

1 **Reduced-complexity air quality intervention modelling** 2 **over China: development of the InMAPv1.6.1-China and** 3 **comparison with the CMAQv5.2 model**

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18 **Abstract.** This paper presents the first development and evaluation of the reduced-complexity air quality
19 model for China. In this study, a reduced-complexity air quality intervention model over China (InMAP-
20 China) is developed by linking a regional air quality model, a reduced-complexity air quality model, an
21 emission inventory database for China, and a health impact assessment model to rapidly estimate the air
22 quality and health impacts of emission sources in China. The modelling system is applied over mainland
23 China for 2017 under various emission scenarios. A comprehensive model evaluation is conducted by
24 comparison against conventional CMAQ simulations and ground-based observations. We found that
25 InMAP-China satisfactorily predicted total PM_{2.5} concentrations in terms of statistical performance.
26 Compared with the observed PM_{2.5} concentrations, the mean bias (MB), normalized mean bias (NMB),
27 and correlations of the total PM_{2.5} concentrations are -8.1 µg/m³, -18%, and 0.6, respectively. The
28 statistical performance is considered to be satisfactory for a reduced-complexity air quality model and
29 remains consistent with that evaluated in the United States. The underestimation of total PM_{2.5}
30 concentrations was mainly caused by its composition, primary PM_{2.5}. In terms of the ability to quantify
31 source contributions of PM_{2.5} concentrations, InMAP-China presents similar results in comparison with

32 those based on the CMAQ model, the difference is mainly caused by the different treatment of secondary
33 inorganic aerosols in the two models. Focusing on the health impacts, the annual PM_{2.5}-related premature
34 mortality estimated using InMAP-China in 2017 was 1.92 million, which was 25 ten thousand deaths
35 lower than that estimated based on CMAQ simulations as a result of underestimation of PM_{2.5}
36 concentrations. This work presents a version of the reduced-complexity air quality model over China,
37 provides a powerful tool to rapidly assess the air quality and health impacts associated with control policy,
38 and to quantify the source contribution attributable to many emission sources.

39 **1 Introduction**

40 With rapid urbanization and industrialization, fine particulate matter pollution less than 2.5 μm in
41 diameter (PM_{2.5}) has become a major environmental issue in China. High PM_{2.5} concentrations can be
42 observed over eastern China from satellite observations (Xiao et al., 2020) and the PM_{2.5} concentrations
43 have been largely decreased since 2013 due to the effective control measures taken by Chinese
44 governments (Zhao et al., 2021). PM_{2.5} can affect air quality, ecosystems, and climate change and
45 damage human health through short-term or long-term exposure. The Global Burden of Disease study
46 reported that 1.1 million premature deaths were caused by long-term PM_{2.5} exposure over China in 2015
47 (Cohen et al., 2017).

48 State-of-the-science three-dimensional air quality models (AQMs) have been widely used in China
49 as tools to simulate regional PM_{2.5} concentrations, quantify the contributions to total PM_{2.5} concentrations
50 resulting from emission sources and assess the benefits associated with control measures (Chang et al.;
51 2019, Li et al., 2015; Zhang et al., 2015; Zhang et al., 2019). The Weather Research and Forecasting
52 model-Community Multiscale Air Quality Modelling System (WRF-CMAQ) (Appel et al., 2017; Chang
53 et al., 2019), the Weather Research and Forecasting model coupled with Chemistry (WRF-Chem)
54 (Reddington et al., 2019), the Weather Research and Forecasting model-Comprehensive Air Quality
55 Model Extension (WRF-CAMx) (Li et al., 2015), and the Global Adjoint model of Atmospheric
56 Chemistry (GEOS-Chem Adjoint) (Zhang et al., 2015) were frequently used in previous studies. To
57 conduct a series of simulations for multiple scenarios or quantify the separate contributions attributable
58 to multiple sources, large computational resources and run time are required while utilizing conventional
59 AQMs. To address these challenges and to improve the availability and accessibility of air quality
60 modelling, a number of reduced-complexity models have been developed by the air quality research

61 community. The three representative reduced-complexity air quality models frequently used are the
62 Estimating Air Pollution Social Impacts Using Regression (EASIUR) model (Heo et al., 2016; Heo et
63 al., 2017), the updated Air Pollution Emission Experiments and Policy (APEEP2) model (Muller et al.,
64 2007; Muller et al., 2011) and the Intervention for Air Pollution model (InMAP) (Tessum et al., 2017).
65 A recent study compares three reduced-complexity models, EASIUR, APEEP2, and InMAP, and the
66 results indicate that these three models are consistent in their assessment of the marginal social cost at
67 the county level (Gilmore et al., 2019). Reduced-complexity air quality models are less computationally
68 intensive and easier to use. However, it is not available for China. Therefore, it is essential to develop a
69 reduced-complexity air quality model over China to quickly predict $PM_{2.5}$ concentrations and the
70 associated health impacts of emission sources.

71 The reduced-complexity intervention model for air pollution, InMAP, was developed by Tessum et
72 al. (Tessum et al., 2017) to rapidly assess the air pollution, health, and economic impacts resulting from
73 marginal changes in air pollutant emissions. Compared with conventional air quality models, InMAP has
74 the advantage of time efficient, can predict annual-average $PM_{2.5}$ concentrations within few hours but
75 with a modest reduction in accuracy compared with CTMs. InMAP reduces the running time by
76 simplifying the physical and chemical process. InMAP has been used to assess marginal health damage
77 of location-specific emission sources (Goodkind et al., 2019), to quantify the health impacts of individual
78 coal-fired power plants in the United States (Thind et al., 2019) and to estimate the health benefits of
79 control policies considering specific locations (Sergi et al., 2020). However, to date, a version of the
80 reduced-complexity air quality intervention model over China is absent.

81 In this work, based on the source code of the version 1.6.1 of InMAP model, a reduced-complexity
82 air quality intervention model over China (InMAP-China) is developed to rapidly predict the air quality
83 and estimate the health impacts of emission sources in China. The total consumed time for a simulation
84 for the year 2017 using the InMAP-China established in this study is approximately an hour with a single
85 CPU of 24 nodes. Therefore, it is convenient when conducting multiple simulations of $PM_{2.5}$
86 concentrations due to air pollutants emissions in 2017. The modelling system is applied over mainland
87 China for 2017 under various emission scenarios to examine model performance. Comparisons against
88 conventional air quality models and surface observations are performed in this study. The model
89 applicability and limitations are also declared.

90 The paper is organized as follows: Section 2.1 presents the components of InMAP-China including
91 the interface development between WRF-CMAQ and InMAP to generate parameters of the base

92 atmospheric state, the preprocessed process of emission input data and the exposure-response functions
93 employed in this model. Section 2.2 introduces the evaluation protocol, including the statistical variables
94 adopted and the simulation design in this study. Section 3 presents the evaluation of InMAP-China's
95 predictions of PM_{2.5} air quality and PM_{2.5}-related health impacts in several simulations. Section 4
96 summarizes the conclusions and limitations of this study.

97 **2 Description of InMAP-China model**

98 **2.1 Model components and configurations**

99 The reduced-complexity intervention model for air pollution, InMAP, was developed by Tessum et
100 al. (Tessum et al., 2017) to rapidly assess the air pollution, health, and economic impacts resulting from
101 marginal changes in air pollutant emissions. The model has been widely used in studies (Sergi et al.,
102 2020; Thind et al., 2019; Goodkind et al., 2019; Dimanchevi et al., 2019) focusing on PM_{2.5} pollution
103 and health, economic impacts resulting from emission sources in the United States. In this model, the
104 continuous equation of atmospheric pollutants is solved at an annual scale, and the run time can be
105 reduced. The parameters used to represent physical and chemical processes for simplified simulation are
106 calculated prior to using CTM output data. PM_{2.5} air quality and PM_{2.5}-related premature mortality are
107 predicted and output in the InMAP model.

108 In this work, a Chinese version of the reduced-complexity air quality intervention model InMAP-
109 China is developed for the purpose of rapidly estimating the PM_{2.5} concentration and associated health
110 impacts of emission sources. Figure 1 shows the model framework. Based on the source code of the
111 InMAP model, three-step development work is conducted to establish InMAP-China. First, we develop
112 a preprocessed interface to calculate physical and chemical process parameters using the WRF-CMAQ
113 output variables to support the simplified simulation in InMAP-China. Second, air pollutant emission
114 data are preprocessed to an appropriate format for the InMAP-China simulation. Third, the exposure-
115 response function of the GEMM model is employed in InMAP-China and replaces the original default
116 function to assess PM_{2.5}-related health impacts.

117 Table 1 presents the basic configurations of InMAP-China. The simulation domain is over East
118 Asia and covers mainland China. The spatial resolution is 36 km. Fourteen vertical layers are used in
119 InMAP-China, ranging from the surface layer to the top level of the tropospheric layer.

