



1 GMD Perspective: the quest to improve the  
2 evaluation of groundwater representation in  
3 continental to global scale models

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37

## Abstract

38 Continental- to global-scale hydrologic and land surface models increasingly include representations of  
39 the groundwater system. Such large-scale models are essential for examining, communicating, and  
40 understanding the dynamic interactions between the Earth System above and below the land surface as  
41 well as the opportunities and limits of groundwater resources. We argue that both large-scale and  
42 regional-scale groundwater models have utility, strengths and limitations so continued modeling at both  
43 scales is essential and mutually beneficial. A crucial quest is how to evaluate the realism, capabilities and  
44 performance of large-scale groundwater models given their modeling purpose of addressing large-scale  
45 science or sustainability questions as well as limitations in data availability and commensurability.  
46 Evaluation should identify if, when or where large-scale models achieve their purpose or where  
47 opportunities for improvements exists so that such models better achieve their purpose. We suggest  
48 that reproducing the spatio-temporal details of regional-scale models and matching local data is not a  
49 relevant goal. Instead, it is important to decide on reasonable model expectations regarding when a  
50 large scale model is performing 'well enough' in the context of its specific purpose. The decision of  
51 reasonable expectations is necessarily subjective even if the evaluation criteria is quantitative. Our  
52 objective is to provide recommendations for improving the evaluation of groundwater representation in  
53 continental- to global-scale models. We describe current modeling strategies and evaluation practices,  
54 and subsequently discuss the value of three evaluation strategies: 1) comparing model outputs with  
55 available observations of groundwater levels or other state or flux variables (observation-based  
56 evaluation); 2) comparing several models with each other with or without reference to actual  
57 observations (model-based evaluation); and 3) comparing model behavior with expert expectations of  
58 hydrologic behaviors in particular regions or at particular times (expert-based evaluation). Based on  
59 evolving practices in model evaluation as well as innovations in observations, machine learning and  
60 expert elicitation, we argue that combining observation-, model-, and expert-based model evaluation



61 approaches, while accounting for commensurability issues, may significantly improve the realism of  
62 groundwater representation in large-scale models. Thus advancing our ability for quantification,  
63 understanding, and prediction of crucial Earth science and sustainability problems. We encourage  
64 greater community-level communication and cooperation on this quest, including among global  
65 hydrology and land surface modelers, local to regional hydrogeologists, and hydrologists focused on  
66 model development and evaluation.

67 **1. INTRODUCTION: why and how is groundwater modeled at continental to global scales?**

68 Groundwater is the largest human- and ecosystem-accessible freshwater storage component of the  
69 hydrologic cycle (UNESCO, 1978; Margat & Van der Gun, 2013; Gleeson et al., 2016). Therefore, better  
70 understanding of groundwater dynamics is critical at a time when the ‘great acceleration’ (Steffen et al.,  
71 2015) of many human-induced processes is increasing stress on water resources (Wagener et al., 2010;  
72 Montanari et al., 2013; Sivapalan et al., 2014; van Loon et al., 2016), especially in regions with limited  
73 data availability and analytical capacity. Groundwater is often considered to be an inherently regional  
74 rather than global resource or system. This is partially reasonable because local to regional peculiarities  
75 of hydrology, politics and culture are paramount to groundwater resource management (Foster et al.  
76 2013) and groundwater dynamics in different continents are less directly connected and coupled than  
77 atmospheric dynamics. Regional-scale analysis and models are essential for addressing local to regional  
78 groundwater issues. Generally, regional scale modeling is a mature, well-established field (Hill &  
79 Tiedeman, 2007; Kresic, 2009; Zhou & Li, 2011; Hiscock & Bense, 2014; Anderson et al. 2015a) with clear  
80 and robust model evaluation guidelines (e.g. ASTM, 2016; Barnett et al., 2012). Regional models have  
81 been developed around the world; for example, Rossman & Zlotnik (2014) and Vergnes et al. (2020)  
82 synthesize regional-scale groundwater models across the western United States and Europe,  
83 respectively.



84

85 Yet, important global aspects of groundwater both as a resource and as part of the Earth System are  
86 emerging (Gleeson et al. 2020). First, our increasingly globalized world trades virtual groundwater and  
87 other groundwater-dependent resources in the food-energy-water nexus, and groundwater often  
88 crosses borders in transboundary aquifers. A solely regional approach can be insufficient to analysing  
89 and managing these complex global interlinkages. Second, from an Earth system perspective,  
90 groundwater is part of the hydrological cycle and connected to the atmosphere, oceans and the deeper  
91 lithosphere. A solely regional approach is insufficient to uncover and understand the complex  
92 interactions and teleconnections of groundwater within the Earth System. Regional approaches  
93 generally focus on important aquifers which underlie only a portion of the world's land mass or  
94 population and do not include many other parts of the land surface that may be important for processes  
95 like surface water-groundwater exchange flows and evapotranspiration. A global approach is also  
96 essential to assess the impact of groundwater depletion on sea level rise, since groundwater storage loss  
97 rate on all continents of the Earth must be aggregated. Thus, we argue that groundwater is  
98 simultaneously a local, regional, and increasingly global resource and system and that examining  
99 groundwater problems, solutions, and interactions at all scales is crucial. As a consequence, we urgently  
100 require predictive understanding about how groundwater, used by humans and connected with other  
101 components of the Earth System, operates at a variety of scales.

102

103 Based on the arguments above for considering global perspectives on groundwater, we see four specific  
104 purposes of representing groundwater in continental- to global-scale hydrological or land surface  
105 models and their climate modeling frameworks:

106 (1) To understand and quantify interactions between groundwater and past, present and future  
107 climate. Groundwater systems can have far-reaching effects on climate affecting modulation of



108 surface energy and water partitioning with a long-term memory (Anyah et al., 2008; Maxwell and  
109 Kollet, 2008; Koirala et al. 2013; Krakauer et al., 2014; Maxwell et al., 2016; Taylor, et al., 2013;  
110 Meixner et et, 2018; Wang et al., 2018; Keune et al., 2018). While there have been significant  
111 advances in understanding the role of lateral groundwater flow on evapotranspiration (Maxwell &  
112 Condon, 2016; Bresciani et al, 2016), the interactions between climate and groundwater over  
113 longer time scales (Cuthbert et al., 2019) as well as between irrigation, groundwater, and climate  
114 (Condon and Maxwell, 2019; Condon et al 2020) remain largely unresolved. Additionally, it is well  
115 established that old groundwater with slow turnover times are common at depth (Befus et al.  
116 2017; Jasechko et al. 2017). Groundwater connections to the atmosphere are well documented in  
117 modeling studies (e.g. Forrester and Maxwell, 2020). Previous studies have demonstrated  
118 connections between the atmospheric boundary layer and water table depth (e.g. Maxwell et al  
119 2007; Rahman et al, 2015), under land cover disturbance (e.g. Forrester et al 2018), under  
120 extremes (e.g. Kuene et al 2016) and due to groundwater pumping (Gilbert et al 2017). While a  
121 number of open source platforms have been developed to study these connections (e.g. Maxwell  
122 et al 2011; Shrestha et al 2014; Sulis, 2017) these platforms are regional to continental in extent.  
123 Recent work has shown global impacts of groundwater on atmospheric circulation (Wang et al  
124 2018), but groundwater is still quite simplified in this study.

125 (2) To understand and quantify two-way interactions between groundwater, the rest of the  
126 hydrologic cycle, and the broader Earth System. As the main storage component of the freshwater  
127 hydrologic cycle, groundwater systems support baseflow levels in streams and rivers, and thereby  
128 ecosystems and agricultural productivity and other ecosystem services in both irrigated and  
129 rainfed systems (Scanlon et al., 2012; Qiu et al., 2019; Visser, 1959; Zipper et al., 2015, 2017).  
130 When pumped groundwater is transferred to oceans (Konikow 2011; Wada et al., 2012; Döll et  
131 al., 2014a; Wada, 2016; Caceres et al., 2020; Luijendijk et al. 2020), resulting sea-level rise can



132 impact salinity levels in coastal aquifers, and freshwater and solute inputs to the ocean (Moore,  
133 2010; Sawyer et al., 2016). Difficulties are complicated by international trade of virtual  
134 groundwater which causes aquifer stress in disparate regions (Dalin et al., 2017)

135 (3) To inform water decisions and policy for large, often transboundary groundwater systems in an  
136 increasingly globalized world (Wada & Heinrich, 2013; Herbert & Döll, 2019). For instance,  
137 groundwater recharge from large-scale models has been used to quantify groundwater resources  
138 in Africa, even though large-scale models do not yet include all recharge processes that are  
139 important in this region (Taylor et al., 2013; Jasechko et al. 2014; Cuthbert et al., 2019; Hartmann  
140 et al., 2017).

141 (4) To create visualizations and interactive opportunities that inform citizens and consumers, whose  
142 decisions have global-scale impacts, about the state of groundwater all around the world such as  
143 the World Resources Institute’s Aqueduct website (<https://www.wri.org/aqueduct>), a decision-  
144 support tool to identify and evaluate global water risks.

