A validation standard for Area of Habitat maps for terrestrial birds and mammals

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- 16 Abstract
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Area of Habitat (AOH) is a deductive model which maps the distribution of suitable habitat at 18 suitable altitudes for a species inside its broad geographical range. AOH maps have been validated 19 using presence-only data for small subsets of species for different taxonomic groups, but no 20 standard validation method exists when absence data are not available. We develop a novel two-step 21 validation protocol for AOH which includes first a model-based evaluation of model prevalence 22 (i.e, the proportion of suitable habitat within a species' range), and second a validation using species 23 point localities (presence-only) data. We applied the protocol to AOH maps of terrestrial birds and 24 mammals. In the first step we built logistic regression models to predict expected model prevalence 25 (the proportion of the range retained as AOH) as a function of each species' elevation range, mid-26 point of elevation range, number of habitats, realm and, for birds, seasonality. AOH maps with large 27 difference between observed and predicted model prevalence were identified as outliers and used to 28 identify a number of sources of systematic error which were then corrected when possible. For the 29 corrected AOH, only 1.7% of AOH maps for birds and 2.3% of AOH maps for mammals were 30 flagged as outliers in terms of the difference between their observed and predicted model 31 prevalence. In the second step we calculated point prevalence, the proportion of point localities of a 32 species falling in pixels coded as suitable in the AOH map. We used 48,336,141 point localities for 33 4889 bird species and 107,061 point localities for 420 mammals. Where point prevalence exceeded 34 model prevalence, the AOH was a better reflection of species' distribution than random. We also 35 found that 4689 out of 4889 (95.9%) AOH maps for birds, and 399 out of 420 (95.0%) AOH maps 36

for mammals were better than random. Possible reasons for the poor performance of a smallproportion of AOH maps are discussed.

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40 Introduction

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An accurate estimate of the distribution of species is central to ecological and conservation research 42 and action. There are three different classes of information on the distribution of species (Rondinini 43 and Boitani, 2006). These are 1) point localities (latitude and longitude) of individuals; 2) 44 geographic ranges, which are derived by mapping the extent of known point localities along with 45 expert knowledge; and 3) species distribution models, which use environmental and other relevant 46 variables associated with the species to refine geographical ranges. Species distribution models are 47 of two types (Stoms et al., 1992). The first are deductive models, which use expert-based 48 information on species' habitat use to model the suitable areas for the species. The second type are 49 inductive models, in which the environmental conditions at point localities where the species were 50 51 recorded are interpolated over wider areas.

Area of Habitat (AOH; also known as Extent of Suitable Habitat, ESH) is a deductive model which maps the distribution of suitable habitat for a species inside its broad geographical range (Brooks et al., 2019). It aims to reduce commission errors present in the range map while minimizing omission errors. Several sets of AOH maps for different taxonomic groups at continental and global scales have already been produced (Rondinini et al., 2005; Rondinini et al., 2006; Catullo et al., 2008; Jenkins and Giri, 2008; Rondinini et al., 2011; Ficetola et al., 2015; Tracewski et al., 2016; Lumbierres et al., 2021b).

Habitat models are prone to two major types of errors: omission errors occur when suitable habitat 59 areas for the species are wrongly mapped as being unsuitable, commission errors occur when areas 60 unsuitable for the species are wrongly mapped as being suitable. Quantification of these errors is 61 one of the key parts of the habitat modeling process and is done by validation. The omission and 62 commission errors could both be quantified only when independent presence and absence data on 63 the species are available. In such cases standard validation metrics such as True Skill Statistics 64 (TSS) (Allouche et al., 2006) and the Boyce Index (Boyce et al., 2002) are used. In case of AOH 65 maps produced for large taxonomic groups when true absence data are not available, no standard 66 validation method exists. 67

Rondinini et al. (2011) and Ficetola et al. (2015) used point localities from GBIF (Global
Biodiversity Information Facility) (www.gbif.org) to validate AOH maps for mammals and
amphibians respectively. AOH maps for South Asian mammals (Catullo et al., 2008) and African

vertebrates (Rondinini et al., 2005) were also validated using point localities. Brooks et al. (2019) recommend using point localities for validation and inclusion of AOH maps for IUCN (International Union for Conservation of Nature) Red List assessment. However, point localities are often not available for many species and are biased towards certain taxonomic group and wellstudied areas.

In this paper, we developed a novel two-step validation protocol for AOH which includes: a) a 76 model-based evaluation of model prevalence (i.e., the proportion of a species' range that comprises 77 AOH), and b) a validation using species point localities (presence-only) data. We demonstrate the 78 use of this approach by validating a new set of AOH maps produced by Lumbierres et al. (2021b) 79 for all terrestrial birds and mammals. The validation method developed here is an iterative process 80 whereby systematic errors in the production of AOH (e.g. in the matching of habitat classes to land 81 cover maps) were identified using logistic regression models, then corrected where possible and a 82 new set of AOH maps produced. Then we employed a point validation analysis for the subset of 83 species for which point localities were available to assess the performance of the AOH maps. 84 85 Finally, we assessed the extent to which the subset of species for which point locality data were available were representative of those for which no point data were available. 86

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88 2. Methods

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90 2.1 Identifying optimal threshold for the habitat-land cover model to produce AOH maps

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The new set of AOH maps (Lumbierres et al., 2021b) was produced at a resolution of 100 m using a 92 novel habitat-land cover model (Lumbierres et al., 2021a) which associated the different land cover 93 classes in the Copernicus land cover map (Buchhorn et al., 2019) with the Level-1 habitat classes of 94 the IUCN habitat classification scheme (IUCN, 2012). The IUCN habitat classification scheme is a 95 hierarchy of habitat classes, and each species assessed in the IUCN Red List is assigned to one or 96 more of these habitat classes, based on available information in the literature, unpublished reports 97 and expert knowledge. The habitat-land cover model (Lumbierres et al., 2021a) has the provision of 98 associating IUCN habitat classes to land cover classes using three different thresholds (1, 2 and 3). 99 Lower thresholds permit weaker associations between land cover and habitat classes. Therefore, 100 with threshold 1 each land cover class is associated with more habitat classes than with threshold 3. 101 Lumbierres et al. (2021b) produced a set of AOH maps for each of the three different thresholds by 102 clipping out of each species' range any cells of land cover that were not linked by the model to the 103

habitat class(es) to which the species was coded, then further clipping out parts of the range fallingoutside the elevation range of the species.

