

1 **Building a machine learning surrogate model for wildfire activities within a global earth**
2 **system model**

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15 **Abstract**

16 Wildfire is an important ecosystem process, influencing land biogeophysical and
17 biogeochemical dynamics and atmospheric composition. Fire-driven loss of vegetation cover, for
18 example, directly modifies the surface energy budget as a consequence of changing albedo,
19 surface roughness, and partitioning of sensible and latent heat fluxes. Carbon dioxide and
20 methane emitted by fires contribute to a positive atmospheric forcing, whereas emissions of
21 carbonaceous aerosols may contribute to surface cooling. Process-based modeling of wildfires in
22 earth system land models is challenging due to limited understanding of human, climate, and
23 ecosystem controls on fire number, fire size, and burned area. Integration of mechanistic wildfire
24 models within Earth system models requires careful parameter calibration, which is
25 computationally expensive and subject to equifinality. To explore alternative approaches, we
26 present a deep neural network (DNN) scheme that surrogates the process-based wildfire model
27 with the Energy Exascale Earth System Model (E3SM) interface. The DNN wildfire model
28 accurately simulates observed burned area with over 90% higher accuracy with a large reduction
29 in parameterization time compared with the current process-based wildfire model. The surrogate
30 wildfire model successfully captured the observed monthly regional burned area during
31 validation period 2011 to 2015 (coefficient of determination, $R^2 = 0.93$). Since the DNN wildfire
32 model has the same input and output requirements as the E3SM process-based wildfire model,
33 our results demonstrate the applicability of machine learning for high accuracy and efficient
34 large-scale land model development and predictions.

35 **1. Introduction**

36 Wildfires burn ~500 million hectares of vegetated land surface each year, which
37 significantly modifies the physical properties and biogeochemical cycles of terrestrial
38 ecosystems [Andela *et al.*, 2017; Bond-Lamberty *et al.*, 2007; Pellegrini *et al.*, 2018; Randerson
39 *et al.*, 2006]. Living vegetation biomass, surface litter, and coarse woody debris are directly
40 combusted and removed by wildfire [Harden *et al.*, 2006; Walker *et al.*, 2019]. It has been
41 suggested that global forest would double if fire were eliminated [Bond *et al.*, 2005; Lasslop *et al.*,
42 2020]. Fire has multiple important consequences for the climate system, including directly
43 releasing greenhouse gases (*e.g.*, CO₂, CH₄) [Kasischke and Bruhwiler, 2002; Ross *et al.*, 2013]
44 and aerosols [Jiang *et al.*, 2020]; changing land surface albedo and energy budgets [French *et al.*,
45 2016; Rother and De Sales, 2020] and land-atmosphere exchanges of heat, mass, and
46 momentum [Chambers and Chapin, 2002]; limiting plant transpiration and regional water
47 recycling [Brando *et al.*, 2020; Holden *et al.*, 2018]; and reshaping forest composition
48 [Mekonnen *et al.*, 2019]. In addition, biomass burning emits a large amount of fine particulate
49 matter that contributes to about 30% of cloud condensation nuclei globally [Day, 2004]. Soil
50 organic matter decomposition, nitrogen mineralization, and the richness and diversity of soil
51 fungal communities [Oliver *et al.*, 2015] could also be influenced by wildfire through modifying
52 litter substrate supply and degraded enzymatic activities [Bowd *et al.*, 2019; Holden *et al.*, 2018;
53 Pellegrini *et al.*, 2018; Pellegrini *et al.*, 2020].

54 Climate change and land use activities have jointly affected fire spatial distribution,
55 frequency, and intensity [Andela *et al.*, 2017; Kelley *et al.*, 2019; Xu *et al.*, 2020] since the pre-
56 industrial era. For example, warmer and drier climate conditions enhance fuel aridity and favor
57 fire occurrence in forest ecosystems where fuels have built up over a period of decades and
58 centuries [Abatzoglou and Williams, 2016; Williams *et al.*, 2019]. Even if annual precipitation
59 does not decline, redistribution of precipitation towards wet season extreme rainfall events could
60 contribute to longer dry periods and thus more severe fire activity [Xu *et al.*, 2020]. Human
61 activities often shape wildfire activity through regulating patterns of ignition and fire occurrence
62 (*e.g.*, powerline ignition) [Keeley and Syphard, 2018] and suppressing wildfire activity by means
63 of land fragmentation, fire management, and livestock grazing [Andela *et al.*, 2017]. In
64 California, fire density is highly associated with population density and the distance to the
65 wildland urban interface (WUI) [Syphard *et al.*, 2007]. At the global scale, along gradients of

66 increasing population density, fire frequency initially increases by up to 20% and then gradually
67 declines in more densely populated areas [Knorr et al., 2014].

68 Although global wildfire burned area has declined over the recent two decades [Andela et
69 al., 2017], many vulnerable ecosystems and geographic regions have experienced significant
70 increases in wildfire activity [Abatzoglou and Williams, 2016; Walker et al., 2019] resulting in
71 large losses of natural resources and economic assets [Papakosta et al., 2017; Stephenson et al.,
72 2013]. Over western U.S. forests, wildfire has dramatically increased, costing billions of dollars
73 each year and gaining wide public attention. This regional wildfire increase is mainly driven by
74 concurrent increases of spring temperature and declining snowpack [Westerling et al., 2006],
75 mid-summer increases in vapor pressure deficit [Williams et al., 2019], and increases in drought
76 stress during fall [Goss et al., 2020]. The enhancement of wet and dry oscillations favors initial
77 vegetation growth and subsequent wildfire activity [Heyerdahl et al., 2002; Saha et al., 2019].

78 Wildfire models have played an important role in many aspects of wildfire research,
79 including monitoring fire spread [Finney, 1998; Radke et al., 2019], analyzing controllers of
80 wildfire short-term and long-term variability [Kelley et al., 2019], predicting severity of the
81 upcoming fire seasons [Preisler and Westerling, 2007] and climate-scale fire variability
82 [Girardin and Mudelsee, 2008; Yue et al., 2013], and understanding the complex climate-
83 wildfire-ecosystem feedbacks [Clark et al., 2004; Mekonnen et al., 2019; Zou et al., 2020]. Two
84 types of wildfire models are widely used: process-based models and data-driven statistical
85 models. Process-based wildfire models consider detail processes related to natural fire ignition
86 [Prentice and Mackerras, 1977], anthropogenic ignition [Venevsky et al., 2002], fire spread and
87 duration [Thonicke et al., 2010], fire suppression [Lenihan and Bachelet, 2015], and fire mass
88 and heat fluxes [F Li et al., 2012]. Process-based wildfire models have been widely used in
89 dynamic vegetation models and coupled earth system models (ESMs) with various complexities
90 of parameterization [Fang Li et al., 2019; Rabin et al., 2017]. As more and more detailed fire
91 processes are considered and parameterized, structural and parametric uncertainties may increase
92 due to incomplete representation of individual processes and imperfect mathematical formulation
93 [Riley and Thompson, 2017]. Historically, data-driven models were often used for fire behavior
94 modeling and aim to track the ignition, spread, duration, and extinction of individual fires
95 [Finney, 1998; Radke et al., 2019] at fine spatial and temporal scales. This group of models are
96 more relevant to operational fire research. While process-based wildfire models in the context of

97 global vegetation models or earth system land models focuses on the gridcell aggregated fire
98 burned area dynamics that are more relevant to researches on large scale patterns and climate
99 scale predictions [Fang Li et al., 2019; Rabin et al., 2017]. This study particularly focuses on the
100 second category of wildfire models.

101 Although explicit processes are simulated, the accuracy of process-based wildfire models
102 are highly dependent on parameterization, which is computationally expensive [Teckentrup et
103 al., 2018; Zhu and Zhuang, 2014]. Data-driven models, however, directly link the driving
104 variables (e.g., climate factors) to the fire activity using simple statistical models or more
105 sophisticated machine learning techniques, ignoring the explicit processes and feedbacks
106 associated with wildfire [Ganapathi Subramanian and Crowley, 2018; Radke et al., 2019; Tonini
107 et al., 2020]. Through training and validation, statistical representations of wildfire dynamics are
108 learned by models using principles from machine learning. Data-driven wildfire models are
109 diverse in terms of driving variables and model structure. For example, many current machine
110 learning wildfire models rely on remote oceanic dynamics (e.g., sea surface temperature
111 variability) and atmospheric teleconnections to simulate land surface fire activities [Chen et al.,
112 2020; Chen et al., 2011; Yu et al., 2020]. Another group of data-driven wildfire models draws
113 more heavily upon regional climate, plant functional type, and human infrastructure driver
114 variables [Coffield et al., 2019; Sayad et al., 2019].

115 In this study, we develop a machine learning wildfire model using the process
116 representation of wildfire in the Energy Exascale Earth System Model (E3SM) land model
117 (ELMv1) [Zhu et al., 2019], five observationally inferred burned area products [Andela et al.,
118 2019; Giglio et al., 2018; Joshua Lizundia-Loiola et al., 2020; J Lizundia-Loiola et al., 2018;
119 Van Der Werf et al., 2017], and a deep neural network approach [Goodfellow et al., 2016]. We
120 implemented a deep learning model that can better capture the complex and non-linear
121 interactions between controlling factors and wildfire activity. The objectives of this study are to
122 surrogate the wildfire parameterization in ELMv1 with the deep neural network and improve the
123 model simulated wildfire burned area across various fire regions [Giglio et al., 2013].

