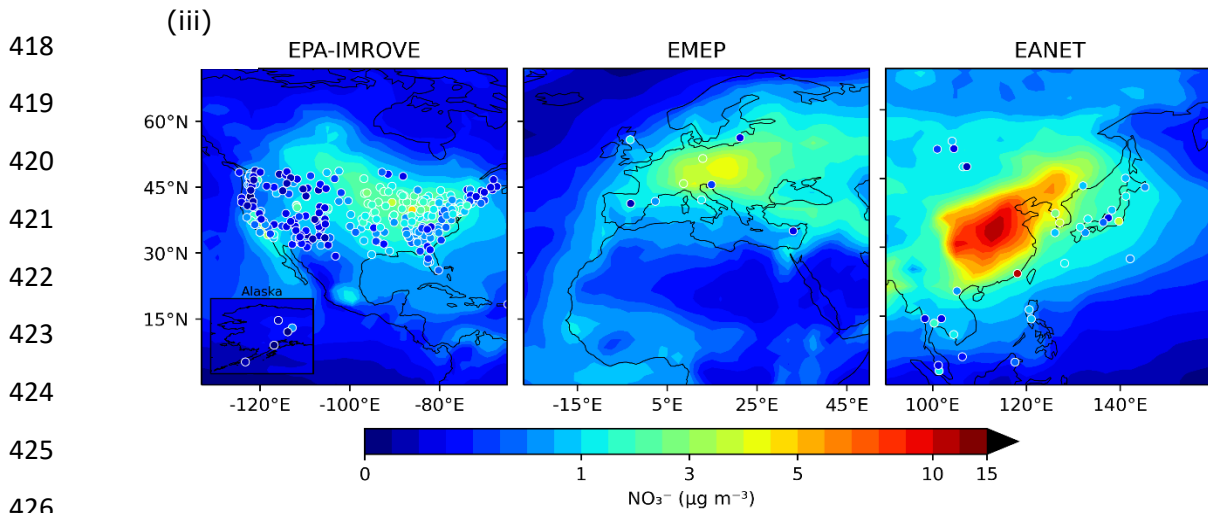
Annual mean SO_4^{2-} concentration (2010), modelled vs. observed



Annual mean NH_4^+ concentration (2010), modelled vs. observed



Annual mean NO_3^- concentration (2010), modelled vs. observed



427 **Figure 3:** Annual mean surface concentrations of $\text{PM}_{2.5}$ i) SO_4^{2-} , ii) NH_4^+ , and iii) NO_3^- as simulated by
428 EMAC using ISORROPIA-lite (shaded contours) versus observations of the same species from the
429 IMPROVE, EMEP and EANET networks (colored circles).

430

431 **Table 3:** Statistical evaluation of EMAC predicted surface concentrations of PM_{2.5} SO₄²⁻ using
 432 ISORROPIA-lite against observations during 2010.

433

Network	Number of datasets	Mean Observed (µg m ⁻³)	Mean Predicted (µg m ⁻³)	MAGE (µg m ⁻³)	MB (µg m ⁻³)	NME (%)	NMB (%)	RMSE (µg m ⁻³)
EPA	1791	2.18	1.28	0.92	-0.90	42	-38	0.93
IMPROVE	1526	1.02	0.92	0.47	-0.10	46	-11	0.73
EMEP	108	1.71	1.27	0.75	-0.44	44	-26	0.91
EANET	353	3.19	1.54	1.95	-1.65	61	-51	2.46

434

435 **Table 4:** Statistical evaluation of EMAC predicted surface concentrations of PM_{2.5} NH₄⁺ using
 436 ISORROPIA-lite against observations during 2010.

Network	Number of datasets	Mean Observed (µg m ⁻³)	Mean Predicted (µg m ⁻³)	MAGE (µg m ⁻³)	MB (µg m ⁻³)	NME (%)	NMB (%)	RMSE (µg m ⁻³)
EPA	1660	1.01	1.01	0.50	0.00	49	0	0.72
IMPROVE	-	-	-	-	-	-	-	-
EMEP	84	1.08	1.44	0.63	0.36	59	34	0.75
EANET	360	0.93	1.25	0.69	0.32	74	34	1.25

437

438 **Table 5:** Statistical evaluation of EMAC predicted surface concentrations of PM_{2.5} NO₃⁻ using
 439 ISORROPIA-lite against observations during 2010.

Network	Number of datasets	Mean Observed (µg m ⁻³)	Mean Predicted (µg m ⁻³)	MAGE (µg m ⁻³)	MB (µg m ⁻³)	NME (%)	NMB (%)	RMSE (µg m ⁻³)
EPA	1762	1.39	1.87	1.06	0.48	76	42	1.65
IMPROVE	1526	0.42	1.18	0.82	0.76	194	175	1.15
EMEP	108	1.15	1.91	1.25	0.76	109	66	1.66
EANET	372	1.32	2.27	1.33	0.95	101	72	2.17

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444 3.4 Computational speed-up metrics

445 The computational efficiency and speed-up that ISORROPIA-lite provides compared to
 446 ISORROPIA II in both stable and metastable modes were quantified. Table 6 contains the total
 447 number of time steps that the EMAC model performed for the same simulation period (i.e., 24 h
 448 of CPU-time using 16 nodes) as well as the real time that was needed per individual time step, for
 449 each ISORROPIA version. The metrics shown in Table 6 concern the average value of each
 450 quantity, along with the corresponding standard deviation, resulting from a total of 18 simulations
 451 (6 for each version). From the difference in the real time required by the model to execute each
 452 individual time step, the speed-up of ISORROPIA-lite was found to be just above 3% compared
 453 to ISORROPIA II in metastable mode and almost 5 % compared to ISORROPIA II in stable mode.
 454 These values are, as expected, lower than the improvement in the computational efficiency that
 455 the ISORROPIA-lite version provides compared to the original version, as found in the offline
 456 evaluation, because EMAC contains several other modules that are quite computationally
 457 expensive. For example, the gas-phase chemistry (MECCA submodel) as well as wet deposition
 458 and liquid phase chemistry (SCAV submodel) are responsible for two thirds of the total
 459 computational cost of the global model. As a comparison, the offline speed-up that ISORROPIA-
 460 lite provided was calculated to be 35% and when utilized in the regional model PMCAMx 3D it
 461 was found to be 10% (Kakavas et al., 2022).

462 **Table 6:** Total number of time steps that EMAC executed in 24 hours of running time and
 463 number of seconds needed for each time step, utilizing ISORROPIA-lite and ISORROPIA II
 464 (both in Stable & Metastable). The computational speed-up refers to how much quicker (in %)
 465 the process is executed by ISORROPIA-lite in comparison to the previous version in both
 466 modes.

Simulation	# Time Steps	# Seconds per Timestep	Computational Speed-Up (%)
ISORROPIA-lite	78,193 ± 116	1.10 ± 0.002	-
ISORROPIA II v2.3 (Metastable)	75,720 ± 242	1.14 ± 0.003	3.3 ± 0.3
ISORROPIA II v2.3 (Stable)	74,599 ± 169	1.16 ± 0.003	4.8 ± 0.3

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470 **4. Comparison of ISORROPIA-lite Against ISORROPIA II in Stable** 471 **Mode**

472 In this section, we present a comparison of the ISORROPIA-lite results in metastable mode against
473 those of the ISORROPIA II results in stable mode. Both versions are now available in the latest
474 version of the EMAC model, and the user has the option to utilize either one. While ISORROPIA-
475 lite always assumes metastable aerosols, ISORROPIA II assumes stable aerosols by default. This
476 comparison is done in an attempt to quantify the effects of using the metastable case in global
477 atmospheric simulations, and to identify the regions and conditions under which the two
478 assumptions have any significant differences. Some discrepancies are expected due to the different
479 physical state of aerosols at low RH, however, the choice between a stable state and a metastable
480 state should not be considered obvious. For example, Fountoukis et al. (2009) and Karydis et al.
481 (2010) have shown that the stable assumption is in better agreement with observations under
482 conditions where RH is consistently below 50%. On the other hand, Ansari and Pandis (2000)
483 emphasize that the metastable assumption must be considered for regions characterized by
484 intermediate RH and low pollutant concentrations (in this case of NO_3^-), while there are no
485 significant differences between the two assumptions over regions with high concentrations. Here,
486 differences in the calculated aerosol acidity by the two modules are also investigated.

487 **4.1 Spatial variability of mean annual aerosol concentrations**

488 For sulfate in TSP the predicted maximum annual average concentration was $7 \mu\text{g m}^{-3}$ found over
489 East Asia highlighting the large anthropogenic impact over that region, while it was also high ($>$
490 $5 \mu\text{g m}^{-3}$) in India, Europe, and the Middle East in both simulations (Fig. 4i). Absolute differences
491 for sulfate in TSP were lower than $0.15 \mu\text{g m}^{-3}$ ($< 3\%$) and found mainly over the polluted northern
492 hemisphere (mainly East USA & Europe) with slightly higher values simulated by ISORROPIA
493 II (Fig. 4ii). This is most likely related to the also higher NO_3^- aerosol predictions by ISORROPIA-
494 lite over the same regions (see below & Fig. 4viii). The higher SO_4^{2-} aerosol concentrations
495 estimated by ISORROPIA II over the Middle East region are mainly due to changes in wet
496 deposition induced by the different physical state of the aerosol due to the higher water content by
497 ISORROPIA-lite. The simulated concentrations of NH_4^+ in TSP had maximum annual average
498 values of $6 \mu\text{g m}^{-3}$ and were found mainly over East Asia, especially around the greater Beijing
499 and Wuhan areas, while India and Europe also exhibited high mean annual values for TSP NH_4^+
500 ($> 3 \mu\text{g m}^{-3}$) (Fig. 4iii). The absolute differences for NH_4^+ in TSP between the two model versions
501 are higher over the Himalayan and East Asian regions (in favor of ISORROPIA II) but apparently
502 weaker over USA, the Middle East and Africa (ISORROPIA-lite predicts higher values), although
503 never higher than $0.5 \mu\text{g m}^{-3}$ ($< 5\%$) (Fig. 4iv). Regarding aerosol NO_3^- concentrations in the coarse
504 mode the maximum annual average of $6 \mu\text{g m}^{-3}$ was predicted at the Arabian Peninsula (Fig. 4v),
505 while in the fine mode the maximum annual average value of $11 \mu\text{g m}^{-3}$ was predicted over the
506 metropolitan areas of Wuhan and Guangzhou (Fig. 4vii). Other high annual average concentrations
507 of fine aerosol NO_3^- are found in the Tibetan Plateau and most prominently in heavy industrial
508 regions such as East US, Eastern Asia and Europe (exceeding $4 \mu\text{g m}^{-3}$ in most of these areas) with
509 the latter two regions contributing high annual average concentrations in the coarse mode as well.