145 The first two purposes are science-focused while the latter two are sustainability-focused. In sum,  
146 continental- to global-scale hydrologic models incorporating groundwater offer a coherent scientific  
147 framework to examine the dynamic interactions between the Earth System above and below the land  
148 surface, and are compelling tools for conveying the opportunities and limits of groundwater resources  
149 to people so that they can better manage the regions they live in, and better understand the world  
150 around them. We consider both large-scale and regional-scale models to be useful practices that should  
151 both continue to be conducted rather than one replacing another. Ideally large-scale and regional-scale  
152 models should benefit from the other since each has strengths and weaknesses and together the two  
153 practices enrich our understanding and support the management of groundwater across scales (Section  
154 2).



155 The challenge of incorporating groundwater processes into continental- or global-scale models is  
156 formidable and sometimes controversial. Some of the controversy stems from unanswered questions  
157 about how best to represent groundwater in the models whereas some comes from skepticism about  
158 the feasibility of modelling groundwater at non-traditional scales. We advocate for the representation of  
159 groundwater stores and fluxes in continental to global models for the four reasons described above. We  
160 do not claim to have all the answers on how best to meet this challenge. We contend, however, that the  
161 hydrologic community needs to work deliberately and constructively towards effective representations  
162 of groundwater in global models.

163

164 Driven by the increasing recognition of the purpose of representing groundwater in continental- to  
165 global-scale models, many global hydrological models and land surface models have incorporated  
166 groundwater to varying levels of complexity depending on the model provenance and purpose. Different  
167 from regional-scale groundwater models that generally focus on subsurface dynamics, the focus of these  
168 models is on estimating either runoff and streamflow (hydrological models) or land-atmosphere water  
169 and energy exchange (land surface models). Simulation of groundwater storages and hydraulic heads  
170 mainly serve to quantify baseflow that affects streamflow during low flow periods or capillary rise that  
171 increases evapotranspiration. Some land-surface models use approaches based on the topographic  
172 index to simulate fast surface and slow subsurface runoff based on the fraction of saturated area in the  
173 grid cell (Clark et al., 2015; Fan et al., 2019); groundwater in these models does not have water storage  
174 or hydraulic heads (Famiglietti & Wood, 1994; Koster et al., 2000; Niu et al., 2003; Takata et al., 2003).

175 In many hydrological models, groundwater is represented as a linear reservoir that is fed by  
176 groundwater recharge and drains to a river in the same grid cell (Müller Schmied et al., 2014; Gascoïn et  
177 al., 2009; Ngo-Duc et al., 2007). Time series of groundwater storage but not hydraulic heads are  
178 computed. This prevents simulation of lateral groundwater flow between grid cells, capillary rise and



179 two-way exchange flows between surface water bodies and groundwater (Döll et al., 2016). However,  
180 representing groundwater as a water storage compartment that is connected to soil and surface water  
181 bodies by groundwater recharge and baseflow and is affected by groundwater abstractions and returns,  
182 enables global-scale assessment of groundwater resources and stress (Herbert and Döll, 2019) and  
183 groundwater depletion (Döll et al., 2014a; Wada et al., 2014; de Graaf et al., 2014). In some land surface  
184 models, the location of the groundwater table with respect to the land surface is simulated within each  
185 grid cell to enable simulation of capillary rise (Niu et al., 2007) but, as in the case of simulating  
186 groundwater as a linear reservoir, lateral groundwater transport or two-way surface water-groundwater  
187 exchange cannot be simulated with this approach.

188

189 Increasingly, models for simulating groundwater flows between all model grid cells in entire countries or  
190 globally have been developed, either as stand-alone models or as part of hydrological models (Vergnes  
191 & Decharme, 2012; Fan et al., 2013; Lemieux et al. 2008; de Graaf et al., 2017; Kollet et al., 2017;  
192 Maxwell et al., 2015; Reinecke et al., 2018, de Graaf et al 2019). The simulation of groundwater in large-  
193 scale models is a nascent and rapidly developing field with significant computational and  
194 parameterization challenges which have led to significant and important efforts to develop and evaluate  
195 individual models. It is important to note that herein ‘large-scale models’ refer to models that are  
196 laterally extensive across multiple regions (hundreds to thousands of kilometers) and generally include  
197 the upper tens to hundreds of meters of subsurface and have resolutions sometimes as small as ~1 km.  
198 In contrast, ‘regional-scale’ models (tens to hundreds of kilometers) have long been developed for a  
199 specific region or aquifer and can include greater depths and resolutions, more complex  
200 hydrostratigraphy and are often developed from conceptual models with significant regional knowledge.  
201 Regional-scale models include a diverse range of approaches from stand-alone groundwater models  
202 (i.e., representing surface water and vadose zone processes using boundary conditions such as recharge)



203 to fully integrated groundwater-surface water models. In the future, large-scale models could be  
204 developed in a number of different directions which we only briefly introduce here to maintain our  
205 primary focus on model evaluation. One important direction is clearer representation of three-  
206 dimensional geology and heterogeneity including karst (Condon et al. in prep) which should be  
207 considered as part of conceptual model development prior to numerical model implementation.  
208

209 Now that a number of models that represent groundwater at continental to global scales have been  
210 developed and will continue evolving, it is equally important that we advance how we evaluate these  
211 models. To date, large-scale model evaluation has largely focused on individual models and lacked the  
212 rigor of regional-scale model evaluation, with inconsistent practices between models and little  
213 community-level discussion or cooperation. Overall, we have only a partial and piecemeal understanding  
214 of the capabilities and limitations of different approaches to representing groundwater in large-scale  
215 models. Our objective is to provide clear recommendations for evaluating groundwater representation  
216 in continental and global models. We focus on model evaluation because this is the heart of model trust  
217 and reproducibility (Hutton et al., 2016) and improved model evaluation will guide how and where it is  
218 most important to focus future model development. We describe current model evaluation practices  
219 (Section 2) and consider diverse and uncertain sources of information, including observations, models  
220 and experts to holistically evaluate the simulation of groundwater-related fluxes, stores and hydraulic  
221 heads (Section 3). We stress the need for an iterative and open-ended process of model improvement  
222 through continuous model evaluation against the different sources of information. We explicitly  
223 contrast the terminology used herein of ‘evaluation’ and ‘comparison’ against terminology such as  
224 ‘calibration’ or ‘validation’ or ‘benchmarking’, which suggests a modelling process that is at some point  
225 complete. We extend previous commentaries advocating improved hydrologic process representation  
226 and evaluation in large-scale hydrologic models (Clark et al. 2015; Melsen et al. 2016) by adding expert-



227 elicitation and machine learning for more holistic evaluation. We also consider model objective and  
228 model evaluation across the diverse hydrologic landscapes which can both uncover blindspots in model  
229 development. It is important to note that we do not consider water quality or contamination, even  
230 though water quality or contamination is important for water resources, management and  
231 sustainability, since large-scale water quality models are in their infancy (van Vliet et al., 2019)

232

233 We bring together somewhat disparate scientific communities as a step towards greater community-  
234 level cooperation on these challenges, including global hydrology and land surface modelers, local to  
235 regional hydrogeologists, and hydrologists focused on model development and evaluation. We see three  
236 audiences beyond those currently directly involved in large-scale groundwater modeling that we seek to  
237 engage to accelerate model evaluation: 1) regional hydrogeologists who could be reticent about global  
238 models, and yet have crucial knowledge and data that would improve evaluation; 2) data scientists with  
239 expertise in machine learning, artificial intelligence etc. whose methods could be useful in a myriad of  
240 ways; and 3) the multiple Earth Science communities that are currently working towards integrating  
241 groundwater into a diverse range of models so that improved evaluation approaches are built directly  
242 into model development.

## 243 **2. CURRENT MODEL EVALUATION PRACTICES**

244 Here we provide a brief overview of the synergies and differences between regional-scale and large-  
245 scale model evaluation and development as well as the imitations of current evaluation practices for  
246 large-scale models.

247

### 248 **2.1 Synergies between regional-scale and large-scales**



249 Regional-scale and large-scale groundwater models are both governed by the same physical equations  
250 and share many of the same challenges. Like large-scale models, some regional-scale models have  
251 challenges with representing important regional hydrologic processes such as mountain block recharge  
252 (Markovich et al. 2019), and data availability challenges (such as the lack of reliable subsurface  
253 parameterization and hydrologic monitoring data) are common. We propose there are largely untapped  
254 potential synergies between regional-scale and large-scale models based on these commonalities and  
255 the inherent strengths and limitations of each scale (Section 1).

256

257 Much can be learned from regional-scale models to inform the development and evaluation of large-  
258 scale groundwater models. Regional-scale models are evaluated using a variety of data types, some of  
259 which are available and already used at the global scale and some of which are not. In general, the most  
260 common data types used for regional-scale groundwater model evaluation match global-scale  
261 groundwater models: hydraulic head and either total streamflow or baseflow estimated using  
262 hydrograph separation approaches (eg. RRCA, 2003; Woolfenden and Nishikawa, 2014; Tolley et al.,  
263 2019). However, numerous data sources unavailable or not currently used at the global scale have also  
264 been applied in regional-scale models, such as elevation of surface water features (Hay et al., 2018),  
265 existing maps of the potentiometric surface (Meriano and Eyles, 2003), and dendrochronology (Schilling  
266 et al., 2014) - these and other 'non-classical' observations (Schilling et al. 2019) could be the inspiration  
267 for model evaluation of large-scale models in the future but are beyond our scope to discuss. Further,  
268 given the smaller domain size of regional-scale models, expert knowledge and local ancillary data  
269 sources can be more directly integrated and automated parameter estimation approaches such as PEST  
270 are tractable (Leaf et al., 2015; Hunt et al., 2013). We directly build upon this practice of integration of  
271 expert knowledge below in Section 3.3.