In order to identify the best threshold among the three thresholds and to validate the set of AOH 106 maps with the best threshold at species level, we quantified two measures: 'model prevalence' and 107 'point prevalence'. Model prevalence is defined as the proportion of pixels inside the range that 108 were retained in the AOH. For example, if 25% of the pixels present in the original range map are 109 clipped out because they contain unsuitable habitat, fall outside the species' elevation range or both, 110 the model prevalence is 0.75. Point prevalence is defined as the proportion of point localities (or 111 their buffers) out of all points inside the range of a species falling inside the suitable pixels. For 112 example the Red-tailed Comet (Sappho sparganurus) had a total of 71 point localities within its 113 range, of which 62 fell in pixels coded as suitable in the species' AOH map, giving a point 114 prevalence of 62/71 = 0.88. 115

Because the number of habitats associated with each land cover class decreases with increasing thresholds, model prevalence is highest for threshold 1 models and lowest for threshold 3 models. With increasing threshold, commission errors are expected to decrease (which is the main purpose of AOH) but omission errors might increase. Our validation protocol therefore aimed to control for omission errors. We did this by calculating point prevalence and model prevalence across the three thresholds and identified the set of AOH maps for which the mean model prevalence was lowest without compromising the mean point prevalence.

123 The point localities for bird species were downloaded from eBird (www.ebird.org), the largest global repository for data on point localities of birds. eBird provides a metadata file called "eBird 124 basic data set" (Cornell Lab of Ornithology, 2020) which is a compilation of all the validated point 125 localities at species level and is updated monthly. These point localities are submitted by citizen 126 scientists as well as experts worldwide and are checked by local experts to remove obvious 127 misidentifications before they are made available for download (Sullivan et al., 2009). We first 128 downloaded the metadata file from eBird updated in January 2020 which was then queried in R (R 129 Core Team, 2018) using the *auk* package (Strimas-Mackey et al., 2018), as recommended by eBird, 130 to extract the point localities at species level. The taxonomy of Birdlife International (BirdLife 131 International and Handbook of the Birds of the World, 2020), which is that followed by the IUCN, 132 was matched with eBird's taxonomy and point localities of only those species common to both were 133 queried and extracted from the metadata. Of the 10,813 species listed in Birdlife International's list 134 for which AOH maps were produced, 9628 species matched by name. Of these 9628 species, 8998 135 species shared the same taxonomic concept and for 730 species the scientific names matched but 136 the taxonomic concept did not. 137

To ensure that only high-accuracy points were used for the validation, we selected the stationary 138 points from eBird's metadata. The stationary points are those that have coordinate uncertainty of 139 less than 30 m. We then applied a temporal filter of 2019-2020 because the point localities from 140 2005-2018 were used to calibrate the habitat-land cover model in Lumbierres et al. (2021a). This 141 ensured there was no overlap between the calibration and validation data. The points were further 142 filtered by the range polygon of the species provided by the IUCN Red List website (IUCN, 2020) 143 to remove the small number of points falling outside the range (many of them likely to be 144 misidentifications). Since the AOH maps in question only include a certain combination of 145 presence, origin and seasonality of the range, we used the same combination to filter the point 146 localities. This ensured that we only included points which fell inside the boundaries of the selected 147 range maps. We also made sure that only one point locality was allowed per pixel of the AOH map 148 to avoid clustering of points. Finally, we excluded species which had fewer than 10 point localities 149 after all the filters were applied. A total of 4889 bird species had 4,836,141 point localities after 150 filtering. For mammals, point localities were downloaded from GBIF (Cold Spring Harbor 151 152 Laboratory, 2021) following the taxonomy of Global Mammal Assessment (which is followed by IUCN) with same temporal and spatial filters as with birds except the filter of coordinate 153 uncertainty which was set to 300 m for mammals. This was done because far too many mammal 154 species would be excluded in the validation if we only considered point localities with coordinate 155 uncertainty of less than 30 m. The rabif package (Chamberlain et al., 2021) in R was used to 156 download the points for mammals. A total of 107,061 point localities for 420 species were available 157 for mammals after applying all the filters. 158

A buffer of 300 m was applied around all the point localities to account for the positional uncertainty of the points and for the fact that the location usually records that of the observer at the time of observation and not the focal animal, following Jung et al. (2020). The buffers of point localities were then overlaid on top of the AOH maps across all three thresholds at species level and if at least one pixel coded to suitable habitat was found inside the buffer, the pixel was considered to be validated at that point locality. The count of validated pixels was used to calculate point prevalence at species level across all three thresholds.

We identified the threshold that produced a set of AOH maps for which the mean model prevalencewas lowest without detriment to the mean point prevalence.

We then employed a two-step approach to validate the set of AOH maps with the optimal threshold. In the first step, we identified potential systematic errors in the AOH maps using a modeling approach that aimed to identify species whose model prevalence was larger or smaller than expected, given the characteristics of the species concerned. In the second step, we validated the AOH maps using point localities following Rondinini et al. (2011).