510 The absolute differences for coarse NO_3^- were similar in magnitude to those of NH_4^+ in TSP with
511 the Middle East yielding higher values by ISORROPIA-lite while the opposite is true for Europe
512 and East USA (Fig. 4vi). The absolute differences for fine NO_3^- are higher than those for coarse
513 NO_3^- reaching up to $1.75 \mu\text{g m}^{-3}$ mainly over the Tibetan Plateau ($\sim 30\%$) with ISORROPIA II
514 predicting the higher values (Fig. 4vii). Higher nitrate concentrations were also predicted by
515 ISORROPIA II mainly close to the West coast of South America and North of Atacama Desert.
516 Around those regions as well as the Tibetan Plateau, the relative humidity is often below 50% and
517 30% respectively (see Fig. 8) and the metastable assumption results in lower nitrate concentrations,
518 in agreement with the findings of Ansari and Pandis (2000). At the same time, ISORROPIA II
519 predicts a higher aerosol fraction for NO_3^- (up to 10%) for the West coast of South America and
520 the Tibetan Plateau. This is not the case for East Asia (Fig. 5ii) although the low sulfate/nitrate
521 ratio of that region, results to an excess of available NH_3 to react with HNO_3 and form ammonium
522 nitrate that would justify the higher fine mode nitrate concentrations by the stable case of
523 ISORROPIA II (Ansari and Pandis, 2000). A higher NO_3^- aerosol fraction (up to 10%) in the
524 Middle East was exhibited by ISORROPIA-lite (Fig. 5ii). This area is characterized by increased
525 mineral ion concentrations and high sulfate to nitrate ratios (Karydis et al., 2016) which led to
526 higher coarse mode nitrate predictions by the metastable case (Ansari and Pandis, 2000), although
527 the maximum difference was only $0.6 \mu\text{g m}^{-3}$ (Fig. 4vi, 4viii). The differences in coarse and fine
528 NO_3^- among the two versions did not display any strong seasonality as they were only slightly
529 higher during autumn (for East Asia) and winter (for India-Himalayas) (not shown). A comparison
530 of the simulated aerosol concentrations at higher altitudes can be found in Figure S3, where the
531 zonal mean annual average concentrations as well as their absolute differences between the two
532 model versions are depicted. The deviations between the results of the two ISORROPIA versions
533 are becoming smaller as the air masses move higher in the atmosphere, until they are practically
534 identical at altitudes above 700hPa. Regarding the behavior of the mineral ions of Ca^{2+} , K^+ , and
535 Mg^{2+} the majority of high concentrations are found around the largest desert regions of the Sahara,
536 Gobi, Atacama and Namib deserts (Figure S4), with Ca^{2+} being evidently the most dominant across
537 all minerals. Furthermore, the absolute difference maps (Fig. S4) show minimal differences in
538 mean annual surface concentrations (mostly less than $0.5 \mu\text{g m}^{-3}$) between the simulations from
539 the two model versions. This is also reflected in the comparison of zonal mean annual average
540 concentrations and their differences, as shown in Figure S5.

541 In the heavily polluted regions (particularly East USA, Europe and East Asia), the
542 particulate NO_3^- dominates compared to the gas phase HNO_3 (Fig. 5i). The fine-mode fraction of
543 the particulate nitrate burden is bigger than the coarse-mode fraction over Eastern Asia, India,
544 Europe, and Eastern USA, while in the large desert areas of the Middle East and the Sahara most
545 of the particulate NO_3^- exists in the coarse mode (Fig. 5iii). The aerosol water fraction is low
546 ($<30\%$) across the most arid regions of Sahara, Atacama, Namib and Gobi, while Europe has the
547 highest continental average aerosol water content in the Northern Hemisphere polluted regions
548 (Fig. 5v). ISORROPIA-lite predicts higher average aerosol water concentration globally since the
549 particles cannot form solids, and the salts remain in a supersaturated metastable solution (Fig. 5vi).

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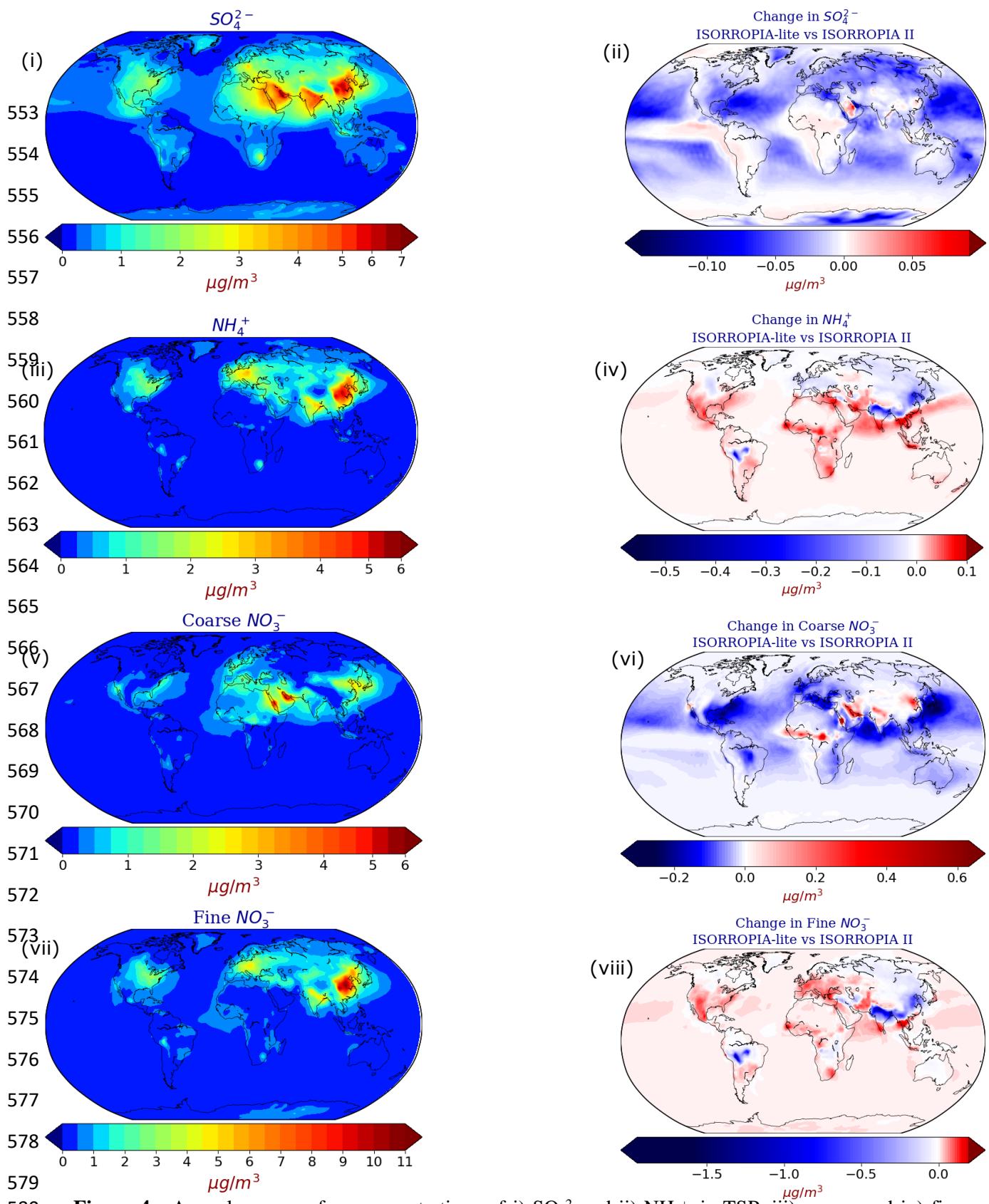
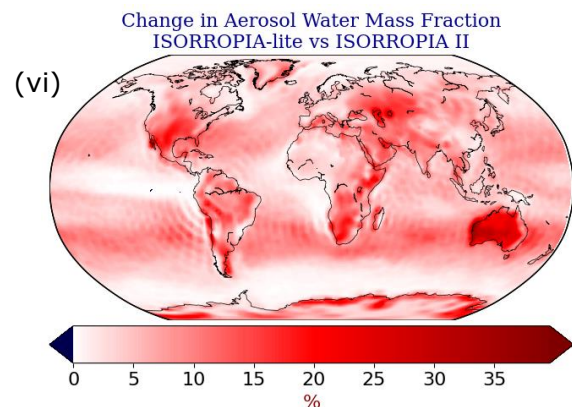
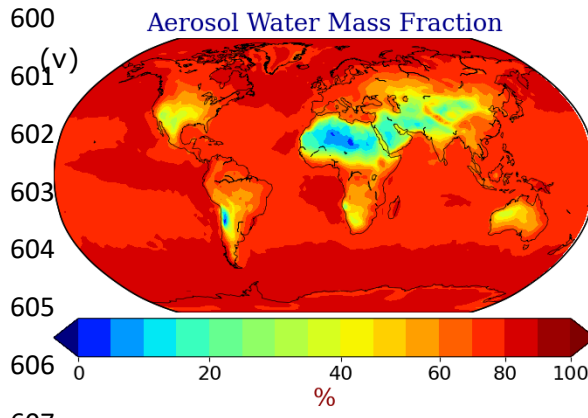
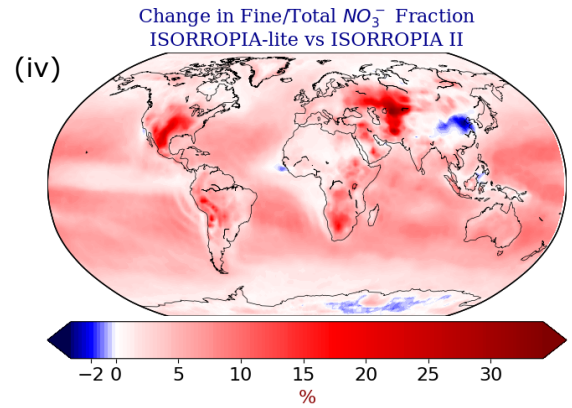
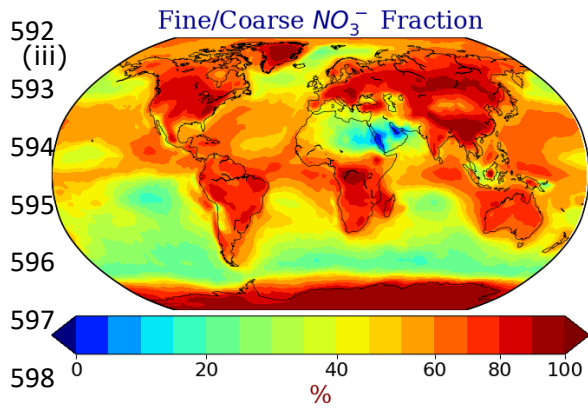
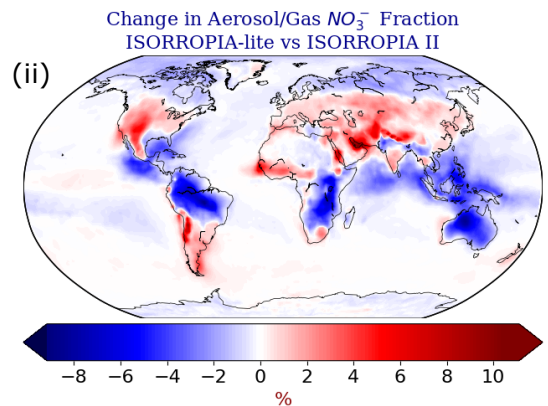
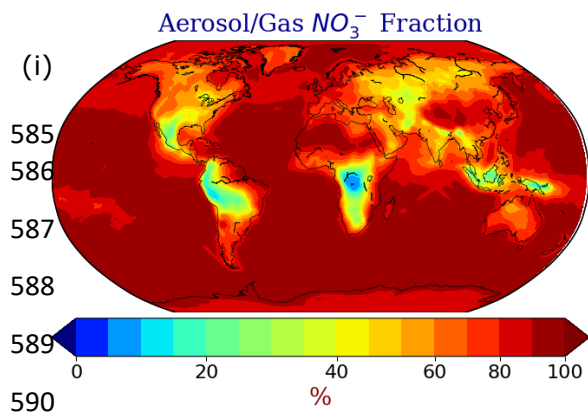


Figure 4 : Annual mean surface concentrations of i) SO_4^{2-} and ii) NH_4^+ in TSP, iii) coarse and iv) fine aerosol NO_3^- as predicted by EMAC using ISORROPIA-lite. Change of the annual mean EMAC-simulated surface concentration of v) NH_4^+ and vi) SO_4^{2-} in TSP, vii) coarse and viii) fine aerosol NO_3^- after employing ISORROPIA II. Positive values in red indicate higher concentrations by ISORROPIA-lite. The models assume different aerosol states.