120 **2.1.1 Parameter interface development for simplified simulation in InMAP-China**

121 We develop a preprocessed interface to calculate physical and chemical process parameters using
122 WRF-CMAQ output variables for simplified simulation in InMAP-China based on the Environmental
123 Protection Agency's (EPA) work (Baker et al., 2020). Two NETCDF files containing the key parameters
124 for simplified simulation are generated by using the parameter interface developed here, one is at 36km
125 resolution across entire mainland of China and another is at 4km resolution over the BTH region. The
126 main step of the preprocessed interface includes meteorological and chemical variable extraction and
127 merging, unit conversion, vertical layer mapping, physical and chemical process parameter calculation
128 and average processing. The hourly chemical and meteorological variable outputs from the WRF-CMAQ
129 modelling system are converted into annual-average physical and chemical process parameters required
130 for simplified simulation.

131 A NETCDF file containing the three-dimensional annually averaged parameters to characterize
132 atmospheric advection, dispersion, mixing, chemical reaction, and deposition is generated. Table 2 shows
133 the relationship between the annual-average parameters for simplified simulation and the original hourly
134 variables. In InMAP-China, the annual averaged component and the deviation of wind speed to represent
135 advection are calculated using hourly elements. The offset of wind vectors in different directions may
136 result in some uncertainties in this process. The parameters of eddy diffusion and convective transport
137 are precalculated using hourly elements, including temperature, pressure, boundary layer height, etc. The
138 annual wet deposition rate is determined by the rainwater mixing ratio and cloud fractions. The annual
139 dry deposition rate of particles and gaseous pollutants at the surface level is precalculated using friction
140 speed, heat flux, radiation flux and land cover. The simplification of chemical reactions is different
141 among pollutants. For NO_x , NH_3 , and volatile organic compound (VOC) precursors, the annual averaged
142 gas-particle partitioning is adopted and calculated before using the output concentrations of species from
143 CMAQ. For SO_2 pollutants, the annual oxidation rate of two major conversion pathways for SO_2 is
144 calculated using concentrations of hydroxyl radical (HO) and hydrogen peroxide (H_2O_2) in CMAQ, and
145 the conversion is estimated in InMAP-China.

146 **2.1.2 Prior WRF-CMAQ simulation**

147 To generate the meteorological and chemical parameters required by InMAP-China, a one-year
148 WRF-CMAQ simulation covering the entire mainland of China is conducted to output hourly
149 meteorological and chemical-related variables in the year 2017. Besides, the nested WRF-CMAQ

150 simulation over the BTH region is also conducted and validated using observed data. The corresponded
151 output data is used to generate the meteorological and chemical parameters required by InMAP-China
152 for the simulations of 4 km resolution in the BTH region. Tables S1 and S2 show the major configurations
153 of the WRF-CMAQ modelling system. The WRF model is driven by the National Centers for
154 Environmental Prediction Final Analysis (NCEP-FNL) (<https://doi.org/10.5065/D6M043C6>) reanalysis
155 data to provide the initial and boundary conditions. The meteorological fields derived from the WRF
156 model is used to drive the CMAQ model (Appel et al., 2016) simulations. The air pollutant emissions
157 used here include anthropogenic emissions over China derived from the MEIC model
158 (<http://meicmodel.org/>), anthropogenic emissions over the region of East Asia outside China derived
159 from the MIX-2010 inventory (Li et al., 2015), and biogenic emissions derived from the MEGANv2.10
160 model. The CB05 chemical mechanism and the AERO6 aerosol module are employed in the model
161 simulation.

162 Table S3 summarizes the performance statistics of meteorological variables, including surface
163 temperature, relative humidity, and wind speed, in China in 2017, as simulated by the WRF model. The
164 hourly observed data of major meteorological variables derived from the National Climate Data Center
165 (NCDC) are utilized here. The results show that the meteorological variables simulated by the WRF
166 model agree well with the surface observations, which is consistent with previous studies (Wu et al.,
167 2019; Zheng et al., 2015; Hong et al., 2017). The model performs well on the predictions of surface
168 temperature, with an MB of -0.7 K, an NMB of -6.1%, and R of 0.9. The predictions of relative humidity
169 at a height of 2 metres are relatively satisfied with an MB of 4.1% and an NMB of 6.1%. The predictions
170 of wind speed at a height of 10 metres are slightly overestimated, with an MB of 0.3 m/s and an NMB
171 of 12.4%, which may be caused by out-of-date USGS land use data employed in the model runs.

172 The SO₂, NO₂ and PM_{2.5} concentrations modelled across the domain agree well with the surface
173 observations in terms of the statistical performance and monthly variations. Table S4 summarizes the
174 performance of the statistics of major air pollutant concentrations. The nationwide annual averaged PM_{2.5}
175 concentration simulated in 2017 in China was 42.1 µg/m³. Compared with the observed PM_{2.5} of 45.9
176 µg/m³, there are slight underpredictions with an MB of 3.7 µg/m³ and NMB of 8.1%. The CMAQ model
177 has moderate underpredictions of the NO₂ concentrations and SO₂ concentrations, which may be related
178 to the uncertainties of emission inputs. For modelled NO₂ concentrations, MB and NMB are -4.6 µg/m³
179 and -13.9%, respectively. For modelled SO₂ concentrations, MB and NMB are -0.8 µg/m³ and -4.5%,

180 respectively. Figure S3 shows the monthly variation. The variation trend of the observed SO₂, NO₂, and
181 PM_{2.5} concentrations can basically be reproduced in the CMAQ simulations.

182 **2.1.3 Preprocessed emission input data**

183 We develop the preprocessed module to generate vector emission input for the InMAP-China
184 simulation. This module can allocate air pollutant emissions vertically and horizontally to supply the
185 missing parameters for the emission file and convert them into a shapefile vector format. The shapefile
186 vector format's emission data of 36km resolution in entire mainland of China and 4km resolution in the
187 BTH region in 2017 are pre-processed by using this module.

188 In this module, the emission data are preprocessed by source and altitude. The anthropogenic
189 emissions of five sectors in China in 2017 from the MEIC inventory (<http://meicmodel.org/>), the
190 anthropogenic emissions over regions outside mainland China in Asia from the MIX-2010 inventory (Li
191 et al., 2015), and the natural emissions estimated using the MEGANv2.10 model (Guenther et al., 2012)
192 are employed in this study.

193 More detailed, the gridded anthropogenic emissions of 0.3 degrees for the residential, transportation,
194 and agricultural sectors are preprocessed and input to the surface layer. The gridded air pollutant
195 emissions of the industrial sector and noncoal power plants are preprocessed for allocation to attitudes
196 ranging from 130 metres to 240 metres and 130 metres to 890 metres, respectively. The emissions of
197 coal-fired power plants (CPPs) are preprocessed as point sources. The air pollutant emissions and the
198 stack attribution of each unit are provided in the emission file. Because the stack attribution of the power
199 unit is missed in the MEIC inventory, we supplied the information in the preprocessed module based on
200 NEI (National Emission Inventory data) data of power units. For stack height/stack diameter, a linear
201 relationship is first established (see Figure S1), and then, supplementation for these two parameters of
202 Chinese power plants is conducted by using the relationships. The fixed value for the other two variables
203 of stack attribution is set here because the PM_{2.5} concentrations attributable to power plants (CPPs-PM_{2.5})
204 are less sensitive to the two variables (see Figure S2). The stack gas exit velocity and stack gas exit
205 temperature of the power unit are 6 m/s and 313 K, respectively. The air pollutant emissions over regions
206 outside mainland China in Asia and the natural emissions simulated by MEGANv2.10 are preprocessed
207 and input to the surface layer.

208 **2.1.4 Exposure-response function from GEMM**

209 To rapidly estimate the premature mortality of PM_{2.5} exposures, we employ the exposure-response
210 function from GEMM to estimate PM_{2.5}-related premature mortality, which was developed by Burnett
211 et al. (Burnett et al., 2018), and calculate the premature mortality using PM_{2.5} concentration predictions
212 of InMAP-China. Premature mortality due to non-communicable diseases (NCDs) and lower respiratory
213 infections (LRIs) was considered in this study. Mortality is determined by the mortality incidence rate,
214 population, and attributable fraction (AF) to certain PM_{2.5} concentrations. The national mortality
215 incidence rate and the population data were derived from the GBD2017 study (Institute for Health
216 Metrics and Evaluation). The spatial distribution of the population in 2015 from the Gridded Population
217 of World Version 4 (Doxsey et al., 2015) was employed to allocate the population in 2017.

218 **2.2 Evaluation protocol**

219 **2.2.1 Evaluation method**

220 In this study, the performances of the InMAP-China predictions are evaluated by comparison
221 against CMAQ simulations and surface observations. Model-to-model comparison and model-to-
222 observation comparison have both been used to evaluate the performance of reduced-complexity air
223 quality models in previous studies (Tessum et al., 2017, Gilmore et al., 2019).

224 The following aspects are considered to make an evaluation. First, we examine the ability of
225 InMAP-China to predict PM_{2.5} concentrations at different emission levels, which will be introduced in
226 Section 3.1. Second, to examine the ability to quantify source contributions to PM_{2.5} concentrations, we
227 compare the InMAP-China's predictions of the sectoral contributions attributable to power, industry,
228 residential, transportation, and agriculture with those based on the CMAQ model, which will be
229 presented in Section 3.2. Third, to comprehensively understand the performance at higher spatial
230 resolution using InMAP-China, we compare the predictions of PM_{2.5} concentrations at 4km spatial
231 resolution in the BTH region both modelled by InMAP-China and conventional CMAQ with the
232 observations, which is displayed in Section 3.3. Fourth, focusing on the health impacts, the PM_{2.5}-related
233 premature mortality predicted by InMAP-China is also compared with mortality estimation based on
234 PM_{2.5} exposure derived from CMAQ, which is presented in Section 3.4.