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174 2.2 A modeling approach to identify outliers

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We used logistic generalized linear models to predict model prevalence of the set of AOH maps 176 produced using the optimal threshold as a function of a number of independent variables, and 177 identified outliers whose observed model prevalence was significantly higher or lower than 178 predicted by the model. Outliers were then examined to identify systematic errors in, for example, 179 the way habitats were coded to land cover classes in the production of the AOH maps, and to 180 identify species that might be coded to the wrong habitats or elevation limits. For example, if a 181 species' range includes a high proportion of a particular land cover type not associated with the 182 suitable habitats of the species in the land cover-habitat association table (Lumbierres et al., 2021b), 183 or if errors in coding species to elevation limits mean that most of the range is outside the species' 184 185 stated limits, the model prevalence would be lower than predicted by the model.

The predictors fitted to the logistic models included: elevation range of the species (upper elevation 186 limit minus lower elevation limit), mid-point of the elevation range, number of habitats to which the 187 species is coded against in the IUCN Red List, seasonality of species (breeding and non-breeding 188 ranges in case of migratory birds) and the geographical realm of the species. In case of migratory 189 birds, Lumbierres et al. (2021b) has three different classes (resident, breeding and non-breeding 190 seasonalities) of AOH maps based on seasonality of the species. We merged resident seasonality to 191 breeding and non breeding seasonalities to have AOH maps with only two seasonalities (breeding 192 and non-breeding). The dependent variable was the model prevalence of the AOH maps. Data from 193 a total of 10475 AOH maps for 9163 bird species (including for some species with separate 194 breeding and non-breeding ranges) and 2758 AOH maps for 2758 mammal species were used to 195 build logistic regression models for birds and mammals separately using the *lme4* (Bates et al., 196 2015) package in R. Data on elevation were lacking for many mammal and bird species which is 197 the reason why not all species could be included in the logistic model. After testing taxonomic 198 genus, family and order as random effects in the model to control the non-independence of closely 199 related taxa, family was selected for fitting as the residual variance was lowest for the models with 200 family as the random effect for both birds and mammals. The predictive power of the model was 201 assessed by calculating marginal R² and conditional R² using the *insight* (Lüdecke et al., 2019) 202 package in R. The marginal R² expresses how much of the variation in data is explained by the fixed 203

204 effects and conditional R² tells how much of the variation in data is explained by both fixed and 205 random effects.

The Tukey fences outlier detection test (Wilcox, 2017) was used to identify outliers based on the difference between the estimated and observed values of model prevalence. This test uses the interquartile ranges to estimate the outliers in a data-set. The outlier test identified mild lower and upper threshold values for the difference between estimated and observed values.

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The AOH maps identified as mild upper outliers have an observed model prevalence much larger than their predicted model prevalence, whereas maps identified as mild lower outliers have an observed model prevalence much smaller than their predicted model prevalence.

217 In order to investigate the sources of errors in the outliers, we produced two more sets of AOH maps for the outliers. One set included AOH maps which were produced by clipping the range of 218 the species by the altitudinal range only (AOH Elevation only). Similarly, the other set included AOH 219 maps which were derived by clipping the range with only suitable habitat of the species (AOH Habitat 220 only). If the model prevalence of an outlier was equal or nearly equal to the model prevalence of its 221 AOH Elevation only, then we concluded that the under-representation of model prevalence could be 222 attributed to errors in elevation range of the species. If the model prevalence of an outlier was equal 223 or nearly equal to the model prevalence of AOH Habitat only, then the source of error could be attributed 224 to the mapping of the habitats inside the range using the habitat-land cover crosswalk (Lumbierres 225 et al., 2021a) or to errors in the species' habitat coding. Furthermore, in some of the outliers the 226 under-representation could result from inclusion of large proportion of habitats which were 227 unsuitable for the species but were inside the range map of the species. Outliers do not necessarily 228 represent errors in AOH, as species might legitimately have very high or low model prevalence, but 229 by identifying suites of outliers sharing common characteristics we were able to identify and correct 230 a number of systematic errors in AOH production. The models also allowed us to identify species 231 whose AOH maps might be unreliable and whose habitat and elevation coding needs to be checked. 232

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237 2.3 Point validation of AOH maps of terrestrial birds and mammals

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We validated 4889 bird and 420 mammal species' AOH maps using the filtered point localities. The point validation was done by comparing the model and point prevalence at species level. If the point prevalence exceeded model prevalence at species level, the AOH maps performed better than random, otherwise they were no better than random. We also calculated the percentage of suitable habitat pixels inside the buffers to ensure that the validation success wasn't due to one or few pixels falling inside the 300 m buffer.

One of the major issues with citizen science data is that there is often a non-representative spread of data across species. It is therefore possible that the species included in the point validation analysis are not representative of the species not included. We assessed how representative the validation sample size was by comparing the representation of variables such as family, order, genus, realm, elevation range, mid-point of the elevation range, range size and extinction risk categories for birds and mammals between species with and without point data. The point validation was done in R and GRASS (GRASS Development Team, 2017).

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253 3. Results

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After comparing point and model prevalence of 4889 birds and 420 mammal species across all the three thresholds, we selected the set of AOH maps derived by using threshold 3 in the habitat-land cover model. At threshold 3, the mean model prevalence decreased as compared to thresholds 1 and 2 with much lower change in the mean point prevalence (Table 1 and 2) for both birds and mammals.

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	Threshold 1	Threshold 2	Threshold 3
Mean model prevalence	0.81 ± 0.21 SD	0.77 ± 0.23 SD	0.65 ± 0.25 SD
Mean point prevalence	0.95 ± 0.14 SD	$0.94 \pm 0.14 \text{ SD}$	$0.90 \pm 0.17 \text{ SD}$

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Table 1: Mean model and point prevalence for AOH maps with standard deviation of 4889 bird
 species across 3 different thresholds.

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	Threshold 1	Threshold 2	Threshold 3
Mean model prevalence	$0.87 \pm 0.21 \text{ SD}$	$0.83 \pm 0.22 \text{ SD}$	0.73 ± 0.24 SD
Mean point prevalence	0.95 ± 0.14 SD	0.95 ± 0.15 SD	$0.93 \pm 0.17 \text{ SD}$

Table 2: Mean model and point prevalence for AOH maps with standard deviation of 420 mammal
species across 3 different thresholds.