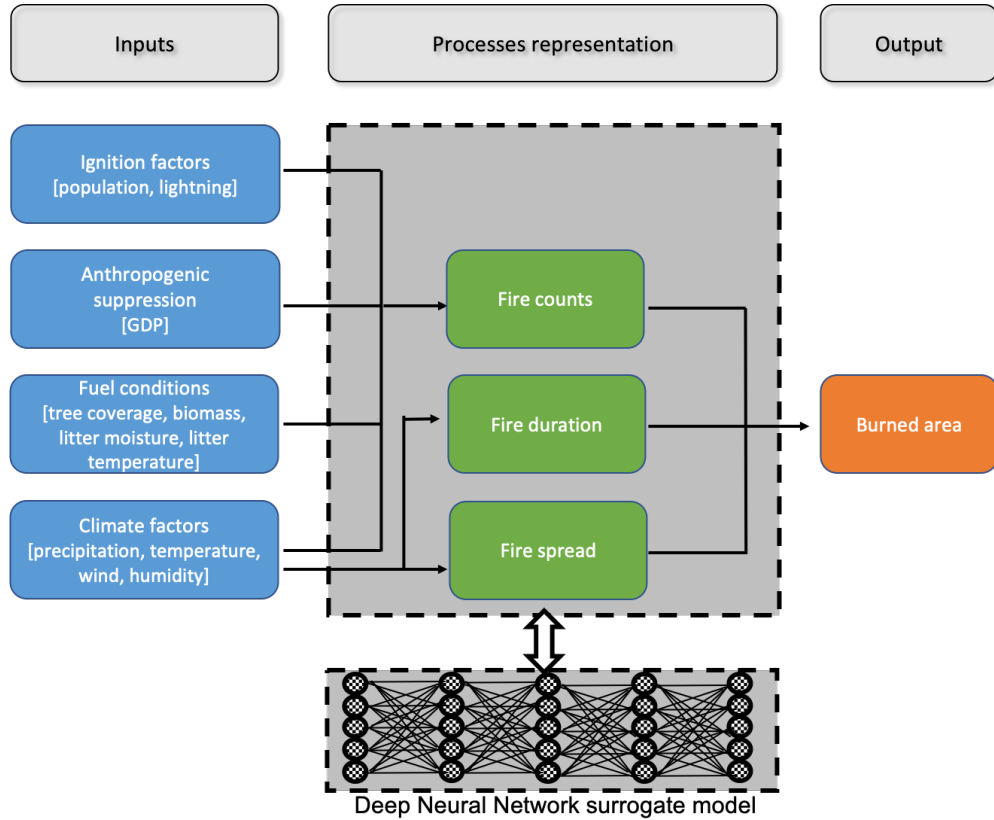
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125 **2. Methodology**

126 **2.1 ELMv1 wildfire model**

127 The process-based wildfire model in ELMv1 originates from the Community Land
128 Model (CLM4.5) [*F Li et al.*, 2012]; we take this wildfire model as the baseline (hereafter refer
129 to as BASE-Fire). BASE-Fire combines information regarding ignition, fuel conditions, surface
130 climate, and anthropogenic suppression to simulate total burned area based on the fire counts and
131 spread area of each fire (Figure 1). The fire count in BASE-Fire is modeled as the sum of
132 anthropogenic ignition and natural ignition, where the latter is proportional to lightning density
133 [*Prentice and Mackerras*, 1977] and the former is determined by population density [*Venevsky et*
134 *al.*, 2002]. Human activity may also intentionally suppress wildfire occurrence if the fire is
135 detected at early stage. For example, developed regions with high population density and gross
136 domestic product are less likely to use fire to remove surface biomass. On the other hand,
137 developed regions more likely suppress fire given more effective fire management policy and
138 suppression capability. Fire count is also affected by surface fuel availability (aboveground
139 biomass) and fuel combustibility (relative humidity, topsoil temperature and moisture). The fire
140 spread area in BASE-Fire is modeled as an elliptical shaped region controlled by wind speed and
141 fuel wetness (using topsoil (0 – 15 cm) moisture as a proxy). The fire duration is set to be one
142 day based on a study that reported years 2001-2004 mean global fire persistence [*Giglio et al.*,
143 2006a]. BASE-Fire also does not explicitly consider roads, rivers, and firefighting activity
144 [*Arora and Boer*, 2005].

145



146

147 **Figure 1.** Schematic representation of the ELMv1 process-based BASE-Fire model and the
 148 components to be surrogated with the Deep Neural Network (DNN) model (dark grey).

149

150 2.2 Deep neural network wildfire surrogate model

151 We developed the new fire model in two steps: (1) surrogating BASE-Fire with a deep
 152 neural network (DNN) approach and (2) improving that surrogate model using five
 153 observationally inferred burned area products (Table S1). First, we surrogated BASE-Fire with a
 154 DNN approach (hereafter refer to as DNN-Fire) that uses the same input and output variables as
 155 BASE-Fire but treats the explicit intermediate processes (*e.g.*, ignition, fire spread) as latent
 156 variables coded by hidden layers in the DNN (Figure 1). DNN-Fire was developed with five
 157 hidden layers and five neurons in each layer for burned area simulation. The DNN approach uses
 158 a fully-connected feedforward neural network [Schmidhuber, 2015] that comprises input, hidden,
 159 and output layers:

$$h_1 = f_1(W_1 I + b_1) \quad (1)$$

$$h_2 = f_2(W_2 h_1 + b_2) \quad (2)$$

$$h_3 = f_3(W_3 h_2 + b_3) \quad (3)$$

$$h_4 = f_4(W_4 h_3 + b_4) \quad (4)$$

$$h_5 = f_5(W_5 h_4 + b_5) \quad (5)$$

$$O = f_6(W_6 h_5 + b_6) \quad (6)$$

160 where I denotes the input layer (*e.g.*, climate factors) with 11 neurons, each corresponding to an
 161 input variable listed in Table 1. h_1 , h_2 , h_3 , h_4 , and h_5 are five hidden vectors that are calculated
 162 with two steps. First is a linear combination of previous layers' input vector (h) and the trainable
 163 weight parameter matrix [W_1 , W_2 , W_3 , W_4 , W_5 , W_6], considering biases b_1 , b_2 , b_3 , b_4 , b_5 , and b_6 .
 164 Then, nonlinear activation functions f_1 , f_2 , f_3 , f_4 , f_5 , and f_6 . are applied to the output from the
 165 previous step. In this study we used *softplus* as the activation function [Zheng *et al.*, 2015] that is
 166 a non-linear transformation of input signals. O denotes the output layer that summarize the latent
 167 variables from the last hidden layer (h_5) and calculate burned area.

168 **Table 1.** Input and output variables of ELMv1 BASE-Fire and surrogate DNN-Fire models

| Variable category | Variable name | Data source and reference |
|----------------------------------|--------------------------|---|
| <i>Input variables</i> | | |
| Fuel conditions | Tree coverage | LUH2 [Hurtt <i>et al.</i> , 2020] |
| | Fuel load | ELMv1 total biomass [Zhu and Riley, 2015; Zhu <i>et al.</i> , 2019] |
| | Fuel wetness | ELMv1 topsoil moisture [Zhu and Riley, 2015; Zhu <i>et al.</i> , 2019] |
| | Fuel temperature | ELMv1 topsoil temperature [Zhu and Riley, 2015; Zhu <i>et al.</i> , 2019] |
| Climate factors | Precipitation | GSWP3 [Dirmeyer <i>et al.</i> , 2006] |
| | Near surface temperature | GSWP3 [Dirmeyer <i>et al.</i> , 2006] |
| | Wind speed | GSWP3 [Dirmeyer <i>et al.</i> , 2006] |
| | Relative humidity | GSWP3 [Dirmeyer <i>et al.</i> , 2006] |
| Ignition | Population density | [Dobson <i>et al.</i> , 2000] |
| | Lightning frequency | NASA-LIS/OTD [Cecil <i>et al.</i> , 2014] |
| Anthropogenic suppression | GDP | [van Vuuren <i>et al.</i> , 2007] |
| | Population density | [Dobson <i>et al.</i> , 2000] |

| <i>output variable</i> | |
|------------------------|--|
| Burned area | ELMv1 percentage burned area [Zhu and Riley, 2015; Zhu et al., 2019] |

169

170 Second, we improved the surrogate DNN-Fire by fine-tuning the weight parameters using
 171 observations (hereafter refer to DNN-Fire-OBS). Between 2001 and 2010, we initialized
 172 DNN-Fire-OBS’s weight parameters (W_1 , W_2 , W_3 , W_4 , W_5 , and W_6) using results from DNN-Fire,
 173 replaced the BASE-Fire burned area by the ensemble mean of five observationally inferred
 174 burned area products including GFEDv4s [Van Der Werf et al., 2017], Fire_CCI51 [Joshua
 175 Lizundia-Loiola et al., 2020], Fire_CCILT11 [J Lizundia-Loiola et al., 2018], MODIS MCD64
 176 [Giglio et al., 2018], and Fire_Atlas [Andela et al., 2019] (Table S1), and adjusted weight
 177 parameters until the model best reproduced the observed burned area. This two-step approach
 178 will also allow rapid parameterization of the Fire model as new fire data and baseline fire model
 179 results become available. DNN-Fire-OBS can be more easily generalized since BASE-Fire
 180 provides explicit physical guidance and a larger-than-observation input and output feature space
 181 for development of the machine learning fire model.

