



1 **A validation standard for Area of Habitat maps for terrestrial** 2 **birds and mammals**

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12 **Abstract**

13 Area of Habitat (AOH) is a deductive model which maps the distribution of suitable habitat at
14 suitable altitudes for a species inside its broad geographical range. AOH maps have been validated
15 using presence-only data for small subsets of species for different taxonomic groups, but no
16 standard validation method exists when absence data are not available. We develop a novel two-step
17 validation protocol for AOH which includes first a model-based evaluation of model prevalence
18 (i.e. the proportion of suitable habitat within a species' range), and second a validation using species
19 point localities (presence-only) data. We applied the protocol to AOH maps of terrestrial birds and
20 mammals. In the first step we built logistic regression models to predict expected model prevalence
21 (the proportion of the range retained as AOH) as a function of each species' elevation range, mid-
22 point of elevation range, number of habitats, realm and, for birds, seasonality. AOH maps with large
23 difference between observed and predicted model prevalence were identified as outliers and used to
24 identify a number of sources of systematic error which were then corrected when possible. For the
25 corrected AOH, only 1.7% of AOH maps for birds and 2.3% of AOH maps for mammals were
26 flagged as outliers in terms of the difference between their observed and predicted model
27 prevalence. In the second step we calculated point prevalence, the proportion of point localities of a
28 species falling in pixels coded as suitable in the AOH map. We used 48,336,141 point localities for
29 4889 bird species and 107,061 point localities for 420 mammals. Where point prevalence exceeded
30 model prevalence, the AOH was a better reflection of species' distribution than random. We also
31 found that 4689 out of 4889 (95.9%) AOH maps for birds, and 399 out of 420 (95.0%) AOH maps



32 for mammals were better than random. Possible reasons for the poor performance of a small
33 proportion of AOH maps are discussed.

34 **Introduction**

35 An accurate estimate of the distribution of species is central to ecological and conservation research
36 and action. There are three different classes of information on the distribution of species (Rondinini
37 and Boitani, 2006). These are 1) point localities (latitude and longitude) of individuals; 2)
38 geographic ranges, which are derived by mapping the extent of known point localities along with
39 expert knowledge; and 3) species distribution models, which use environmental and other relevant
40 variables associated with the species to refine geographical ranges. Species distribution models are
41 of two types (Stoms et al., 1992). The first are deductive models, which use expert-based
42 information on species' habitat use to model the suitable areas for the species. The second type are
43 inductive models, in which the environmental conditions at point localities where the species were
44 recorded are interpolated over wider areas.

45 Area of Habitat (AOH; also known as Extent of Suitable Habitat, ESH) is a deductive model which
46 maps the distribution of suitable habitat for a species inside its broad geographical range (Brooks et
47 al., 2019). It aims to reduce commission errors present in the range map while minimizing omission
48 errors. Several sets of AOH maps for different taxonomic groups at continental and global scales
49 have already been produced (Rondinini et al., 2005; Rondinini et al., 2006; Catullo et al., 2008;
50 Jenkins and Giri, 2008; Rondinini et al., 2011; Ficetola et al., 2015; Tracewski et al., 2016;
51 Lumbierres et al., 2021b).

52 Habitat models are prone to two major types of errors: omission errors occur when suitable habitat
53 areas for the species are wrongly mapped as being unsuitable, commission errors occur when areas
54 unsuitable for the species are wrongly mapped as being suitable. Quantification of these errors is
55 one of the key parts of the habitat modeling process and is done by validation. The omission and
56 commission errors could both be quantified only when independent presence and absence data on
57 the species are available. In such cases standard validation metrics such as True Skill Statistics
58 (TSS) (Allouche et al., 2006) and the Boyce Index (Boyce et al., 2002) are used. In case of AOH
59 maps produced for large taxonomic groups when true absence data are not available, no standard
60 validation method exists.

61 Rondinini et al. (2011) and Ficetola et al. (2015) used point localities from GBIF (Global
62 Biodiversity Information Facility) (www.gbif.org) to validate AOH maps for mammals and
63 amphibians respectively. AOH maps for South Asian mammals (Catullo et al., 2008) and African



64 vertebrates (Rondinini et al., 2005) were also validated using point localities. Brooks et al. (2019)
65 recommend using point localities for validation and inclusion of AOH maps for IUCN
66 (International Union for Conservation of Nature) Red List assessment. However, point localities are
67 often not available for many species and are biased towards certain taxonomic group and well-
68 studied areas.

69 In this paper, we developed a novel two-step validation protocol for AOH which includes: a) a
70 model-based evaluation of model prevalence (i.e., the proportion of a species' range that comprises
71 AOH), and b) a validation using species point localities (presence-only) data. We demonstrate the
72 use of this approach by validating a new set of AOH maps produced by Lumbierres et al. (2021b)
73 for all terrestrial birds and mammals. The validation method developed here is an iterative process
74 whereby systematic errors in the production of AOH (e.g. in the matching of habitat classes to land
75 cover maps) were identified using logistic regression models, then corrected where possible and a
76 new set of AOH maps produced. Then we employed a point validation analysis for the subset of
77 species for which point localities were available to assess the performance of the AOH maps.
78 Finally, we assessed the extent to which the subset of species for which point locality data were
79 available were representative of those for which no point data were available.

80 **2. Methods**

81 The new set of AOH maps (Lumbierres et al., 2021b) was produced at a resolution of 100 m using
82 a novel habitat-land cover model (Lumbierres et al., 2021a) which associated the different land
83 cover classes in the Copernicus land cover map (Buchhorn et al., 2019) with the Level-1 habitat
84 classes of the IUCN habitat classification scheme (IUCN, 2012). The IUCN habitat classification
85 scheme is a hierarchy of habitat classes, and each species assessed in the IUCN Red List is assigned
86 to one or more of these habitat classes, based on available information in the literature, unpublished
87 reports and expert knowledge. The habitat-land cover model (Lumbierres et al., 2021a) has the
88 provision of associating IUCN habitat classes to land cover classes using three different thresholds
89 (1, 2 and 3). Lower thresholds permit weaker associations between land cover and habitat classes.
90 Therefore, with threshold 1 each land cover class is associated with more habitat classes than with
91 threshold 3. Lumbierres et al. (2021b) produced a set of AOH maps for each of the three different
92 thresholds by clipping out of each species' range any cells of land cover that were not linked by the
93 model to the habitat class(es) to which the species was coded, then further clipping out parts of the
94 range falling outside the elevation range of the species.