We also assessed the relative contribution of elevation range, habitat, and both in reducing the range to AOH. For both birds and mammals, most of the pixels removed from the range were because either the habitat or the elevation were unsuitable, with a relatively small proportion being removed because both were unsuitable (Figs. 1,2). The proportion of the range that was clipped out on the basis of having unsuitable habitat at suitable elevations increased as model prevalence decreased, whereas there was little change across the same axis in the proportion of the range that was excluded on the basis of having suitable habitat at unsuitable elevations (Figs. 1,2). The number of both bird and mammal species peaked at model prevalence of 95-100% and gradually decreased as the model prevalence decreased.

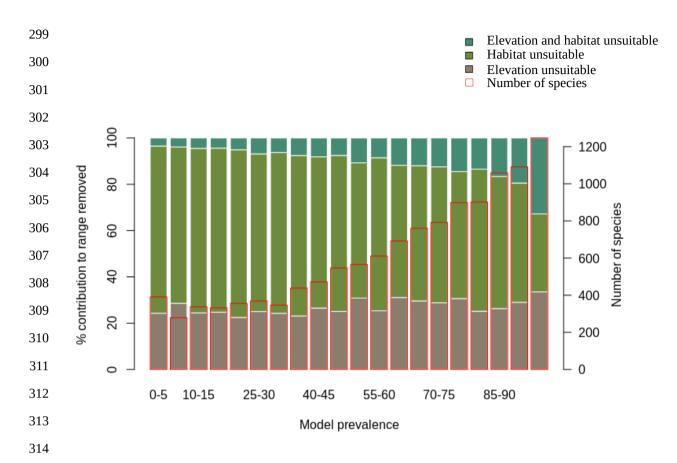


Figure 1: Percentage contribution of elevation range, habitat and both in clipping the IUCN range to produce AOH maps for birds. Each bar represents a 5% bin of model prevalence, divided to show how much of the range was clipped out due to unsuitable habitat at suitable elevations ("Habitat unsuitable"), by suitable habitat at unsuitable elevations ("Elevation unsuitable") and by unsuitable habitat at unsuitable elevations ("Elevation and habitat unsuitable"). The red blocks correspond to the second *y*-axis and show the number of species falling into each 5 % bin of model prevalence.

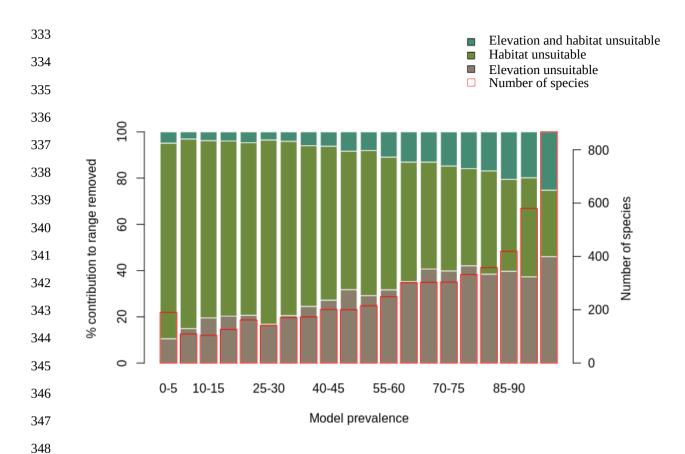


Figure 2: Percentage contribution of elevation range, habitat and both in clipping the IUCN range
to AOH for mammals. See caption to Fig. 1 for interpretation.

For birds, the logistic model identified 178 AOH maps (1.7%) as lower outliers and 118 AOH maps (1.1%) as upper outliers out of 10475 AOH maps for 9163 terrestrial bird species. Similarly for mammals, the logistic model was applied to the AOH maps of 2758 species and identified 64 (2.3%) as lower outliers and 21 (0.8%) as upper outliers.

The mean of mid-point of elevation of the bird and mammal species identified as upper outliers was 356 2725 m and 3193 m respectively while the mid-point of elevation for species which were not 357 identified as upper outliers was 1261 m for birds and 1289 m for mammals. This suggests that 358 species identified as upper outliers were those found in higher elevation. These species were 359 identified as upper outliers because the logistic models predicted low model prevalence at higher 360 elevations. Also, the range maps for high-altitude species are drawn using contour maps, therefore 361 most of the range is within the correct attitudinal band leading to high model prevalence for these 362 species. 363

The lower outliers indicate where model prevalence was possibly underestimated due to potential errors in habitat mapping/coding and elevation range of the species. We found that the habitats "Shrubland" and "Savannah" in the habitat-land cover crosswalk were not associated with the land

cover class "Herbaceous cover", leading to under-representation of these habitat types and hence 367 lower model prevalence than estimated by the logistic model (Fig. A1). We also found mismatch in 368 the elevation range and geographical range for the lower outliers (Fig. A2). There were few cases 369 where the range included large proportion of a particular land cover type which was not associated 370 with the suitable habitat of the species (Fig. A3). Moreover, we found that there was no land cover 371 information in the Copernicus land cover map for very small range polygons located on oceanic 372 islands which caused the AOH maps for these species to be empty. Furthermore, the land cover 373 class "open forest unknown" was discarded in the habitat land cover model. This led to low model 374 prevalence of AOH maps for some species whose ranges included this land cover. This was 375 corrected and a new set of AOH maps produced. 376

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378 **Point validation**

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Out of 4889 bird species (45% of all bird species) for which point data were available, 4689 (95.9%) had higher point prevalence than model prevalence and 200 species had lower point prevalence than model prevalence (Fig. 3). The mean percentage of pixels coded as suitable inside the 300 m buffers of point localities of 4889 species of birds was 62% (Fig. A5).