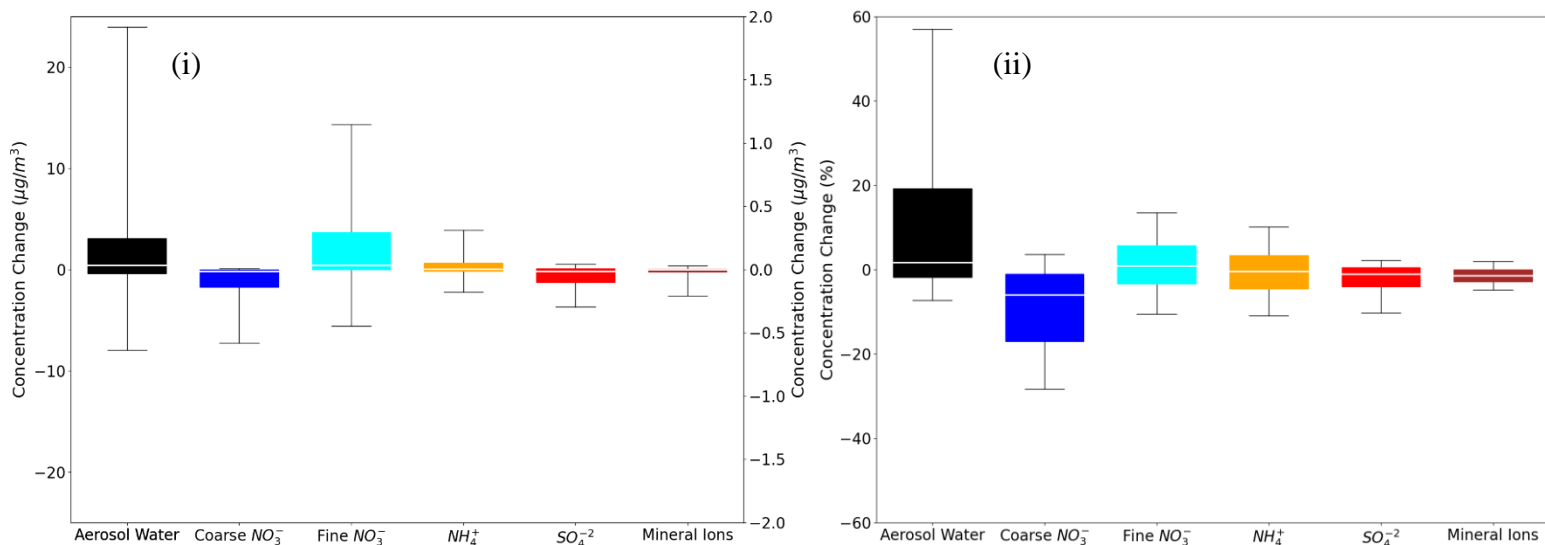


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609 **Figure 5 :** Annual mean surface fractions of i) aerosol/gas NO_3^- , ii) fine/total-aerosol NO_3^- and iii) aerosol
610 water mass as calculated by EMAC using ISORROPIA-lite. Change of the annual mean EMAC-simulated
611 surface fractions of aerosol/gas iv) NO_3^- , v) fine/total-aerosol NO_3^- , and vi) aerosol water mass after
612 employing ISORROPIA II. Positive values in red indicate higher fractions by ISORROPIA-lite. The models
613 assume different aerosol states.

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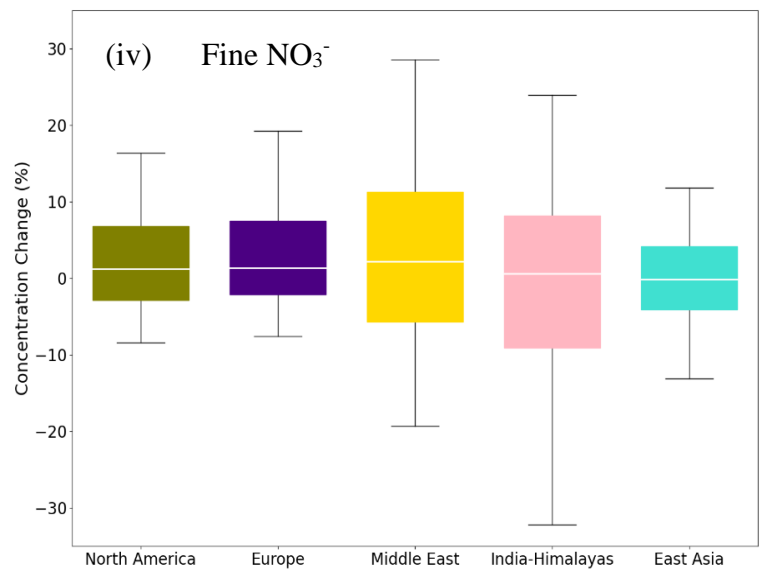
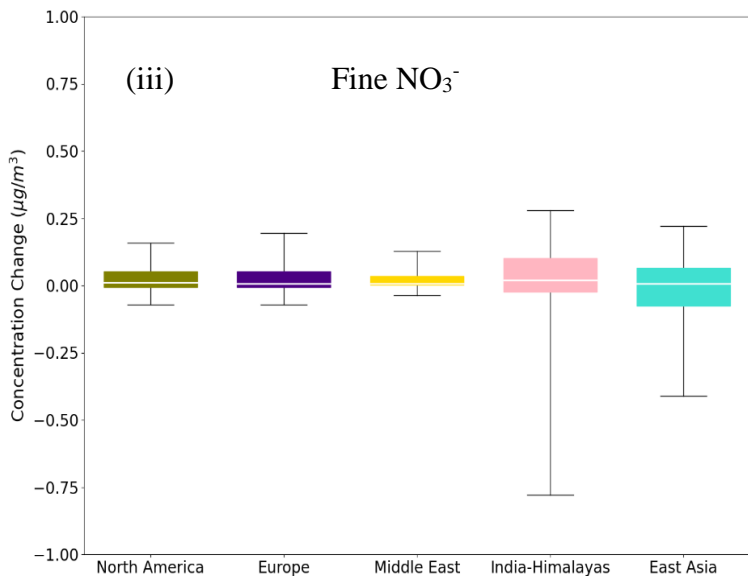
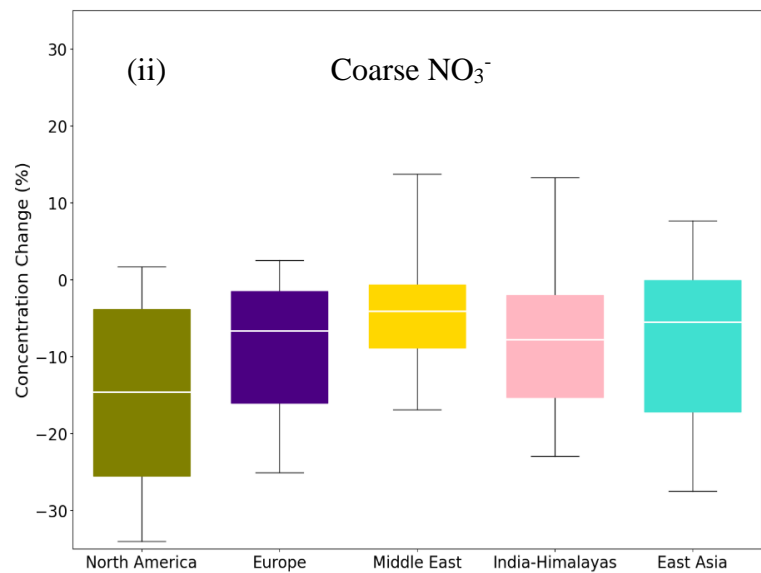
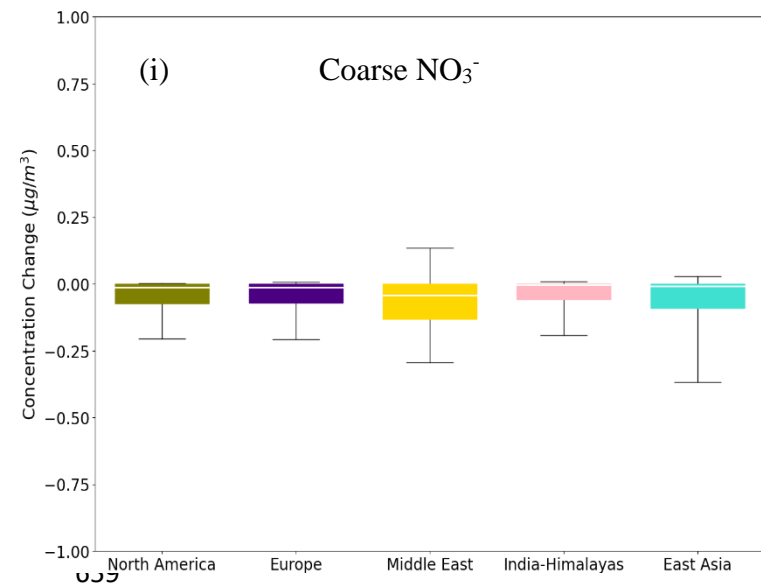
618 The absolute differences in global daily mean concentrations are mostly less than $0.3 \mu\text{g m}^{-3}$
 619 m^{-3} for all species (NH_4^+ , SO_4^{2-} and Mineral Cations in TSP as well as coarse and fine aerosol NO_3^-
 620) except aerosol water in TSP (Figure 6). In that case the absolute differences for the 25th and 75th
 621 percentiles are less than $5 \mu\text{g m}^{-3}$. This translates to fractional differences for the 25th and 75th
 622 percentiles mostly below 20 % for aerosol water in TSP and coarse NO_3^- aerosol, and mostly
 623 below 5% for all the remaining species.

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635
 636 **Figure 6 :** Bar chart plots depicting the 25th, 50th and 75th percentiles (box) of the i) difference and ii)
 637 fractional difference in global daily mean surface concentrations of aerosol water (left y-axis), mineral ions,
 638 NH_4^+ and SO_4^{2-} in TSP as well as coarse and fine aerosol NO_3^- (right y-axis) , as predicted by EMAC using
 639 ISORROPIA-lite and ISORROPIA II. The models assume different aerosol states at low RH and a positive
 640 change corresponds to higher concentrations by ISORROPIA-lite.

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674 **Figure 7:** Bar chart plots depicting the 25th, 50th and 75th percentiles (box) of the difference in the global
 675 daily mean surface concentrations of i) coarse and ii) fine aerosol NO_3^- for the regions of North America,
 676 Europe, Middle East, India-Himalayas and East Asia, as predicted by EMAC using ISORROPIA-lite and
 677 ISORROPIA II. The fractional differences in global daily mean surface concentrations of iii) coarse and
 678 iv) fine aerosol NO_3^- for the same regions are also shown. The models assume different aerosol states at
 679 low RH and a positive change corresponds to higher concentrations by ISORROPIA-lite.