182 **2.3 Model setup and simulation protocol**

183 We ran ELMv1 with BASE-Fire at 1.9° by 2.5° spatial resolution [Zhu et al., 2020; Zhu
 184 et al., 2016] to generate training and testing datasets for the DNN wildfire model. BASE-Fire
 185 was first spun up for 600 years with accelerated soil decomposition followed by 200 years
 186 regular spinup with regular soil decomposition [Koven et al., 2013]. The spinup simulations were
 187 forced with constant atmospheric CO₂ concentration (285 ppmv) and 1901-1920 repeated
 188 climate forcing from GSWP3 (Global Soil Wetness Project) [Dirmeyer et al., 2006]. The purpose
 189 of the spinup was to initialize ecosystem carbon pools and stabilize plant and soil carbon and
 190 water fluxes. Restarting from the “spinup” conditions, a transient simulation was then conducted
 191 from 1901 to 2015 with GSWP3 transient climate forcing, atmospheric CO₂ concentrations, and
 192 nitrogen and phosphorus deposition [Lamarque et al., 2005; Mahowald et al., 2008]. Wildfire
 193 associated variables were selected for output with a monthly temporal resolution (Table 1).

194 BASE-Fire output from years 1981 to 2010 were used to train, test, and fine-tune
 195 DNN-Fire. We developed 14 region-specific models, corresponding to 14 widely used GFED
 196 regions. For each region, all land gridcells (comprising no fire history, infrequent fire, and

197 repeated fire) were concatenated into one data matrix (where rows consist of the number of
 198 samples and columns of the number of variables). 80% of the data matrix was randomly sampled
 199 for the training dataset and the remaining 20% of the data were reserved for testing. Furthermore,
 200 the random sampling was stratified in order to reduce the risk of sampling, e.g., adjacent high
 201 fire grid cells. All grid cells were first divided into three “strata”: low burn (0-33% percentile),
 202 median burn (33%-66% percentile), and high burn (67-100% percentile) grid cells based on the
 203 magnitude of the burn. The stratified random sample assured the sampled grid cells for training
 204 and testing had the same ratios of low/medium/high burn, thus eliminating the sampling bias
 205 from spatial autocorrelation [Wang *et al.*, 2012]. In addition to random sampling, we also
 206 investigated the impacts of data choice on the model performance, by sampling the testing
 207 datasets within specific years (*e.g.*, 2001-2002, 2003-2004, 2005-2006, 2007-2008, 2009-2010)
 208 and used the rest of the years for training. We found neglected differences among the models
 209 (Figure S1) indicating the choice of training/testing data years were not impactful. Therefore, we
 210 will discuss the results with stratified random sampling approach as the major results throughout
 211 the paper.

212 All training and testing datasets were normalized to the range [0, 1] with the following
 213 scaler:

$$X = \frac{X - X_{min}}{X_{max} - X_{min}} \quad (7)$$

214 where X is the variable vector of interest and X_{min} and X_{max} are minimum and maximum values of
 215 X , respectively. During the training stage, we randomly initialized the weighting parameters (Eq.
 216 1-6) and optimized them using the Adaptive Moment Estimation method [Kingma and Ba,
 217 2014], which is a variant of the gradient descent optimization method but considers adaptive
 218 learning rate and momentum-like exponentially decaying gradients. The parameter optimization
 219 aimed to minimize a mean squared error cost function:

$$J = \frac{1}{n} \sum_{i=1}^n (y_i^{DNN} - y_i^{BASE})^2 \quad (8)$$

220 where y_i^{DNN} and y_i^{BASE} are DNN-Fire and BASE-Fire generated burned area, respectively. Cost
 221 function J summarizes the overall magnitude of the error between the surrogate DNN-Fire and
 222 BASE-Fire. We then evaluated model performance using metrics of mean absolute error (Eqn.
 223 9), Pearson correlation (Eqn. 10), and coefficient of determination (Eqn. 11).

$$MAE = \frac{1}{n} \sum_{i=1}^n |y_i^{DNN} - y_i^{BASE}| \quad (9)$$

225
$$p = \frac{\text{covariance}(y^{DNN}, y^{BASE})}{\text{variance}(y^{DNN})\text{variance}(y^{BASE})} \quad (10)$$

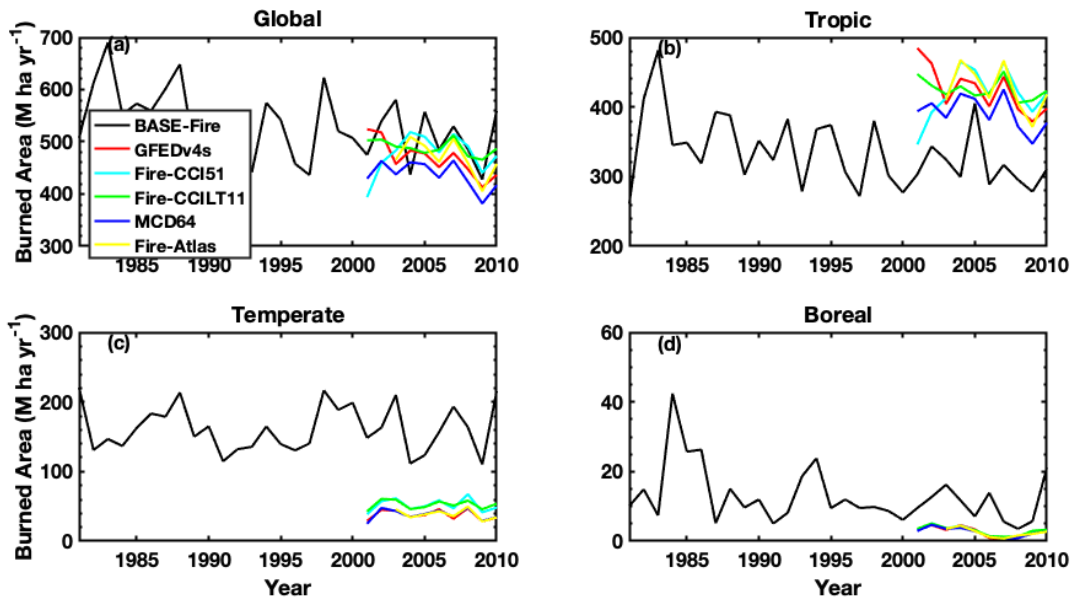
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$$R^2 = 1 - \frac{\sum_{i=1}^n (y_i^{DNN} - y_i^{BASE})^2}{\sum_{i=1}^n (y_i^{BASE} - y_{mean}^{BASE})^2} \quad (11)$$

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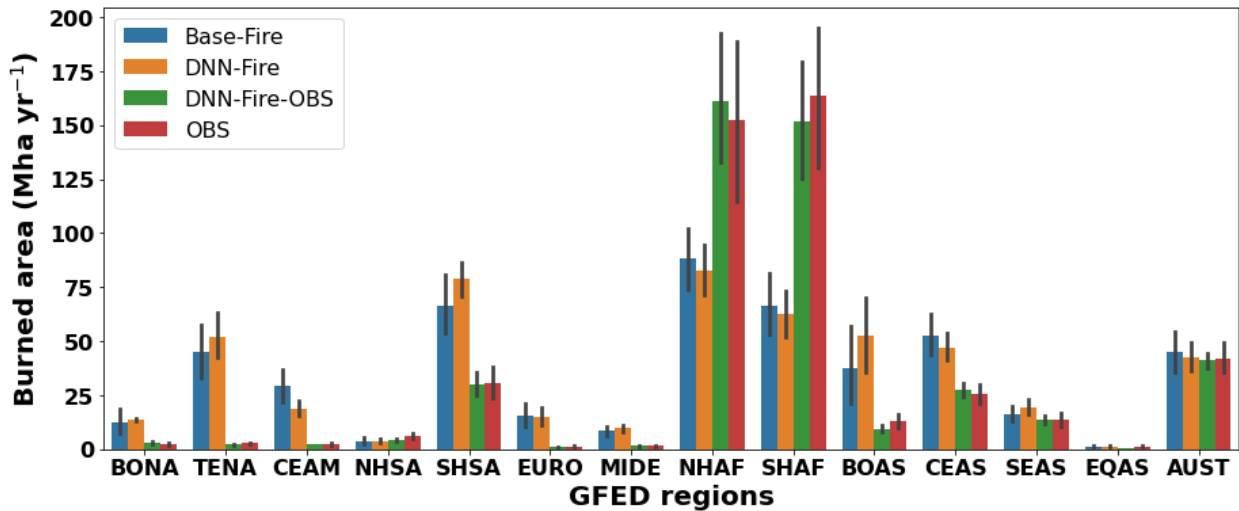
228 **3. Results and discussion**