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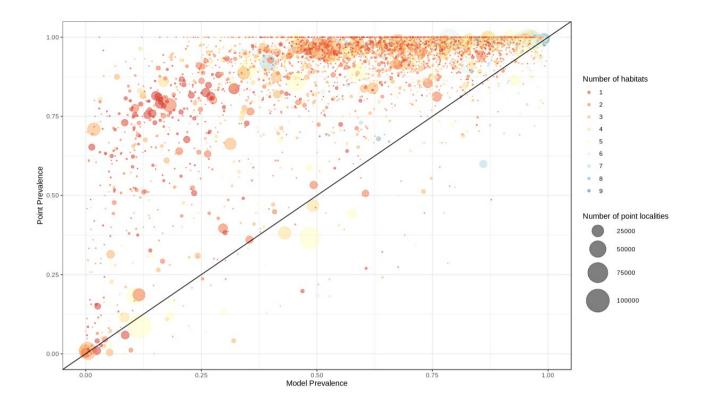
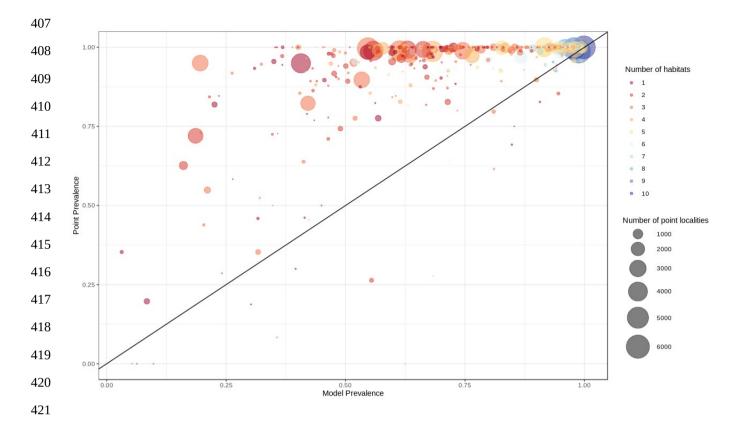
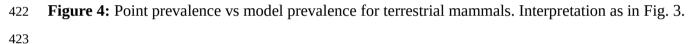


Figure 3: Point prevalence vs model prevalence for terrestrial birds. Colors indicate the number of
habitats each species is coded to, size of circles indicates the number of point localities.

Out of 420 mammal species (8% of all mammal species) for which point data were available, 399 (95.0%) had point prevalence higher than model prevalence (Fig. 4). The mean percentage of pixels coded as suitable inside the 300 m buffers of point localities of 420 species of mammals was 78% (Fig. A5).





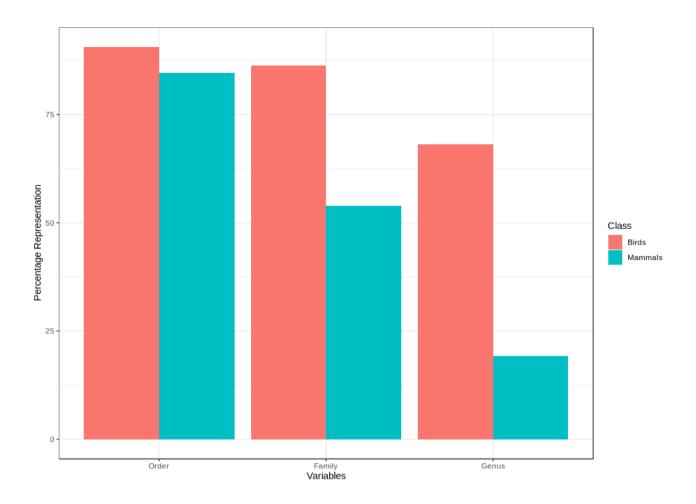
424 **Representativeness of validation sample**

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We found that for birds over 60% all families, genera and orders were represented in the sample included in the point validation and species from all biomes were represented but representation for mammals was lower, as expected due to the much lower proportion of mammal species for which point locality data were available (Fig. 5).

The validation points were spread across all of the variables and majority of their sub-classes (Fig. A6, Fig. A7). Species with validation points tended to have larger range sizes, wider elevation ranges and to be coded to more habitat classes than those without. Furthermore, validation points were not available for any critically endangered or endangered mammals as these species are rare in the wild.

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442 **Figure 5:** Taxonomic representativeness of validation sample for birds and mammals.

444 Discussion

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On comparing our point validation results with previous validation analysis of AOH maps, we 446 found that validation results are similar to or better than previous exercises. For mammals, 447 Rondinini et al. (2011) evaluated AOH maps for 263 species at 300 m resolution, of which 241 448 (91.6 %) were better than random as compared to 95.0% in our analysis. However, it should be 449 noted that the mean model prevalence for AOH maps of Rondinini et al. (2011) was 54.8 ± 21.5 SD 450 as compared to 65.16 ± 25.42 for our AOH maps. The ratio of mean point prevalence to mean 451 model prevalence for Rondinini et al. (2011) was 1.4 compared to 1.38 in our case. Ficetola et al. 452 (2015) found that AOH for 94% of 115 amphibian species used in the validation analysis were 453 better than random with the mean model prevalence for species with validation points being 0.79 \pm 454 0.21 SD. The ratio of mean point prevalence to mean model prevalence was 1.18 in this case. 455

456 Moreover, Catullo et al. (2008) found that 140 AOH maps out of 190 (73.7 %) South Asian 457 mammal species gave positive validation results while Rondinini et al. (2005) found the mean proportion of suitable habitats correctly mapped inside the range for 181 species of African vertebrates was 0.55 ± 0.01 SE using presence-absence data sets. The high validation success in our analyses could be attributed to the use of novel habitat-land cover model (Lumbierres et al., 2021a), the use of logistic regression models to identify systematic errors and the larger validation sample as compared with previous exercises. Furthermore, the underlying land cover map used in Lumbierres et al. (2021b), has the highest resolution among the global land cover maps providing more detailed land cover classification.