95 In order to identify the best threshold among the three thresholds and to validate the set of AOH
96 maps with the best threshold at species level, we quantified two measures: ‘model prevalence’ and
97 ‘point prevalence’. Model prevalence is defined as the proportion of pixels inside the range that
98 were retained in the AOH. For example, if 25% of the pixels present in the original range map are
99 clipped out because they contain unsuitable habitat, fall outside the species’ elevation range or both,
100 the model prevalence is 0.75. Point prevalence is defined as the proportion of point localities (or
101 their buffers) out of all points inside the range of a species falling inside the suitable pixels. For
102 example the Red-tailed Comet (*Sappho sparganurus*) had a total of 71 point localities within its
103 range, of which 62 fell in pixels coded as suitable in the species’ AOH map, giving a point
104 prevalence of $62/71 = 0.88$.

105 Because the number of habitats associated with each land cover class decreases with increasing
106 thresholds, model prevalence is highest for threshold 1 models and lowest for threshold 3 models.
107 With increasing threshold, commission errors are expected to decrease (which is the main purpose
108 of AOH) but omission errors might increase. Our validation protocol therefore aimed to control for
109 omission errors. We did this by calculating point prevalence and model prevalence across the three
110 thresholds and identified the set of AOH maps for which the mean model prevalence was lowest
111 without compromising the mean point prevalence.

112 The point localities for bird species were downloaded from eBird (www.ebird.org), the largest
113 global repository for data on point localities of birds. eBird provides a metadata file called “eBird
114 basic data set” (Cornell Lab of Ornithology, 2020) which is a compilation of all the validated point
115 localities at species level and is updated monthly. These point localities are submitted by citizen
116 scientists as well as experts worldwide and are checked by local experts to remove obvious
117 misidentifications before they are made available for download (Sullivan et al., 2009). We first
118 downloaded the metadata file from eBird updated in January 2020 which was then queried in R (R
119 Core Team, 2018) using the *auk* package (Strimas-Mackey et al., 2018), as recommended by eBird,
120 to extract the point localities at species level. The taxonomy of Birdlife International (BirdLife
121 International and Handbook of the Birds of the World, 2020), which is that followed by the IUCN,
122 was matched with eBird’s taxonomy and point localities of only those species common to both were
123 queried and extracted from the metadata. Of the 10,813 species listed in Birdlife International’s list
124 for which AOH maps were produced, 9628 species matched by name. Of these 9628 species, 8998
125 species shared the same taxonomic concept and for 730 species the scientific names matched but
126 the taxonomic concept did not.

127 To ensure that only high-accuracy points were used for the validation, we selected the stationary
128 points from eBird’s metadata. The stationary points are those that have coordinate uncertainty of



129 less than 30 m. We then applied a temporal filter of 2019-2020 because the point localities from
130 2005-2018 were used to calibrate the habitat-land cover model in Lumbierres et al. (2021a). This
131 ensured there was no overlap between the calibration and validation data. The points were further
132 filtered by the range polygon of the species provided by the IUCN Red List website (IUCN, 2020)
133 to remove the small number of points falling outside the range (many of them likely to be
134 misidentifications). Since the AOH maps in question only include a certain combination of
135 presence, origin and seasonality of the range, we used the same combination to filter the point
136 localities. This ensured that we only included points which fell inside the boundaries of the selected
137 range maps. We also made sure that only one point locality was allowed per pixel of the AOH map
138 to avoid clustering of points. Finally, we excluded species which had fewer than 10 point localities
139 after all the filters were applied. A total of 4889 bird species had 4,836,141 point localities after
140 filtering. For mammals, point localities were downloaded from GBIF (Cold Spring Harbor
141 Laboratory, 2021) following the taxonomy of Global Mammal Assessment (which is followed by
142 IUCN) with same temporal and spatial filters as with birds except the filter of coordinate
143 uncertainty which was set to 300 m for mammals. This was done because far too many mammal
144 species would be excluded in the validation if we only considered point localities with coordinate
145 uncertainty of less than 30 m. The *rgbif* package (Chamberlain et al., 2021) in R was used to
146 download the points for mammals. A total of 107,061 point localities for 420 species were available
147 for mammals after applying all the filters.

148 A buffer of 300 m was applied around all the point localities to account for the positional
149 uncertainty of the points and for the fact that the location usually records that of the observer at the
150 time of observation and not the focal animal, following Jung et al. (2020). The buffers of point
151 localities were then overlaid on top of the AOH maps across all three thresholds at species level and
152 if at least one pixel coded to suitable habitat was found inside the buffer, the pixel was considered to
153 be validated at that point locality. The count of validated pixels was used to calculate point
154 prevalence at species level across all three thresholds.

155 We identified the threshold that produced a set of AOH maps for which the mean model prevalence
156 was lowest without detriment to the mean point prevalence.

157 We then employed a two-step approach to validate the set of AOH maps with the optimal threshold.
158 In the first step, we identified potential systematic errors in the AOH maps using a modeling
159 approach that aimed to identify species whose model prevalence was larger or smaller than
160 expected, given the characteristics of the species concerned. In the second step, we validated the
161 AOH maps using point localities following Rondinini et al. (2011).



162 2.1 A modeling approach to identify outliers

163 We used logistic generalized linear models to predict model prevalence of the set of AOH maps
164 produced using the optimal threshold as a function of a number of independent variables, and
165 identified outliers whose observed model prevalence was significantly higher or lower than
166 predicted by the model. Outliers were then examined to identify systematic errors in, for example,
167 the way habitats were coded to land cover classes in the production of the AOH maps, and to
168 identify species that might be coded to the wrong habitats or elevation limits. For example, if a
169 species' range includes a high proportion of a particular land cover type not associated with the
170 suitable habitats of the species in the land cover-habitat association table (Lumbierres et al., 2021b),
171 or if errors in coding species to elevation limits mean that most of the range is outside the species'
172 stated limits, the model prevalence would be lower than predicted by the model.

173 The predictors fitted to the logistic models included: elevation range of the species (upper elevation
174 limit minus lower elevation limit), mid-point of the elevation range, number of habitats to which the
175 species is coded against in the IUCN Red List, seasonality of species (breeding and non-breeding
176 ranges in case of migratory birds) and the geographical realm of the species. In case of migratory
177 birds, Lumbierres et al. (2021b) has three different classes (resident, breeding and non-breeding
178 seasonalities) of AOH maps based on seasonality of the species. We merged resident seasonality to
179 breeding and non-breeding seasonalities to have AOH maps with only two seasonalities (breeding
180 and non-breeding). The dependent variable was the model prevalence of the AOH maps. Data from
181 a total of 10475 AOH maps for 9163 bird species (including for some species with separate
182 breeding and non-breeding ranges) and 2758 AOH maps for 2758 mammal species were used to
183 build logistic regression models for birds and mammals separately using the *lme4* (Bates et al.,
184 2015) package in R. Data on elevation were lacking for many mammal and bird species which is
185 the reason why not all species could be included in the logistic model. After testing taxonomic
186 genus, family and order as random effects in the model to control the non-independence of closely
187 related taxa, family was selected for fitting as the residual variance was lowest for the models with
188 family as the random effect for both birds and mammals. The predictive power of the model was
189 assessed by calculating marginal R^2 and conditional R^2 using the *insight* (Lüdecke et al., 2019)
190 package in R. The marginal R^2 expresses how much of the variation in data is explained by the fixed
191 effects and conditional R^2 tells how much of the variation in data is explained by both fixed and
192 random effects.