229 **3.1 Evaluation of wildfire surrogate model**

230 BASE-Fire performed reasonably well for total global burned area (508 ± 53 Mha yr⁻¹
 231 (million hector per year) between years 2001 and 2010 compared with the observational long-
 232 term average of 424~484 Mha yr⁻¹; Figure 2, Table S1). BASE-Fire also captured the global
 233 declining trend of wildfire burned area over this time period, attributed to a decrease in tropical
 234 fires [Andela *et al.*, 2017]. At the regional scale, however, BASE-Fire underestimated tropical
 235 (S23.5° - N23.5°) burned area and overestimated temperate (N23.5° - N67.5°) and boreal (N67.5°
 236 above) burned area (Figure 2). Large spatial heterogeneity existed for BASE-Fire regional bias.
 237 For example, over tropical GFED regions, BASE-Fire overestimated wildfire burned area over
 238 Southern Hemisphere South America (SHSA), but underestimated wildfire burned area over both
 239 Southern and Northern Hemisphere Africa regions (SHAF and NHAF), despite an overall
 240 underestimation over the tropical region (Figure 3). In contrast, consistent overestimation
 241 occurred over all temperate GFED regions. For example, wildfire burned was overestimated by
 242 about a factor of 16 (~1 versus 16 Mha yr⁻¹) over the Europe GFED region (EURO) (Figure 3).
 243 Although there is room to improve BASE-Fire performance, the parameterization would involve
 244 large ensemble simulations and computational resources. Instead, we first used BASE-Fire
 245 generated data as training and testing datasets to parameterize DNN-Fire, then we fine-tuned the
 246 DNN-Fire model against observed burned area.



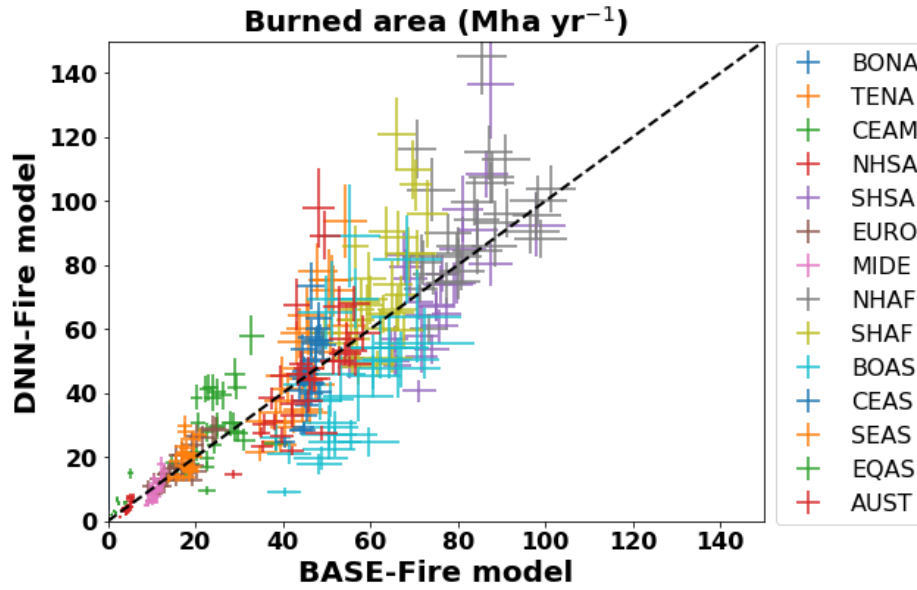
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 248 **Figure 2.** ELMv1 process-based model (BASE-Fire) simulated and five observationally inferred
 249 burned area products (Table S1) at (a) global scale; (b) Tropical (S23.5° -N23.5°); (c) Temperate
 250 (N23.5° - N 67.5°); and (d) Boreal (north of N 67.5°) regions.
 251



252
 253 **Figure 3.** A comparison of wildfire burned area between estimates from the ELMv1 process-
 254 based model (BASE-Fire), Deep Neural Network wildfire model (DNN-Fire), Deep Neural
 255 Network wildfire model fine-tuned with observed burned area (DNN-Fire-OBS), and
 256 observations over 14 GFED fire regions.
 257

258 Next we parameterized and compared DNN-Fire with BASE-Fire outputs. Using BASE-
259 Fire generated $1.9^\circ \times 2.5^\circ$ resolution datasets of surface fuel conditions (fuel load (vegetation
260 biomass), fuel temperature (topsoil temperature), and fuel wetness (topsoil moisture)) with
261 gridded climate forcing (GSWP3) [Dirmeyer *et al.*, 2006], land use (LUH2 dataset) [Hurtt *et al.*,
262 2020], and social economic [Dobson *et al.*, 2000; van Vuuren *et al.*, 2007] factors, DNN-Fire
263 captured the spatial pattern of BASE-fire predicted wildfire activity (Figure 4, Figure S2).
264 Across all GFED regions, mean absolute error of DNN-Fire was 4.4 Mha yr^{-1} (<1% of total burn
265 area), with median and maximum errors of 1.8 and 13.0 Mha yr^{-1} , respectively (Figure 3).
266 Equatorial Asia (EQAS), Northern Hemisphere South America (NHSA), Central America
267 (CEAS), and Europe (EURO) regions had the lowest DNN-Fire errors (< 1.0 Mha yr^{-1}), while
268 Southern Hemisphere Africa (SHAF), and Boreal Asia (BOAS) had the largest errors (10-13
269 Mha yr^{-1}). Overall, the correlation coefficient between BASE-Fire and DNN-Fire simulated
270 burned area was 0.91 (p value < 0.01) and the coefficient of determination (R^2) was 0.79. Across
271 seasons, DNN-Fire also reasonably captured the BASE-Fire peak fire months (June to October),
272 which were dominated by Southern Hemisphere Africa and Southern Hemisphere South
273 America (Figure 5).

274 By surrogating BASE-Fire, DNN-Fire is expected to have similar biases and
275 uncertainties. The deficiency of BASE-Fire model will propagate to DNN-Fire. In our future
276 work we will overcome such limitation by training multiple DNN-Fire models with ensemble
277 simulations of BASE-Fire models that differ in critical parameters and vary in model structures.
278



279

280 **Figure 4.** The performance of the Deep Neural Network wildfire model (DNN-Fire), compared
 281 with the original ELMv1 process-based wildfire model (BASE-Fire) over 14 GFED regions
 282 between years 2001 and 2010.

283

284 3.2 Calibrating the wildfire surrogate model using observations

285 Although the global pattern was reasonably captured, BASE-Fire had relatively large
 286 biases in several GFED regions, as discussed above. Since DNN-Fire was trained and validated
 287 only with BASE-Fire generated inputs (*e.g.*, fuel conditions) and outputs (burned area), we
 288 expect that, at best, DNN-Fire would have comparable biases as BASE-Fire. Starting from
 289 DNN-Fire, we further calibrated the model weighting parameters to create DNN-Fire-OBS and
 290 validated DNN-Fire-OBS performance using observed burned area from five existing burned
 291 area products (Table S1) between years 2001 and 2010. The calibration time cost several minutes
 292 with Intel Xeon Phi Processor 7250 processor.

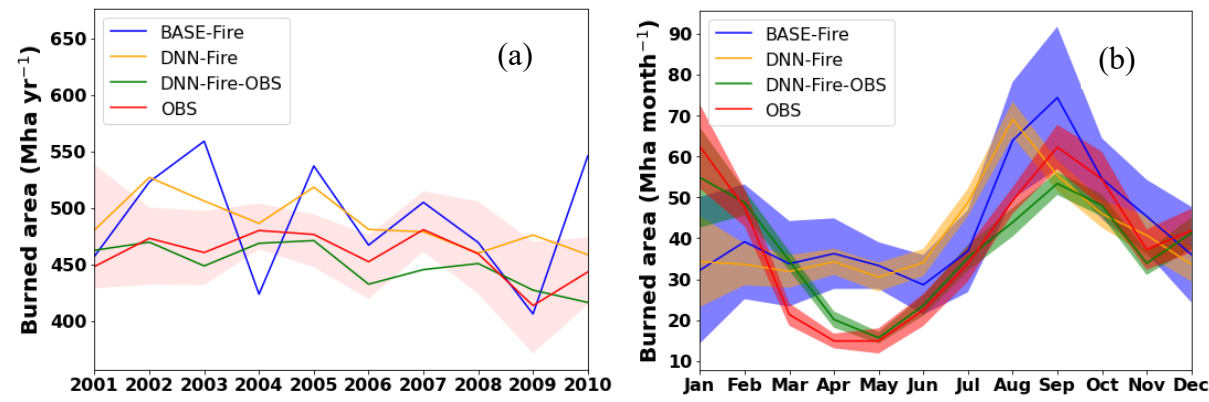
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294 Dramatic improvements were found in most of the 14 regions simulated by DNN-Fire-
 295 OBS (Figure 3). Overall, DNN-Fire-OBS simulated global long-term average burned area was
 296 458 Mha yr⁻¹ (compared with observational average 467 Mha yr⁻¹). Averaged across 14 regions,
 297 73% reduction of mean absolute error was achieved by DNN-Fire-OBS, compared with the
 298 BASE-Fire model. Pearson correlation coefficient between the DNN-Fire-OBS simulated and
 299 observational burned area was 0.98 (p value < 0.001) with an R^2 of 0.97. Bias reduction was
 disproportionately distributed across the GFED regions (Figure 3). For example, severely burned

300 regions, including Southern and Northern Hemisphere Africa (SHAF and NHAF) and Southern
 301 Hemisphere South America (SHSA) greatly benefited from the tuning and their regional biases
 302 were reduced by 88, 65, and 51 Mha yr⁻¹ (or 88%, 89%, 98% reduction), respectively. Although
 303 Temperate Northern America (TENA) and Europe (EURO) wildfire burned area is relatively
 304 small (1-3 Mha yr⁻¹), the impacts of wildfire activity were significant due to their high population
 305 densities. DNN-Fire tended to overestimate the burned area in TENA and EURO by 47 and 13
 306 Mha yr⁻¹, while DNN-Fire-OBS significantly reduced biases in both regions to less than 0.3 Mha
 307 yr⁻¹ (a 97-98%% reduction).