The point validation identified a small proportion of AOH maps which were no better than random. 465 Some of these had high model prevalence. In such cases, point prevalence must be exceptionally 466 high for the models to be better than random since even if a majority of point localities fall within 467 the AOH these maps may perform no better than random. For the AOH maps which were no better 468 than random and had low point prevalence, this was usually due to an apparent error in the coding 469 of elevation range of the species, the areas inside the range of the species where the point localities 470 fell being clipped out by what was assumed to be an erroneous elevation range. A list of species 471 472 with probably erroneous elevation coding will be forwarded to IUCN Red List team for future corrections. 473

AOH maps aim to minimize the commission errors known to be present in species ranges without increasing omission errors (Rondinini and Boitani, 2006). One of the limitations of this validation analysis is the inability to quantify the commission errors of the AOH maps as we don't have the true absence data of the species. Therefore, some uncertainty remains in AOH maps regarding the commission errors.

Also, there are some intrinsic errors in the models as identified by the logistic regression analysis. The species which are coded only to habitats like "Shrubland" might have under-represented model prevalence as discussed above. However, the number of AOH maps identified as lower outliers by the application of the logistic model was low for birds (178/10475) and for mammals (64/2758), indicating that for the majority of AOH maps the observed model prevalence was fairly close to that predicted by the model.

The AOH maps validated in this paper is the largest validation done till date in terms of number of species validated for birds and mammals. These maps will be freely available after the publication of Lumbierres et al. (2021b). We have also provided the metadata for all the species along with validation statistics in this paper which can be used as a guideline by the users while using the AOH maps.

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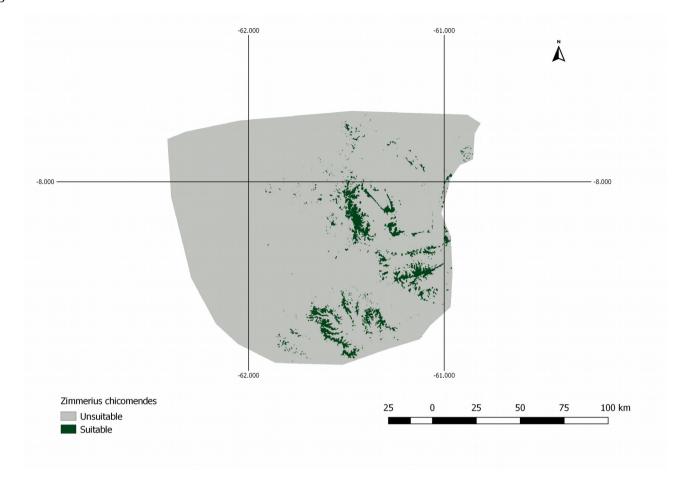


Figure A1: AOH map for species *Zimmerius chicomendesi*. The species is coded against "Forest" and "Shrubland" habitats and the elevation range falls inside the IUCN range. However, the land cover inside this range map includes a high proportion of "Herbaceous cover" land cover type which is not associated with "Shrubland" habitat in the habitat – land cover association table. Therefore, the model prevalence of this AOH is much lower than expected.

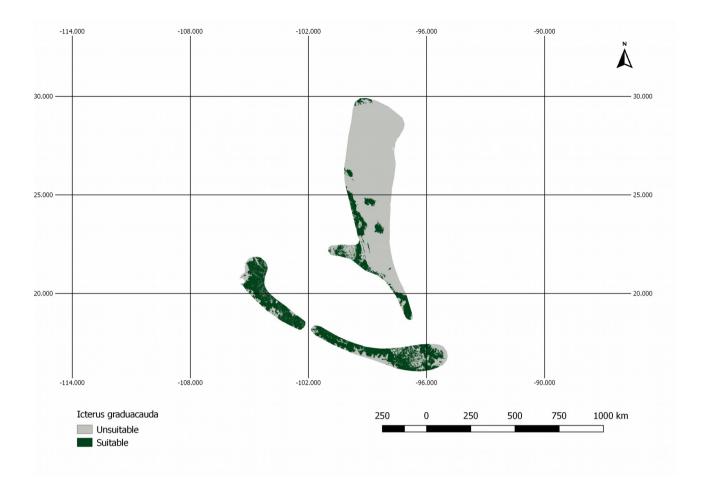


Figure A2: AOH map for the species *Icterus graduacauda*. The IUCN range of the species doesn't
cover much of the elevation range. Therefore, the model prevalence of this species is lower than
estimated.

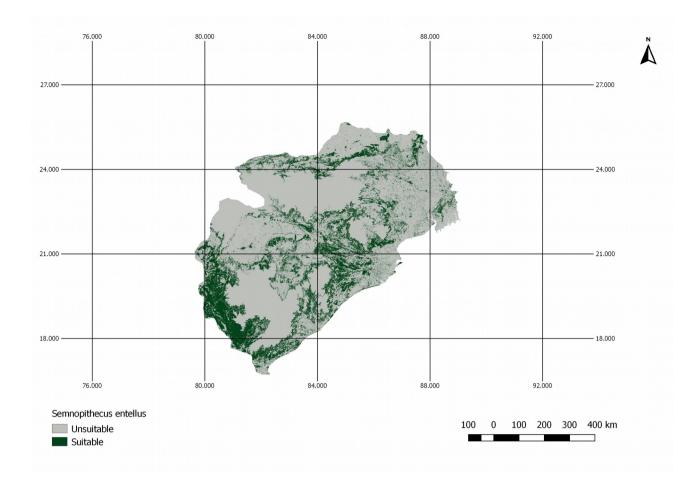


Figure A3: AOH for the species *Semnopithecus entellus*. There is a large proportion of land cover class "Cropland" inside the range map of this species. However, this species is not coded to habitats that are associated with the land cover "Cropland". Therefore, the model prevalence is lower than estimated.