193 The Tukey fences outlier detection test (Wilcox, 2017) was used to identify outliers based on the
194 difference between the estimated and observed values of model prevalence. This test uses the



195 interquartile ranges to estimate the outliers in a data-set. The outlier test identified mild lower and
196 upper threshold values for the difference between estimated and observed values.

197 *Mild upper threshold = (interquartile range * 1.5) + upper quartile*

198 *Mild lower threshold = lower quartile - (interquartile range * 1.5)*

199 The AOH maps identified as mild upper outliers have an observed model prevalence much larger
200 than their predicted model prevalence, whereas maps identified as mild lower outliers have an
201 observed model prevalence much smaller than their predicted model prevalence.

202 In order to investigate the sources of errors in the outliers, we produced two more sets of AOH
203 maps for the outliers. One set included AOH maps which were produced by clipping the range of
204 the species by the altitudinal range only (AOH_{Elevation only}). Similarly, the other set included AOH
205 maps which were derived by clipping the range with only suitable habitat of the species (AOH_{Habitat}
206_{only}). If the model prevalence of an outlier was equal or nearly equal to the model prevalence of its
207 AOH_{Elevation only}, then we concluded that the under-representation of model prevalence could be
208 attributed to errors in elevation range of the species. If the model prevalence of an outlier was equal
209 or nearly equal to the model prevalence of AOH_{Habitat only}, then the source of error could be attributed
210 to the mapping of the habitats inside the range using the habitat-land cover crosswalk (Lumbierres
211 et al., 2021a) or to errors in the species' habitat coding. Furthermore, in some of the outliers the
212 under-representation could result from inclusion of large proportion of habitats which were
213 unsuitable for the species but were inside the range map of the species. Outliers do not necessarily
214 represent errors in AOH, as species might legitimately have very high or low model prevalence, but
215 by identifying suites of outliers sharing common characteristics we were able to identify and correct
216 a number of systematic errors in AOH production. The models also allowed us to identify species
217 whose AOH maps might be unreliable and whose habitat and elevation coding needs to be checked.

218 2.2 Point validation of AOH maps of terrestrial birds and mammals

219 We validated 4889 bird and 420 mammal species' AOH maps using the filtered point localities. The
220 point validation was done by comparing the model and point prevalence at species level. If the point
221 prevalence exceeded model prevalence at species level, the AOH maps performed better than
222 random, otherwise they were no better than random. We also calculated the percentage of suitable



223 habitat pixels inside the buffers to ensure that the validation success wasn't due to a one off pixel
224 falling inside the 300 m buffer.
225 One of the major issues with citizen science data is that there is often a non-representative spread of
226 data across species. It is therefore possible that the species included in the point validation analysis
227 are not representative, in terms of the ratio between point prevalence and model prevalence, of the
228 species not included. We assessed how representative the validation sample size was by comparing
229 the representation of variables such as family, order, genus, realm, elevation range, mid-point of the
230 elevation range, range size and extinction risk categories for birds and mammals between species
231 with and without point data. The point validation was done in R and GRASS (GRASS Development
232 Team, 2017).

233 3. Results

234 After comparing point and model prevalence of 4889 birds and 420 mammal species across all the
235 three thresholds, we selected the set of AOH maps derived by using threshold 3 in the habitat-land
236 cover model. At threshold 3, the mean model prevalence decreased as compared to thresholds 1 and
237 2 with much lower change in the mean point prevalence (Table 1 and 2) for both birds and
238 mammals.

	Threshold 1	Threshold 2	Threshold 3
Mean model prevalence	0.81 ± 0.21 SD	0.77 ± 0.23 SD	0.65 ± 0.25 SD
Mean point prevalence	0.95 ± 0.14 SD	0.94 ± 0.14 SD	0.90 ± 0.17 SD

239 **Table 1:** Mean model and point prevalence for AOH maps with standard deviation of 4889 bird
240 species across 3 different thresholds.

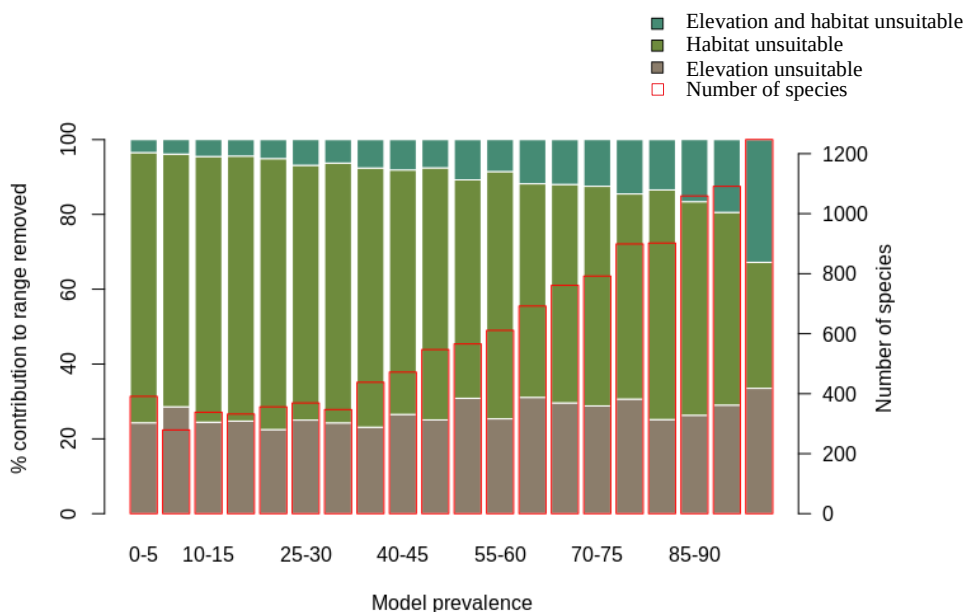
	Threshold 1	Threshold 2	Threshold 3
Mean model prevalence	0.87 ± 0.21 SD	0.83 ± 0.22 SD	0.73 ± 0.24 SD
Mean point prevalence	0.95 ± 0.14 SD	0.95 ± 0.15 SD	0.93 ± 0.17 SD

241 **Table 2:** Mean model and point prevalence for AOH maps with standard deviation of 420 mammal
242 species across 3 different thresholds.