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681 The model results in the regions with the highest mean annual loads of fine and coarse aerosol
 682 NO_3^- concentrations (see Section 3.1) as well as the most significant differences in estimated
 683 aerosol water and aerosol acidity (see Section 4.3), were further analyzed to determine whether
 684 the phase state assumption has a large effect on simulated aerosol nitrate formation (Figure 7). For
 685 both coarse and fine daily mean NO_3^- concentrations, Europe and North America are clearly the
 686 regions with the smallest differences between the two versions. On the other hand, East Asia and
 687 especially the India-Himalayas regions are areas where the differences are the highest with
 688 ISORROPIA II predicting higher fine aerosol NO_3^- concentrations while in the Middle East,
 689 ISORROPIA-lite is predicting higher coarse mode aerosol NO_3^- concentrations. However, even
 690 for these areas the differences are typically below $0.25 \mu\text{g m}^{-3}$ (25th and 75th percentiles) with the

691 higher differences not exceeding $0.8 \mu\text{g m}^{-3}$ (10th and 90th percentiles). This translates to fractional
692 differences below 25 % (25th and 75th percentiles) for all regions, reaching up to 30 % (10th and
693 90th percentiles) mainly in the Tibetan Plateau and the Middle East.

694 Table 7 contains the statistics for the comparisons of the global daily average surface
695 concentrations calculated by the two simulations. While all the aerosol component concentrations,
696 except for aerosol water, are higher for ISORROPIA II, the differences are still quite low.
697 Furthermore, despite the different aerosol phase state assumption by the two versions, the
698 normalized mean absolute difference remains low for all species (on average < 11 %) except
699 HNO₃. The overall statistics support the conclusion that on the global scale, the phase state
700 assumption for low RH does not have a significant impact on the predicted tropospheric aerosol
701 load. More specifically, ISORROPIA-lite produces a slightly higher tropospheric burden for
702 aerosol NO₃⁻ than ISORROPIA II (0.875 Tg versus 0.861 Tg, respectively) while the opposite was
703 the case for HNO₃ (0.921 Tg versus 0.935 Tg). The higher burden of ISORROPIA-lite is due to
704 the fact that the higher aerosol water content favors the partitioning of HNO₃ to the particulate
705 phase.

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723 **Table 7:** Statistical analysis of EMAC-simulated mean daily surface concentrations by
 724 employing ISORROPIA-lite in **metastable mode** versus ISORROPIA II in **stable mode**.
 725 Bias is given as ISORROPIA-lite – ISORROPIA II.

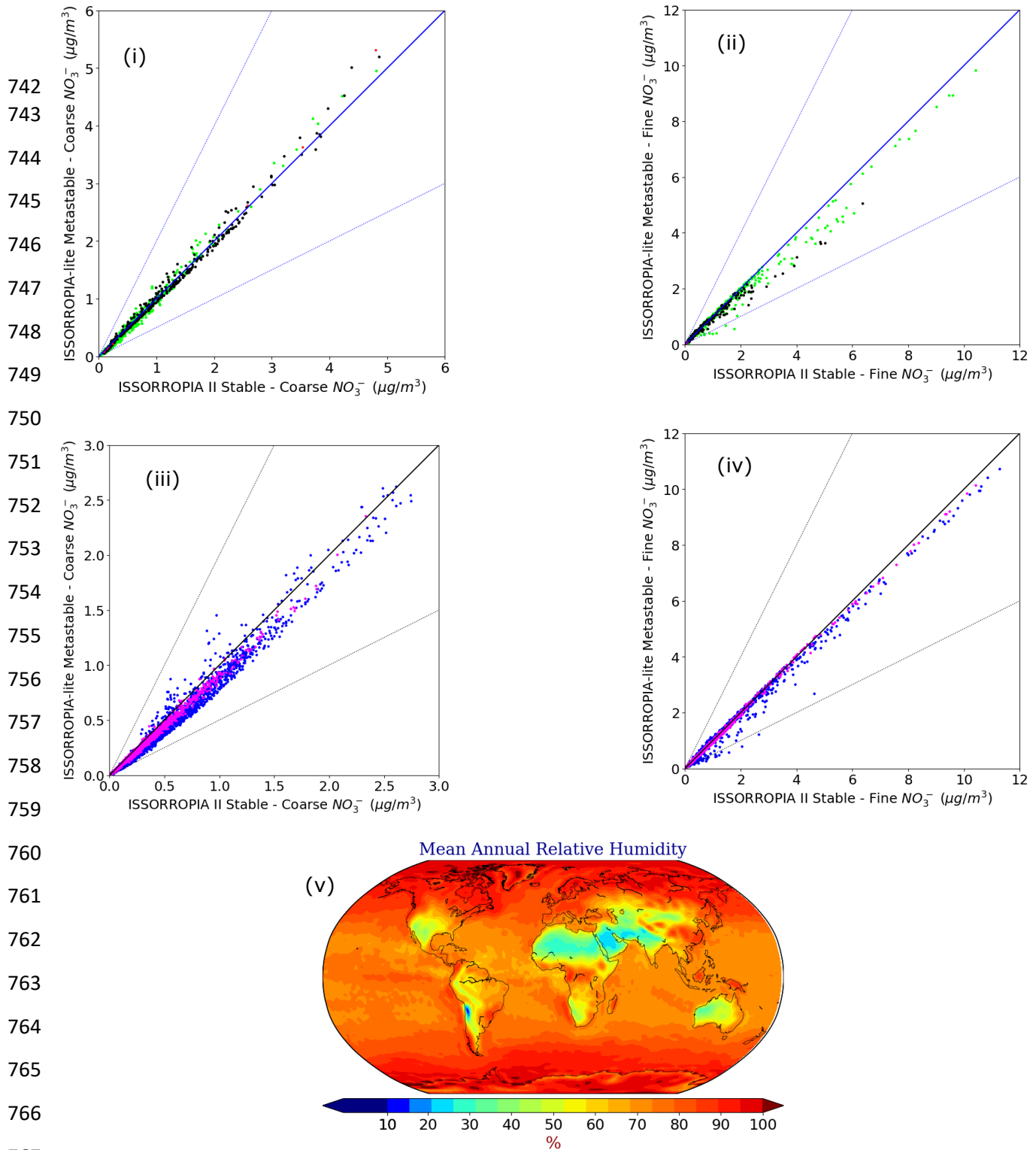
	Mean Difference ($\mu\text{g}/\text{m}^3$)	Normalized Mean Absolute Difference (%)
Coarse NO_3^-	-0.026	9.1
Fine NO_3^-	-0.044	9.8
HNO_3	-0.002	10.3
NH_4^+	-1.8×10^{-4}	8.0
SO_4^{2-}	-0.020	4.8
Na^+	-0.081	8.6
Ca^{2+}	-0.005	1.7
K^+	-0.002	1.8
Mg^+	-0.002	1.7
Cl^-	-0.120	9.4
H_2O	2.717	10.8
H^+	-4.7×10^{-4}	6.1
pH Accumulation	-0.06 (pH)	2.3
pH Coarse	0.03 (pH)	2.3

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727 **4.2 Relative humidity dependent behavior of NO_3^- aerosols**

728 The dependence of the differences in nitrate predictions on relative humidity was examined both
 729 for fine and coarse particles (Figure 8). The differences between ISORROPIA II and ISORROPIA-
 730 lite are higher at intermediate RH ranging from 20% to 60% being more evident in the fine mode
 731 aerosol NO_3^- and for high annual mean concentrations of coarse mode aerosol NO_3^- ($> 4 \mu\text{g m}^{-3}$).
 732 In this RH range, solid salts can precipitate when the stable equilibrium state is assumed (Seinfeld
 733 and Pandis, 2016), while in the metastable state all these salts remain dissolved in water. A region
 734 that has often RH in the 20 - 60% range is the Tibetan Plateau which leads to discrepancies of the
 735 fine mode particulate nitrate predictions of the two models in this area, while higher coarse mode
 736 particulate nitrate concentrations are predicted by ISORROPIA-lite in the Middle East, an area
 737 that is also often characterized by intermediate RH. The differences found for coarse mode
 738 particulate nitrate in the higher RH ranges of 60 – 100 %, can account for the respective differences
 739 that occurred in areas characterized by such RH values (East USA, Europe and East Asia) but
 740 concern lower annual mean concentration values ($< 3 \mu\text{g m}^{-3}$).

741



768 **Figure 8:** Scatterplots comparing the annual mean surface concentrations of coarse (i, iii) and fine aerosol
 769 NO_3^- (ii, iv) for relative humidity ranges of 20-60 % (i, ii) and 60-100 % (iii, iv) as predicted by EMAC
 770 using ISSORROPIA-lite versus ISSORROPIA II. The models assume different aerosol states at low RH.
 771 Black points represent the 20-40 % RH range, green points the 40-60 % range, blue points the 60-80 %
 772 range and pink points the 80-100 % range. v) Mean annual relative humidity as calculated by EMAC
 773 using ISSORROPIA-lite.

774 4.3 Comparison of the estimated aerosol acidity

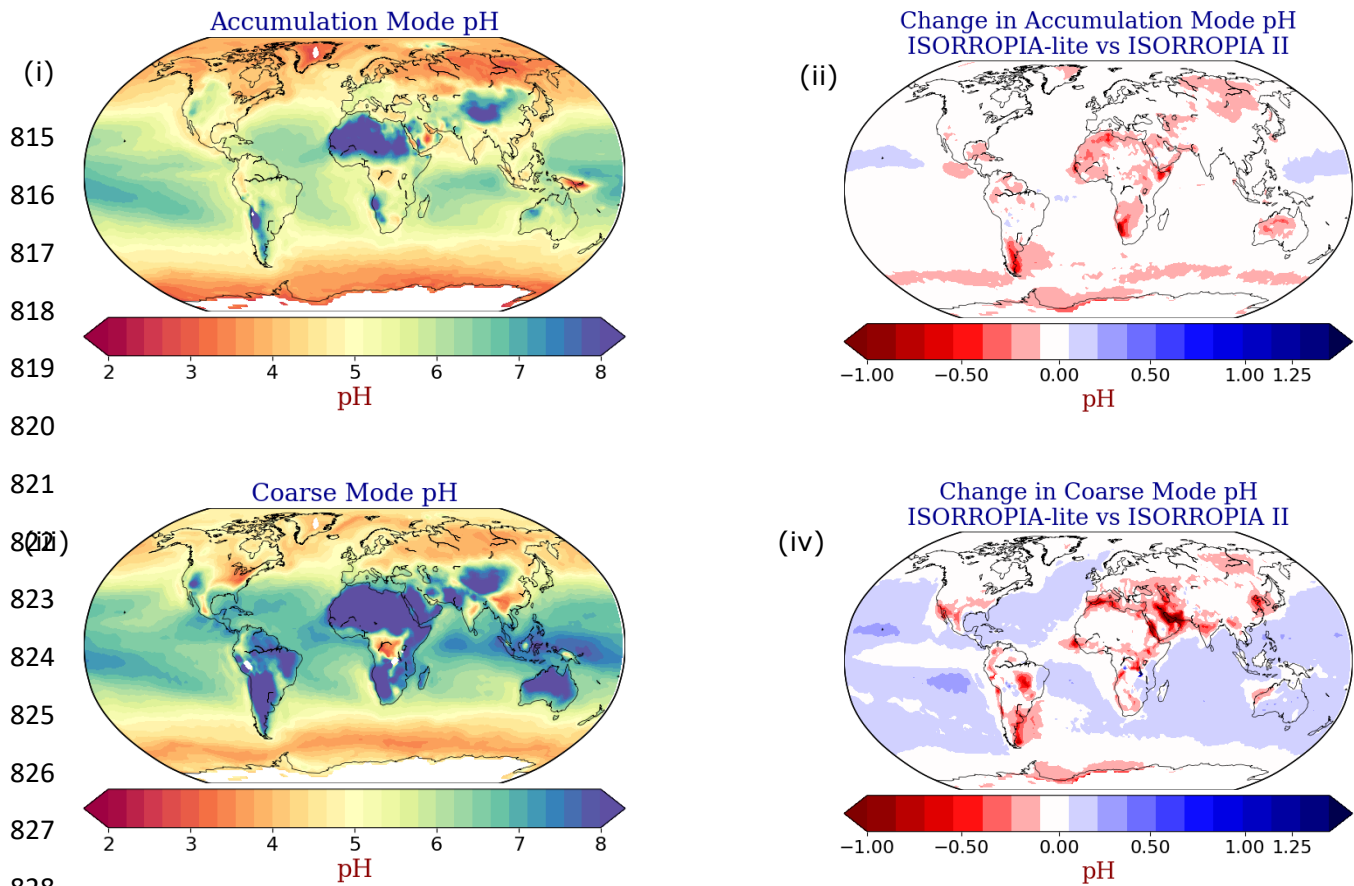
775 The estimated aerosol acidity by the two model versions was compared separately for the
776 accumulation and coarse size modes. This comparison aims at verifying the credibility of the
777 estimated inorganic aerosol acidity of ISORROPIA-lite, as the first results of its implementation
778 in the EMAC model are presented here. Since this capability is well established for ISORROPIA
779 II (Karydis et al., 2021), it is of interest to examine any potential, but otherwise expected,
780 differences between the two versions. The pH was computed for the fine and coarse particles, by:
781