235 For the observed PM_{2.5} concentration data, the annual averaged observed PM_{2.5} concentrations in
236 2017 were calculated using hourly concentration data from the China National Environmental
237 Monitoring Center, CNEMC (<http://www.cnemc.cn/>). More than 1400 national monitoring sites for air

238 pollutant concentrations are included in the simulation domain. The statistical parameters used in this
239 study include the correlation coefficient (R), mean bias (MB), mean error (ME), normalized mean bias
240 (NMB), normalized mean error (NME), and root mean square error (RMSE). The statistical analyses on
241 the performance of InMAP-China are similar to our previous evaluation of conventional CTMs (Zheng
242 et al., 2015; Wu et al., 2019).

243 **2.2.2 Experimental design**

244 We design twelve simulations to examine the model ability of InMAP-China in this study. Table 3
245 shows the sequence of simulations.

246 InMAP_TOT represents the baseline simulation with maximum emissions input, in which five
247 sectoral anthropogenic emissions are derived from the MEIC inventory, natural emissions are derived
248 from the MEGANv2.10 model, and Asian emissions outside mainland China are derived from the MIX-
249 2010 inventory are combined as emission inputs. Five sectoral and five abatement simulations are also
250 conducted to examine the ability of InMAP-China to predict concentration changes in response to
251 sectoral emissions and abatement emissions. The emission inputs for these ten simulations have been
252 declared in Table 3. The annual averaged physical and chemical process parameters are calculated based
253 on the output variables of WRF-CMAQ model, which has already been mentioned in Section 2.1.2.
254 Based on the above input, the particle continuity equations are solved by InMAP-China model to obtain
255 the annual averaged PM_{2.5} concentrations at the steady-state of the atmosphere. The above simulations
256 are all conducted at 36km spatial resolution across the entire mainland of China. Besides, another
257 simulation represented by InMAP-BTH is conducted at 4km spatial resolution over the BTH region, with
258 the anthropogenic emission input data at 4km resolution derived from the MEIC inventory and natural
259 emissions derived from the MEGANv2.10 model is utilized in this simulation.

260 In order to make a comparison with the InMAP-China simulations, eleven CMAQ simulations are
261 also performed under the same emission inputs. The hourly PM_{2.5} concentrations simulated by CMAQ
262 in 2017 are averaged at obtain the annual averaged PM_{2.5} concentrations. Due to limited computational
263 resources, each simulation is conducted for four representative months (January, April, July, and October)
264 in 2017.

265 3 Results and Discussion

266 3.1 Model performance of PM_{2.5} concentrations in China

267 3.1.1 Total PM_{2.5} concentrations

268 Figure 3 shows the performance evaluation of total PM_{2.5} concentrations in the InMAP_TOT
269 simulations. Compared with the observed annual averaged PM_{2.5} concentrations, the total PM_{2.5}
270 concentrations are moderately underpredicted by InMAP-China with an MB of -8.1 μg/m³ and an NMB
271 of -18.1%. Compared with the CMAQ predictions, the total PM_{2.5} concentrations are also underpredicted,
272 with an MB of -5.3 μg/m³ due to the underprediction of primary PM_{2.5}. Consistent air pollutant emissions
273 are employed in the CMAQ and InMAP-China simulations. Therefore, the underpredictions are caused
274 by the different mechanisms in the two models. Basically, InMAP-China reproduces the spatial pattern
275 of total PM_{2.5} concentrations simulated by CMAQ. Notably, significant overpredictions of PM_{2.5}
276 concentrations can be observed over mountain areas across Northern China, and the complex terrain and
277 large emission intensity increase the challenge of predicting PM_{2.5} concentrations using the reduced-
278 complexity air quality model in this region.

279 Figure 4 shows a comparison of PM_{2.5} compositions. Compared with the CMAQ results, the
280 InMAP-China predictions of PM_{2.5} compositions are satisfactory, with NMBs for SO₄²⁻, NO₃⁻, NH₄⁺, and
281 primary PM_{2.5} equal to 13%, -8%, -10%, and -23%, respectively. The predictions of SO₄²⁻, NO₃⁻, and
282 NH₄⁺ perform better than those of primary PM_{2.5}. Figure 5 and Figure 6 compare the spatial distribution
283 of PM_{2.5} compositions, and similar over-predictions of PM_{2.5} compositions can be observed in the
284 mountain area in Northern China.

285 The ability of InMAP-China to predict PM_{2.5} compositions is also examined at various emission
286 levels. Figure 7 compares the concentrations of PM_{2.5} compositions and the proportions of secondary
287 inorganic aerosols (hereafter, SNA) in total PM_{2.5} concentrations in different scenarios by two models.
288 In the InMAP_TOT scenario, the proportion of SNA is 56%, which is extremely close to the 50%
289 proportion in the WRF-CMAQ simulations. In five emission abatement simulations, the proportion was
290 approximately equal to that in the baseline scenario because the linearly treated chemical reaction
291 relationship of SNA was employed in InMAP-China. However, focusing on the simulations of five
292 sectoral emission scenarios, a significant difference can be observed, which is mainly caused by the
293 difference in chemical treatments in InMAP-China and CMAQ. In this situation, the impacts on PM_{2.5}
294 concentrations are distinct due to the nonlinear emission-concentration process.

295 3.1.2 Marginal change in PM_{2.5} concentrations

296 Figure 8 compares the InMAP-China and CMAQ predictions of population-weighted PM_{2.5}
297 concentrations and PM_{2.5} compositions for eleven emission scenarios. Marginal changes in air pollutant
298 concentrations are defined as 1 µg/m³ by normalizing the population-weighted air pollutant
299 concentrations of each scenario using the largest value among all scenarios modelled by CMAQ. The
300 InMAP-China reproduces CMAQ predictions on the marginal change in population-weighted PM_{2.5}
301 concentrations, with a NMB of -12% and correlations of 0.98, as shown in Figure 8(a). This performance
302 is similar to that predicted by InMAP in the United States (Tessum et al., 2017).

303 Figure 8(b)-(f) compares the predictions of PM_{2.5} compositions. The InMAP-China predictions of
304 SO₄²⁻, NO₃⁻, NH₄⁺ and primary PM_{2.5} agree well with the CMAQ results, but the predictions of secondary
305 organic aerosol (SOA) are the poorest. The marginal changes in NO₃⁻ and primary PM_{2.5} concentrations
306 are moderately underpredicted by InMAP-China, with NMB values of -13% and -21%, respectively.
307 Conversely, the marginal change in SO₄²⁻ concentrations is overpredicted with an NMB of 23%. The
308 marginal change in NH₄⁺ predicted by InMAP-China agrees well with the CMAQ predictions. Because
309 few reaction pathways of SOA are included in the CB05 mechanism in the CMAQ simulations, SOAs
310 are underpredicted in the entire modelling system.

311 The regional performance of the changes in PM_{2.5} and its compositions for eleven emission
312 scenarios is also examined in this study. Figures S4-S7 show the regional results. Four regions, including
313 the Beijing-Tianjin-Hebei region (BTH), Yangtze River Delta (YRD), Pearl River Delta (PRD), and
314 Fen Wei Plain (FWP), are analysed here (see Figure 2). At the regional level, the CMAQ predicted
315 marginal changes in population-weighted PM_{2.5} concentrations, and its composition can be reproduced
316 by InMAP-China, which is similar to the nationwide performance. However, the marginal change in
317 SO₄²⁻ concentrations over the BTH is significantly overpredicted by InMAP-China, with an NMB of
318 135%, which is expected to be improved by optimizing the representation of the annual sulfate oxidation
319 rate in this region.

320 3.2 Model performance of source contributions in China

321 Figure 9 shows the contribution of each sector to PM_{2.5} concentrations nationwide and at the regional
322 scale, and Table 4 displays the proportion value of sectoral contribution based on two models. The
323 predictions of the source contributions of PM_{2.5} concentrations in InMAP-China are basically reliable
324 compared with those based on the CMAQ model, and the difference can be explained.

325 The results based on the two models indicate that the industrial and residential sectors are the first
326 and second contributors among the five sectors. The contribution of the electricity sector is comparable
327 when using the two models, while the contributions of transportation and agriculture are moderately
328 different, which is mainly due to the difference in the model mechanism and the treatment of secondary
329 inorganic aerosols in the two models. At the regional scale, the difference in the sectoral contribution
330 caused by the mechanism in the two models is more significant than at the national scale.

331 **3.3 Model performance of PM_{2.5} predictions at higher resolution in the BTH region**

332 We also conducted a simulation with higher spatial resolution of 4 km in the BTH region by using
333 InMAP-China model and make a comparison with the WRF-CMAQ nested simulation at the same area
334 in the BTH region. Figure 10 and Figure 11 show the performance evaluation of total PM_{2.5} concentration
335 and the composition in the InMAP_BTH scenario. Compared with the observed annual averaged PM_{2.5}
336 concentrations, the total PM_{2.5} concentrations are moderately overpredicted in InMAP_BTH with an
337 NMB of 41.3% and an R of 0.5.