272



273 We propose that there may also be potential benefits of large-scale models for the development of  
274 regional-scale models. For instance, the boundary conditions of some regional-scale models could be  
275 improved with large-scale model results. The boundary conditions of regional-scale models are often  
276 assumed, calibrated or derived from other models or data. In a regional-scale model, increasing the  
277 model domain (moving the boundary conditions away from region of interests) or incorporating more  
278 hydrologic processes (for example, moving the boundary condition from recharge to the land surface  
279 incorporating evapotranspiration and infiltration) both can reduce the impact of boundary conditions on  
280 the region and problem of interest. Another potential benefit of large-scale models for regional-scale  
281 models is the more fulsome inclusion of large-scale hydrologic and human processes that could further  
282 enhance the ability of regional-scale models to address both the science-focused and sustainability-  
283 focused purposes described in Section 1. For example, the stronger representation of large-scale  
284 atmospheric processes means that the downwind impact of groundwater irrigation on  
285 evapotranspiration on precipitation and streamflow can be assessed (DeAngelis et al., 2010; Kustu et al.,  
286 2011). Or, the effects of climate change and increased water use that affect the inflow of rivers into the  
287 regional modelling domain can be taken from global scale analyses (Wada and Bierkens, 2014 ). Also,  
288 regional groundwater depletion might be largely driven by virtual water trade which can be better  
289 represented in global analysis and models than regional-scale models (Dalin et al. 2017). Therefore the  
290 processes and results of large-scale models could be used to make regional-scale models even more  
291 robust and better address key science and sustainability questions.

292

293 Given the strengths of regional models, a potential alternative to development of large-scale  
294 groundwater models would be combining or aggregating multiple regional models in a patchwork  
295 approach (as in Zell and Sanford, 2020) to provide global coverage. This would have the advantage of  
296 better respecting regional differences but potentially create additional challenges because the regional



297 models would have different conceptual models, governing equations, boundary conditions etc. in  
298 different regions. Some challenges of this patchwork approach include 1) the required collaboration of a  
299 large number of experts from all over the world over a long period of time; 2) regional groundwater flow  
300 models alone are not sufficient, they need to be integrated into a hydrological model so that  
301 groundwater-soil water and the surface water-groundwater interactions can be simulated; 3) the extent  
302 of regional aquifers does not necessarily coincide with the extent of river basins; and 4) the bias of  
303 regional groundwater models towards important aquifers which as described above, underlie only a  
304 portion of the world's land mass or population and may bias estimates of fluxes such as surface water-  
305 groundwater exchange or evapotranspiration. Given these challenges, we argue that a patchwork  
306 approach of integrating multiple regional models is a compelling idea but likely insufficient to achieve  
307 the purposes of large-scale groundwater modeling described in Section 1. Although this nascent idea of  
308 aggregating regional models is beyond the scope of this manuscript, we consider this an important  
309 future research avenue, and encourage further exploration and improvement of regional-scale model  
310 integration from the groundwater modeling community.

311

## 312 **2.2 Differences between regional-scale and large-scales**

313 Although there are important similarities and potential synergies across scales, it is important to  
314 consider how or if large-scale models are fundamentally different to regional-scale models, especially in  
315 ways that could impact evaluation. The primary differences between large-scale and regional-scale  
316 models are that large-scale models (by definition) cover larger areas and, as a result, typically include  
317 more data-poor areas and are generally built at coarser resolution. These differences impact evaluations  
318 in at least five relevant ways:

319 1) Commensurability errors (also called 'representativeness' errors) occur either when modelled grid  
320 values are interpolated and compared to an observation 'point' or when aggregation of observed



321 'point' values are compared to a modelled grid value (Beven, 2005; Tustison et al., 2001; Beven,  
322 2016; Pappenberger et al., 2009; Rajabi et al., 2018). For groundwater models in particular,  
323 commensurability error will depend on the number and locations of observation points, the  
324 variability structure of the variables being compared such as hydraulic head and the interpolation or  
325 aggregation scheme applied (Tustison et al., 2001; Pappenberger et al., 2009; Reinecke et al., 2020).  
326 Commensurability is a problem for most scales of modelling, but likely more significant the coarser  
327 the model. Regional-scale groundwater models typically have fewer (though not insignificant)  
328 commensurability issues due to smaller grid cell sizes compared to large-scale models.

329 2) Specificity to region, objective and model evaluation criteria because regional-scale models are  
330 developed specifically for a certain region and modeling or management objective whereas large-  
331 scale models are often more general and include different regions. As a result, large-scale models  
332 often have greater heterogeneity of processes and parameters, may not adopt the same calibration  
333 targets and variables, and are not subject to the policy or litigation that sometimes drives model  
334 evaluation of regional-scale models.

335 3) Computational requirements can be immense for large-scale models which leads to challenges with  
336 uncertainty and sensitivity analysis. While some regional-scale models also have large  
337 computational demands, large-scale models cover larger domains and are therefore more  
338 vulnerable to this potential constraint.

339 4) Data availability for large-scale models can be limited because they typically include data-poor  
340 areas, which leads to challenges when only using observations for model evaluation. While data  
341 availability also affects regional-scale models, they are often developed for regions with known  
342 hydrological challenges based on existing data and/or modeling efforts are preceded by significant  
343 regional data collection from detailed sources (such as local geological reports) that are not often  
344 included in continental to global datasets used for large-scale model parameterization.



345 5) Subsurface detail in regional-scale models routinely include heterogeneous and anisotropic  
346 parameterizations which could be improved in future large-scale models. For example, intense  
347 vertical anisotropy routinely induces vertical flow dynamics from vertical head gradients that are  
348 tens to thousands of times greater than horizontal gradients which profoundly alter the meaning of  
349 the deep and shallow groundwater levels, with only the latter remotely resembling the actual water  
350 table. In contrast, currently most large-scale models use a single vertically homogeneous value for  
351 each grid cell, or at best have two layers (de Graaf et al., 2017)

352

### 353 **2.3 Limitations of current evaluation practices for large-scale models**

354 Evaluation of large-scale models has often focused on streamflow or evapotranspiration observations  
355 but joint evaluation together with groundwater-specific variables is appropriate and necessary (e.g.  
356 Maxwell et al. 2015; Maxwell and Condon, 2016). Groundwater-specific variables useful for evaluating  
357 the groundwater component of large-scale models include a) hydraulic head or water table depth; b)  
358 groundwater storage and groundwater storage changes which refer to long-term, negative or positive  
359 trends in groundwater storage where long-term, negative trends are called groundwater depletion; c)  
360 groundwater recharge; d) flows between groundwater and surface water bodies; and e) human  
361 groundwater abstractions and return flows to groundwater. It is important to note that groundwater  
362 and surface water hydrology communities often have slightly different definitions of terms like recharge  
363 and baseflow (Barthel, 2014); we therefore suggest trying to precisely define the meanings of such  
364 words using the actual hydrologic fluxes which we do below. Table 1 shows the availability of  
365 observational data for these variables but does not evaluate the quality and robustness of observations.  
366 Overall there are significant inherent challenges of commensurability and measurability of groundwater  
367 observations in the evaluation of large-scale models. We describe the current model evaluation  
368 practices for each of these variables here:



369

370 a) Simulated hydraulic heads or water table depth in large scale models are frequently compared  
371 to well observations, which are often considered the crucial data for groundwater model  
372 evaluation. Hydraulic head observations from a large number groundwater wells (>1 million)  
373 have been used to evaluate the spatial distribution of steady-state heads (Fan et al., 2013, de  
374 Graaf et al., 2015; Maxwell et al., 2015; Reinecke et al., 2019a, 2020). Transient hydraulic heads  
375 with seasonal amplitudes (de Graaf et al. 2017), declining heads in aquifers with groundwater  
376 depletion (de Graaf et al. 2019) and daily transient heads (Tran et al 2020) have also been  
377 compared to well observations. All evaluation with well observations is severely hampered by  
378 the incommensurability of point values of observed head with simulated heads that represent  
379 averages over cells of a size of tens to hundreds square kilometers; within such a large cell, land  
380 surface elevation, which strongly governs hydraulic head, may vary a few hundred meters, and  
381 average observed head strongly depends on the number and location of well within the cell  
382 (Reinecke et al., 2020). Additional concerns with head observations are the 1) strong sampling  
383 bias of wells towards accessible locations, low elevations, shallow water tables, and more  
384 transmissive aquifers in wealthy, generally temperate countries (Fan et al., 2019); 2) the impacts  
385 of pumping which may or may not be well known; 3) observational errors and uncertainty (Post  
386 and von Asmuth, 2013; Fan et al., 2019); and 4) that heads can reflect the poro-elastic effects of  
387 mass loading and unloading rather than necessarily aquifer recharge and drainage (Burgess et al,  
388 2017). To date, simulated hydraulic heads have more often been compared to observed heads  
389 (rather than water table depth) which results in lower relative errors (Reinecke et al., 2020)  
390 because the range of heads (10s to 1000s m head) is much larger than the range of water table  
391 depths (<1 m to 100s m).