	Red-tailed Comet	×Sila Tas	1.Point prevalence:
		<u> </u>	Number of points falling in suitable pixels
	All and the second second		Total number of point locations
			Red-tailed comet: 62/71= 0.88
			2.Model prevalence:
			Number of suitable habitat pixels
			Total pixels inside the range
			Red-tailed comet: 0.61
		If point prevalen random (Rondini	ce>model prevalence , the model is better than ni et al. 2010)
525	Unsuitable habitats		
526	Suitable habitats at threshold 3		
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528	Figure A4: Point validation of the AOH m	hap of the spec	ies Red-tailed Comet using model and
529	point prevalence. The yellow circles repres	ent the buffere	d point localities of Red-tailed Comet.
530	Image credit: Andres Vasquez Noboa, Macau	ılay Library MI	L 239910751
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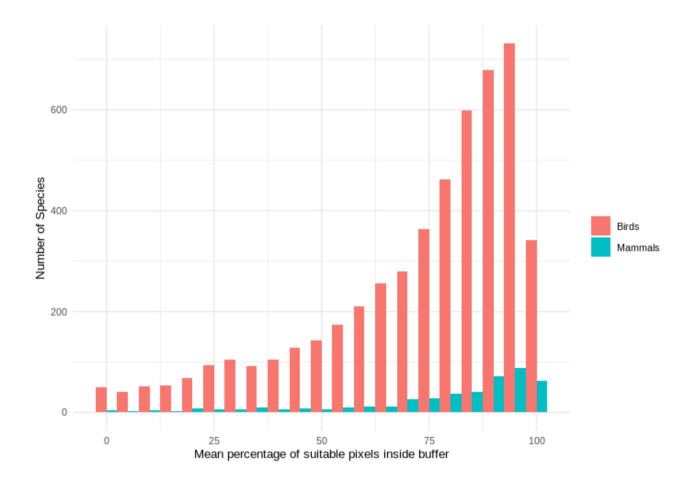


Figure A5: Histogram of mean percentage of suitable AOH pixels inside the 300 m buffer for
mammals and birds species used in point validation.

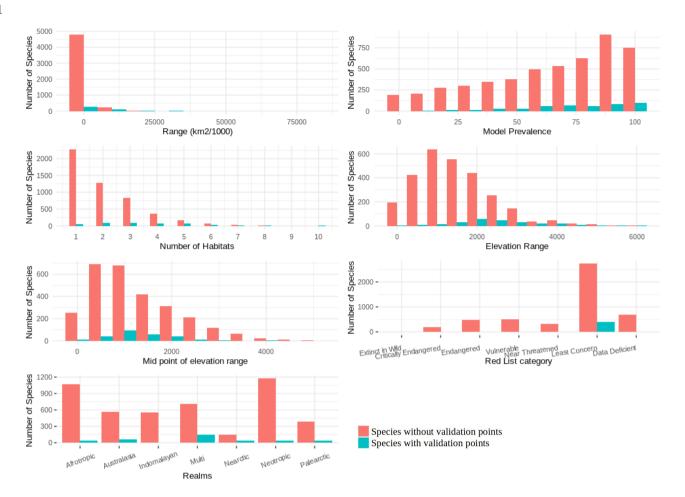
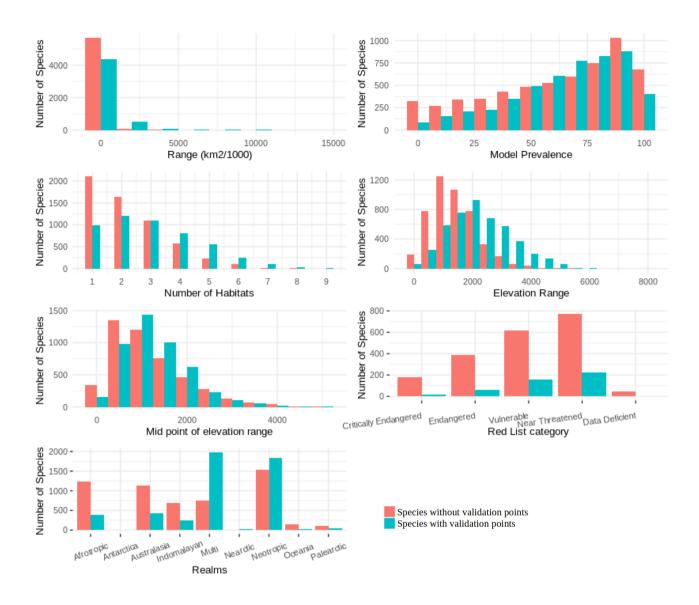


Figure A6: Comparison of species with and without validation points for mammals.



578 **Figure A7:** Comparison of species with and without validation points for birds.

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580 Data and code availability

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The point localities used in the validation analyses along with the metadata tables summarizing the validation analyses can be found at http://doi.org/10.5281/zenodo.5109073. The same DOI can be used to access the code used for validation and to also access some sample AOH maps which were validated.

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587 Author contribution

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589 PRD PFD and CR conceptualized the idea. PRD and ML curated and did the formal data analysis.