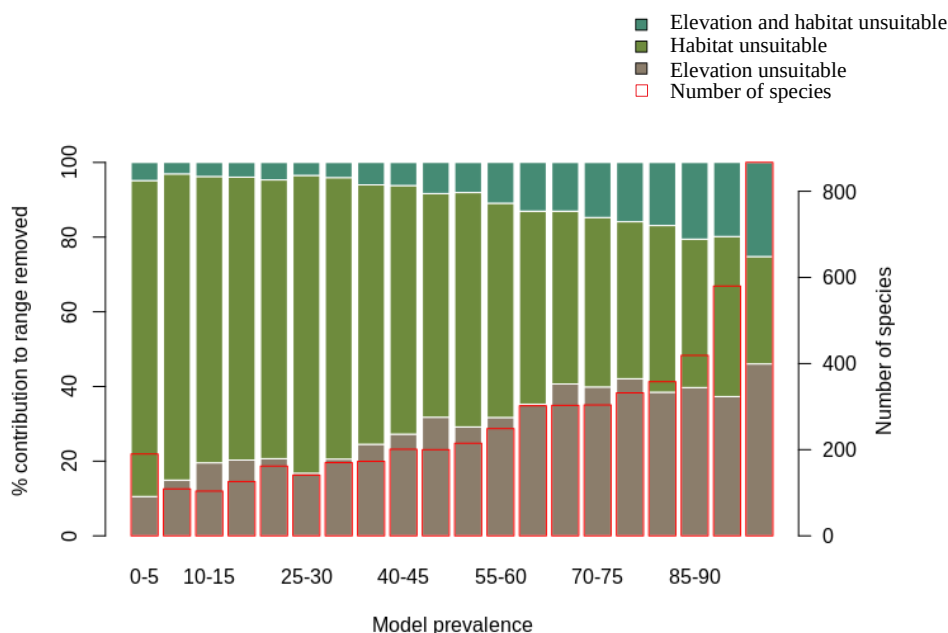
243 We also assessed the relative contribution of elevation range, habitat, and both in reducing the range



244 to AOH. For both birds and mammals, most of the pixels removed from the range were because
245 either the habitat or the elevation were unsuitable, with a relatively small proportion being removed
246 because both were unsuitable (Figs. 1,2). The proportion of the range that was clipped out on the
247 basis of having unsuitable habitat at suitable elevations increased as model prevalence decreased,
248 whereas there was little change across the same axis in the proportion of the range that was
249 excluded on the basis of having suitable habitat at unsuitable elevations (Figs. 1,2). The number of
250 both bird and mammal species peaked at model prevalence of 95-100% and gradually decreased as
251 the model prevalence decreased.



252 **Figure 1:** Percentage contribution of elevation range, habitat and both in clipping the IUCN range
253 to produce AOH maps for birds. Each bar represents a 5% bin of model prevalence, divided to show
254 how much of the range was clipped out due to unsuitable habitat at suitable elevations (“Habitat
255 unsuitable”), by suitable habitat at unsuitable elevations (“Elevation unsuitable”) and by unsuitable
256 habitat at unsuitable elevations (“Elevation and habitat unsuitable”). The red blocks correspond to
257 the second y-axis and show the number of species falling into each 5 % bin of model prevalence.



258 **Figure 2:** Percentage contribution of elevation range, habitat and both in clipping the IUCN range
259 to AOH for mammals. See caption to Fig. 1 for interpretation.

260 For birds, the logistic model identified 178 AOH maps (1.7%) as lower outliers and 118 AOH maps
261 (1.1%) as upper outliers out of 10475 AOH maps for 9163 terrestrial bird species. Similarly for
262 mammals, the logistic model was applied to the AOH maps of 2758 species and identified 64
263 (2.3%) as lower outliers and 21 (0.8%) as upper outliers.

264 The mean of mid-point of elevation of the bird and mammal species identified as upper outliers was
265 2725 m and 3193 m respectively while the mid-point of elevation for species which were not
266 identified as upper outliers was 1261 m for birds and 1289 m for mammals. This suggests that
267 species identified as upper outliers were those found in higher elevation. These species were
268 identified as upper outliers because the logistic models predicted low model prevalence at higher
269 elevations. Also, the range maps for high-altitude species are drawn using contour maps, therefore
270 most of the range is within the correct altitudinal band leading to high model prevalence for these
271 species.

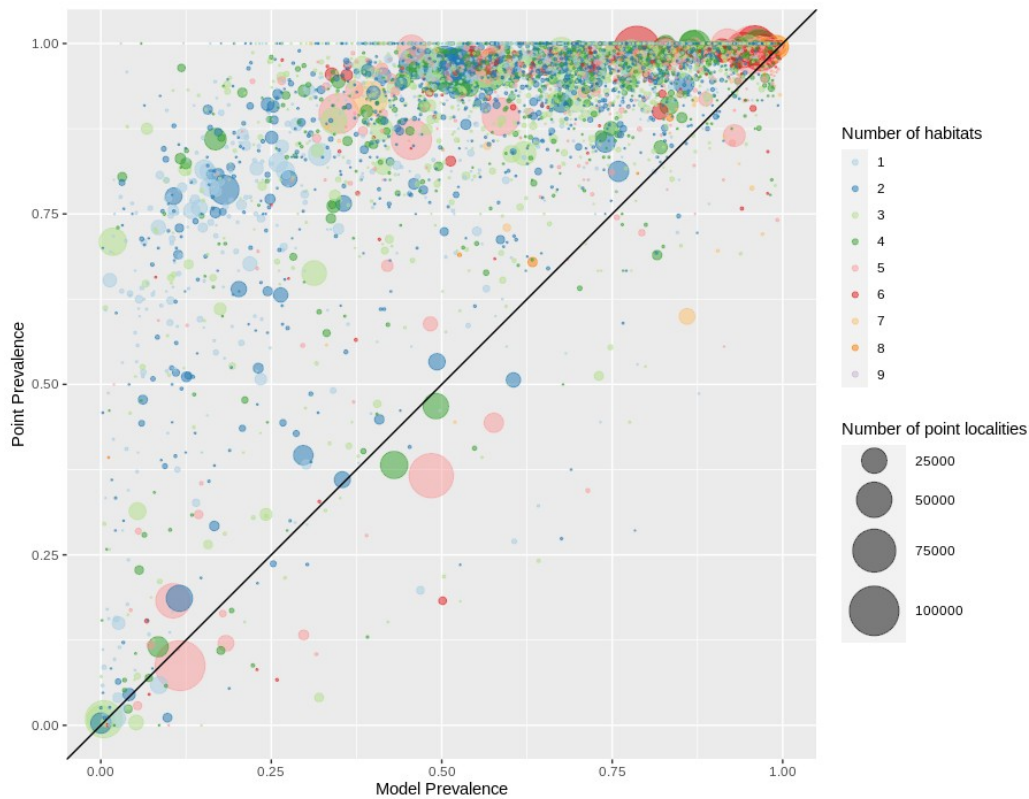
272 The lower outliers indicate where model prevalence was possibly underestimated due to potential
273 errors in habitat mapping/coding and elevation range of the species. We found that the habitats



274 “Shrubland” and “Savannah” in the habitat-land cover crosswalk were not associated with the land
275 cover class “Herbaceous cover”, leading to under-representation of these habitat types and hence
276 lower model prevalence than estimated by the logistic model (Fig. A1). We also found mismatch in
277 the elevation range and geographical range for the lower outliers (Fig. A2). There were few cases
278 where the range included large proportion of a particular land cover type which was not associated
279 with the suitable habitat of the species (Fig. A3). Moreover, we found that there was no land cover
280 information in the Copernicus land cover map for very small range polygons located on oceanic
281 islands which caused the AOH maps for these species to be empty. Furthermore, the land cover
282 class “open forest unknown” was discarded in the habitat land cover model. This led to low model
283 prevalence of AOH maps for some species whose ranges included this land cover. This was
284 corrected and a new set of AOH maps produced.