392



393 b) Simulated groundwater storage trends or anomalies in large-scale hydrological models have  
394 been evaluated using observations of groundwater well levels combined with estimates of  
395 storage parameters, such as specific yield; local-scale groundwater modeling; and translation of  
396 regional total water storage trends and anomalies from satellite gravimetry (GRACE: Gravity  
397 Recovery And Climate Experiment) to groundwater storage changes by estimating changes in  
398 other hydrological storages (Döll et al., 2012; 2014a). Groundwater storage changes volumes  
399 and rates have been calculated for numerous aquifers, primarily in the United States, using  
400 calibrated groundwater models, analytical approaches, or volumetric budget analyses (Konikow,  
401 2010). Regional-scale models have also been used to simulate groundwater storage trends  
402 untangling the impacts of water management during drought (Thatch et al. 2020). Satellite  
403 gravimetry (GRACE) is important but has limitations (Alley and Konikow, 2015). First, monthly  
404 time series of very coarse-resolution groundwater storage are indirectly estimated from  
405 observations of total water storage anomalies by satellite gravimetry (GRACE) but only after  
406 model- or observation-based subtraction of water storage changes in glaciers, snow, soil and  
407 surface water bodies (Lo et al., 2016; Rodell et al., 2009; Wada, 2016). As soil moisture, river or  
408 snow dynamics often dominate total water storage dynamics, the derived groundwater storage  
409 dynamics can be so uncertain that severe groundwater drought cannot be detected in this way  
410 (Van Loon et al., 2017). Second, GRACE cannot detect the impact of groundwater abstractions  
411 on groundwater storage unless groundwater depletion occurs (Döll et al., 2014a,b). Third, the  
412 very coarse resolution can lead to incommensurability but in the opposite direction of well  
413 observations. It is important to note that the focus is on storage trends or anomalies since total  
414 groundwater storage to a specific depth (Gleeson et al., 2016) or in an aquifer (Konikow, 2010)  
415 can be estimated but the total groundwater storage in a specific region or cell cannot be  
416 simulated or observed unless the depth of interest is specified (Condon et al., 2020).



417

418 c) Simulated large-scale groundwater recharge (vertical flux across the water table) has been  
419 evaluated using compilations of point estimates of groundwater recharge, results of regional-  
420 scale models, baseflow indices, and expert opinion (Döll and Fiedler, 2008; Hartmann et al.,  
421 2015) or compared between models (e.g. Wada et al. 2010). In general, groundwater recharge is  
422 not directly measurable except by meter-scale lysimeters (Scanlon et al., 2002), and many  
423 groundwater recharge methods such as water table fluctuations and chloride mass balance also  
424 suffer from similar commensurability issues as water table depth data. Although sometimes an  
425 input or boundary condition to regional-scale models, recharge in many large-scale groundwater  
426 models is simulated and thus can be evaluated.

427

428 d) The flows between groundwater and surface water bodies (rivers, lakes, wetlands) are  
429 simulated by many models but are generally not evaluated directly against observations of such  
430 flows since they are very rare and challenging. Baseflow (the slowly varying portion of  
431 streamflow originating from groundwater or other delayed sources) or streamflow ‘low flows’  
432 (when groundwater or other delayed sources predominate), generally cannot be used to directly  
433 quantify the flows between groundwater and surface water bodies at large scales. Groundwater  
434 discharge to rivers can be estimated from streamflow observations only in the very dense gauge  
435 network and/or if streamflow during low flow periods is mainly caused by groundwater  
436 discharge and not by water storage in upstream lakes, reservoirs or wetlands. These conditions  
437 are rarely met in case of streamflow gauges with large upstream areas that can be used for  
438 comparison to large-scale model output. de Graaf et al. (2019) compared the simulated timing  
439 of changes in groundwater discharge to observations and regional-scale models, but only  
440 compared the fluxes directly between the global- and regional-scale models. Due to the



441 challenges of directly observing the flows between groundwater and surface water bodies at  
442 large scales, this is not included in the available data in Table 1; instead in Section 3 we highlight  
443 the potential for using baseflow or the spatial distribution of perennial, intermittent and  
444 ephemeral streams in the future.

445

446 e) Groundwater abstractions have been evaluated by comparison to national, state and county  
447 scale statistics in the U.S. (Wada et al. 2010, Döll et al., 2012, 2014a, de Graaf et al. 2014).  
448 Irrigation is the dominant groundwater use sector in many regions; however, irrigation pumpage  
449 is generally estimated from crop water demand and rarely metered although GRACE and other  
450 remote sensing data have been used to estimate the irrigation water demand (Anderson et al.  
451 2015b). The lack of records or observations of abstraction introduces significant uncertainties  
452 into large-scale models and is simulated and thus can be evaluated. Human groundwater  
453 abstractions and return flows as well as groundwater recharge and the flows between  
454 groundwater and surface water bodies are necessary to simulate storage trends (described  
455 above). But each of these are considered separate observations since they each have different  
456 data sources and assumptions. Groundwater abstraction data at the well scale are severely  
457 hampered by the incommensurability like hydraulic head and recharge described above.

### 458 3. HOW TO IMPROVE THE EVALUATION OF LARGE-SCALE GROUNDWATER MODELS

459 Based on Section 2, we argue that the current model evaluation practices are insufficient to robustly  
460 evaluate large-scale models. We therefore propose evaluating large-scale models using at least three  
461 strategies (pie-shapes in Figure 1): observation-, model-, and expert-driven evaluation which are  
462 potentially mutually beneficial because each strategy has its strengths and weaknesses. We are not  
463 proposing a brand new evaluation method here but rather separating strategies to consider the problem



464 of large-scale model evaluation from different but highly interconnected perspectives. All three  
465 strategies work together for the common goal of ‘improved model large-scale model evaluation’ which  
466 is what is the centre of Figure 1.

467

468 When evaluating large-scale models, it is necessary to first consider reasonable expectations or how to  
469 know a model is ‘well enough’. Reasonable expectations should be based on the modeling purpose,  
470 hydrologic process understanding and the plausibly achievable degree of model realism. First, model  
471 evaluation should be clearly linked to the four science- or sustainability-focused purposes of  
472 representing groundwater in large-scale models (Section 1) and second, to our understanding of  
473 relevant hydrologic processes. The objective of large-scale models cannot be to reproduce the spatio-  
474 temporal details that regional-scale models can reproduce. Determining the reasonable expectations is  
475 necessarily subjective, but can be approached using observation-, model-, and expert-driven evaluation.  
476 As a simple first step in setting realistic expectations, we propose that three physical variables can be  
477 used to form more convincing arguments that a large-scale model is well enough: change in  
478 groundwater storage, water table depth, and regional fluxes between groundwater and surface water.  
479 Below we explore in more detail additional variables and approaches that can support this simple  
480 approach.

481

482 Across all three model evaluation strategies of observation-, model-, and expert-driven evaluation, we  
483 advocate three principles underpinning model evaluation (base of Figure 1), none of which we are the  
484 first to suggest but we highlight here as a reminder: 1) model objectives, such as the groundwater  
485 science or groundwater sustainability objective summarised in Section 1, are important to model  
486 evaluation because they provide the context through which relevance of the evaluation outcome is set;  
487 2) all sources of information (observations, models and experts) are uncertain and this uncertainty



488 needs to be quantified for robust evaluation; and 3) regional differences are likely important for large-  
489 scale model evaluation - understanding these differences is crucial for the transferability of evaluation  
490 outcomes to other places or times.

491

492 We stress that we see the consideration and quantification of uncertainty as an essential need across all  
493 three types of model evaluation we describe below, so we discuss it here rather than with model-driven  
494 model evaluation (Section 3.2) where uncertainty analysis more narrowly defined would often be  
495 discussed. We further note that large-scale models have only been assessed to a very limited degree  
496 with respect to understanding, quantifying, and attributing relevant uncertainties. Expanding computing  
497 power, developing computationally frugal methods for sensitivity and uncertainty analysis, and  
498 potentially employing surrogate models can enable more robust sensitivity and uncertainty analysis  
499 such as used in regional-scale models (Habets et al., 2013; Hill, 2006; Hill & Tiedeman, 2007; Reinecke et  
500 al., 2019b). For now, we suggest applying computationally frugal methods such as the elementary effect  
501 test or local sensitivity analysis (Hill, 2006; Morris, 1991; Saltelli et al., 2000). Such sensitivity and  
502 uncertainty analyses should be applied not only to model parameters and forcings but also to model  
503 structural properties (e.g. boundary conditions, grid resolution, process simplification, etc.) (Wagener  
504 and Pianosi, 2019). This implies that the (independent) quantification of uncertainty in all model  
505 elements (observations, parameters, states, etc.) needs to be improved and better captured in available  
506 metadata.