$$782 \quad pH = -\log_{10}\left(\frac{[H^+]}{[H_2O]}\right) \quad (1)$$

783 The calculations were performed neglecting the water associated with the organic fraction of
784 aerosols, as they are handled by other parts of the aerosol microphysics submodel GMXe. The
785 average pH was calculated based on the instantaneous H^+ and H_2O values estimated every 5 hours.
786 This is because utilizing daily average values for H^+ and H_2O can result in a low-biased predicted
787 pH of ~ 2 units globally (Karydis et al., 2021). The 5 hour interval provides a frequent output of
788 values at different times of the day to account for the diurnal variability of pH, since a selection of
789 6 or 8 hour intervals would result in instantaneous H^+ and H_2O values at identical times on different
790 days. pH calculations are performed only in cases where there is enough water in the aerosol
791 (instantaneous values exceeding $0.05 \mu\text{g m}^{-3}$).

792 ISORROPIA-lite predicts slightly more acidic particles mainly in the coarse mode (Fig.
793 9iv). The most significant differences (up to 1 unit) in that size range are located over the Middle
794 East and Arabian Peninsula, while smaller differences can be found in limited parts of the
795 Himalayan and the East Asian regions as well as West USA and the Amazon Basin. These regions
796 are characterized by high mineral cation concentrations and/or low RH. Therefore, the stable state
797 results in increased pH values due to the precipitation of insoluble salts out of the aqueous phase.
798 On the other hand, in the metastable state all anions remain in the aqueous phase lowering the
799 particle pH. Differences in accumulation mode particle acidity are not as high (Fig. 9ii).
800 ISORROPIA-lite predicts that accumulation mode particles over heavy industrialized regions such
801 as Southeast Asia, Europe and Eastern USA are moderately acidic with mean pH values in the
802 range of 4 - 5 while exhibiting alkaline behavior in desert areas where the increased levels of
803 mineral ions elevate the pH above 7 (Fig. 9i). Coarse mode particles are in general more alkaline
804 than those in the accumulation mode, with a few exceptions over east US, central Europe, north
805 India and SE Asia (Fig. 9iii). These regions are characterized by high NH_3 concentrations from
806 agricultural activities.

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829 **Figure 9:** Annual mean EMAC-simulated i) accumulation and ii) coarse mode aerosol pH using
 830 ISORROPIA-lite. Change of the annual mean EMAC-simulated iii) accumulation and iv) coarse mode
 831 aerosol pH after using ISORROPIA II, with negative values in red indicating lower pH by ISORROPIA-
 832 lite. The models assume different aerosol states.

833
834

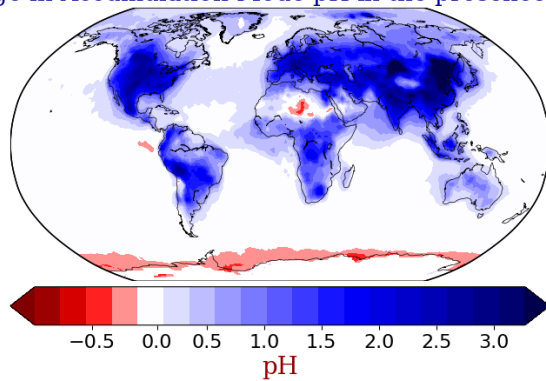
835 A sensitivity test was performed by reducing all NH_3 emissions by half to investigate if
 836 there is a buffering mechanism that controls the pH of the accumulation mode particles more than
 837 in the coarse mode. Figure 10 shows the difference of the mean annual calculated aerosol pH
 838 between the base case (NH_3 emissions present) and the sensitivity case (half NH_3 emissions).
 839 When NH_3 emissions are switched off, the pH of fine PM decreases by up to 3 units and the
 840 particles become a lot more acidic (Fig. 10i). For the coarse mode this effect is not that strong (pH
 841 reduction of up to 1.5 units) (Fig. 10ii). As expected, this buffering mechanism is mainly observed
 842 across the aforementioned regions where NH_3 concentrations are high, but also over areas affected
 843 by natural NH_3 emissions. This is consistent with the results of Karydis et al. (2021) who found
 844 that in the absence of NH_3 , aerosol particles would be extremely acidic in most of the world.

845 The differences in the accumulation mode pH calculated by ISORROPIA-lite and
 846 ISORROPIA II are extremely small (i.e., mean difference of 0.06 pH units or 2.3%), and even
 847 smaller for coarse mode pH (Table 7), indicating an overall good agreement between the two
 848 model versions.

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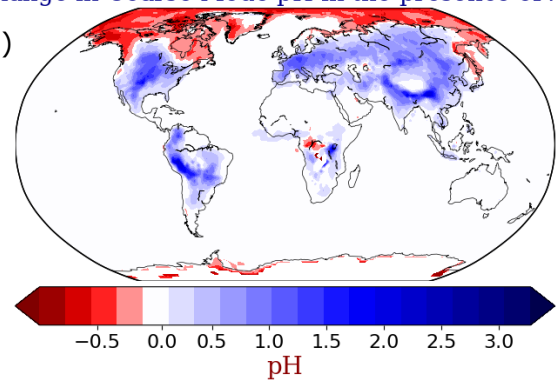
Change in Accumulation Mode pH in the presence of NH_3

(i)



Change in Coarse Mode pH in the presence of NH_3

(ii)



857

858 **Figure 10:** Absolute change of the annual mean EMAC-simulated i) accumulation and ii) coarse mode
859 aerosol pH using ISORROPIA-lite after reducing the NH_3 emissions by half. Positive values in blue
860 indicate higher aerosol pH when NH_3 is present.

861

862 5. Conclusions

863 This study presents the first results of the implementation of the ISORROPIA-lite thermodynamic
864 module in the EMAC global chemistry and climate model, and is compared to the previous version,
865 ISORROPIA II v2.3, after the latter has successfully replaced ISORROPIA II v1 to improve pH
866 predictions close to neutral conditions.

867 The results of ISORROPIA II versions 1 and 2.3 both in stable mode, had insignificant
868 differences (<3%) concerning the global predictions of NH_4^+ , SO_4^{2-} , mineral ions and aerosol water
869 in TSP concentrations as well as fine and coarse mode aerosol NO_3^- . The comparison of results
870 from ISORROPIA-lite against ISORROPIA II v2.3 in metastable mode, showed also negligible
871 differences (<7%) between all the examined aerosol components on a global scale. The
872 comparison of the ISORROPIA-lite results for $PM_{2.5}$ NH_4^+ , SO_4^{2-} , and NO_3^- versus observations
873 from the IMPROVE, EMEP and EANET networks reveals that East Asia is the area with the
874 largest discrepancies. There was satisfactory agreement in Europe and over the US for NH_4^+ and
875 SO_4^{2-} , while ISORROPIA-lite predicted lower concentrations around Hong Kong with a maximum
876 difference of $1.5 \mu g m^{-3}$ (~20 %) for these two species. For NO_3^- , the discrepancy was up to $3 \mu g$
877 m^{-3} (~30 %) in the same region, while a difference of about $1.5 \mu g m^{-3}$ (~25 %) was found over
878 Central Europe with ISORROPIA-lite predicting the higher values. With the exception of Hong
879 Kong, the model in general overpredicted the concentrations of all three aerosol components over
880 the East Asian region.

881 A comparison between ISORROPIA-lite in the metastable state and ISORROPIA II in the
882 stable state was performed to identify potential discrepancies in the inorganic aerosol
883 concentrations simulated by EMAC. Although differences between the two model versions are to
884 be expected due to the different physical state of aerosols at low RH, it is of interest to examine
885 under which conditions these differences occur so that potential users are informed of the strengths
886 and weaknesses of using either model version depending on the application. Both modules are now
887 available as different options in the EMAC model. The agreement between the two versions was

888 generally quite good for global daily mean surface concentrations of inorganic aerosols, mineral
889 ions and aerosol water. More specifically mineral ions, SO_4^{2-} and NH_4^+ in TSP had the smallest
890 differences overall, less than $0.5 \mu\text{g m}^{-3}$ even in localized extreme cases but in the vast majority
891 less than $0.1 \mu\text{g m}^{-3}$ (or less than 5%). For coarse NO_3^- aerosols the absolute differences were of
892 similar magnitude with the higher concentrations simulated by ISORROPIA-lite in the Middle
893 East being the most notable. In the case of fine NO_3^- aerosols, the differences were larger (up to ~
894 $1.75 \mu\text{g m}^{-3}$ in local extremes), mainly over the West coast of South America (North of Atacama
895 Desert), the Tibetan Plateau and Eastern Asia regions with higher concentrations simulated by
896 ISORROPIA II but still within ~30%. In Europe and the US, the corresponding differences were
897 less than $0.25 \mu\text{g m}^{-3}$. The most important difference was the higher aerosol water calculated by
898 ISORROPIA-lite, especially for relative humidity in the 20% to 60 % range. However, this was
899 less than $5 \mu\text{g m}^{-3}$ or 20 % in most cases. Therefore, even though local differences are expected in
900 regions where the relative humidity is often in this range, on a global scale choosing a different
901 physical state of the aerosol at lower RH does not have such a big impact.

902 When the relative humidity ranged from 20 % to 60 %, differences in coarse and fine NO_3^-
903 concentrations predictions among the two versions increased. The highest discrepancies were
904 found in the Tibetan Plateau and the Middle East regions both of which are dominated by such RH
905 values during most of the year. In the first region, the combination of those RH values with mid-
906 range temperatures does not favor nitrate aerosol formation if the aerosol is in the metastable state
907 (ISORROPIA-lite). In the second region, the low RH values result in very low aerosol water
908 predictions for the stable state assumed by ISORROPIA II which hinder the condensation of HNO_3
909 into the aerosol phase.