308 BASE-Fire tended to overestimate inter-annual variability (IAV) and had opposite burned
 309 area anomalies between years 2001 and 2005. DNN-Fire dampened BASE-Fire's IAV, but had
 310 systematic overestimation of burned area. DNN-Fire-OBS agreed well with the observed IAV
 311 between years 2001 and 2010 (Figure 5a). The seasonal cycle was also improved in DNN-Fire-
 312 OBS in terms of reducing BASE-Fire's overestimation of burned area during peak fire seasons
 313 (Figure 5b, Figure S3), although we note that DNN-Fire-OBS is biased high during low fire
 314 seasons (March and April).

315



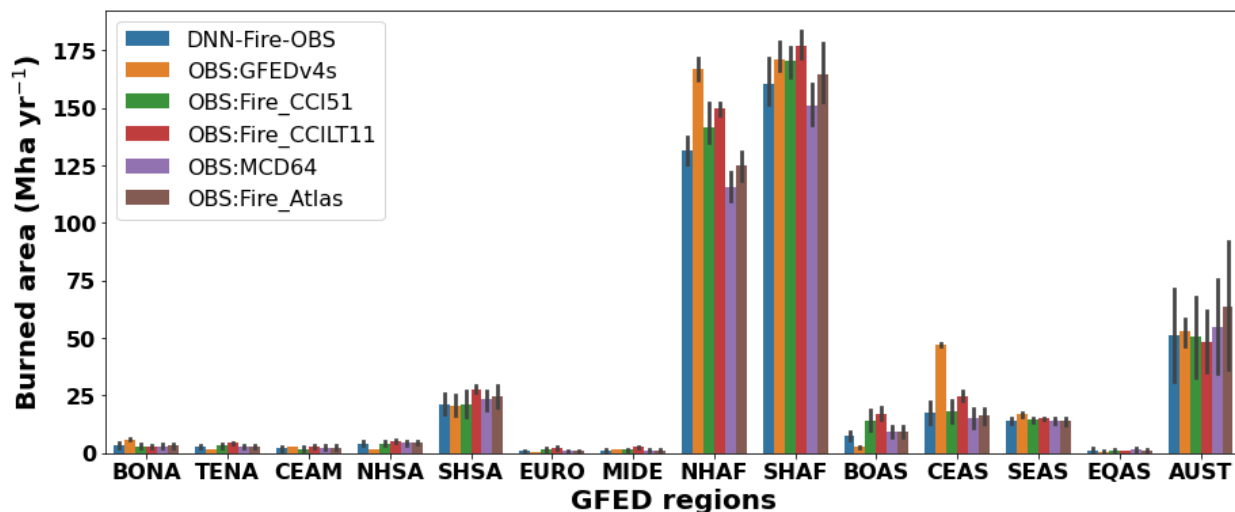
316
 317 **Figure 5.** Inter-annual variation of burned area from years 2001 to 2010 (a) and the averaged
 318 seasonal cycle (b) of burned area estimated by the ELMv1 process-based wildfire model (BASE-
 319 Fire), Deep Neural Network wildfire model (DNN-Fire), Deep Neural Network wildfire model
 320 fine-tuned with observations (DNN-Fire-OBS), and observations.

321

322 3.3 Prognostic simulation and limitations

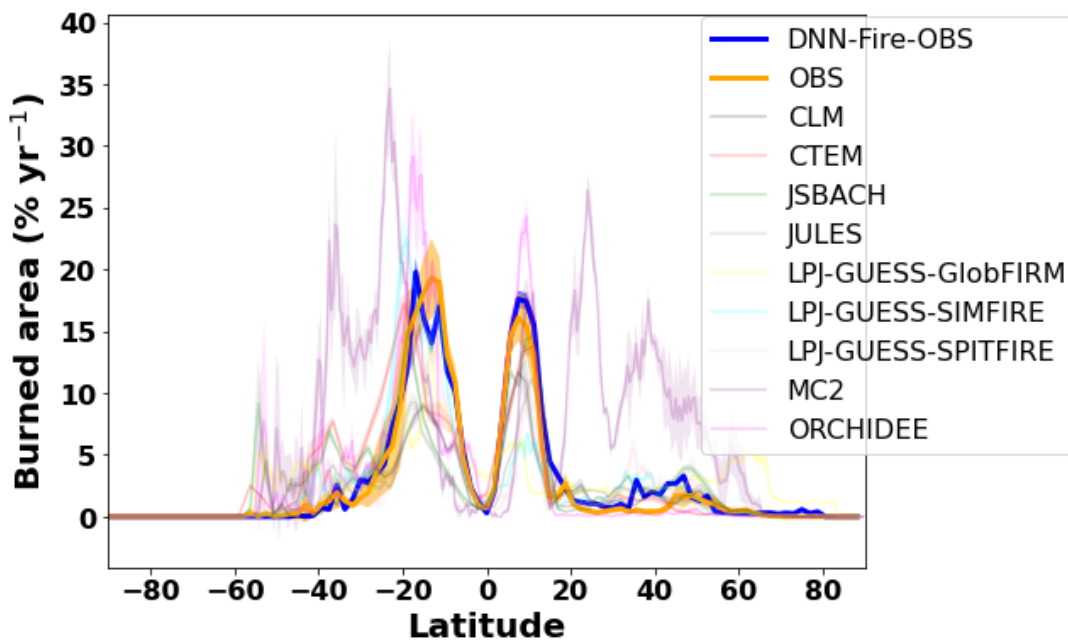
323 We next evaluated the DNN-Fire-OBS model against observations for the period 2011 to
 324 2015, using data which were not used to train and validate the model. Overall, DNN-Fire-OBS

325 simulated 469-514 Mha yr⁻¹ global burned area, compared with observations 349-509 Mha yr⁻¹.
 326 Note that the large observational ranges were mainly due to the differences among the five
 327 burned area products rather than the inter-annual variability (Figure 6). Regionally, DNN-Fire-
 328 OBS overestimated NHAF, SHAF and SHSA annual burned area by 8, 6, 2 Mha yr⁻¹,
 329 respectively (Figure 6) compared with the observational mean. Averaged latitudinal distribution
 330 of simulated burned area during this period showed that global wildfire activity peaked around
 331 S10°- S15° and N5°-N10°, together accounting for burning 12-16% of the land surface (Figure
 332 7). These two peaks were dominated by large burned area over Southern (SHAF) and Northern
 333 Hemisphere Africa (NHAF) fire regions. Compared with observations, DNN-Fire-OBS
 334 simulated reasonable burned area latitudinal distributions (Figure 7). We also compared the nine
 335 FireMIP models [*Rabin et al., 2017; Teckentrup et al., 2018*] and found diverse latitudinal
 336 distribution of burned area. The across model differences were much larger than the inter-annual
 337 variation simulated by each individual model, which indicated large model structural
 338 uncertainties. Validation was also conducted for the historical period 1981-2000, when most of
 339 the satellite based burned area data were not available. Compared with charcoal index inferred
 340 burned area during 1981-2000 (Figure S4), DNN-Fire-OBS model reasonably captured the
 341 declining of burned area from ~530 Mha yr⁻¹ to 490 Mha yr⁻¹. In summary, DNN-Fire-OBS
 342 simulation is reasonably accurate and: (1) improved the simulated wildfire spatial and temporal
 343 distributions in ELMv1; (2) enabled effective and efficient parameterization of fires at regional
 344 scale.
 345



346

347 **Figure 6.** Prognostic simulation of annual wildfire burned area with the Deep Neural Network
 348 wildfire model fine-tuned with observations (DNN-Fire-OBS) compared with five burned area
 349 products (Table S1) over 2011-2015 for 14 GFED regions.
 350



351
 352 **Figure 7.** Prognostic simulation of wildfire burned area (2011-2015) with the Deep Neural
 353 Network wildfire model fine-tuned with observations (DNN-Fire-OBS) compared with
 354 observations and nine FireMIP models outputs.
 355