507

508 We advocate for considering regional differences more explicitly in model evaluation since likely no  
509 single model will perform consistently across the diverse hydrologic landscapes of the world (Van  
510 Werkhoven et al., 2008). Considering regional differences in large-scale model evaluation is motivated  
511 by recent model evaluation results and is already starting to be practiced. Two recent sensitivity



512 analyses of large-scale models reveal how sensitivities to input parameters vary in different regions for  
513 both hydraulic heads and flows between groundwater and surface water (de Graaf et al. 2019; Reinecke  
514 et al., 2020). In mountain regions, large-scale models tend to underestimate steady-state hydraulic  
515 head, possibly due to over-estimated hydraulic conductivity in these regions, which highlights that  
516 model performance varies in different hydrologic landscapes. (de Graaf et al., 2015; Reinecke et al.  
517 2019b). Additionally, there are significant regional differences in performance with low flows for a  
518 number of large-scale models (Zaherpour et al. 2018) likely because of diverse implementations of  
519 groundwater and baseflow schemes. Large-scale model evaluation practice is starting to shift towards  
520 highlighting regional differences as exemplified by two different studies that explicitly mapped  
521 hydrologic landscapes to enable clearer understanding of regional differences. Reinecke et al. (2019b)  
522 identified global hydrological response units which highlighted the spatially distributed parameter  
523 sensitivities in a computationally expensive model, whereas Hartmann et al. (2017) developed and  
524 evaluated models for karst aquifers in different hydrologic landscapes based on different a priori system  
525 conceptualizations. Considering regional differences in model evaluation suggests that global models  
526 could in the future consider a patchwork approach of different conceptual models, governing equations,  
527 boundary conditions etc. in different regions. Although beyond the scope of this manuscript, we  
528 consider this an important future research avenue.

### 529 **3.1 Observation-based model evaluation**

530 Observation-based model evaluation is the focus of most current efforts and is important because we  
531 want models to be consistent with real-world observations. Section 2 and Table 1 highlight both the  
532 strengths and limitations of current practices using observations. Despite existing challenges, we foresee  
533 significant opportunities for observation-based model evaluation and do not see data scarcity as a  
534 reason to exclude groundwater in large-scale models or to avoid evaluating these models. It is important



535 to note that most so-called ‘observations’ are modeled or derived quantities, and often at the wrong  
536 scale for evaluating large-scale models (Table 1; Beven, 2019). Given the inherent challenges of direct  
537 measurement of groundwater fluxes and stores especially at large scales, herein we consider the word  
538 ‘observation’ loosely as any measurements of physical stores or fluxes that are combined with or filtered  
539 through models for an output. For example, GRACE gravity measurements are combined with model-  
540 based estimates of water storage changes in glaciers, snow, soil and surface water for ‘groundwater  
541 storage change observations’ or streamflow measurements are filtered through baseflow separation  
542 algorithms for ‘baseflow observations’. The strengths and limitations as well as the data availability and  
543 spatial and temporal attributes of different observations are summarized in Table 1 which we hope will  
544 spur more systematic and comprehensive use of observations.

545

546 Here we highlight nine important future priorities for improving evaluation using available observations.  
547 The first five priorities focus on current observations (Table 1) whereas the latter four focus on new  
548 methods or approaches:

549 1) Focus on transient observations of the water table depth rather than hydraulic head  
550 observations that are long-term averages or individual times (often following well  
551 drilling). Water table depth are likely more robust evaluation metrics than hydraulic  
552 head because water table depth reveals great discrepancies and is a complex function of  
553 the relationship between hydraulic head and topography that is crucial to predicting  
554 system fluxes (including evapotranspiration and baseflow). Comparing transient  
555 observations and simulations instead of long-term averages or individual times  
556 incorporates more system dynamics of storage and boundary conditions as temporal  
557 patterns are more important than absolute values (Heudorfer et al. 2019). For regions  
558 with significant groundwater depletion, comparing to declining water tables is a useful



559 strategy (de Graaf et al. 2019), whereas in aquifers without groundwater depletion,  
560 seasonally varying water table depths are likely more useful observations (de Graaf et  
561 al. 2017).

562 2) Use baseflow, the slowly varying portion of streamflow originating from groundwater or  
563 other delayed sources. Döll and Fiedler (2008) included the baseflow index in evaluating  
564 recharge and baseflow has been used to calibrate the groundwater component of a land  
565 surface model (Lo et al. 2008, 2010). But the baseflow index (BFI), linear and nonlinear  
566 baseflow recession behavior or baseflow fraction (Gnann et al., 2019) have not been  
567 used to evaluate any large-scale model that simulates groundwater flows between all  
568 model grid cells. There are limitations of using BFI and baseflow recession characteristics  
569 to evaluate large-scale models (Table 1). Using baseflow only makes sense when the  
570 baseflow separation algorithm is better than the large-scale model itself, which may not  
571 be the case for some large-scale models and only in time periods that can be assumed  
572 to be dominated by groundwater discharge. Similarly, using recession characteristics is  
573 dependent on an appropriate choice of recession extraction methods. But this remains  
574 available and obvious data derived from streamflow or spring flow observations that has  
575 been under-used to date.

576 3) Use the spatial distribution of perennial, intermittent, and ephemeral streams as an  
577 observation, which to our best knowledge has not been done by any large-scale model  
578 evaluation. The transition between perennial and ephemeral streams is an important  
579 system characteristic in groundwater-surface water interactions (Winter et al. 1998), so  
580 we suggest that this might be a revealing evaluation criteria although there are similar  
581 limitations to using baseflow. The results of both quantifying baseflow and mapping  
582 perennial streams depend on the methods applied, they are not useful for quantifying



583 groundwater-surface water interactions when there is upstream surface water storage,  
584 and they do not directly provide information about fluxes between groundwater and  
585 surface water.

586 4) Use data on land subsidence to infer head declines or aquifer properties for regions  
587 where groundwater depletion is the main cause of compaction (Bierkens and Wada,  
588 2019). Lately, remote sensing methods such as GPS, airborne and space borne radar and  
589 lidar are frequently used to infer land subsidence rates (Erban et al., 2014). Also, a  
590 number of studies combine geomechanical modelling (Ortega-Guerrero et al 1999;  
591 Minderhoud et al 2017) and geodetic data to explain the main drivers of land  
592 subsidence. A few papers (e.g. Zhang and Burbey 2016) use a geomechanical model  
593 together with a withdrawal data and geodetic observations to estimate hydraulic and  
594 geomechanical subsoil properties.

595 5) Consider using socio-economic data for improving model input. For example, reported  
596 crop yields in areas with predominant groundwater irrigation could be used to evaluate  
597 groundwater abstraction rates. Or using well depth data (Perrone and Jasechko, 2019)  
598 to assess minimum aquifer depths or in coastal regions and deltas, the presence of  
599 deeper fresh groundwater under semi-confining layers.

600 6) Derive additional new datasets using meta-analysis and/or geospatial analysis such as  
601 gaining or losing stream reaches (e.g., from interpolated head measurements close to  
602 the streams), springs and groundwater-dependent surface water bodies, or tracers.  
603 Each of these new data sources could in principle be developed from available data  
604 using methods already applied at regional scales but do not currently have an 'off the  
605 shelf' global dataset. For example, some large-scale models have been explicitly  
606 compared with residence time and tracer data (Maxwell et al., 2016) which have also



607                    been recently compiled globally (Gleeson et al., 2016; Jasechko et al., 2017). This could  
608                    be an important evaluation tool for large-scale models that are capable of simulating  
609                    flow paths, or can be modified to do although a challenge of this approach is the  
610                    conservativity of tracers. Future meta-analyses data compilations should report on the  
611                    quality of the data and include possible uncertainty ranges as well as the mean  
612                    estimates.

613                    7) Use machine learning to identify process representations (e.g. Beven, 2020) or  
614                    spatiotemporal patterns, for example of perennial streams, water table depths or  
615                    baseflow fluxes, which might not be obvious in multi-dimensional datasets and could be  
616                    useful in evaluation. For example, Yang et al. (2019) predicted the state of losing and  
617                    gaining streams in New Zealand using random forests. A staggering variety of machine  
618                    learning tools are available and their use is nascent yet rapidly expanding in geoscience  
619                    and hydrology (Reichstein et al., 2019; Shen, 2018; Shen et al., 2018; Wagener et al.,  
620                    2020). While large-scale groundwater models are often considered ‘data-poor’, it may  
621                    seem strange to propose using data-intensive machine learning methods to improve  
622                    model evaluation. But some of the data sources are large (e.g over 2 million water level  
623                    measurements in Fan et al. 2013 although biased in distribution) whereas other  
624                    observations such as evapotranspiration (Jung et al., 2011) and baseflow (Beck et al.  
625                    2013) are already interpolated and extrapolated using machine learning. Moving  
626                    forwards, it is important to consider commensurability while applying machine learning  
627                    in this context.

628                    8) Consider comparing models against hydrologic signatures - indices that provide insight  
629                    into the functional behavior of the system under study (Wagener et al., 2007; McMillan,  
630                    2020). The direct comparison of simulated and observed variables through statistical



631 error metrics has at least two downsides. One, the above mentioned unresolved  
632 problem of commensurability, and two, the issue that such error metrics are rather  
633 uninformative in a diagnostic sense - simply knowing the size of an error does not tell  
634 the modeller how the model needs to be improved, only that it does (Yilmaz et al.,  
635 2009). One way to overcome these issues, is to derive hydrologically meaningful  
636 signatures from the original data, such as the signatures derived from transient  
637 groundwater levels by Heudorfer et al. (2019). For example, recharge ratio (defined as  
638 the ratio of groundwater recharge to precipitation) might be hydrologically more  
639 informative than recharge alone (Jasechko et al., 2014) or the water table ratio and  
640 groundwater response time (Cuthbert et al. 2019; Opie et al., 2020) which are spatially-  
641 distributed signatures of groundwater systems dynamics. Such signatures might be used  
642 to assess model consistency (Wagener & Gupta, 2005; Hrachowitz et al.2014) by looking  
643 at the similarity of patterns or spatial trends rather than the size of the aggregated  
644 error, thus reducing the commensurability problem.