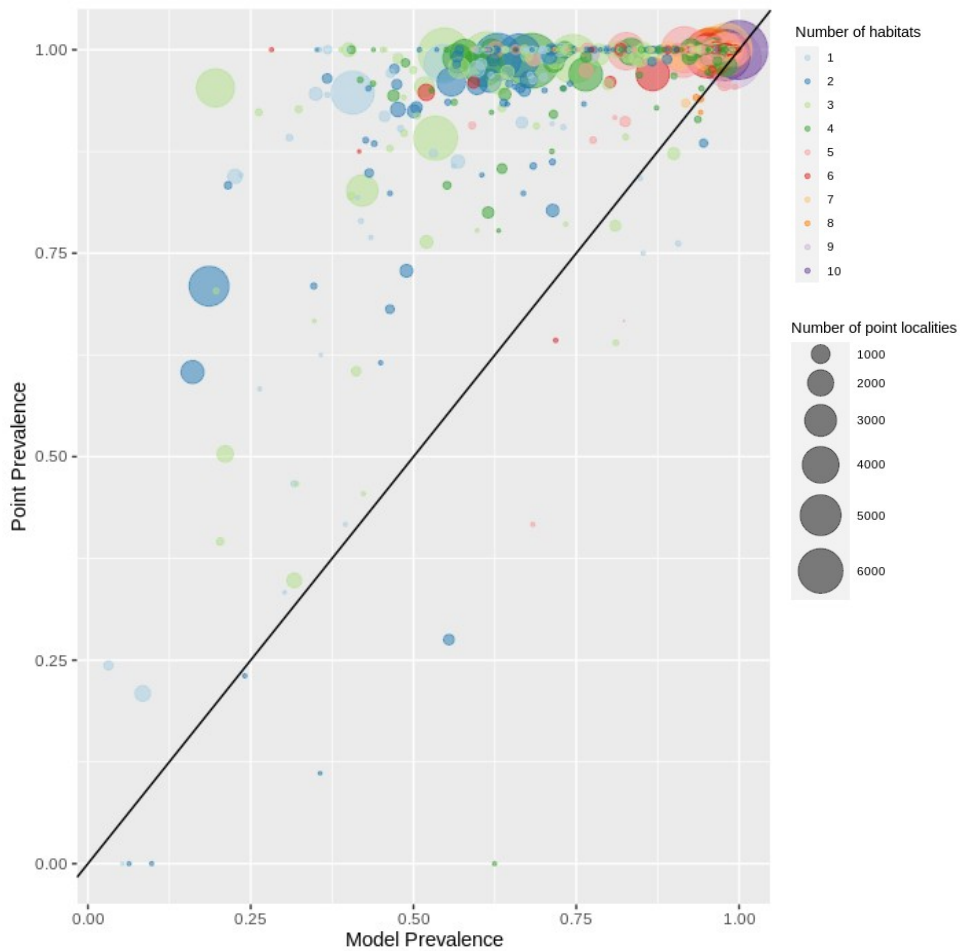
285 **Point validation**

286 Out of 4889 bird species (45% of all bird species) for which point data were available, 4689
287 (95.9%) had higher point prevalence than model prevalence and 200 species had lower point
288 prevalence than model prevalence (Fig. 3). The mean percentage of pixels coded as suitable inside
289 the 300 m buffers of point localities of 4889 species of birds was 62% (Fig. A5).



290 **Figure 3:** Point prevalence vs model prevalence for terrestrial birds. Colors indicate the number of
291 habitats each species is coded to, size of circles indicates the number of point localities.

292 Out of 420 mammal species (8% of all mammal species) for which point data were available, 399
293 (95.0%) had point prevalence higher than model prevalence (Fig. 4). The mean percentage of pixels
294 coded as suitable inside the 300 m buffers of point localities of 420 species of mammals was 78%
295 (Fig. A5).



296 **Figure 4:** Point prevalence vs model prevalence for terrestrial mammals. Interpretation as in Fig. 3.

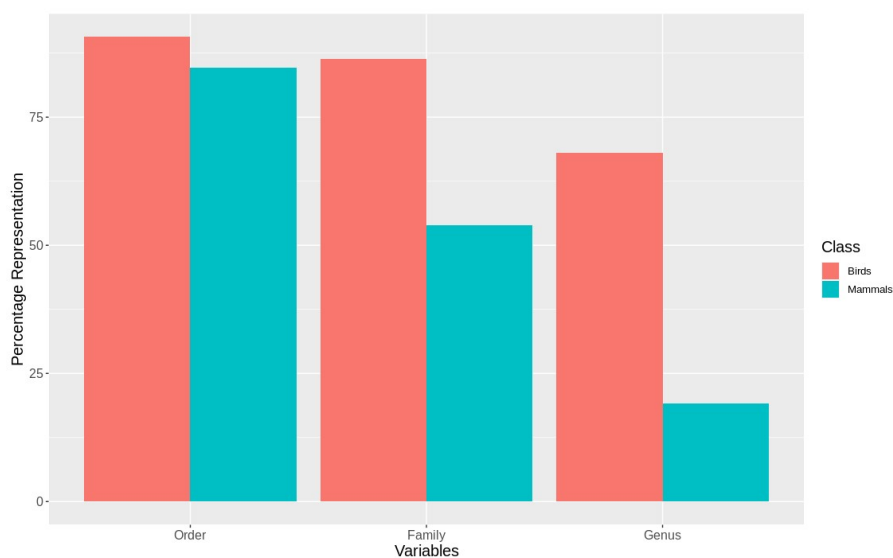
297 **Representativeness of validation sample**

298 We found that for birds over 60% all families, genera and orders were represented in the sample
299 included in the point validation and species from all biomes were represented but representation for
300 mammals was lower, as expected due to the much lower proportion of mammal species for which
301 point locality data were available (Fig. 5).

302 The validation points were spread across all of the variables and majority of their sub-classes (Fig.
303 A6, Fig. A7). Species with validation points tended to have larger range sizes, wider elevation



304 ranges and to be coded to more habitat classes than those without. Furthermore, validation points
305 were not available for any critically endangered or endangered mammals as these species are rare in
306 the wild.