910 Investigation of the differences in the estimated inorganic aerosol acidity between the two
911 versions, due to the different assumed aerosol phase states, is of great interest for potential future
912 use of ISORROPIA-lite in global climate simulations. ISORROPIA-lite produces slightly more
913 acidic coarse mode aerosols (in comparison to ISORROPIA II) but by less than 1 pH unit on
914 average. The most important differences were found mainly in the Middle East and Arabian
915 Peninsula due to the presence of high mineral cation concentrations. The stable state considered
916 by ISORROPIA II allows the precipitation of insoluble salts and removes anions from the aqueous
917 phase that would otherwise deplete the pH, while this is not the case for the metastable aerosol
918 state considered by ISORROPIA-lite. Furthermore, NH_3 is found to control the aerosol acidity of
919 both fine and coarse mode particles, however, it provides significantly larger buffering capacity to
920 the accumulation mode than to the coarse. This results in slightly more basic accumulation
921 particles than coarse in regions with high NH_3 presence from agricultural activities and low
922 mineral cation concentrations (e.g., Europe).

923 Finally concerning the computational efficiency that ISORROPIA-lite provides when used
924 by the EMAC global model, a speed-up of more than 3% was achieved compared to ISORROPIA
925 II in metastable state and nearly 5% compared to ISORROPIA II in stable state.

926

927 **Code and Data Availability**

928 The usage of MESSy (Modular Earth Submodel System) and access to the source code is licensed
929 to all affiliates of institutions which are members of the MESSy Consortium. Institutions can
930 become a member of the MESSy Consortium by signing the MESSy Memorandum of
931 Understanding. More information can be found on the MESSy Consortium Website
932 <http://www.messy-interface.org>. The code developed in this study and all relevant features,
933 including the ISORROPIA II v2.3 and ISORROPIA-lite v1.0 thermodynamic equilibrium codes
934 as part of the MESSy system, are archived with a restricted access DOI
935 (<https://doi.org/10.5281/zenodo.8379120>) and have already been incorporated into the official
936 development branch of the EMAC modelling system and will therefore be part of all future
937 released versions. The data produced in the study are available from the author upon request.

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946 **Competing Interests**

947 HT acts as a topical editor for GMD but has no further competing interests.

948 **Author Contributions**

949 AM and VAK wrote the paper with contributions from all coauthors. VAK planned the research
950 with contributions from AKS, SNP and AN. AN and SNP provided the ISORROPIA-lite model.
951 AM and HT performed the implementation in EMAC. AM performed the simulations and
952 analyzed the results assisted by VAK and APT. APT provided the observations and performed the
953 model evaluation. All the authors discussed the results and contributed to the manuscript.

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958 References

- 959 Alvarez, D.: JUWELS cluster and booster: Exascale pathfinder with modular supercomputing architecture
960 at juelich supercomputing Centre. *Journal of large-scale research facilities JLSRF*, 7, A183-A183,
961 <https://doi.org/10.17815/jlsrf-7-183>, 2021.
- 962 Andreae, M. O., Jones, C. D., and Cox, P. M.: Strong present-day aerosol cooling implies a hot
963 future. *Nature*, 435(7046), 1187-1190, <https://doi.org/10.1038/nature03671>, 2005.
- 964 Ansari, A. S. and Pandis, S. N.: The effect of metastable equilibrium states on the partitioning of nitrate
965 between the gas and aerosol phases. *Atmospheric Environment*, 34(1), 157-168,
966 [https://doi.org/10.1016/S1352-2310\(99\)00242-3](https://doi.org/10.1016/S1352-2310(99)00242-3), 2000.
- 967 Astitha, M., Lelieveld, J., Abdel Kader, M., Pozzer, A., and De Meij, A: Parameterization of dust emissions
968 in the global atmospheric chemistry-climate model EMAC: impact of nudging and soil
969 properties. *Atmospheric Chemistry and Physics*, 12(22), 11057-11083,
970 <https://doi.org/10.5194/acp-12-11057-2012>, 2012.
- 971 Bacer, S., Sullivan, S. C., Karydis, V. A., Barahona, D., Kramer, M. and co-authors: Implementation of a
972 comprehensive ice crystal formation parameterization for cirrus and mixed-phase clouds in the
973 EMAC model (based on MESSy 2.53). *Geoscientific model development*, 11(10), 4021-4041,
974 <https://doi.org/10.5194/gmd-11-4021-2018>, 2018 .
- 975 Bassett, M. and Seinfeld, J. H.: Atmospheric equilibrium model of sulfate and nitrate
976 aerosols. *Atmospheric Environment (1967)*, 17(11), 2237-2252, [https://doi.org/10.1016/0004-](https://doi.org/10.1016/0004-6981(83)90221-4)
977 [6981\(83\)90221-4](https://doi.org/10.1016/0004-6981(83)90221-4), 1983.
- 978 Bouwman, A. F., Lee, D. S., Asman, W. A., Dentener, F. J., Van Der Hoek, K. W., and Olivier, J. G. J.: A global
979 high-resolution emission inventory for ammonia. *Global biogeochemical cycles*, 11(4), 561-587,
980 <https://doi.org/10.1029/97GB02266>, 1997.
- 981 Bromley, L. A.: Thermodynamic properties of strong electrolytes in aqueous solutions. *AIChE*
982 *journal*, 19(2), 313-320, <https://doi.org/10.1002/aic.690190216>, 1973.
- 983 Brook, R. D., Rajagopalan, S., Pope, C. A., Brook, J. R., Bhatnagar, A. and co-authors: Particulate matter air
984 pollution and cardiovascular disease: an update to the scientific statement from the American
985 Heart Association. *Circulation*, 121(21), 2331-2378,
986 <https://doi.org/10.1161/CIR.0b013e3181d8bece1>, 2010.
- 987 Chen, Y., Shen, H., and Russell, A. G.: Current and future responses of aerosol pH and composition in the
988 US to declining SO₂ emissions and increasing NH₃ emissions. *Environmental Science &*
989 *Technology*, 53(16), 9646-9655, <https://doi.org/10.1021/acs.est.9b02005>, 2019.
- 990 Clegg, S. L., Seinfeld, J. H., and Edney, E. O.: Thermodynamic modelling of aqueous aerosols containing
991 electrolytes and dissolved organic compounds. II. An extended Zdanovskii–Stokes–Robinson
992 approach. *Journal of aerosol science*, 34(6), 667-690, [https://doi.org/10.1016/S0021-](https://doi.org/10.1016/S0021-8502(03)00019-3)
993 [8502\(03\)00019-3](https://doi.org/10.1016/S0021-8502(03)00019-3), 2003.
- 994 Crippa, M., Guizzardi, D., Muntean, M., Schaaf, E., Dentener, F. and co-authors: Gridded emissions of air
995 pollutants for the period 1970–2012 within EDGAR v4. 3.2. *Earth Syst. Sci. Data*, 10(4), 1987-2013,
996 <https://doi.org/10.5194/essd-10-1987-2018>, 2018.
- 997 Dentener, F., Kinne, S., Bond, T., Boucher, O., Cofala, J. and co-authors: Emissions of primary aerosol and
998 precursor gases in the years 2000 and 1750 prescribed data-sets for AeroCom. *Atmospheric*
999 *Chemistry and Physics*, 6(12), 4321-4344, <https://doi.org/10.5194/acp-6-4321-2006>, 2006.
- 1000 Doney, S. C., Mahowald, N., Lima, I., Feely, R. A., Mackenzie, F. T., Lamarque, J. F., and Rasch, P. J.: Impact
1001 of anthropogenic atmospheric nitrogen and sulfur deposition on ocean acidification and the
1002 inorganic carbon system. *Proceedings of the National Academy of Sciences*, 104(37), 14580-
1003 14585, <https://doi.org/10.1073/pnas.0702218104>, 2007.
- 1004 EMEP Programme Air Pollutant Monitoring Data, available at : <https://projects.nilu.no/cc/index.html>

1005 Fang, T., Guo, H., Zeng, L., Verma, V., Nenes, A., and Weber, R. J.: Highly acidic ambient particles, soluble
1006 metals, and oxidative potential: a link between sulfate and aerosol toxicity. *Environmental science*
1007 & *technology*, 51(5), 2611-2620, <https://doi.org/10.1021/acs.est.6b06151>, 2017.

1008 Fountoukis, C. and Nenes, A.: ISORROPIA II: a computationally efficient thermodynamic equilibrium model
1009 for K^+ – Ca^{2+} – Mg^{2+} – NH_4^+ – Na^+ – SO_4^{2-} – NO_3^- – Cl^- – H_2O aerosols. *Atmospheric Chemistry and*
1010 *Physics*, 7(17), 4639-4659, <https://doi.org/10.5194/acp-7-4639-2007>, 2007.

1011 Fountoukis, C., Nenes, A., Sullivan, A., Weber, R., Van Reken, T., Fischer, M., Matías, E., Moya, M., Farmer,
1012 D., and Cohen, R. C.: Thermodynamic characterization of Mexico City aerosol during MILAGRO
1013 2006. *Atmospheric Chemistry and Physics*, 9(6), 2141-2156, [https://doi.org/10.5194/acp-9-2141-](https://doi.org/10.5194/acp-9-2141-2009)
1014 [2009](https://doi.org/10.5194/acp-9-2141-2009), 2009.

1015 Fu, X., Wang, S., Xing, J., Zhang, X., Wang, T., and Hao, J.: Increasing ammonia concentrations reduce the
1016 effectiveness of particle pollution control achieved via SO_2 and NO_x emissions reduction in east
1017 China. *Environmental Science & Technology Letters*, 4(6), 221-227,
1018 <https://doi.org/10.1021/acs.estlett.7b00143>, 2017.

1019 Grewe, V., Brunner, D., Dameris, M., Grenfell, J. L., Hein, R., Shindell, D., and Staehelin, J.: Origin and
1020 variability of upper tropospheric nitrogen oxides and ozone at northern mid-
1021 latitudes. *Atmospheric Environment*, 35(20), 3421-3433, [https://doi.org/10.1016/S1352-](https://doi.org/10.1016/S1352-2310(01)00134-0)
1022 [2310\(01\)00134-0](https://doi.org/10.1016/S1352-2310(01)00134-0), 2001.

1023 Guo, H., Sullivan, A. P., Campuzano-Jost, P., Schroder, J. C., Lopez-Hilfiker, F. D., Dibb, J. E., Jimenez, J. L.,
1024 Thornton, J. A., Brown, S. S., Nenes, A., and Weber, R. J.: Fine particle pH and the partitioning of
1025 nitric acid during winter in the northeastern United States. *Journal of Geophysical Research:*
1026 *Atmospheres*, 121, 10, 355-310, 376, <https://doi.org/10.1002/2016JD025311>, 2016.

1027 He, K., Yang, F., Ma, Y., Zhang, Q., Yao, X., Chan, C. K., Cadle, S., Chan, T., and Mulawa, P.: The
1028 characteristics of $PM_{2.5}$ in Beijing, China, *Atmospheric Environment*, 35(29), 4959-4970,
1029 [https://doi.org/10.1016/S1352-2310\(01\)00301-6](https://doi.org/10.1016/S1352-2310(01)00301-6), 2001.

1030 Héroux, M. E., Anderson, H. R., Atkinson, R., Brunekreef, B., Cohen, A., Forastiere, F. and co-authors:
1031 Quantifying the health impacts of ambient air pollutants: recommendations of a WHO/Europe
1032 project. *International journal of public health*, 60, 619-627, [https://doi.org/10.1007/s00038-015-](https://doi.org/10.1007/s00038-015-0690-y)
1033 [0690-y](https://doi.org/10.1007/s00038-015-0690-y), 2015.

1034 Hersbach, H., Bell, B., Berrisford, P., Hirahara, S., Horanyi, A. and co-authors: The ERA5 global
1035 reanalysis. *Quarterly Journal of the Royal Meteorological Society*, 146(730), 1999-2049,
1036 <https://doi.org/10.1002/qj.3803>, 2020.