645 9) Understand and quantify commensurability error issues better so that a fairer  
646 comparison can be made across scales using existing data. As described above,  
647 commensurability errors will depend on the number and locations of observation  
648 points, the variability structure of the variables being compared such as hydraulic head  
649 and the interpolation or aggregation scheme applied. While to some extent we may  
650 appreciate how each of these factors affect commensurability error in theory, in  
651 practice their combined effects are poorly understood and methods to quantify and  
652 reduce commensurability errors for groundwater model purposes remain largely  
653 undeveloped. As such, quantification of commensurability error in (large-scale)  
654 groundwater studies is regularly overlooked as a source of uncertainty because it cannot



655 be satisfactorily evaluated (Tregoning et al., 2012). Currently, evaluation of simulated  
656 groundwater heads is plagued by, as yet, poorly quantified uncertainties stemming from  
657 commensurability errors and we therefore recommend future studies focus on  
658 developing solutions to this problem. An additional, subtle but important and  
659 unresolved commensurability issue can stem from conceptual models. Different  
660 hydrogeologists examining different scales, data or interpreting geology differently can  
661 produce quite different conceptual models of the same region (Troldborg et al. 2007).

662 We recommend evaluating models with a broader range of currently available data sources (with  
663 explicit consideration of data uncertainty and regional differences) while also simultaneously working to  
664 derive new data sets. Using data (such as baseflow, land subsidence, or the spatial distribution of  
665 perennial, intermittent, and ephemeral streams) that is more consistent with the scale modelled grid  
666 resolution will hopefully reduce the commensurability challenges. However, data distribution and  
667 commensurability issues will likely still be present, which underscores the importance of the two  
668 following strategies.

### 669 3.2. Model-based model evaluation

670 Model-based model evaluation, which includes model intercomparison projects (MIP) and model  
671 sensitivity and uncertainty analysis, can be done with or without explicitly using observations. We  
672 describe both inter-model and inter-scale comparisons which could be leveraged to maximize the  
673 strengths of each of these approaches.

674

675 The original MIP concept offers a framework to consistently evaluate and compare models, and  
676 associated model input, structural, and parameter uncertainty under different objectives (e.g., climate  
677 change, model performance, human impacts and developments). Early model intercomparisons of



678 groundwater models focused on nuclear waste disposal (SKI, 1984). Since the Project for the  
679 Intercomparison of Land-Surface Parameterization Schemes (PILPS; Sellers et al., 1993), the first large-  
680 scale MIP, the land surface modeling community has used MIPs to deepen understanding of land  
681 physical processes and to improve their numerical implementations at various scales from regional (e.g.,  
682 Rhône-aggregation project; Boone et al., 2004) to global (e.g., Global Soil Wetness Project; Dirmeyer,  
683 2011). Two examples of recent model intercomparison efforts illustrate the general MIP objectives and  
684 practice. First, ISIMIP (Schewe et al., 2014; Warszawski et al., 2014) assessed water scarcity at different  
685 levels of global warming. Second, IH-MIP2 (Kollet et al., 2017) used both synthetic domains and an  
686 actual watershed to assess fully-integrated hydrologic models because these cannot be validated easily  
687 by comparison with analytical solutions and uncertainty remains in the attribution of hydrologic  
688 responses to model structural errors. Model comparisons have revealed differences, but it is often  
689 unclear whether these stem from differences in the model structures, differences in how the  
690 parameters were estimated, or from other modelling choices (Duan et al., 2006). Attempts for modular  
691 modelling frameworks to enable comparisons (Wagener et al., 2001; Leavesley et al., 2002; Clark et al.,  
692 2008; Fenicia et al., 2011; Clark et al., 2015) or at least shared explicit modelling protocols and boundary  
693 conditions (Refsgaard et al., 2007; Ceola et al., 2015; Warszawski et al., 2014) have been proposed to  
694 reduce these problems.

695

696 Inter-scale model comparison - for example, comparing a global model to a regional-scale model - is a  
697 potentially useful approach which is emerging for surface hydrology models (Hattermann et al., 2017;  
698 Huang et al., 2017) and could be applied to large-scale models with groundwater representation. For  
699 example, declining heads and decreasing groundwater discharge have been compared between a  
700 calibrated regional-scale model (RRCA, 2003) and a global model (de Graaf et al., 2019). A challenge to  
701 inter-scale comparisons is that regional-scale models often have more spatially complex subsurface



702 parameterizations because they have access to local data which can complicate model inter-  
703 comparison. Another approach which may be useful is running large-scale models over smaller  
704 (regional) domains at a higher spatial resolution (same as a regional-scale model) so that model  
705 structure influences the comparison less. In the future, various variables that are hard to directly  
706 observe at large scales but routinely simulated in regional-scale models such as baseflow or recharge  
707 could be used to evaluate large-scale models. In this way, the output fluxes and intermediate spatial  
708 scale of regional models provide a bridge across the “river of incommensurability” between highly  
709 location-specific data such as well observations and the coarse resolution of large-scale models. It is  
710 important to consider that regional-scale models are not necessarily or inherently more accurate than  
711 large-scale models since problems may arise from conceptualization, groundwater-surface water  
712 interactions, scaling issues, parameterization etc.

713

714 In order for a regional-scale model to provide a useful evaluation of a large-scale model, there are  
715 several important documentation and quality characteristics it should meet. At a bare minimum, the  
716 regional-scale model must be accessible and therefore meet basic replicability requirements including  
717 open and transparent input and output data and model code to allow large-scale modelers to run the  
718 model and interpret its output. Documentation through peer review, either through a scientific journal  
719 or agency such as the US Geological Survey, would be ideal. It is particularly important that the  
720 documentation discusses limitations, assumptions and uncertainties in the regional-scale model so that  
721 a large-scale modeler can be aware of potential weaknesses and guide their comparison accordingly.

722 Second, the boundary conditions and/or parameters being evaluated need to be reasonably comparable  
723 between the regional- and large-scale models. For example, if the regional-scale model includes human  
724 impacts through groundwater pumping while the large-scale model does not, a comparison of baseflow  
725 between the two models may not be appropriate. Similarly, there needs to be consistency in the time



726 period simulated between the two models. Finally, as with data-driven model evaluation, the purpose of  
727 the large-scale model needs to be consistent with the model-based evaluation; matching the hydraulic  
728 head of a regional-scale model, for instance, does not indicate that estimates of stream-aquifer  
729 exchange are valid. Ideally, we recommend developing a community database of regional-scale models  
730 that meet this criteria. It is important to note that Rossman & Zlotnik (2014) review 88 regional-scale  
731 models while a good example of such a repository is the California Groundwater Model Archive  
732 ([https://ca.water.usgs.gov/sustainable-groundwater-management/california-groundwater-](https://ca.water.usgs.gov/sustainable-groundwater-management/california-groundwater-modeling.html)  
733 [modeling.html](https://ca.water.usgs.gov/sustainable-groundwater-management/california-groundwater-modeling.html)).

734

735 In addition to evaluating whether models are similar in terms of their outputs, e.g. whether they  
736 simulate similar groundwater head dynamics, it is also relevant to understand whether the influence of  
737 controlling parameters are similar across models. This type of analysis provides insights into process  
738 controls as well as dominant uncertainties. Sensitivity analysis provides the mathematical tools to  
739 perform this type of model evaluation (Saltelli et al., 2008; Pianosi et al., 2016; Borgonovo et al., 2017).  
740 Recent applications of sensitivity analysis to understand modelled controls on groundwater related  
741 processes include the study by Reinecke et al. (2019b) trying to understand parametric controls on  
742 groundwater heads and flows within a global groundwater model. Maples et al. (2020) demonstrated  
743 that parametric controls on groundwater recharge can be assessed for complex models, though over a  
744 smaller domain. As highlighted by both of these studies, more work is needed to understand how to  
745 best use sensitivity analysis methods to assess computationally expensive, spatially distributed and  
746 complex groundwater models across large domains (Hill et al., 2016). In the future, it would be useful to  
747 go beyond parameter uncertainty analysis (e.g. Reinecke et al. 2019b) to begin to look at all of the  
748 modelling decisions holistically such as the forcing data (Weiland et al., 2015) and digital elevation  
749 models (Hawker et al., 2018). Addressing this problem requires advancements in statistics (more



750 efficient sensitivity analysis methods), computing (more effective model execution), and access to large-  
751 scale models codes (Hutton et al. 2016), but also better utilization of process understanding, for  
752 example to create process-based groups of parameters which reduces the complexity of the sensitivity  
753 analysis study (e.g. Hartmann et al., 2015; Reinecke et al., 2019b).