356 This study focuses on design, development, and parameterization of the DNN fire model
 357 within the E3SM model interface. In this way the DNN model can be readily coupled in the
 358 future and iteratively simulate climate, ecosystem fuel conditions, and fire dynamics. This study
 359 is an important step towards fully coupling E3SM and the DNN-Fire models in the future. We
 360 acknowledge several challenges and limitations in our modeling framework. First, the DNN
 361 model uncertainty was subject to the accuracy of climate forcings as well as other physical
 362 driving variables simulated by the physical wildfire model (ELMv1). For example, in addition to
 363 the default GSWP3 climate forcings dataset used in the study, CRU-JRA [Onogi *et al.*, 2007]
 364 and NCEP-DOE2 [Kanamitsu *et al.*, 2002] reanalysis forcings were also widely used and
 365 potentially different from GSWP3 forcings. ELMv1 used climate forcing (*e.g.*, temperature,
 366 precipitation, wind speed, relative humidity) to simulate soil temperature, soil moisture, fuel load

367 and so on. These simulated variables served as inputs for the DNN model and could also result in
368 prediction uncertainty. It was challenging to eliminate the forcing uncertainties in this work, but
369 we could at least evaluate the magnitude of these uncertainties. We ran the DNN-Fire-OBS
370 model with alternative forcings of CRU-JRA, NCEP-DOE2, and CDAS soil moisture from 2001
371 to 2010 and compared the results with DNN-Fire-OBS driven by default inputs (GSWP3 climate
372 and ELMv1 simulated soil moisture) (Figure S5). The results showed relatively larger
373 uncertainties from climate forcing than that from soil moisture forcing particularly over the
374 major fire regions (e.g., SHSA, SHAF, and NHAF). For fuel load, although no transient dataset
375 of global living biomass existed yet, we directly compared the ELM model simulated biomass
376 with the global estimate (GEOCARBON \sim 455 Pg C). We found that the modeled present-day
377 biomass continuously increased from 425 to 470 Pg C and compared reasonably well with the
378 global benchmark (Figure S6). Future work will focus on evaluating the uncertainties from dead
379 fuel load and fuel temperature variables.

380 Second, the original ELMv1 wildfire model has a unified mathematical representation of
381 how fuel, climate, and social-economic conditions control wildfire burned area [*F Li et al.*,
382 2012]. However, training one single DNN wildfire model across the globe will produce a model
383 dominated by gridcells that have high burned area (e.g., Africa). The performance of the trained
384 DNN model, therefore, will likely have larger biases over the low fire gridcells although the
385 globally aggregated burned area could be reasonable. We partly overcame this challenge by
386 applying the widely used 14 GFED fire regions that assume unique and relatively uniform
387 dynamics over each region [*Giglio et al.*, 2006b], and employed stratified random sampling
388 method for training and testing datasets. Although the regionally specific wildfire model
389 introduces additional complexity, it better represents distinct characteristics of wildfire activity
390 over different climate regimes and biomes [*Zhu and Zhuang*, 2013; *Zou et al.*, 2019] and allows
391 for future analyses of how the relevant controllers vary across the globe.

392 Finally, our GFED region-based parameterization strategy relied on the combination of
393 climate and biome types, while an alternative parameterization strategy for DNN-Fire model
394 could be based on plant functional type distributions. Based on our analysis, the PFT-based
395 DNN-Fire model had similar performance compared with the GFED-based model (Figure S7,
396 S8). Since the current version of the E3SM land model does not allow PFT changes driven by

397 climate, both GFED-based and PFT-based models may not fully capture the changes of fire
398 dynamics due to longer-time scale fire regimes changes.

399

400 **4. Conclusions**

401 In this study, we first surrogated the baseline ELMv1 wildfire model with a Deep Neural
402 Network (DNN) approach (Pearson correlation coefficient = 0.91 (p value < 0.01), $R^2 = 0.79$).
403 The development was based on inputs and outputs from the baseline ELMv1 wildfire simulation,
404 which is process-based and reasonably simulates global burned area, although regional biases
405 existed. We then calibrated the neural network weights using the years 2001-2010
406 observationally inferred burned area. The final calibrated DNN wildfire model (DNN-Fire-OBS)
407 was shown to be more accurate over the 14 GFED regions. For example, reductions in absolute
408 error over Africa, South America, and Europe were by ~90%. More importantly, the DNN-Fire-
409 OBS model parameters could be calibrated within minutes, compared with traditional ELMv1
410 parameterization ensemble simulations that consume a large amount of computational time. The
411 improved DNN-Fire-OBS model also accurately prognosed global and regional burned area in
412 the five-year period following the training period from 2011 to 2015 (modeled 469-514 Mha yr⁻¹).
413 We conclude that the improved surrogate wildfire model (DNN-Fire-OBS) developed in this
414 study can serve as an effective alternative to the process-based fire model currently used in
415 ELMv1. More broadly, we conclude that machine learning techniques can facilitate earth system
416 model development, parameterization, and uncertainty reduction with high efficiency and
417 accuracy.

418

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426

427 **Author contribution**

428 Q.Z., W.J.R, designed the study, Q.Z., W.J.R, L.X., and J.T.R designed model experiments,
429 Q.Z. and F.L. wrote code and run experiments, L.Z, K.Y, H.W., J.G all contribute to the results
430 interpretation, and writing.

431

432 **Code availability**

433 https://github.com/qzhu-lbl/ANN_wildfire

434 **Data availability**

435 GFEDv4s: https://daac.ornl.gov/VEGETATION/guides/fire_emissions_v4.html

436 Fire_CCI51: https://geogra.uah.es/fire_cci/firecci51.php

437 Fire_CCILT11: https://geogra.uah.es/fire_cci/fireccilt11.php

438 MCD64: https://modis-fire.umd.edu/files/MODIS_C6_Fire_User_Guide_C.pdf

439 Fire_Atlas: <https://www.globalfiredata.org/fireatlas.html>

440 FireMIP model outputs: <https://zenodo.org/record/3555562/accessrequest>

441

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693 **Supplementary Material**

694

695 Table S1. Burned area datasets used in this study

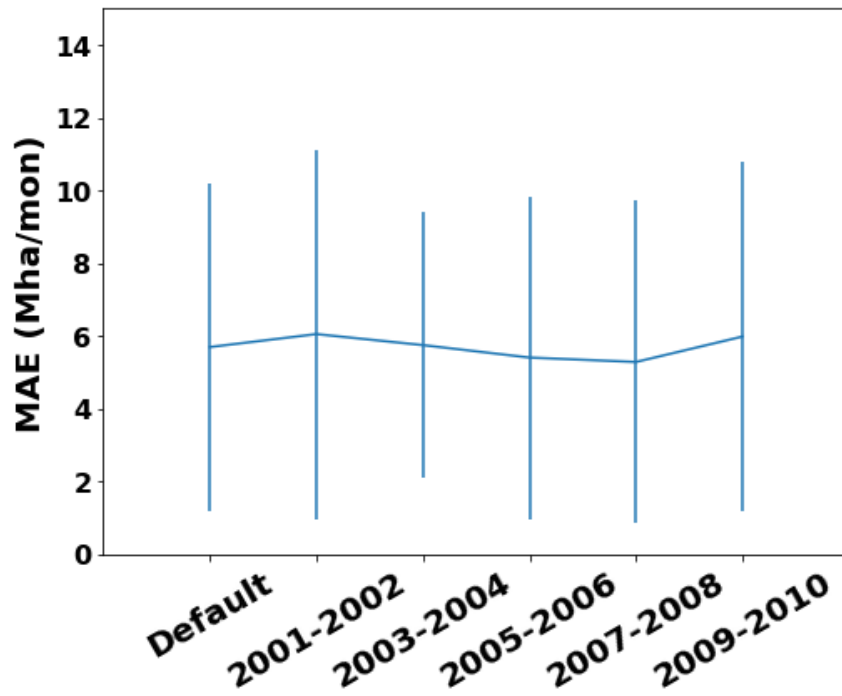
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| Dataset name | Temporal range | Spatial resolution | Burned area, mean (std) | Citations |
|--------------|----------------|--------------------|-------------------------|--|
| GFEDv4s | 1997-2015 | 0.25 degree | 455(39) | (Van Der Werf, Randerson et al. 2017) |
| Fire_CCI51 | 2001-2019 | 0.25 degree | 476(26) | (Lizundia-Loiola, Otón et al. 2020) |
| Fire_CCILT11 | 1982-2018 | 0.25 degree | 484(20) | (Lizundia-Loiola, Pettinari et al. 2018) |
| MCD64 | 2001-2019 | 0.25 degree | 424(35) | (Giglio, Boschetti et al. 2018) |
| Fire_Atlas | 2003-2016 | 0.25x0.25 degree | 459(43) | (Andela, Morton et al. 2019) |

697 **Note:** the long-term average global burned area was calculated using data with the same
698 overlapping temporal range (2003-2015), unit Mha yr⁻¹