1037 Honour, S. L., Bell, J. N. B., Ashenden, T. W., Cape, J. N., and Power, S. A.: Responses of herbaceous plants
1038 to urban air pollution: effects on growth, phenology and leaf surface
1039 characteristics. *Environmental pollution*, 157(4), 1279-1286,
1040 <https://doi.org/10.1016/j.envpol.2008.11.049>, 2009.

1041 Interagency Monitoring of Protected Visual Environment (IMPROVE), available at :
1042 <http://vista.cira.colostate.edu/Improve/improve-data/>

1043 Jacobson, M. Z., Tabazadeh, A., and Turco, R. P.: Simulating equilibrium within aerosols and
1044 nonequilibrium between gases and aerosols. *Journal of Geophysical Research:*
1045 *Atmospheres*, 101(D4), 9079-9091, <https://doi.org/10.1029/96JD00348>, 1996.

1046 Jöckel, P., Sander, R., Kerkweg, A., Tost, H., and Lelieveld, J.: the modular earth submodel system (MESSy)-
1047 a new approach towards earth system modeling. *Atmospheric Chemistry and Physics*, 5(2), 433-
1048 444, <https://doi.org/10.5194/acp-5-433-2005>, 2005.

1049 Jöckel, P., Tost, H., Pozzer, A., Bruhl, C., Buchholz, J. and co-authors: The atmospheric chemistry general
1050 circulation model ECHAM5/MESSy1: consistent simulation of ozone from the surface to the
1051 mesosphere. *Atmospheric Chemistry and Physics*, 6(12), 5067-5104, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-6-5067-2006)
1052 [6-5067-2006](https://doi.org/10.5194/acp-6-5067-2006), 2006.

1053 Jockel, P., Tost, H., Pozzer, A., Kunze, M., Kirner, O. and co-authors: Earth system chemistry integrated
1054 modelling (ESCiMo) with the modular earth submodel system (MESSy) version 2.51. *Geoscientific*
1055 *Model Development*, 9(3), 1153-1200, <https://doi.org/10.5194/gmd-9-1153-2016>, 2016.

1056 Kakavas, S., Pandis, S. N., and Nenes, A.: ISORROPIA-Lite: A Comprehensive Atmospheric Aerosol
1057 Thermodynamics Module for Earth System Models. *Tellus B: Chemical and Physical*
1058 *Meteorology*, 74(2022), <https://doi.org/10.16993/tellusb.33>, 2022.

1059 Kakavas, S., Pandis, S., and Nenes, A.: Effects of Secondary Organic Aerosol Water on fine PM levels and
1060 composition over US. *Atmospheric Chemistry and Physics [preprint]*, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-2022-815)
1061 [2022-815](https://doi.org/10.5194/acp-2022-815), 16 January 2023.

1062 Karydis, V. A., Tsimpidi, A. P., Fountoukis, C., Nenes, A., Zavala, M., Lei, W. F., Molina, L. T., and Pandis, S.
1063 N.: Simulating the fine and coarse inorganic particulate matter concentrations in a polluted
1064 megacity, *Atmospheric Environment*, 44(5), 608-620 ,
1065 <https://doi.org/10.1016/j.atmosenv.2009.11.023> , 2010.

1066 Karydis, V. A., Tsimpidi, A. P., Lei, W., Molina, L. T., and Pandis, S. N.: Formation of semivolatile inorganic
1067 aerosols in the Mexico City Metropolitan Area during the MILAGRO campaign. *Atmospheric*
1068 *Chemistry and Physics*, 11(24), 13305-13323, <https://doi.org/10.5194/acp-11-13305-2011>, 2011.

1069 Karydis, V. A., Tsimpidi, A. P., Pozzer, A., Astitha, M., and Lelieveld, J.: Effects of mineral dust on global
1070 atmospheric nitrate concentrations. *Atmospheric Chemistry and Physics*, 16(3), 1491-1509,
1071 <https://doi.org/10.5194/acp-16-1491-2016>, 2016.

1072 Karydis, V. A., Tsimpidi, A. P., Bacer, S., Pozzer, A., Nenes, A., and Lelieveld, J.: Global impact of mineral
1073 dust on cloud droplet number concentration. *Atmospheric Chemistry and Physics*, 17(9), 5601-
1074 5621, <https://doi.org/10.5194/acp-17-5601-2017>, 2017.

1075 Karydis, V. A., Tsimpidi, A. P., Pozzer, A., and Lelieveld, J.: How alkaline compounds control atmospheric
1076 aerosol particle acidity. *Atmospheric Chemistry and Physics*, 21(19), 14983-15001,
1077 <https://doi.org/10.5194/acp-21-14983-2021>, 2021.

1078 Kerckweg, A., Buchholz, J., Ganzeveld, L., Pozzer, A., Tost, H., and Jöckel, P.: An implementation of the dry
1079 removal processes DRY DEPosition and SEDimentation in the Modular Earth Submodel System
1080 (MESSy). *Atmospheric Chemistry and Physics*, 6(12), 4617-4632, [https://doi.org/10.5194/acp-6-](https://doi.org/10.5194/acp-6-4617-2006)
1081 [4617-2006](https://doi.org/10.5194/acp-6-4617-2006), 2006.

1082 Kim, Y. P., Seinfeld, J. H., and Saxena, P.: Atmospheric gas-aerosol equilibrium I. Thermodynamic
1083 model. *Aerosol Science and Technology*, 19(2), 157-181,
1084 <https://doi.org/10.1080/02786829308959628>, 1993.

1085 Klingmüller, K., Metzger, S., Abdelkader, M., Karydis, V. A., Stenchikov, G. L., Pozzer, A., and Lelieveld, J.:
1086 Revised mineral dust emissions in the atmospheric chemistry–climate model EMAC (MESSy 2.52
1087 DU_Astitha1 KKDU2017 patch). *Geoscientific Model Development*, 11(3), 989-1008,
1088 <https://doi.org/10.5194/gmd-11-989-2018>, 2018.

1089 Klingmüller, K., Lelieveld, J., Karydis, V. A., and Stenchikov, G. L.: Direct radiative effect of dust–pollution
1090 interactions. *Atmospheric chemistry and physics*, 19(11), 7397-7408,
1091 <https://doi.org/10.5194/acp-19-7397-2019>, 2019.

1092 Klingmüller, K., Karydis, V. A., Bacer, S., Stenchikov, G. L., and Lelieveld, J.: Weaker cooling by aerosols due
1093 to dust–pollution interactions. *Atmospheric Chemistry and Physics*, 20(23), 15285-15295,
1094 <https://doi.org/10.5194/acp-20-15285-2020>, 2020 .

1095 Kusik, C. L. and Meissner H.P.: Electrolyte Activity Coefficients in Inorganic Processing. AIChE Symp. Series,
1096 173, 14-20, 1978.

1097 Lelieveld, J., Evans, J. S., Fnais, M., Giannadaki, D., and Pozzer, A.: The contribution of outdoor air pollution
1098 sources to premature mortality on a global scale. *Nature*, 525(7569), 367-371,
1099 <https://doi.org/10.1038/nature15371>, 2015.

1100 Lohmann, U. and Ferrachat, S.: Impact of parametric uncertainties on the present-day climate and on the
1101 anthropogenic aerosol effect. *Atmospheric Chemistry and Physics*, 10(23), 11373-11383,
1102 <https://doi.org/10.5194/acp-10-11373-2010>, 2010.

1103 Manisalidis, I., Stavropoulou, E., Stavropoulos, A., and Bezirtzoglou, E.: Environmental and health impacts
1104 of air pollution: a review. *Frontiers in public health*, 14,
1105 <https://doi.org/10.3389/fpubh.2020.00014>, 2020.

1106 Marais, E. A., Jacob, D. J., Jimenez, J. L., Campuzano-Jost, P., Day, D. A. and co-authors: Aqueous-phase
1107 mechanism for secondary organic aerosol formation from isoprene: application to the southeast
1108 United States and co-benefit of SO₂ emission controls. *Atmospheric Chemistry and Physics*, 16(3),
1109 1603-1618, <https://doi.org/10.5194/acp-16-1603-2016>, 2016.

1110 Meissner, H. P. and Peppas, N. A.: Activity coefficients – aqueous solutions of polybasic acids and
1111 their salts, *AIChE Journal*, 19(4), 806–809, <https://doi.org/10.1002/aic.690190419>, 1973.

1112 Metzger, S., Dentener, F., Pandis, S., and Lelieveld, J., Gas/aerosol partitioning, 1, A computationally
1113 efficient model, *Journal of Geophysical Research: Atmospheres*, 107(D16),
1114 <https://doi.org/10.1029/2001JD001102>, 2002.

1115 Miinalainen, T., Kokkola, H., Lehtinen, K. E., and Kühn, T.: Comparing the radiative forcings of the
1116 anthropogenic aerosol emissions from Chile and Mexico. *Journal of Geophysical Research:
1117 Atmospheres*, 126(10), <https://doi.org/10.1029/2020JD033364>, 2021.

1118 Myhre, G., Samset, B. H., Schulz, M., Balkanski, Y., Bauer, S., Berntsen, T. K. and co-authors: Radiative
1119 forcing of the direct aerosol effect from AeroCom Phase II simulations. *Atmospheric Chemistry
1120 and Physics*, 13(4), 1853-1877, <https://doi.org/10.5194/acp-13-1853-2013>, 2013.

1121 Nenes, A., Pandis, S. N., and Pilinis, C.: ISORROPIA: A new thermodynamic equilibrium model for
1122 multiphase multicomponent inorganic aerosols. *Aquatic geochemistry*, 4, 123-152,
1123 <https://doi.org/10.1023/A:1009604003981>, 1998.

1124 Nenes, A., Pandis, S. N., Weber, R. J., and Russell, A.: Aerosol pH and liquid water content determine when
1125 particulate matter is sensitive to ammonia and nitrate availability. *Atmospheric Chemistry and
1126 Physics*, 20(5), 3249-3258, <https://doi.org/10.5194/acp-20-3249-2020>, 2020.

1127 Pilinis, C. and Seinfeld, J. H.: Continued development of a general equilibrium model for inorganic
1128 multicomponent atmospheric aerosols. *Atmospheric Environment (1967)*, 21(11), 2453-2466,
1129 [https://doi.org/10.1016/0004-6981\(87\)90380-5](https://doi.org/10.1016/0004-6981(87)90380-5), 1987.

1130 Pope, C. A., Burnett, R. T., Turner, M. C., Cohen, A., Krewski, D. and co-authors: Lung cancer and
1131 cardiovascular disease mortality associated with ambient air pollution and cigarette smoke: shape
1132 of the exposure–response relationships. *Environmental health perspectives*, 119(11), 1616-1621,
1133 <https://doi.org/10.1289/ehp.1103639>, 2011.

1134 Pozzer, A., Jöckel, P., Sander, R., Williams, J., Ganzeveld, L., and Lelieveld, J.: The MESSy-submodel AIRSEA
1135 calculating the air-sea exchange of chemical species. *Atmospheric Chemistry and Physics*, 6(12),
1136 5435-5444, <https://doi.org/10.5194/acp-6-5435-2006>, 2006.

1137 Pringle, K. J., Tost, H., Message, S., Steil, B., Giannadaki, D. and co-authors: Description and evaluation of
1138 GMXe: a new aerosol submodel for global simulations (v1). *Geoscientific Model
1139 Development*, 3(2), 391-412, <https://doi.org/10.5194/gmd-3-391-2010>, 2010.

1140 Pringle, K. J., Tost, H., Message, S., Steil, B., Giannadaki, D. and co-authors: Corrigendum to “Description
1141 and evaluation of GMXe: a new aerosol submodel for global simulations (v1)”. *Geoscientific Model
1142 Development*, 3(2), 413-413, <https://doi.org/10.5194/gmd-3-413-2010>, 2010.