### 754 3.3 Expert-based model evaluation

755 A path much less traveled is expert-based model evaluation which would develop hypotheses of  
756 phenomena (and related behaviors, patterns or signatures) we expect to emerge from large-scale  
757 groundwater systems based on expert knowledge, intuition, or experience. In essence, this model  
758 evaluation approach flips the traditional scientific method around by using hypotheses to test the  
759 simulation of emergent processes from large-scale models, rather than using large-scale models to test  
760 our hypotheses about environmental phenomena. This might be an important path forward for regions  
761 where available data is very sparse or unreliable. The recent discussion by Fan et al. (2019) shows how  
762 hypotheses about large-scale behavior might be derived from expert knowledge gained through the  
763 study of smaller scale systems such as critical zone observatories. While there has been much effort to  
764 improve our ability to make hydrologic predictions in ungauged locations through the regionalization of  
765 hydrologic variables or of model parameters (Bloeschl et al., 2013), there has been much less effort to  
766 directly derive expectations of hydrologic behavior based on our perception of the systems under study.  
767

768 Large-scale models could then be evaluated against such hypotheses, thus providing a general  
769 opportunity to advance how we connect hydrologic understanding with large-scale modeling - a strategy  
770 that could also potentially reduce epistemic uncertainty (Beven et al., 2019), and which may be  
771 especially useful for groundwater systems given the data limitations described above. Developing  
772 appropriate and effective hypotheses is crucial and should likely focus on large-scale controlling factors



773 or relationships between controlling factors and output in different parts of the model domain;  
774 hypotheses that are too specific may only be able to be tested by certain model complexities or in  
775 certain regions. To illustrate the type of hypotheses we are suggesting, we list some examples of  
776 hypotheses drawn from current literature:

- 777 • water table depth and lateral flow strongly affect transpiration partitioning (Famiglietti and  
778 Wood, 1994; Salvucci and Entekhabi, 1995; Maxwell & Condon, 2016);
- 779 • the percentage of inter-basinal regional groundwater flow increases with aridity or decreases  
780 with frequency of perennial streams (Gleeson & Manning, 2008; Goderniaux et al, 2013; Schaller  
781 and Fan, 2008); or
- 782 • human water use systematically redistributes water resources at the continental scale via non-  
783 local atmospheric feedbacks (Al-Yaari et al., 2019; Keune et al., 2018).

784 Alternatively, it might be helpful to also include hypotheses that have been shown to be incorrect since  
785 models should also not show relationships that have been shown to not exist in nature. For example of  
786 a hypotheses that has recently been shown to be incorrect is that the baseflow fraction (baseflow  
787 volume/precipitation volume) follows the Budyko curve (Gnann et al. 2019) . As yet another alternative,  
788 hydrologic intuition could form the basis of model experiments, potentially including extreme model  
789 experiments (far from the natural conditions). For example, an experiment that artificially lowers the  
790 water table by decreasing precipitation (or recharge directly) could hypothesize the spatial variability  
791 across a domain regarding how 'the drainage flux will increase and evaporation flux will decrease as the  
792 water table is lowered'. These hypotheses are meant only for illustrative purposes and we hope future  
793 community debate will clarify the most appropriate and effective hypotheses. We believe that the  
794 debate around these hypotheses alone will lead to advance our understanding, or, at least highlight  
795 differences in opinion.

796



797 Formal approaches are available to gather the opinions of experts and to integrate them into a joint  
798 result, often called expert elicitation (Aspinall, 2010; Cooke, 1991; O’Hagan, 2019). Expert elicitation  
799 strategies have been used widely to describe the expected behavior of environmental or man-made  
800 systems for which we have insufficient data or knowledge to build models directly. Examples include  
801 aspects of future sea-level rise (Bamber and Aspinall, 2013), tipping points in the Earth system (Lenton  
802 et al., 2018), or the vulnerability of bridges to scour due to flooding (Lamb et al., 2017). In the  
803 groundwater community, expert opinion is already widely used to develop system conceptualizations  
804 and related model structures (Krueger et al., 2012; Rajabi et al., 2018; Refsgaard et al., 2007), or to  
805 define parameter priors (Ross et al., 2009; Doherty and Christensen, 2011; Brunner et al., 2012;  
806 Knowling and Werner, 2016; Rajabi and Ataie-Ashtiani, 2016). The term expert opinion may be  
807 preferable to the term expert knowledge because it emphasizes a preliminary state of knowledge  
808 (Krueger et al., 2012).

809

810 A critical benefit of expert elicitation is the opportunity to bring together researchers who have  
811 experienced very different groundwater systems around the world. It is infeasible to expect that a single  
812 person could have gained in-depth experience in modelling groundwater in semi-arid regions, in cold  
813 regions, in tropical regions etc. Being able to bring together different experts who have studied one or a  
814 few of these systems to form a group would certainly create a whole that is bigger than the sum of its  
815 parts. If captured, it would be a tremendous source of knowledge for the evaluation of large-scale  
816 groundwater models. Expert elicitation also has a number of challenges including: 1) formalizing this  
817 knowledge in such a way that it is still usable by third parties that did not attend the expert workshop  
818 itself; and 2) perceived or real differences in perspectives, priorities and backgrounds between regional-  
819 scale and large-scale modelers.

820



821 So, while expert opinion and judgment play a role in any scientific investigation (O'Hagan, 2019),  
822 including that of groundwater systems, we rarely use formal strategies to elicit this opinion. It is also less  
823 common to use expert opinion to develop hypotheses about the dynamic behavior of groundwater  
824 systems, rather than just priors on its physical characteristics. Yet, it is intuitive that information about  
825 system behavior can help in evaluating the plausibility of model outputs (and thus of the model itself).  
826 This is what we call expert-based evaluation herein. Expert elicitation is typically done in workshops with  
827 groups of a dozen or so experts (e.g. Lamb et al., 2018). Upscaling such expert elicitation in support of  
828 global modeling would require some web-based strategy and a formalized protocol to engage a  
829 sufficiently large number of people. Contributors could potentially be incentivized to contribute to the  
830 web platform by publishing a data paper with all contributors as co-authors and a secondary analysis  
831 paper with just the core team as coauthors. We recommend the community develop expert elicitation  
832 strategies to identify effective hypotheses that directly link to the relevant large-scale hydrologic  
833 processes of interest.

#### 834 **4. CONCLUSIONS: towards a holistic evaluation of groundwater representation in large-scale models**

835 Ideally, all three strategies (observation-based, model-based, expert-based) should be pursued  
836 simultaneously because the strengths of one strategy might further improve others. For example,  
837 expert- or model-based evaluation may highlight and motivate the need for new observations in certain  
838 regions or at new resolutions. Or observation-based model evaluation could highlight and motivate  
839 further model development or lead to refined or additional hypotheses. We thus recommend the  
840 community significantly strengthens efforts to evaluate large-scale models using all three strategies.  
841 Implementing these three model evaluation strategies may require a significant effort from the scientific  
842 community, so we therefore conclude with two tangible community-level initiatives that would be



843 excellent first steps that can be pursued simultaneously with efforts by individual research groups or  
844 collaborations of multiple research groups.

845

846 First, we need to develop a ‘Groundwater Modeling Data Portal’ that would both facilitate and  
847 accelerate the evaluation of groundwater representation in continental to global scale models (Bierkens,  
848 2015). Existing initiatives such as IGRAC’s Global Groundwater Monitoring Network ([https://www.un-](https://www.un-igrac.org/special-project/ggm-global-groundwater-monitoring-network)  
849 [igrac.org/special-project/ggm-global-groundwater-monitoring-network](https://www.un-igrac.org/special-project/ggm-global-groundwater-monitoring-network)) and HydroFrame  
850 ([www.hydroframe.org](http://www.hydroframe.org)), are an important first step but were not designed to improve the evaluation of  
851 large-scale models and the synthesized data remains very heterogeneous - unfortunately, even  
852 groundwater level time series data often remains either hidden or inaccessible for various reasons. This  
853 open and well documented data portal should include:

- 854 a) observations for evaluation (Table 1) as well as derived signatures (Section 3.1);
- 855 b) regional-scale models that meet the standards described above and could facilitate inter-scale  
856 comparison (Section 3.2) and be a first step towards linking regional models (Section 2.1);
- 857 c) Schematizations, conceptual or perceptual models of large-scale models since these are the  
858 basis of computational models; and
- 859 d) Hypothesis and other results derived from expert elicitation (Section 3.3).

860 Meta-data documentation, data tagging, aggregation and services as well as consistent data structures  
861 using well-known formats (netCDF, .csv, .txt) will be critical to developing a useful, dynamic and evolving  
862 community resource. The data portal should be directly linked to harmonized input data such as forcings  
863 (climate, land and water use etc.) and parameters (topography, subsurface parameters etc.), model  
864 codes, and harmonized output data. Where possible, the portal should follow established protocols,  
865 such as the Dublin Core Standards for metadata (<https://dublincore.org>) and ISIMIP protocols for  
866 harmonizing data and modeling approach, and would ideally be linked to or contained within an existing



867 disciplinary repository such as HydroShare (<https://www.hydroshare.org/>) to facilitate discovery,  
868 maintenance, and long-term support. Additionally, an emphasis on model objective, uncertainty and  
869 regional differences as highlighted (Section 3) will be important in developing the data portal. Like  
870 expert-elicitation, contribution to the data portal could be incentivized through co-authorship in data  
871 papers and by providing digital object identifiers (DOIs) to submitted data and models so that they are  
872 citable. By synthesizing and sharing groundwater observations, models, and hypotheses, this portal  
873 would be broadly useful to the hydrogeological community beyond just improving global model  
874 evaluation.