307 **Figure 5:** Taxonomic representativeness of validation sample for birds and mammals.

308 Discussion

309 On comparing our point validation results with previous validation analysis of AOH maps, we
310 found that validation results are similar to or better than previous exercises. For mammals,
311 Rondinini et al. (2011) evaluated AOH maps for 263 species at 300 m resolution, of which 241
312 (91.6 %) were better than random as compared to 95.0% in our analysis. However, it should be
313 noted that the mean model prevalence for AOH maps of Rondinini et al. (2011) was 54.8 ± 21.5 SD
314 as compared to 65.16 ± 25.42 for our AOH maps. The ratio of mean point prevalence to mean
315 model prevalence for Rondinini et al. (2011) was 1.4 compared to 1.38 in our case. Ficetola et al.
316 (2015) found that AOH for 94% of 115 amphibian species used in the validation analysis were
317 better than random with the mean model prevalence for species with validation points being $0.79 \pm$
318 0.21 SD. The ratio of mean point prevalence to mean model prevalence was 1.18 in this case.
319 Moreover, Catullo et al. (2008) found that 140 AOH maps out of 190 (73.7 %) South Asian
320 mammal species gave positive validation results while Rondinini et al. (2005) found the mean
321 proportion of suitable habitats correctly mapped inside the range for 181 species of African



322 vertebrates was 0.55 ± 0.01 SE using presence-absence data sets. The high validation success in our
323 analyses could be attributed to the use of novel habitat-land cover model (Lumbierres et al., 2021a),
324 the use of logistic regression models to identify systematic errors and the larger validation sample
325 as compared with previous exercises. Furthermore, the underlying land cover map used in
326 Lumbierres et al. (2021b), has the highest resolution among the global land cover maps providing
327 with more detailed land cover classification.

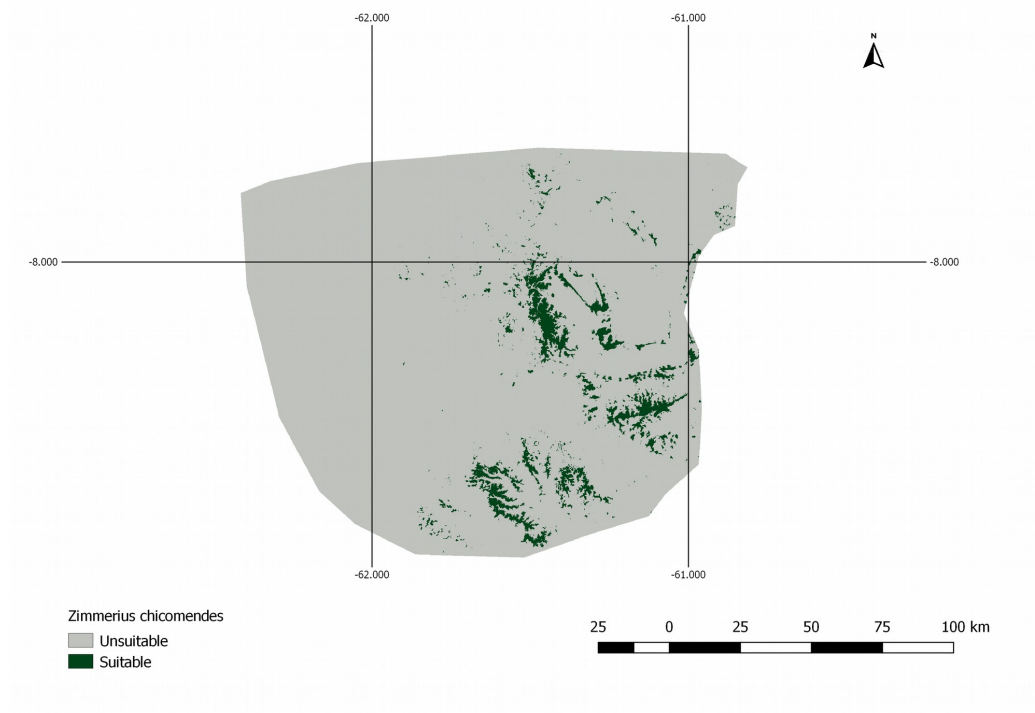
328 The point validation identified a small proportion of AOH maps which were no better than random.
329 Some of these had high model prevalence. In such cases, point prevalence must be exceptionally
330 high for the models to be better than random since even if a majority of point localities fall within
331 the AOH these maps may perform no better than random. For the AOH maps which were no better
332 than random and had low point prevalence, this was usually due to an apparent error in the coding
333 of elevation range of the species, the areas inside the range of the species where the point localities
334 fell being clipped out by what was assumed to be an erroneous elevation range. A list of species
335 with probably erroneous elevation coding will be forwarded to IUCN Red List team for future
336 corrections.

337 AOH maps aim to minimize the commission errors known to be present in species ranges without
338 increasing omission errors (Rondinini and Boitani, 2006). One of the limitations of this validation
339 analysis is the inability to quantify the commission errors of the AOH maps as we don't have the
340 true absence data of the species. Therefore, some uncertainty remains in AOH maps regarding the
341 commission errors.

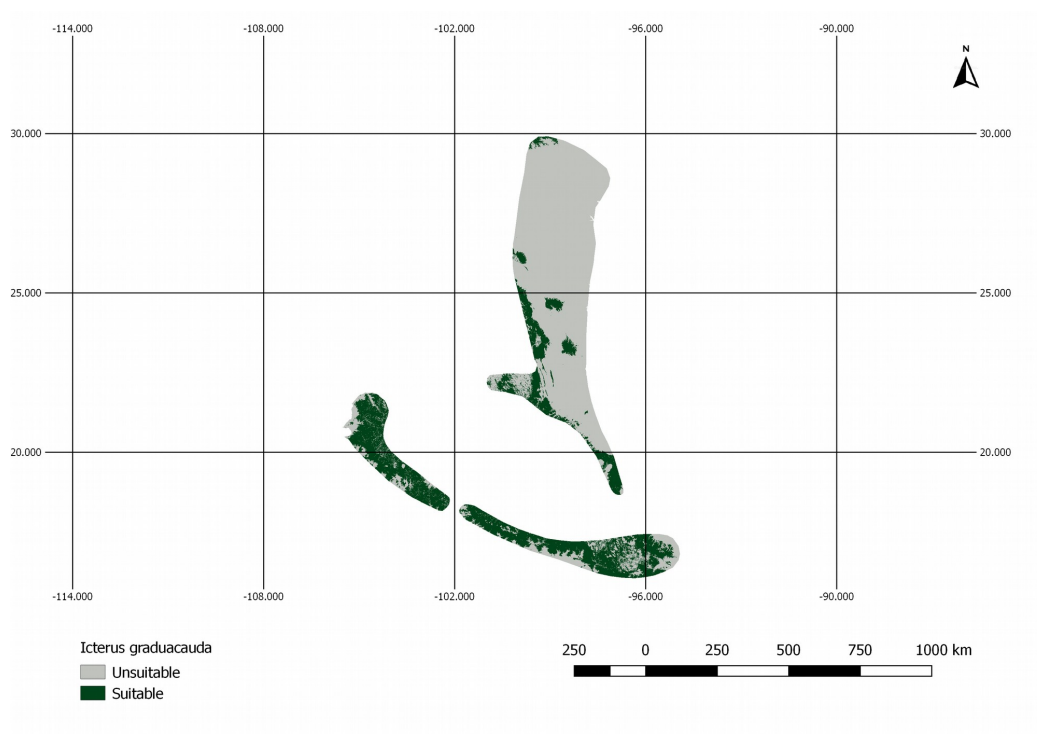
342 Also, there are some intrinsic errors in the models as identified by the logistic regression analysis.
343 The species which are coded only to habitats like "Shrubland" might have under-represented model
344 prevalence as discussed above. However, the number of AOH maps identified as lower outliers by
345 the application of the logistic model was low for birds (178/10475) and for mammals (64/2758),
346 indicating that for the majority of AOH maps the observed model prevalence was fairly close to that
347 predicted by the model.