1143 Putaud, J. P., Van Dingenen, R., Alastuey, A., Bauer, H., Birmili, W. and co-authors: A European aerosol
1144 phenomenology–3: Physical and chemical characteristics of particulate matter from 60 rural,
1145 urban, and kerbside sites across Europe. *Atmospheric Environment*, 44(10), 1308-1320,
1146 <https://doi.org/10.1016/j.atmosenv.2009.12.011>, 2010.

1147 Roeckner, E., Brokopf, R., Esch, M., Giorgetta, M., Hagemann, S. and co-authors: Sensitivity of simulated
1148 climate to horizontal and vertical resolution in the ECHAM5 atmosphere model. *Journal of*
1149 *Climate*, 19(16), 3771-3791, <https://doi.org/10.1175/JCLI3824.1>, 2006.

1150 Saiz-Lopez, A. and von Glasow, R.: Reactive halogen chemistry in the troposphere. *Chemical Society*
1151 *Reviews*, 41(19), 6448-6472, <https://doi.org/10.1039/C2CS35208G>, 2012.

1152 Sander, R., Baumgaertner, A., Cabrera-Perez, D., Frank, F., Gromov, S. and co-authors: The community
1153 atmospheric chemistry box model CAABA/MECCA-4.0. *Geoscientific model development*, 12(4),
1154 1365-1385, <https://doi.org/10.5194/gmd-12-1365-2019>, 2019.

1155 Savoie, D. L. and Prospero, J.: Particle size distribution of nitrate and sulfate in the marine
1156 atmosphere. *Geophysical Research Letters*, 9(10), 1207-1210,
1157 <https://doi.org/10.1029/GL009i010p01207>, 1982.

1158 Saxena, P., Hudischewskyj, A. B., Seigneur, C., and Seinfeld, J. H.: A comparative study of equilibrium
1159 approaches to the chemical characterization of secondary aerosols. *Atmospheric Environment*
1160 (1967), 20(7), 1471-1483, [https://doi.org/10.1016/0004-6981\(86\)90019-3](https://doi.org/10.1016/0004-6981(86)90019-3), 1986.

1161 Seinfeld, J. H. and Pandis, S. N.: *Atmospheric chemistry and physics: from air pollution to climate change*.
1162 John Wiley & Sons, ISBN 1118947401, 2016.

1163 Silva, P. J., Vawdrey, E. L., Corbett, M., and Erupe, M.: Fine particle concentrations and composition during
1164 wintertime inversions in Logan, Utah, USA. *Atmospheric Environment*, 41(26), 5410-5422,
1165 <https://doi.org/10.1016/j.atmosenv.2007.02.016>, 2007.

1166 Song, S., Gao, M., Xu, W., Shao, J., Shi, G., Wang, S. and co-authors: Fine-particle pH for Beijing winter
1167 haze as inferred from different thermodynamic equilibrium models. *Atmospheric Chemistry and*
1168 *Physics*, 18(10), 7423-7438, <https://doi.org/10.5194/acp-18-7423-2018>, 2018.

1169 Szopa, S., Naik, V., Adhikary, B., Artaxo, P., Berntsen, T. and co-authors. 2021. Short-lived climate forcers,
1170 AGU Fall Meeting 2021, held in New Orleans, LA, 13-17 December 2021, id. U13B-06,
1171 2021AGUFM.U13B..06S, 2021.

1172 Tang, Y. S., Flechard, C. R., Dammggen, U., Vidic, S., Djuricic, V. and co-authors: Pan-European rural
1173 monitoring network shows dominance of NH₃ gas and NH₄NO₃ aerosol in inorganic
1174 atmospheric pollution load. *Atmospheric Chemistry and Physics*, 21(2), 875-914, .
1175 <https://doi.org/10.5194/acp-21-875-2021>, 2021.

1176 Tarin-Carrasco, P., Im, U., Geels, C., Palacios-Peña, L., and Jiménez-Guerrero, P.: Contribution of fine
1177 particulate matter to present and future premature mortality over Europe: A non-linear
1178 response. *Environment international*, 153, 106517,
1179 <https://doi.org/10.1016/j.envint.2021.106517>, 2021.

1180 The Acid Deposition Monitoring Network in East Asia, available at :
1181 <https://monitoring.eanet.asia/document/public/index>

1182 Tost, H., Jöckel, P., Kerkweg, A., Sander, R., and Lelieveld, J.: A new comprehensive SCAVenging submodel
1183 for global atmospheric chemistry modelling. *Atmospheric Chemistry and Physics*, 6(3), 565-574,
1184 <https://doi.org/10.5194/acp-6-565-2006>, 2006.

1185 Tost, H., Jöckel, P., Kerkweg, A., Pozzer, A., Sander, R., and Lelieveld, J.: Global cloud and precipitation
1186 chemistry and wet deposition: tropospheric model simulations with
1187 ECHAM5/MESy1. *Atmospheric Chemistry and Physics*, 7(10), 2733-2757,
1188 <https://doi.org/10.5194/acp-7-2733-2007>, 2007.

1189 Tsimpidi, A. P., Karydis, V. A., and Pandis, S. N.: Response of Inorganic Fine Particulate Matter to Emission
1190 Changes of Sulfur Dioxide and Ammonia: The Eastern United States as a Case Study, *Journal of the*
1191 *Air & Waste Management Association*, 57(12), 1489-1498, [https://doi.org/10.3155/1047-](https://doi.org/10.3155/1047-3289.57.12.1489)
1192 [3289.57.12.1489](https://doi.org/10.3155/1047-3289.57.12.1489), 2007.

1193 Tsimpidi, A. P., Karydis, V. A., Pozzer, A., Pandis, S. N., and Lelieveld, J.: ORACLE (v1. 0): module to simulate
1194 the organic aerosol composition and evolution in the atmosphere. *Geoscientific Model*
1195 *Development*, 7(6), 3153-3172, <https://doi.org/10.5194/gmd-7-3153-2014>, 2014.

1196 Tsimpidi, A. P., Karydis, V. A., Pozzer, A., Pandis, S. N., and Lelieveld, J.: ORACLE 2-D (v2.0): an efficient
1197 module to compute the volatility and oxygen content of organic aerosol with a global chemistry–
1198 climate model, *Geoscientific Model Development*, 11(8), 3369-3389,
1199 <https://doi.org/10.5194/gmd-11-3369-2018>, 2018.

1200 U.S. Environmental Protection Agency Clean Air Markets Division
1201 *Clean Air Status and Trends Network (CASTNET)*, available at : <https://www.epa.gov/castnet>

1202 van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M. and co-authors: Global fire emissions
1203 and the contribution of deforestation, savanna, forest, agricultural, and peat fires (1997–
1204 2009). *Atmospheric chemistry and physics*, 10(23), 11707-11735, [https://doi.org/10.5194/acp-](https://doi.org/10.5194/acp-10-11707-2010)
1205 [10-11707-2010](https://doi.org/10.5194/acp-10-11707-2010), 2010.

1206 Vignati, E., Wilson, J., and Stier, P.: M7: An efficient size-resolved aerosol microphysics module for large-
1207 scale aerosol transport models. *Journal of Geophysical Research: Atmospheres*, 109(D22),
1208 <https://doi.org/10.1029/2003JD004485>, 2004.

1209 Weagle, C. L., Snider, G., Li, C., Van Donkelaar, A., Philip, S., Bissonnette, P., Burke, J., Jackson, J., Latimer,
1210 R., and Stone, E.: Global sources of fine particulate matter: interpretation of PM_{2.5} chemical
1211 composition observed by SPARTAN using a global chemical transport model, *Environmental*
1212 *science & technology*, 52(20), 11670-11681, <https://doi.org/10.1021/acs.est.8b01658>, 2018.

1213 Wexler, A. S. and Clegg, S. L.: Atmospheric aerosol models for systems including the ions H⁺, NH₄⁺, Na⁺,
1214 SO₄²⁻, NO₃⁻, Cl⁻, Br⁻, and H₂O. *Journal of Geophysical Research: Atmospheres*, 107(D14), ACH-
1215 14, <https://doi.org/10.1029/2001JD000451>, 2002.

1216 Wexler, A. S. and Seinfeld, J. H.: Second-generation inorganic aerosol model. *Atmospheric Environment.*
1217 *Part A. General Topics*, 25(12), 2731-2748, [https://doi.org/10.1016/0960-1686\(91\)90203-J](https://doi.org/10.1016/0960-1686(91)90203-J), 1991.

1218 Wolff, G. T.: On the nature of nitrate in coarse continental aerosols. *Atmospheric Environment*
1219 (1967), 18(5), 977-981, [https://doi.org/10.1016/0004-6981\(84\)90073-8](https://doi.org/10.1016/0004-6981(84)90073-8), 1984.

1220 World Health Organization. Ambient (outdoor) air pollution: [https://www.who.int/news-room/fact-](https://www.who.int/news-room/fact-sheets/detail/ambient-(outdoor)-air-quality-and-health)
1221 [sheets/detail/ambient-\(outdoor\)-air-quality-and-health](https://www.who.int/news-room/fact-sheets/detail/ambient-(outdoor)-air-quality-and-health), last access: 19 December 2022.

1222 Xu, G., Zhang, Q., Yao, Y., and Zhang, X.: Changes in PM_{2.5} sensitivity to NO_x and NH₃ emissions due to
1223 a large decrease in SO₂ emissions from 2013 to 2018. *Atmospheric and Oceanic Science*
1224 *Letters*, 13(3), 210-215, <https://doi.org/10.1080/16742834.2020.1738009>, 2020.

1225 Yan, W., Topphoff, M., Rose, C., and Gmehling, J.: Prediction of vapor–liquid equilibria in mixed-solvent
1226 electrolyte systems using the group contribution concept. *Fluid Phase Equilibria*, 162(1-2), 97-113,
1227 [https://doi.org/10.1016/S0378-3812\(99\)00201-0](https://doi.org/10.1016/S0378-3812(99)00201-0), 1999.

1228 Yienger, J. J. and Levy, H.: Empirical model of global soil-biogenic NO_x emissions. *Journal of Geophysical*
1229 *Research: Atmospheres*, 100(D6), 11447-11464, <https://doi.org/10.1029/95JD00370>, 1995.

1230 Zakoura, M. and Pandis, S. N.: Overprediction of aerosol nitrate by chemical transport models: The role of
1231 grid resolution. *Atmospheric Environment*, 187, 390-400,
1232 <https://doi.org/10.1016/j.atmosenv.2018.05.066>, 2018.

1233 Zuend, A., Marcolli, C., Luo, B. P., and Peter, T.: A thermodynamic model of mixed organic-inorganic
1234 aerosols to predict activity coefficients. *Atmospheric Chemistry and Physics*, 8(16), 4559-4593,
1235 <https://doi.org/10.5194/acp-8-4559-2008>, 2008.

1236 Zuend, A., Marcolli, C., Booth, A. M., Lienhard, D. M., Soonsin, V., Krieger, U. K., Topping, D. O., McFiggans,
1237 G., Peter, T., and Seinfeld, J. H.: New and extended parameterization of the thermodynamic model
1238 AIOMFAC: calculation of activity coefficients for organic-inorganic mixtures containing carboxyl,
1239 hydroxyl, carbonyl, ether, ester, alkenyl, alkyl, and aromatic functional groups, *Atmospheric*
1240 *Chemistry and Physics*, 11(17), 9155-9206, <https://doi.org/10.5194/acp-11-9155-2011>, 2011.