338 Further compared with the nested CMAQ predictions, the total PM_{2.5} concentrations are also over-
339 predicted by InMAP-China model. The predictions of PM_{2.5} compositions in the InMAP_BTH scenario
340 are partially satisfactory, **except for SO₄²⁻**, with NMBs for SO₄²⁻, NO₃⁻, NH₄⁺, and primary PM_{2.5} equal
341 to 178%, 36%, 33%, and 27%, respectively. Figure 12 further shows the comparison of the spatial
342 distribution of PM_{2.5} compositions in the BTH region. The overall spatial distribution pattern of PM_{2.5}
343 compositions is similarly modeled by two models, however, an obvious difference can be observed
344 across the mountain area in the BTH region, for instance, the over-predictions of PM_{2.5} compositions,
345 especially, SO₄²⁻ and NO₃⁻ observed near the Taihang mountain area.

346 **3.4 Model performance of PM_{2.5}-related premature mortality in China**

347 To examine the performance of the predictions of PM_{2.5}-related premature mortality, a comparison
348 of premature mortality using the PM_{2.5} predictions from InMAP-China and CMAQ, separately, is
349 performed here. Figure 13 shows the comparison based on two models for all provinces. The results
350 demonstrate that, compared with the premature mortality based on CMAQ, the relative difference is
351 ranging from -44% to 15% at the provincial level due to the difference of PM_{2.5} concentrations in the two
352 models.

353 At the provincial level, the PM_{2.5}-related premature mortality in Beijing city, Tianjin city, Hebei
354 province, and Shanghai city is slightly over-predicted by InMAP-China, with the relative difference

355 ranging from 4% to 15%. Conversely, for the other majority of provinces, PM_{2.5}-related premature
356 mortality is under-predicted by InMAP-China, with the relative difference ranging from -3% to -44%.
357 Overall, the PM_{2.5}-related premature mortality estimated using InMAP-China was 1.92 million people in
358 2017. Compared with the CMAQ-based estimations, 25 ten thousand deaths are under-predicted by
359 InMAP-China because of underestimation of total PM_{2.5} concentrations in the baseline simulation.

360 **4 Conclusions**

361 This work develops a reduced-complexity air quality intervention model over China and presents a
362 comprehensive evaluation by comparing CMAQ simulations and surface observations. The InMAP-
363 China aims at providing a simplified modeling tool to rapidly predict the PM_{2.5} concentrations due to
364 emission change as well as health impact of emission sources in China. After the model is established,
365 the total consumed time for a new simulation under the atmosphere condition in the year 2017 across the
366 mainland of China using InMAP-China is merely an hour with a single CPU of 24 nodes. Therefore, it
367 is time-efficient when conduct new simulations of PM_{2.5} concentrations in China. Notably, the running
368 of WRF-CMAQ simulations is merely necessary in our developing stage of InMAP-China. For the
369 application of InMAP-China, we recommend users to select InMAP-China as a prior tool with extensive
370 simulation demands, for instance, to quantify the PM_{2.5} concentrations due to hundreds of pollution
371 emitters or to rapidly estimate the PM_{2.5} concentrations caused by dozens of control policies, separately.
372 Besides, the variable grid can also be set in InMAP-China to allow high spatial resolution of 1km or even
373 higher in certain urban area.

374 InMAP-China has moderately satisfactory performance in this study, however, this model has
375 reductions in accuracy compared with conventional CTMs. Overall, InMAP-China satisfactorily predicts
376 total PM_{2.5} concentrations in the baseline simulation in terms of statistical performance. Compared with
377 the observed PM_{2.5} concentrations, the MB, NMB, and correlations of the total PM_{2.5} concentrations are
378 -8.1 µg/m³, -18%, and 0.6, respectively. The statistical performance is satisfactory for a reduced-
379 complexity air quality model and remains consistent with the performance evaluation in the United States.
380 The underestimation of total PM_{2.5} mainly comes from the primary PM_{2.5}. Moreover, the spatial pattern
381 of total PM_{2.5} concentrations can be reproduced in InMAP-China, while an overestimation over the
382 mountain area in Northern China can be observed. The large emission intensity and complex terrain over
383 this region increase the difficulty of modelling concentrations in this area. The predictions of source
384 contributions to PM_{2.5} concentrations by InMAP-China are comparable with those based on the CMAQ

385 model, and the difference is mainly caused by the uncertainty of the simplification of chemical process
386 in the InMAP-China. The global version of reduced-complexity air quality model (Global-InMAP) is
387 also developed and preprint recently (Thakrar et al., 2021), our results of InMAP-China can provide
388 more accurate result in the mainland of China.

389 This study is subject to some limitations and uncertainties. In InMAP-China, the annual-average
390 chemical and physical processes parameters are calculated using hourly parameters from WRF-CMAQ.
391 Complicated seasonal and daily variations affecting the formation and transportation of particulate matter
392 are challenging to retain. The intensity of advection of the air mass is supposed to be weakened due to
393 the offset of the wind vector in the averaging process, which was also pointed out in a previous study.
394 Moreover, InMAP-China has difficulty predicting SOA concentrations because reaction pathways for
395 SOA are insufficient in this modelling system. Further research work is suggested to improve the model
396 performance. For instance, the combination of machine learning with the simplified simulation may need
397 to research to promote the reduced-complexity air quality modeling over China.

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414 **Code and data availability**

415 The source code for the localized version of reduced-complexity air quality model over China (InMAP-
416 China), which is developed based on the original InMAP model over the United states. The data related
417 to this study as well as the user manual are available at <https://doi.org/10.5281/zenodo.5111961>.

418 **Author contributions**

419 RL. Wu and Q. Zhang designed the research and RL. Wu carried them out. RL. Wu, CW. Tessum and
420 Y. Zhang contributed to model development. RL. Wu prepared the manuscript with contributions from
421 all co-authors.

422 **Competing interests**

423 The authors declare no competing interests.

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608 **Table 1. Model configurations in InMAP-China.**

Category	Parameters	Configurations
Basic	Research area and period	China, 2017
	Spatial resolution	36 km × 36 km
	Vertical layers	14 layers
	Run type	Steady run
	Variable grid	Static grid
	Projection	Lambert
	Grid numbers	305816
Input	Meteorological and chemical parameters	Calculated using variables from WRFv3.8-CMAQv5.2
	Anthropogenic emissions	MEIC, MIX, MEGAN
	Population data	GPW 2015 and GBD 2017
	Baseline mortality rate	GBD 2017
Output	Air pollutants	PM _{2.5} and its composition concentrations
	Mortality	PM _{2.5} -related premature mortality

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Table 2 The relationship between parameters for simplified simulation and original variables.

WRF-CMAQ's Variables	Descriptions	InMAP-China's Parameters	Descriptions
U, V, W	Wind fields	UAvg, UDeviation VAvg, VDeviation WAvg, WDeviation	Advection and mixing coefficients
PH, PHB	Base state of geopotential and perturbation geopotential	Dz	Layer heights
PBLH	Planetary boundary layer height	M2d, M2u, Kxxyy, Kzz	Mixing coefficients
T	Potential Temperature	SO ₂ Oxidation, PlumeHeight	Chemical reaction rates and plume rise
P, PB	Base state pressure plus perturbation pressure		Chemical reaction rates and plume rise
QRAIN	Mixing ratio of rain	ParticleWetdep, GasWetdep	Wet deposition
QCLOUD	Cloud mixing ratio	SO ₂ Oxidation	Aqueous-phase chemical reaction rates
CLDFRA	Fraction of grid cell covered by clouds	ParticleWetdep, GasWetdep	Wet deposition
SWDOW N, GLW	Downward shortwave and longwave radiative flux at ground level	GasDrydep, ParticleWetdep	Dry deposition
HFX	Surface heat flux	M2d, M2u, Kxxyy, Kzz, Drydep	Mixing and dry deposition
UST	Friction velocity		Mixing and dry deposition
LU_INDE X	Land use type	M2d, M2u, Kxxyy, Kzz	Mixing
DENS	Inverse air density		Mixing and convert between mixing ratio and mass concentration
aVOC	Anthropogenic VOCs that are SOA precursors	aOrgPartitioning	VOCs/SOA partitioning
aSOA	Anthropogenic SOA		
OH, H ₂ O ₂	Hydroxyl radical and hydrogen peroxide concentrations	SO ₂ Oxidation	Oxidation rates
pNO	ANO ₃ I, ANO ₃ J	NOPartitioning	

gNO	NO and NO ₂		NO _x /pNO ₃ partitioning
pNH	ANH ₄ I, ANH ₄ J	NHPartitioning	NH ₃ /pNH ₄ partitioning
gNH	NH ₃		

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644 **Table 3 Simulation experiments conducted using InMAP-China.**