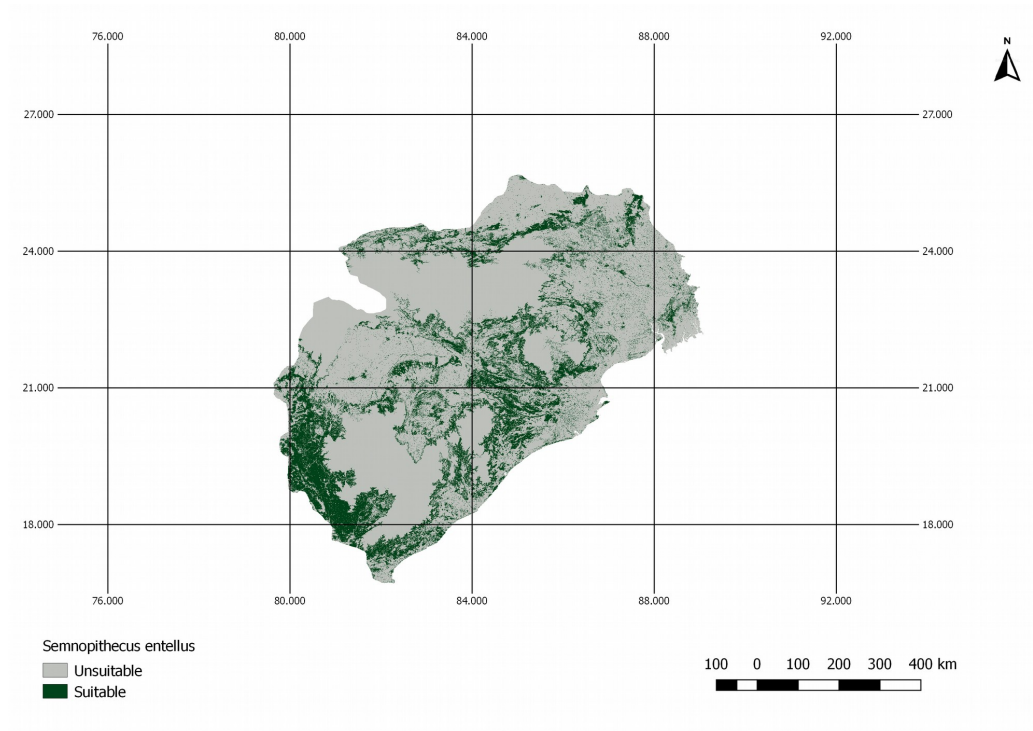
348 **Appendix A**



349 **Figure A1:** AOH map for species *Zimmerius chicomendesi*. The species is coded against “Forest”
350 and “Shrubland” habitats and the elevation range falls inside the IUCN range. However, the land
351 cover inside this range map includes a high proportion of “Herbaceous cover” land cover type
352 which is not associated with “Shrubland” habitat in the habitat – land cover association table.
353 Therefore, the model prevalence of this AOH is much lower than expected.



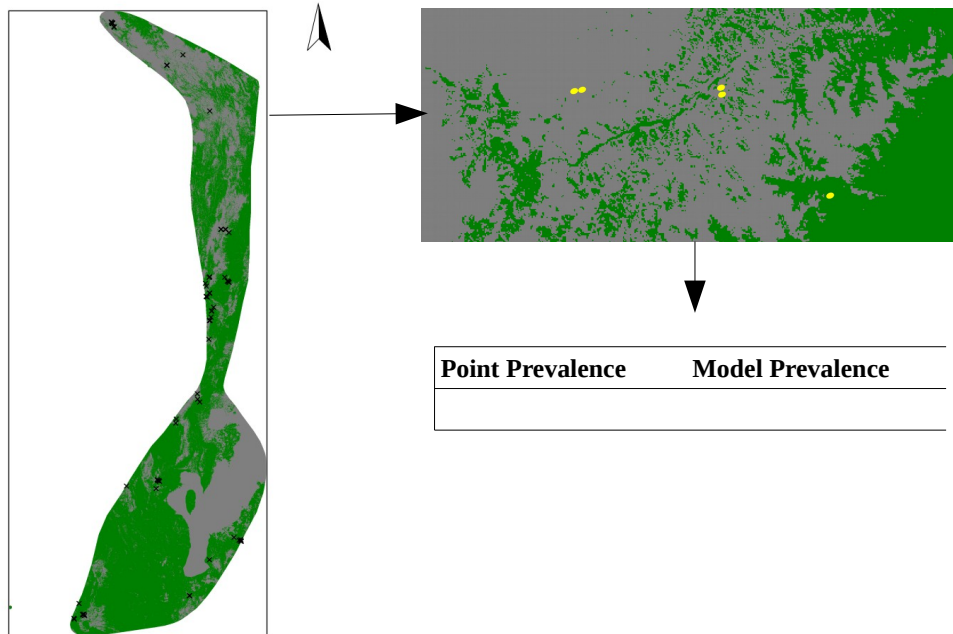
354 **Figure A2:** AOH map for the species *Icterus graduacauda*. The IUCN range of the species doesn't
355 cover much of the elevation range. Therefore, the model prevalence of this species is lower than
356 estimated.