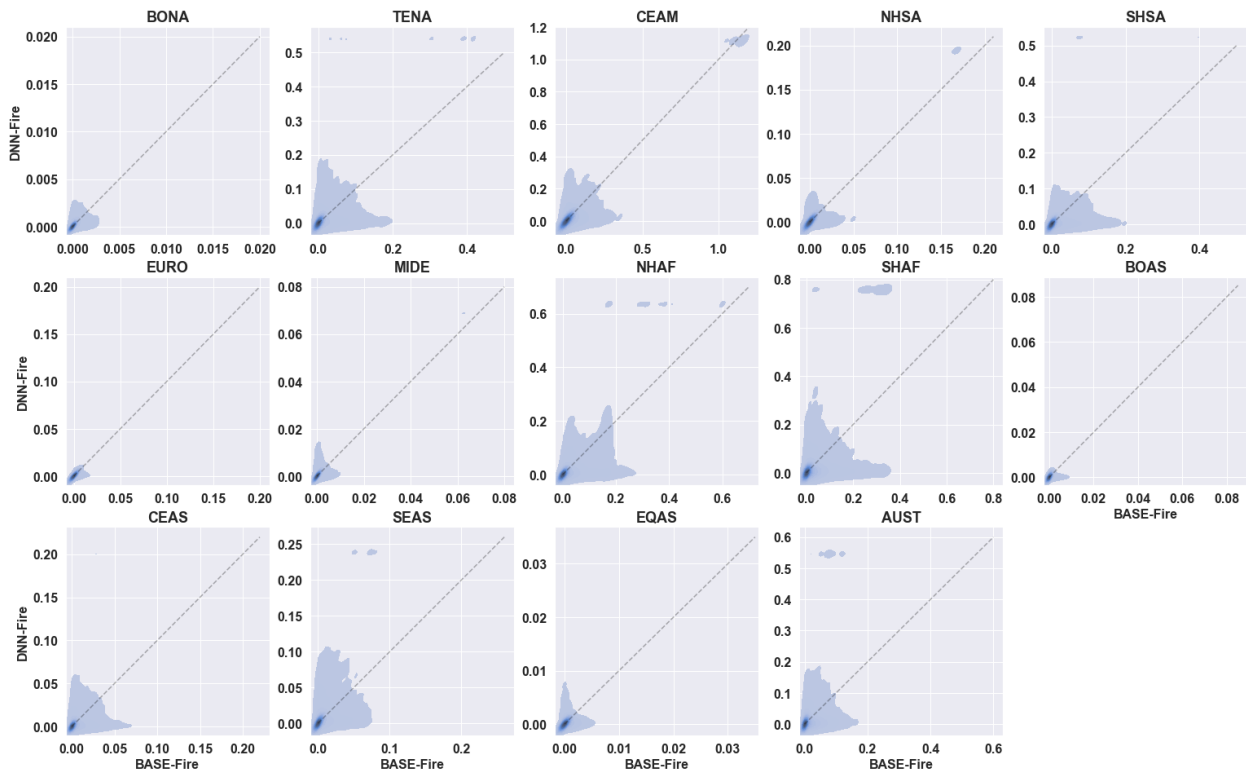
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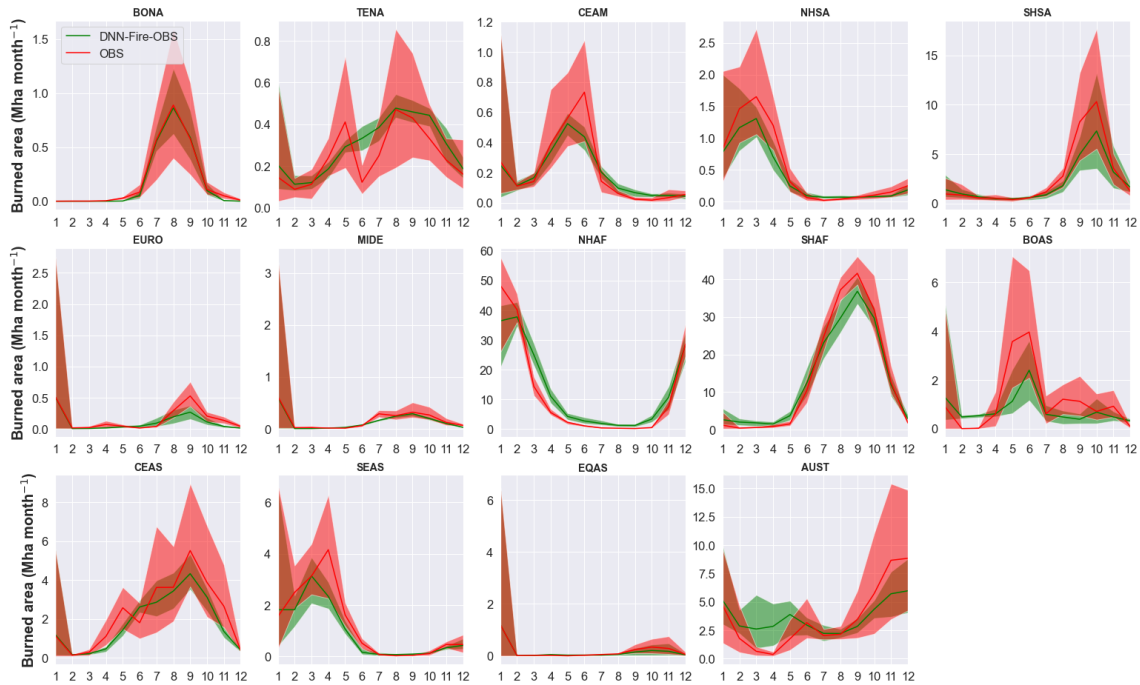
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Figure S1. Model performance evaluated with testing datasets of default (20% randomly selected samples), or fixed to 2001-2002 period, 2003-2004 period, 2005-2006 period, 2007-2008 period, and 2009-2010 periods (the rest of the dataset was used as a training dataset.).

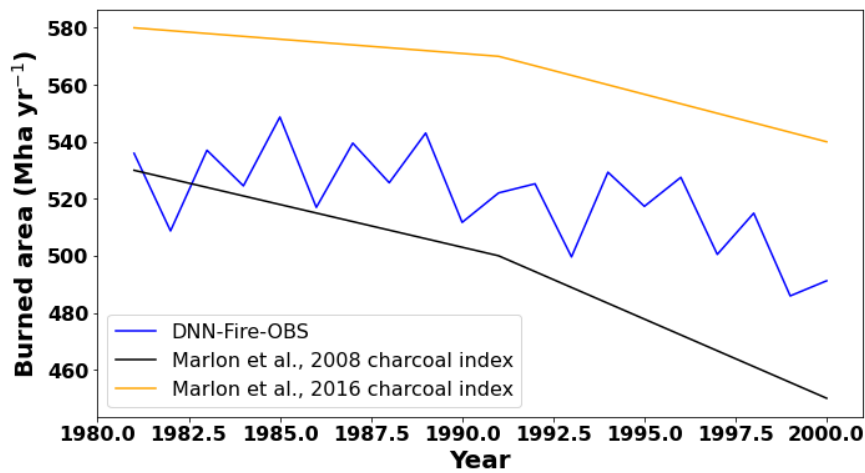


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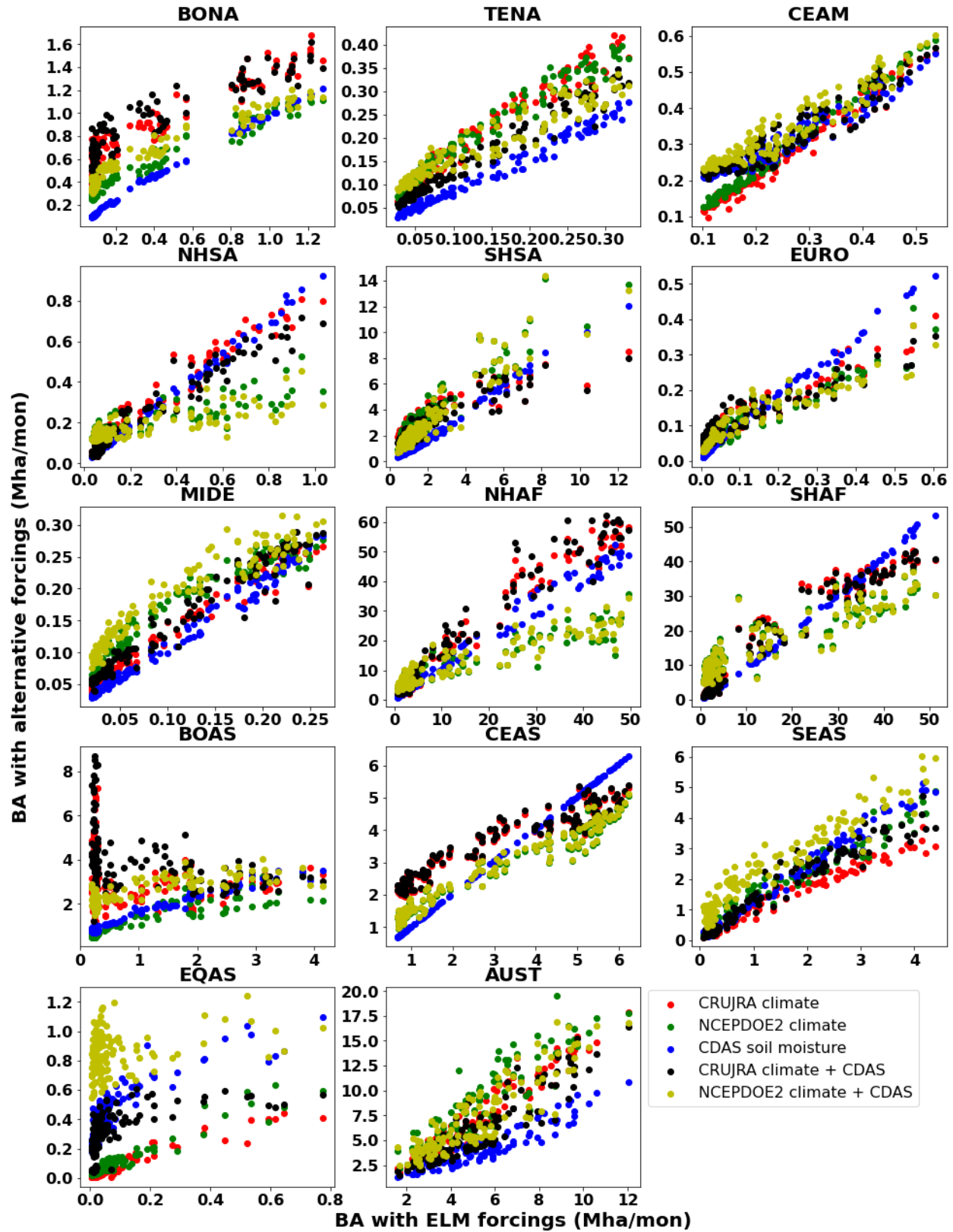
707 **Figure S2.** Performance of surrogate model (DNN-Fire) compared with ELMv1 process-based
 708 model (BASE-Fire).
 709