Class	Simulations	Emission input	Physical and chemical parameter input
Base	InMAP_TOT	Five sectoral anthropogenic emissions and natural emissions	
High_re	InMAP_BTH	Five sectoral anthropogenic emissions and natural emissions with 4km resolution at BTH region	
Sec1	InMAP_POW	Power plants emissions	
Sec2	InMAP_INDUS	Industrial emissions	
Sec3	InMAP_TRANS	Transportation emissions	
Sec4	InMAP_RESI	Residential emissions	
Sec5	InMAP_AGRI	Agricultural emissions	
Aba1	InMAP_RE10	Reduce the air pollutants emissions by 10% based on InMAP_TOT emissions	
Aba2	InMAP_RE30	Reduce the air pollutants emissions by 30% based on InMAP_TOT emissions	Converted using WRF-CMAQv5.2 simulation data in the year of 2017;
Aba3	InMAP_RE50	Reduce the air pollutants emissions by 50% based on InMAP_TOT emissions	Remain the same in all simulations.
Aba4	InMAP_RE70	Reduce the air pollutants emissions by 70% based on InMAP_TOT emissions	
Aba5	InMAP_RE90	Reduce the air pollutants emissions by 90% based on InMAP_TOT emissions	

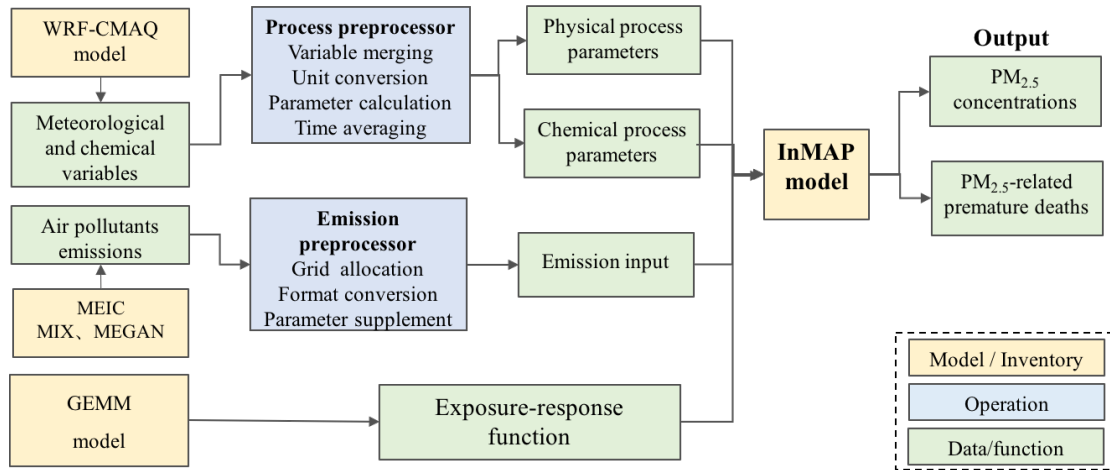
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653 **Table 4 Comparison of the proportions of sectoral contributions to PM_{2.5} concentrations using InMAP-**
 654 **China and CMAQ.**

Sector	National		BTH		YRD		PRD		FWPY	
	CMAQ	InMAP- P- China	CMAQ	InMAP- P- China	CMAQ	InMAP- P- China	CMAQ	InMAP- P- China	InMAP- P- China	
Power	6.9%	8.1%	6.2%	9.4%	7.4%	8.6%	10.4%	8.2%	7.0%	10.0%
Industry	30.8%	35.0%	30.2%	38.2%	33.3%	39.1%	37.5%	35.4%	27.7%	31.9%
Residential	25.9%	28.1%	24.7%	28.2%	17.9%	20.8%	19.5%	28.4%	30.0%	33.8%
Transportation	14.0%	17.3%	13.4%	15.6%	15.7%	21.2%	17.1%	17.5%	13.2%	15.0%
Agriculture	22.5%	11.5%	25.5%	10.4%	25.7%	12.4%	15.4%	11.6%	22.0%	9.4%

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InMAP-China model



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Figure 1 Model framework of InMAP-China.

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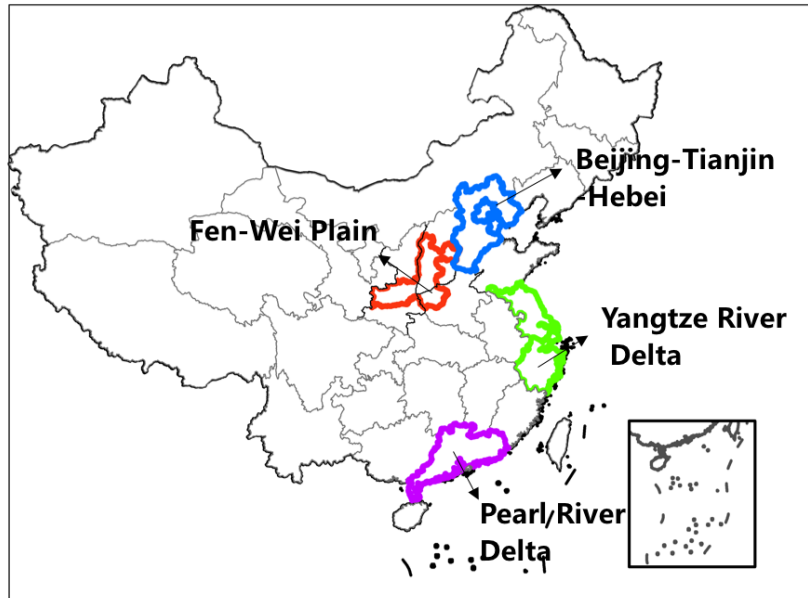
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685 **Figure 2 Four key regions defined in this study, including the Beijing-Tianjin-Hebei region, Yangtze River**
686 **Delta region, Pearl River Delta region and Fen Wei Plain region.**

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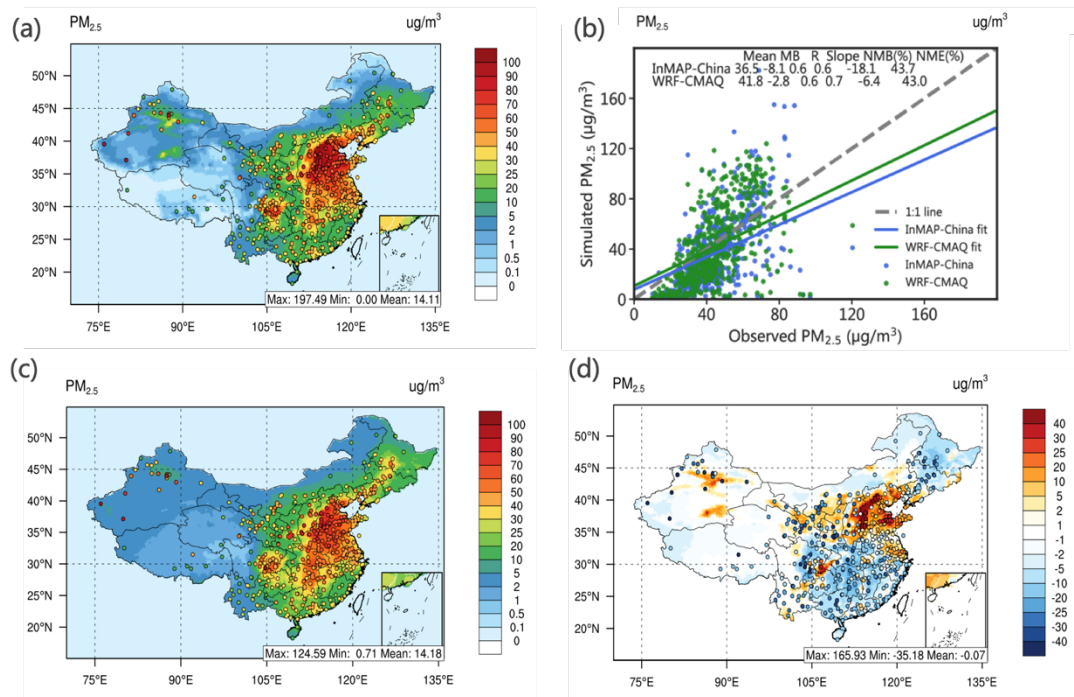
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699 **Figure 3 The spatial pattern and statistical metrics of total PM_{2.5} concentrations predicted by InMAP-China**

700 **and WRF-CMAQ.** Panels (a) and (c) display the spatial patterns of total PM_{2.5} concentrations predicted by InMAP-