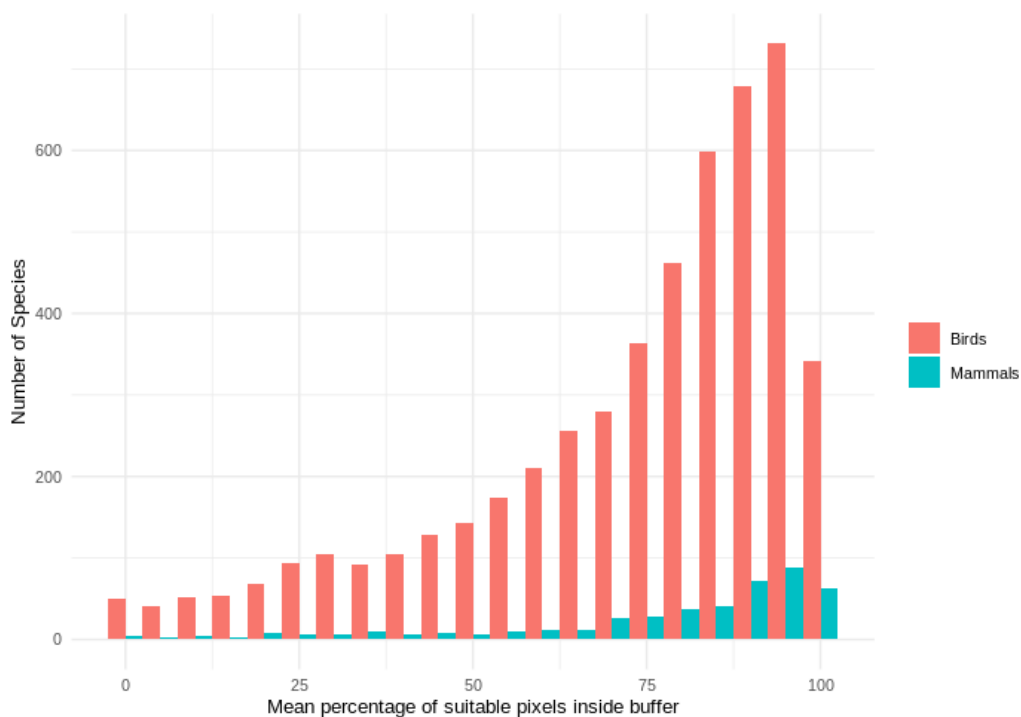
357 **Figure A3:** AOH for the species *Semnopithecus entellus*. There is a large proportion of land cover
358 class “Cropland” inside the range map of this species. However, this species is not coded to habitats
359 that are associated with the land cover “Cropland”. Therefore, the model prevalence is lower than
360 estimated.



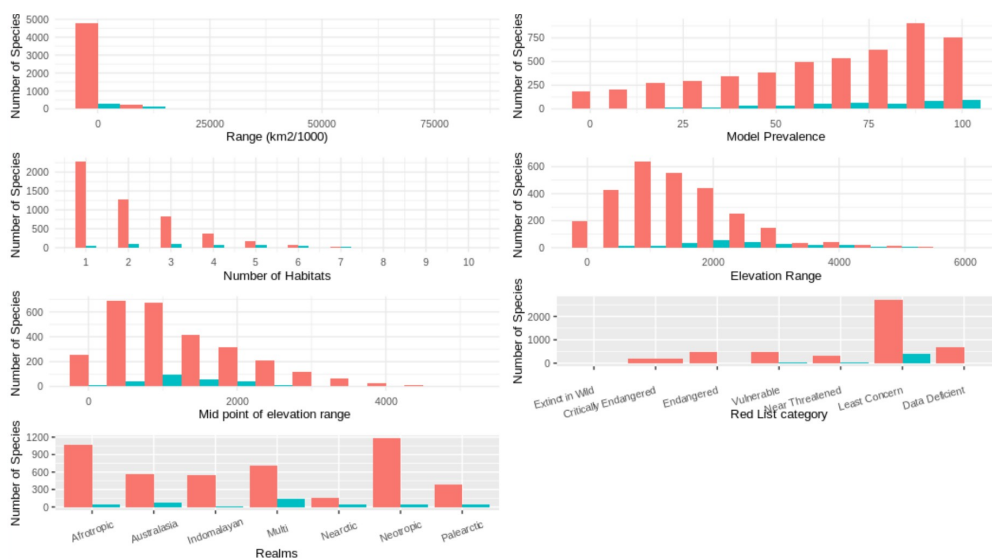
361



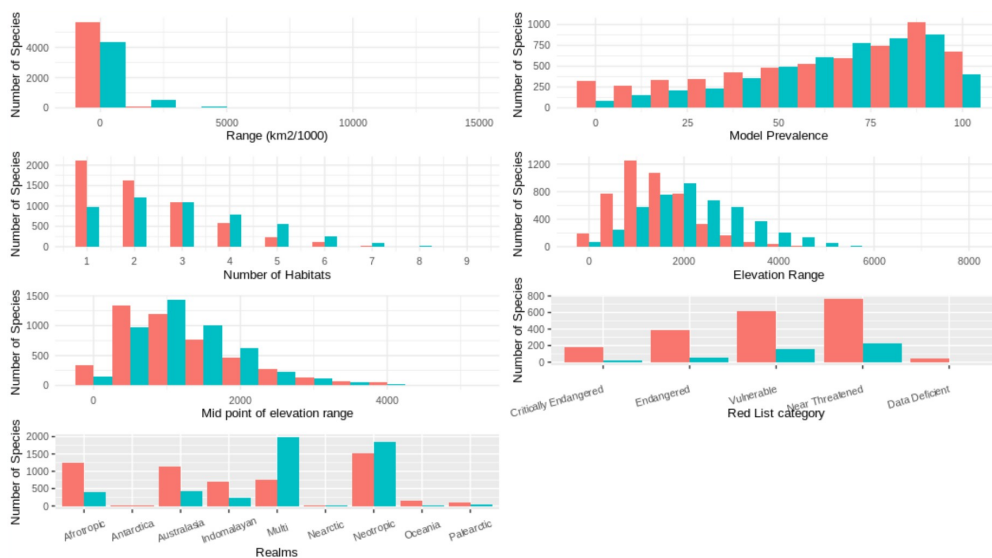
362 **Figure A4:** Point validation of the AOH maps using model and point prevalence.



363 **Figure A5:** Histogram of mean percentage of suitable AOH pixels inside the 300 m buffer for
364 mammals and birds species used in point validation.



365 **Figure A6:** Comparison of species with and without validation points for mammals. Colours as in
 366 A5.



367 **Figure A7:** Comparison of species with and without validation points for birds. Colours as in A5.



368 **Data and code availability**

369 The point localities used in the validation analyses along with the metadata tables summarizing the
370 validation analyses can be found at <http://doi.org/10.5281/zenodo.5109073>. The same DOI can be
371 used to access the code used for validation and to also access some sample AOH maps which were
372 validated.

373 **Author contribution**

374 PRD PFD and CR conceptualized the idea. PRD and ML curated and did the formal data analysis.
375 PRD led the manuscript writing with contributions from all the authors. PFD CR SHMB supervised
376 the whole process.

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