710 **Figure S3.** Seasonal cycles of fine-tuned Deep Neural Network wildfire model (DNN-Fire-OBS)
 711 and observations over 14 GFED fire regions.
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 713
 714



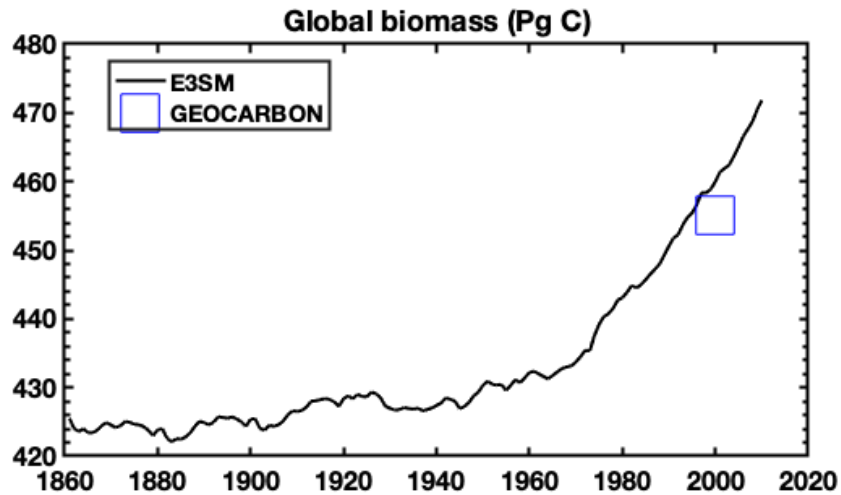
715 **Figure S4.** Comparison of DNN-Fire-OBS model simulated global burned area during 1981-
 716 1999 with two charcoal index inferred burned area.
 717



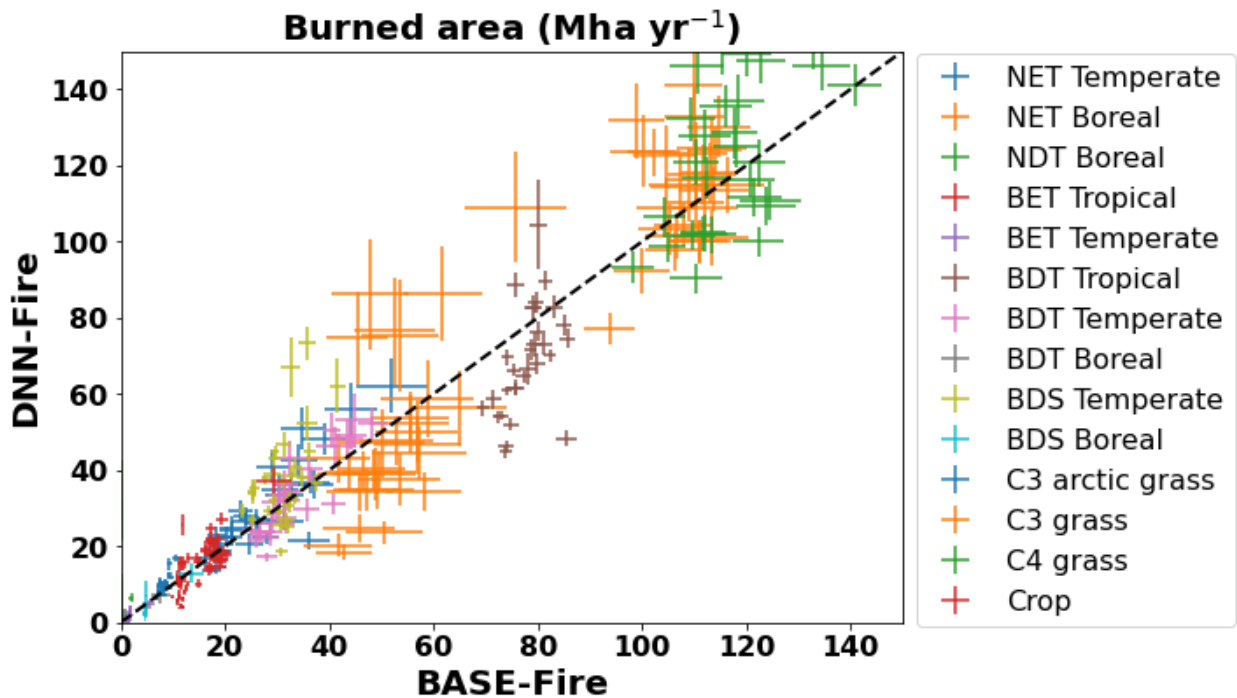
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Figure S5. Sensitivity of modeled burned area (2001-2010 long-term averaged) to climate forcings (including temperature, precipitation, wind speed, relative humidity) and soil moisture.

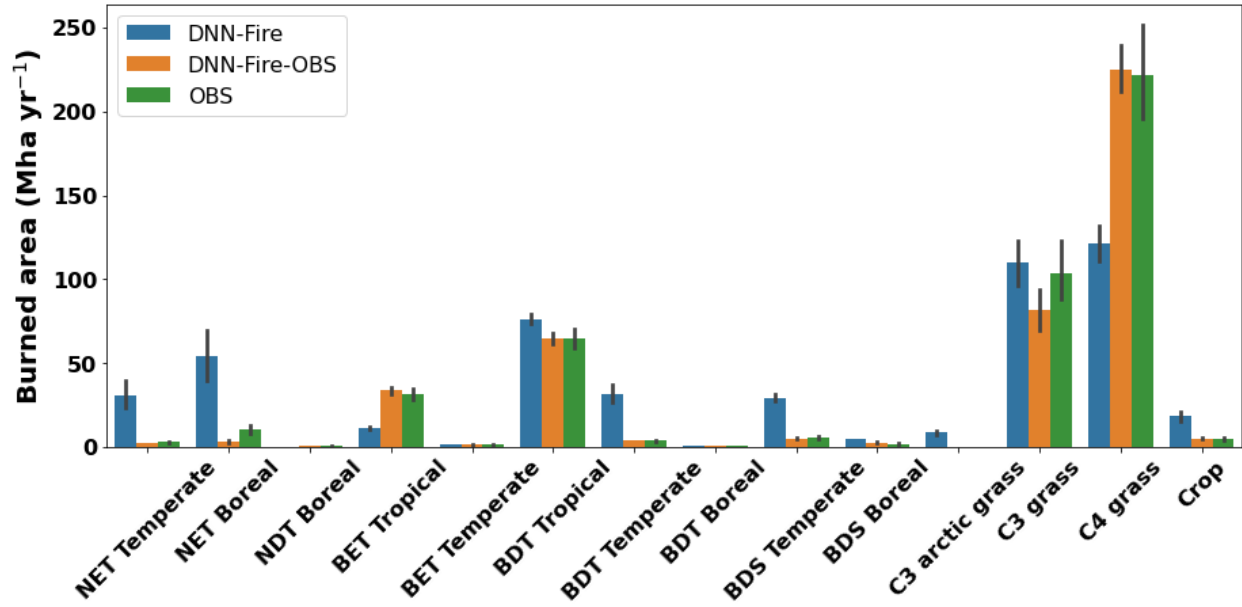
721 X-axis was burned area simulated by the default model using GSWP3 climate forcing and
 722 ELMv1 simulated soil moisture. Y-axis were models with alternative climate forcing (CRUJRA,
 723 NCEPDOE2) and soil moisture product (NCEP CDAS soil moisture).
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 727 **Figure S6.** 3SM simulated global vegetation biomass [425-472 PgC] and observational based
 728 estimate of present-day living biomass (455 PgC GEOCARBON).
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 731 **Figure S7.** The performance of the Deep Neural Network wildfire model (DNN-Fire), compared
 732 with the original ELMv1 process-based wildfire model (BASE-Fire) aggregated over 14 plant
 733 functional types between years 2001 and 2010.
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Figure S8. A comparison of wildfire burned area among Deep Neural Network wildfire model (DNN-Fire), Deep Neural Network wildfire model fine-tuned with observed burned area (DNN-Fire-OBS), and observations for 14 plant functional types.