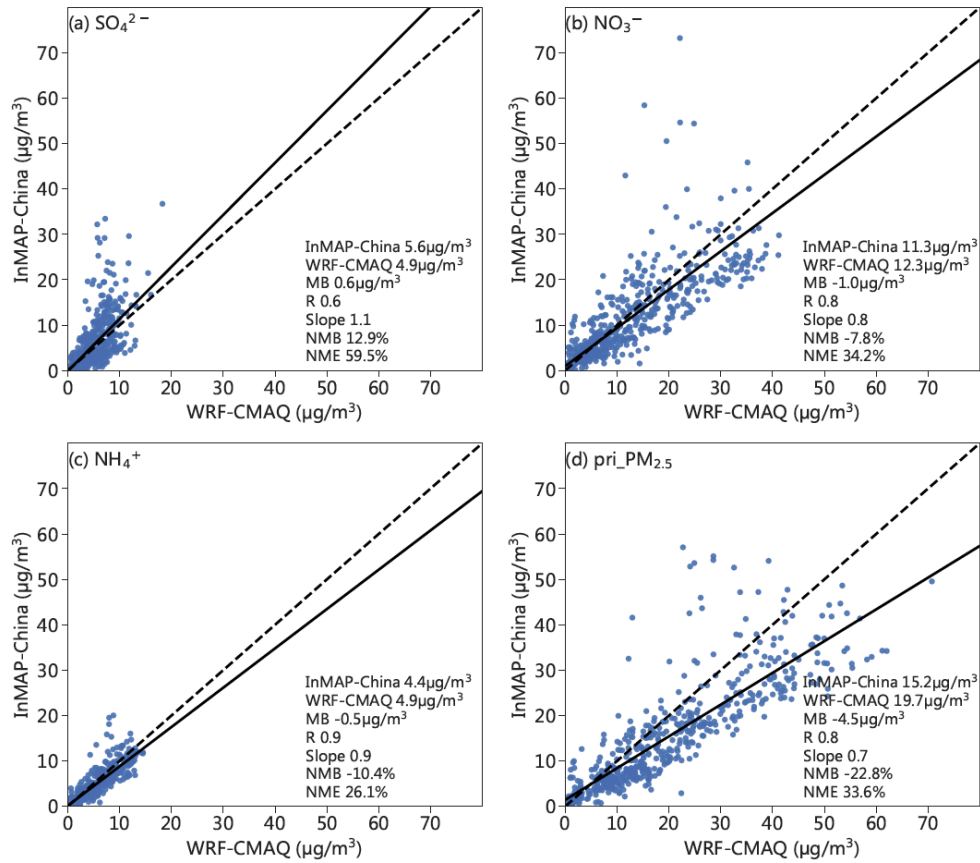
701 China and WRF-CMAQ, respectively. Panel (d) presents the difference in the spatial distribution of the total PM_{2.5}

702 concentrations predicted by the two models. Panel (b) shows the statistical metrics between the simulated and

703 observed PM_{2.5}. The observed total PM_{2.5} concentrations are marked as circles in panel (a) and panel (c). In panel

704 (d), the circle shows the difference between the PM_{2.5} simulated by InMAP-China and the observed PM_{2.5}. The same

705 colorbar is utilized in the contour and the marked circle.

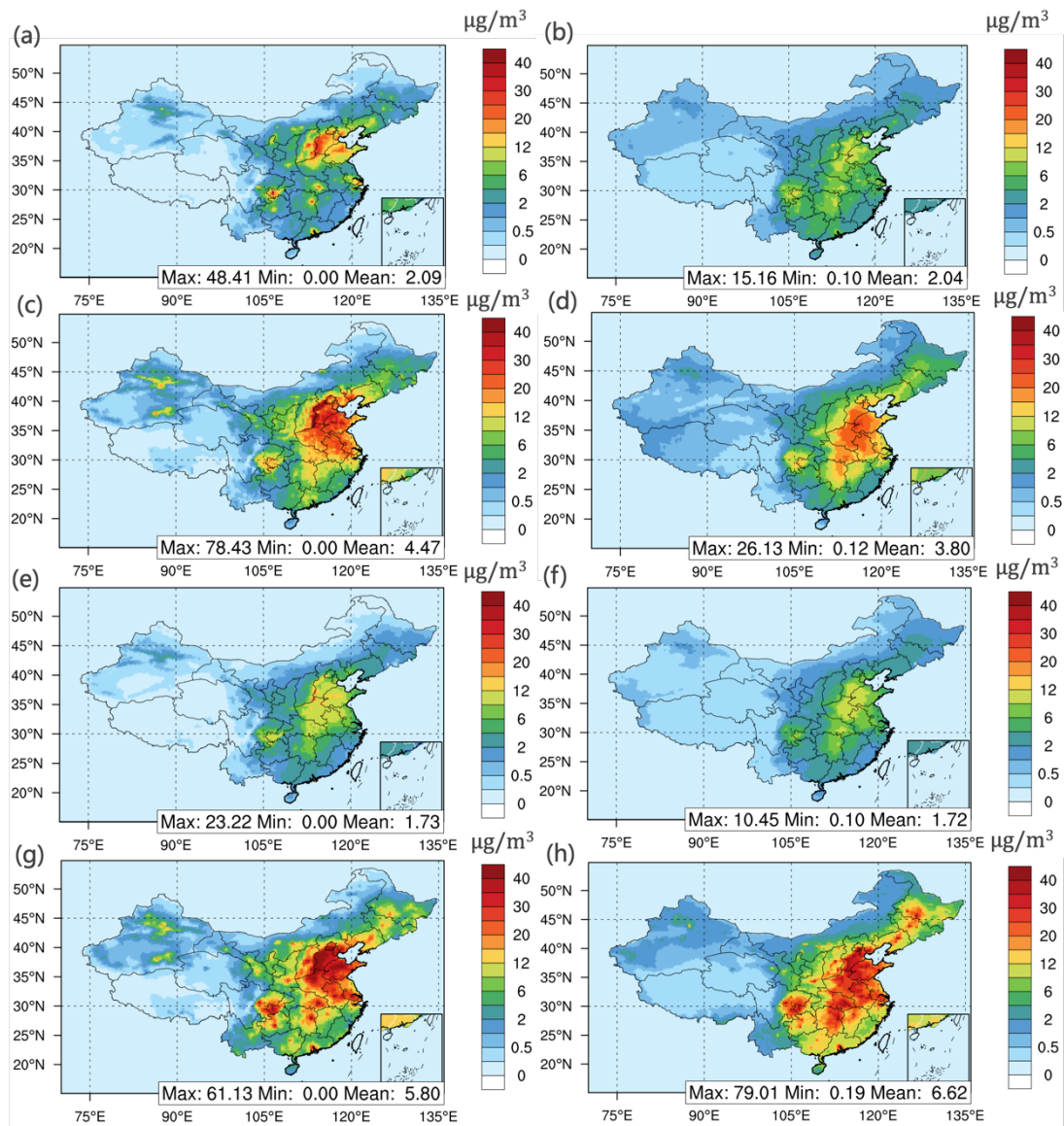


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707 **Figure 4 Scatter plot comparing the PM_{2.5} composition concentration modelled by the InMAP-China and**

708 **WRF-CMAQ models.** Panels (a), (b), (c) and (d) display sulfate, nitrate, ammonium, and primary PM_{2.5},

709 respectively. The statistical metrics are labelled in the lower right corner of each panel.



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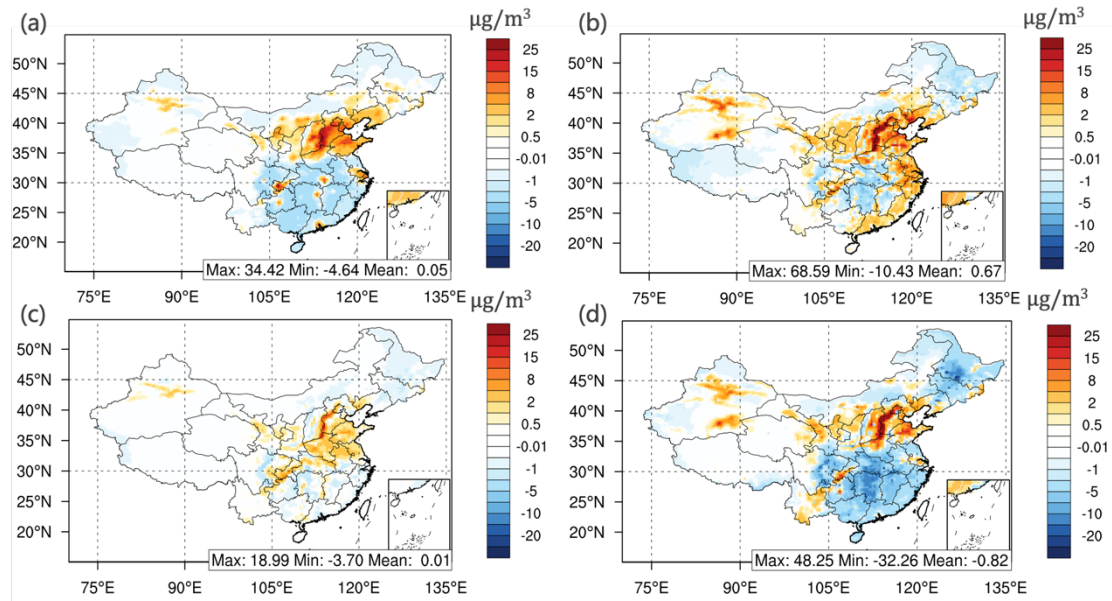
Figure 5 The spatial pattern of PM_{2.5} compositions modelled by the InMAP-China and WRF-CMAQ models.

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Panels (a), (c), (e), and (g) present the sulfate, nitrate, ammonium, and primary PM_{2.5}, respectively, simulated by

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InMAP-China in the InMAP-TOT scenario. Panels (b), (d), (f), and (h) present the results modelled by WRF-CMAQ.



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Figure 6 The difference in the spatial pattern of PM_{2.5} compositions between InMAP-China and WRF-CMAQ.

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Panels (a), (b), (c), and (d) display sulfate, nitrate, ammonium, and primary PM_{2.5}, respectively.