875

876 Second, we suggest ISIMIP, or a similar model intercomparison project, could be harnessed as a  
877 platform to improve the evaluation of groundwater representation in continental to global scale models.  
878 For example, in ISIMIP (Warszawski et al., 2014), modelling protocols have been developed with an  
879 international network of climate-impact modellers across different sectors (e.g. water, agriculture,  
880 energy, forestry, marine ecosystems) and spatial scales. Originally, ISIMIP started with multi-model  
881 comparison (model-based model evaluation), with a focus on understanding how model projections  
882 vary across different sectors and different climate change scenarios (ISIMIP Fast Track). However, more  
883 rigorous model evaluation came to attention more recently with ISIMIP2a, and various observation data,  
884 such as river discharge (Global Runoff Data Center), terrestrial water storage (GRACE), and water use  
885 (national statistics), have been used to evaluate historical model simulation (observation-based model  
886 evaluation). To better understand model differences and to quantify the associated uncertainty sources,  
887 ISIMIP2b includes evaluating scenarios (land use, groundwater use, human impacts, etc) and key  
888 assumptions (no explicit groundwater representation, groundwater availability for the future, water  
889 allocation between surface water and groundwater), highlighting that different types of hypothesis  
890 derived as part of the expert-based model evaluation could possibly be simulated as part of the ISIMIP



891 process in the future. While there has been a significant amount of research and publications on MIPs  
892 including surface water availability, limited multi-model assessments for large-scale groundwater  
893 studies exist. Important aspects of MIPs in general could facilitate all three model evaluation strategies:  
894 community-building and cooperation with various scientific communities and research groups, and  
895 making the model input and output publicly available in a standardized format.

896

897 Large-scale hydrologic and land surface models increasingly represent groundwater, which we envision  
898 will lead to a better understanding of large-scale water systems and to more sustainable water resource  
899 use. We call on various scientific communities to join us in this effort to improve the evaluation of  
900 groundwater in continental to global models. As described by examples above, we have already started  
901 this journey and we hope this will lead to better outcomes especially for the goals of including  
902 groundwater in large-scale models that we started with above: improving our understanding of Earth  
903 system processes; and informing water decisions and policy. Along with the community currently  
904 directly involved in large-scale groundwater modeling, above we have made pointers to other  
905 communities who we hope will engage to accelerate model evaluation: 1) regional hydrogeologists, who  
906 would be useful especially in expert-based model evaluation (Section 3.3); 2) data scientists with  
907 expertise in machine learning, artificial intelligence etc. whose methods could be useful especially for  
908 observation- and model-based model evaluation (Sections 3.1 and 3.2); and 3) the multiple Earth  
909 Science communities that are currently working towards integrating groundwater into a diverse range of  
910 models so that improved evaluation approaches are built directly into model development. Together we  
911 can better understand what has always been beneath our feet, but often forgotten or neglected.

912

913

914



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916

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923

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927 and then ordered in reverse alphabetical order for all remaining coauthors.

928

929 **Code and data availability:** This Perspective paper does not present any computational results. There is

930 therefore no code or data associated with this paper.

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936 **Table 1. Available observations for evaluating the groundwater component of large-scale models**  
 937

| Data type   | Strengths   | Limitations  | Data availability and spatial resolution  |
|---|---|--|---|
| <b>Available observations already used to evaluate large-scale models</b> |   |  |   |
| Hydraulic heads or water table depth (averages or single times)           | Direct observation of groundwater levels and storage                                      | observations biased towards North America and Europe; non- commensurable with large-scale models; mixture of observation times | <a href="#">IGRAC Global Groundwater Monitoring Network</a> ; Fan et al., 2013; USGS<br>Point measurements at existing wells  |
| Hydraulic heads or water table depth (transient)                          | Direct observation of changing groundwater levels and storage                             | As above   | time-series available in a few regions, especially through USGS and <a href="#">European Groundwater Drought Initiative</a><br>Point measurements at existing wells |
| Total water storage anomalies (GRACE)                                     | Globally available and regionally integrated signal of water storage trends and anomalies | Groundwater changes are uncertain model remainder; very coarse spatial resolution and limited period                           | Various mascons gridded with resolution of ~100,000 km <sup>2</sup> (Scanlon et al. 2016) which are then processed as groundwater storage change                    |
| Storage change (regional aquifers)  | Regionally integrated response of aquifer   | Bias towards North America and Europe  | Konikow 2011<br>Döll et al., 2014a<br>Regional aquifers (10,000s to 100,000s km <sup>2</sup> )  |
| Recharge  | Direct inflow of groundwater system   | Challenging to measure and upscale   | Döll and Fiedler, 2008; Hartmann et al. 2017; Mohan et al. 2018; Moeck et al. 2020<br>Point to small basin  |
| Abstractions  | Crucial for groundwater depletion and sustainability studies                              | National scale data highly variable in quality; downscaling uncertain  | de Graaf et al. 2014<br>Döll et al. 2014<br>National-scale data down-scaled to grid   |
| Streamflow or spring flow observations                                    | Widely available at various scales; low flows can be related to groundwater               | Challenging to quantify the flows between groundwater and surface water from streamflow  | Global Runoff Data Centre (GRDC) or other <a href="#">data sources</a> ; large to small basin; Olariño et al. 2020<br>point measurements of spring flow             |

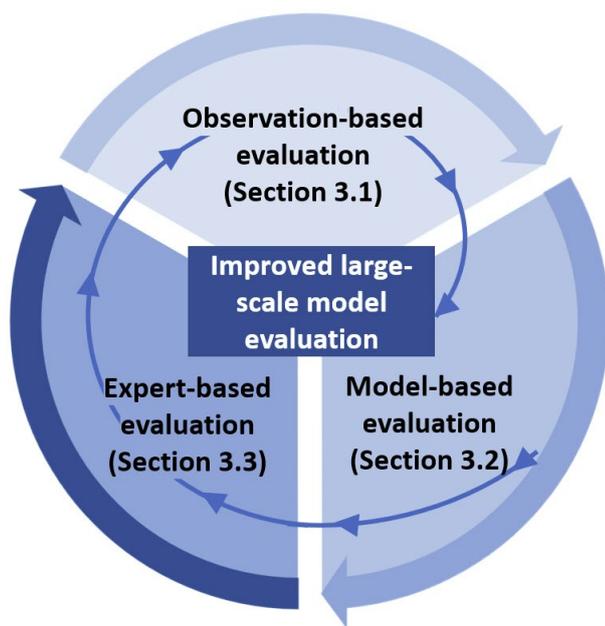


|   |   |   |  |
|---|---|---|--|
| Evapotranspiration  | Widely available; related to groundwater recharge or discharge (for shallow water tables)   | Not a direct groundwater observations   | Various datasets (Miralles et al., 2016); gridded  |
| <b>Available observations not being used to evaluate large-scale models</b> |   |   |  |
| Baseflow index (BFI) or (non-)linear baseflow recession behavior            | Possible integrator of groundwater contribution to streamflow over a basin  | BFI and k values vary with method; baseflow may be dominated by upstream surface water storage rather than groundwater inflow; can not identify losing river conditions                                       | Beck et al. (2013)<br>Point observations extrapolated by machine learning  |
| Perennial stream map  | Ephemeral streams are losing streams, whereas perennial streams could be gaining (or impacted by upstream surface water storage)  | Mapping perennial streams requires arbitrary streamflow and duration cutoffs; not all perennial streams reaches are groundwater-influenced; does not provide information about magnitude of inflows/outflows. | Schneider et al. (2017)<br>Cuthbert et al. (2019);<br>Spatially continuous along stream networks   |
| Gaining or losing stream reaches  | Multiple techniques for measurement (interpolated head measurements, streamflow data, water chemistry). Constrains direction of fluxes at groundwater system boundaries | Relevant processes occur at sub-grid-cell resolution.   | Not globally available but see Bresciani et al. (2018) for a regional example;<br>Spatially continuous along stream networks   |
| Springs and groundwater-dependent surface water bodies                      | Constrains direction of fluxes at groundwater system boundaries   | Relevant processes occur at sub-grid-cell resolution.   | Springs available for various regions (e.g. Springer, & Stevens, 2009) but not globally;<br>Point measurements at water feature locations  |
| Tracers (heat, isotopes or other geochemical)                               | Provides information about temporal aspects of groundwater systems (e.g. residence time)  | No large-scale models simulate transport processes (Table S1)   | Isotopic data compiled (Gleeson et al., 2016; Jasechko et al., 2017) but no global data for heat or other chemistry;<br>Point measurements at existing wells or surface water features |



|   |   |  |   |
|---|---|--|---|
| Surface elevation data (leveling, GPS, radar/lidar) an in particular land subsidence observations | Provides information about changes in surface elevation that are related to groundwater head variations or groundwater head decline | Provides indirect information and needs a geomechanical model to translate to head. Introduces additional uncertainty of geomechanical properties. | Leveling data, GPS data and lidar observations mostly limited to areas of active subsidence (e.g. Minderhoud et al., 2019,2020) and not always open. Global data on elevation change are available from the Sentinel 1 mission. |
|---|---|--|---|

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**Improved model evaluation rests of three core principles:**

- 1) Modelling purpose or objective are paramount
- 2) All sources of information are uncertain
- 3) Regional differences are important

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**Figure 1: Improved large-scale model evaluation rests on three pillars: observation-, model-, and expert-based model evaluation. We argue that each pillar is an essential strategy so that all three should be simultaneously pursued by the scientific community. The three pillars of model evaluation all rest on three core principles related to 1) model objectives, 2) uncertainty and 3) regional differences.**



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