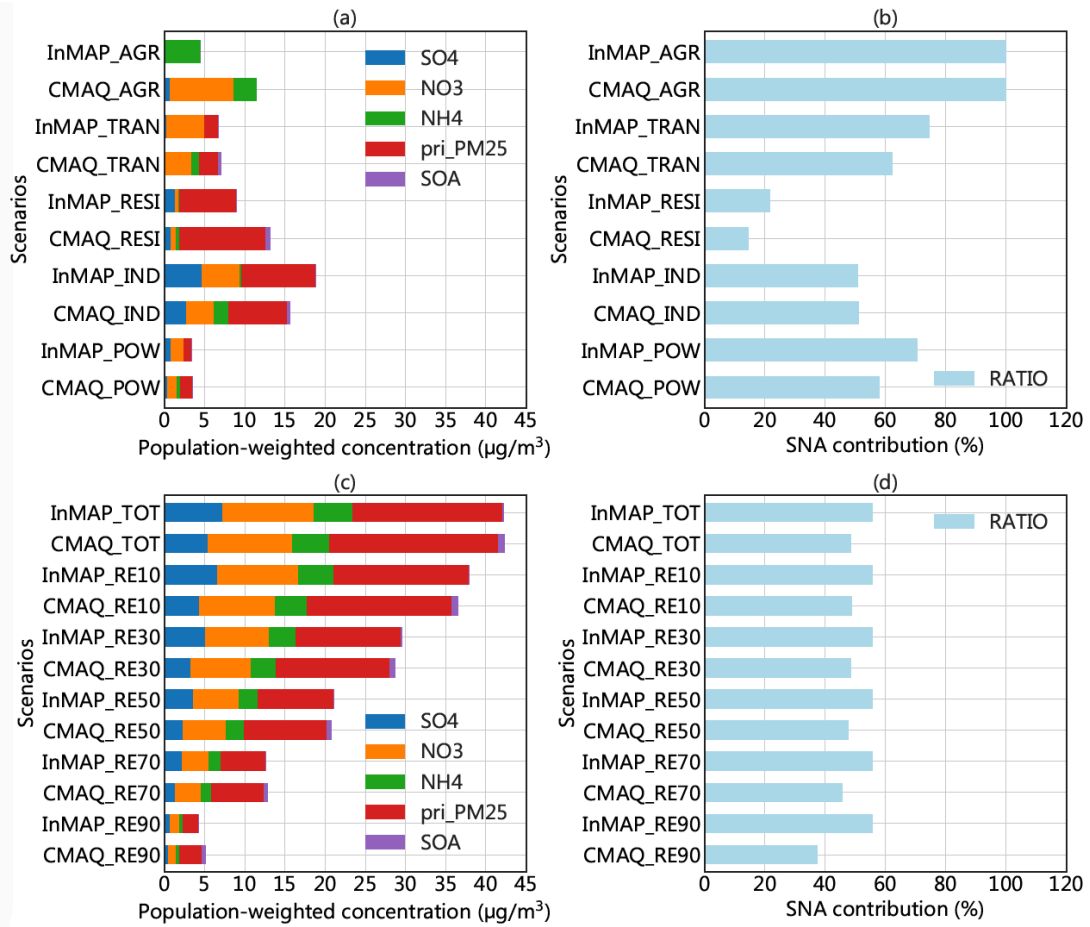
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Figure 7 Comparison of PM_{2.5} component concentrations and SNA contributions in these eleven simulations.

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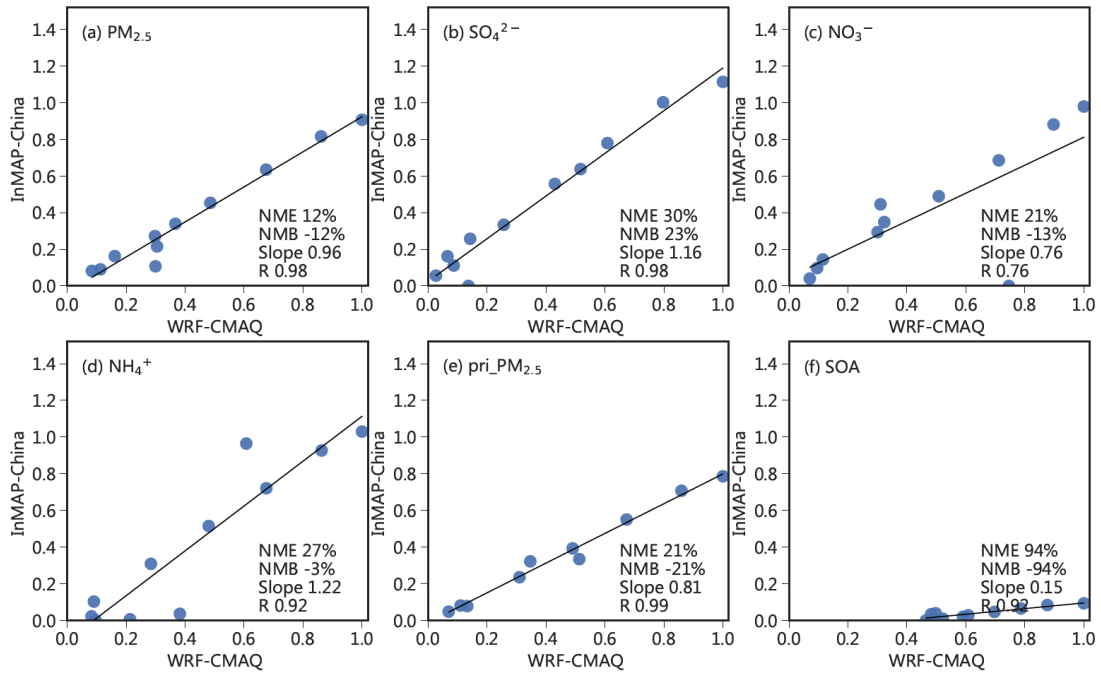
(a) and (c) show the modelled PM_{2.5} compositions. Panel (a) presents the results of sectoral emission scenarios, and

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panel (c) presents the results of the baseline and emission abatement scenarios. Panels (b) and (d) present the SNA

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contribution (%) for each scenario.



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Figure 8 Marginal change in nationwide annual average population-weighted PM_{2.5} concentration and its

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composition as modelled by InMAP-China and WRF-CMAQ for eleven emissions scenarios. The population-

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weighted pollutant concentration for each scenario is normalized using the largest value among all scenarios

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modelled by CMAQ. The eleven dots represent the eleven scenarios, and the statistical metrics are labelled in the

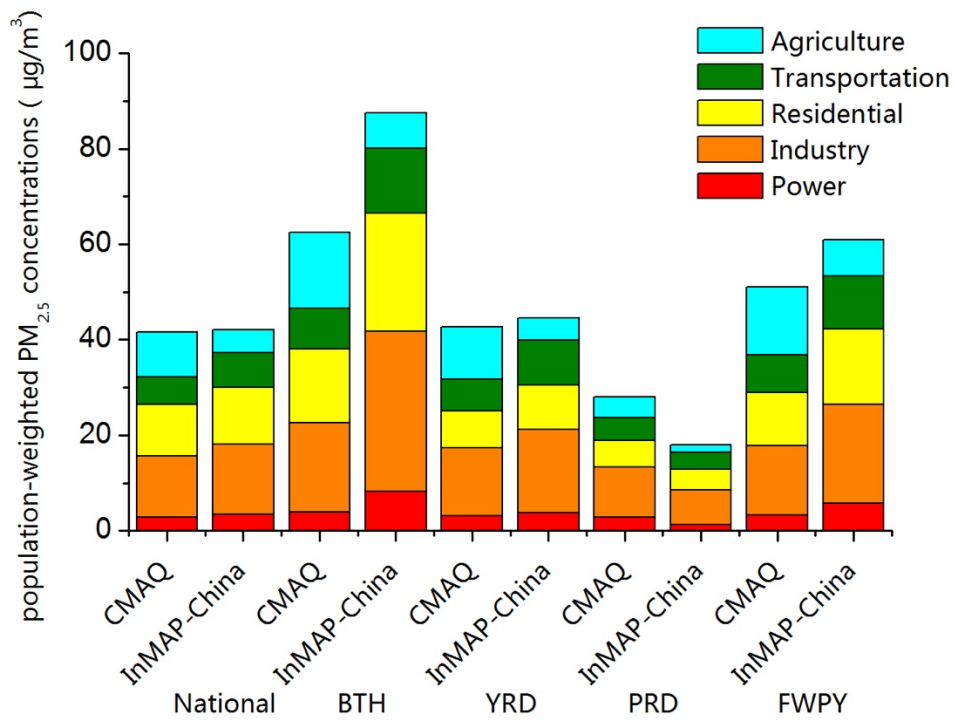
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lower right corner for each panel.

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737 **Figure 9 Comparison of source contributions to population-weighted PM_{2.5} concentrations estimated by the**
 738 **two models.**

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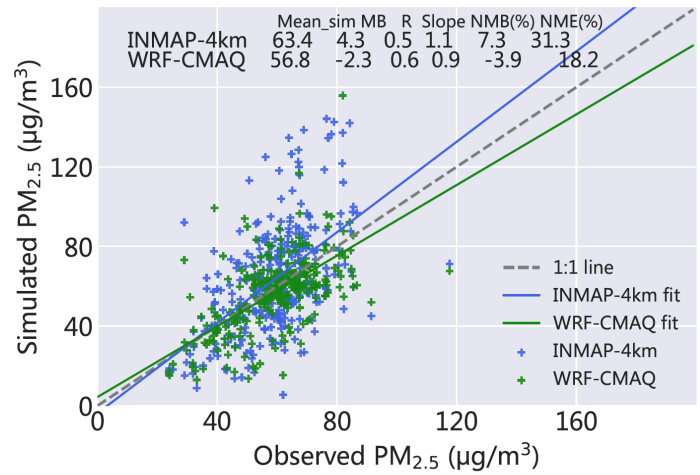
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750 **Figure 10** Scatter plot comparing the $PM_{2.5}$ concentration modeled in the BTH region with 4 km spatial

751 **resolution by the InMAP-China and WRF-CMAQ.** The value of statistical metrics is labeled in the panel.

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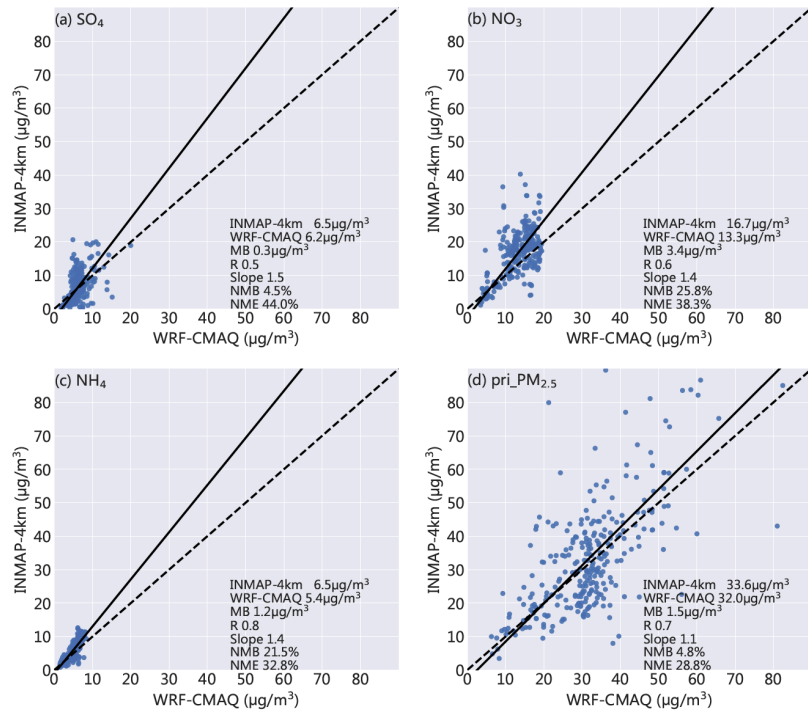
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762 **Figure 11 Scatter plot comparing the $\text{PM}_{2.5}$ composition concentration modeled at BTH region with 4km**

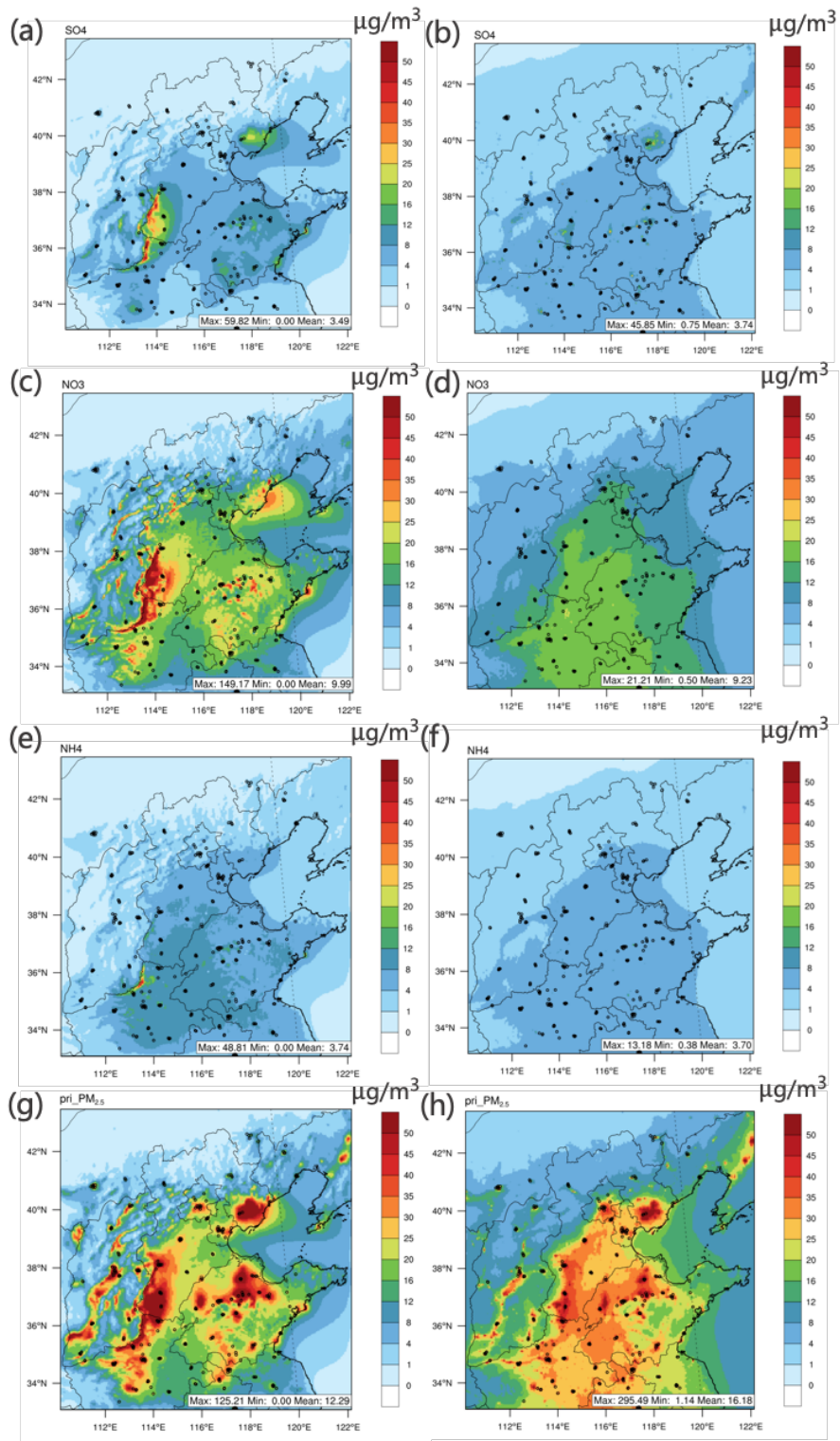
763 **spatial resolution by the InMAP-China and WRF-CMAQ. Panels (a), (b), (c) and (d) display the sulfate, nitrate,**

764 **ammonium, and primary $\text{PM}_{2.5}$, respectively. The statistical metrics are labeled in the lower right corner of each**

765 **panel.**

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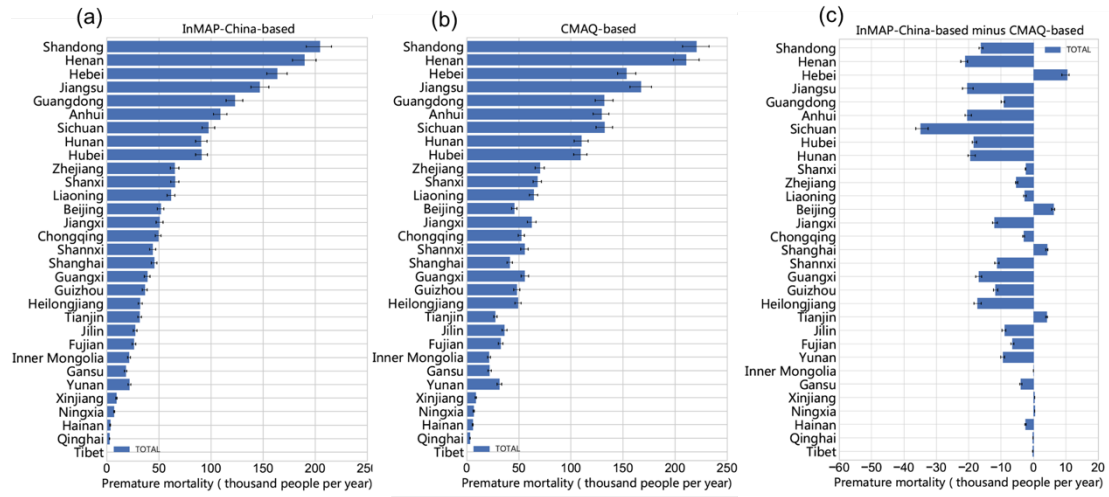
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769 **Figure 12 The spatial pattern of PM_{2.5} compositions simulated in the BTH region with 4km spatial resolution**

770 **by the InMAP-China and WRF-CMAQ. Panels (a), (c), (e), and (g) present the sulfate, nitrate, ammonium, and**

771 **primary PM_{2.5}, respectively, simulated by InMAP-China. Panels (b), (d), (f), and (h) present the corresponding**

772 **results simulated by WRF-CMAQ.**



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774 **Figure 13 Comparison of PM_{2.5}-related premature mortality using the PM_{2.5} predictions from two models.**

775 (a) InMAP-China-based; (b) CMAQ-based; and (c) difference between the two models.

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