PRD led the manuscript writing with contributions from all the authors. PFD CR SHMB supervised 590 the whole process. 591 592 Acknowledgment 593 594 This research is part of the Inspire4Nature Innovative Training Network, funded by the European 595 Union's Horizon 2020 research and innovation program under the Marie Skłodowska Curie grant 596 597 agreement no. 766417. 598 599 **References:** 600 Allouche, O., A. Tsoar, and R. Kadmon.: Assessing the accuracy of species distribution models: 601 prevalence, kappa and the true skill statistic (TSS), J APPL ECOL., 43, 6, 1223 – 1232, DOI: 602 603 10.1111/j.1365-2664.2006.01214.x, 2006 604 Bates, D., M. Mächler, B. Bolker, and S. Walker.: Fitting Linear Mixed-Effects Models Using lme4, 605 J STAT SOFTW., 1406, 1, DOI: 10.18637/jss.v067.i01, 2015 606 607 BirdLife International and Handbook of the Birds of the World.: Bird species distribution maps of 608 the world., http://datazone.birdlife.org/species/requestdis, 2019 609 610 BirdLife International and Handbook of the Birds of the World.: Handbook of the Birds of the 611 World and BirdLife International digital checklist of the birds of the world. Version 5, 612 url:http://datazone.birdlife.org/userfiles/file/Species/Taxonomy/HBWBirdLife_Checklist_v5_Dec2 613 0.zip, 2020 614 615 Boyce, M. S., P. R. Vernier, S. E. Nielsen, and F. K. Schmiegelow.: Evaluating resource selection 616 functions, Ecological Modelling., 157, 281-300, DOI: 10.1016/S0304-3800(02)00200-4 617 618 Brooks, T. M., S. L. Pimm, H. R. Akçakaya, G. M. Buchanan, S. H. Butchart, W. Foden, C. Hilton-619 Taylor, M. Hoffmann, C. N. Jenkins, L. Joppa, B. V. Li, V. Menon, N. Ocampo-Peñuela, and C. 620 Rondinini.: Measuring Terrestrial Area of Habitat (AOH) and Its Utility for the IUCN Red List, 621 TRENDS ECOL EVOL., 34, 11, 977-986, DOI: https://doi.org/10.1016/j.tree.2019.06.009, 2019. 622 623

624	Buchhorn, M., B. Smets, L. Bertels, M. Lesiv, N. Tsendbazar, M. Herold and S. Fritz.: Copernicus
625	Global Land Service: Land Cover 100m: epoch 2015: Globe, Dataset of the global component of
626	the Copernicus Land Monitoring Service, doi:10.5281/zenodo.3243509, 2019
627	
628	Catullo, G., M. Masi, A. Falcucci, L. Maiorano, C. Rondinini, and L. Boitani.: A gap analysis of
629	Southeast Asian mammals based on habitat suitability models, BIOL CONSERV., 141, 11, 2730-
630	2744, DOI: 10.1016/j.biocon.2008.08.019, 2008
631	
632	Chamberlain, S., V. Barve, D. Mcglinn, D. Oldoni, P. Desmet, L. Geffert, K. Ram.: rgbif:Interface
633	to the Global Biodiversity Information Facility API. R package version 3.5.2, https://CRAN.R-
634	project.org/package=rgbif, 2021
635	
636	Cold Spring Harbor Laboratory.: Data used in Dahal PR, Lumbierres M, Butchart SHM, Donald PF
637	and Rondinini C (2021) A validation standard for Area of Habitat maps for terrestrial birds and
638	mammals, Available at: https://doi.org/10.1101/2021.07.02.450824, 2021
639	
640	Cornell Lab of Ornithology.: eBird Basic Dataset. Version: EBD_Jan 2020, Ithaca, New York, 2020
641	
642	Dahal, P. R., M. Lumbierries, S. H. M. Butchart, P. F. Donald, & C. Rondinini.: Data used,
643	summary and codes: A validation standard for Area of Habitat maps for terrestrial birds and
644	mammals [Data set]. Zenodo. http://doi.org/10.5281/zenodo.5109073, 2021
645	
646	Ficetola, G. F., C. Rondinini, A. Bonardi, D. Baisero, and E. Padoa Schioppa.: Habitat availability
647	for amphibians and extinction threat: A global analysis, DIVERS DISTRIB., 21, 3, DOI:
648	10.1111/ddi.12296, 2015
649	
650	GRASS Development Team.: Geographic Resources Analysis Support System (GRASS) Software,
651	Version 7.2. Open Source Geospatial Foundation. Electronic document:. http://grass.osgeo.org,
652	2017
653	
654	Habitats Classification Scheme (Version 3.1).: IUCN, 2012
655	
656	https://ebird.org, last access: 1 st January 2020
657	

- 658 https://www.gbif.org, https://doi.org/10.15468/dd.mezbz7
- 659
- Jenkins, C. N. and C. Giri.: Protection of mammal diversity in Central America, CONSERV BIOL.,
 22, 4, 1037-44, DOI: 10.1111/j.1523-1739.2008.00974.x, 2008

- Jung, M., P. R. Dahal, S. H. M. Butchart, P. F. Donald, X. De Lamo, M. Lesiv, V. Kapos, C.
- Rondinini, and P. Visconti.: A global map of terrestrial habitat types, Scientific data., 7, 1, 256, DOI:
 10.1038/s41597-020-00599-8, 2020
- 666
- Lüdecke, D., P. D. Waggoner, and D. Makowski.: insight: A Unified Interface to Access Information
 from Model Objects in R, Journal of Open Source Software, 4, 38, 2019
- 669
- 670 Lumbierres, M., P. R. Dahal, M. Di Marco, S. H. Butchart, P. F. Donald, and C. Rondinini.: Area of
- Habitat maps for the world's terrestrial birds and mammals, in preparation., 2021b
- 672
- Lumbierres, M., P. R. Dahal, M. Di Marco, S. H. Butchart, P. F. Donald, and C. Rondinini.: A
- habitat class to land cover translation model for mapping Area of Habitat of terrestrial vertebrates,
- 675 bioRxiv [pre-print], doi: https://doi.org/10.1101/2021.06.08.447053, 2021a
- 676
- R Core Team.: R: A language and environment for statistical computing, R Foundation for
 Statistical Computing, https://www.R-project.org/, 2018
- 679
- Rondinini, C., S. Stuart, and L. Boitani.: Habitat suitability models and the shortfall in conservation
 planning for African vertebrates, CONSERV BIOL., 19, 5, 1488 1497, DOI: 10.1111/j.15231739.2005.00204.x, 2005
- 683
- Rondinini C.& Boitani L.: Differences in the umbrella effects of African amphibians and mammals
 based on two estimators of the area of occupancy, CONSERV BIOL., 20, 170-179, DOI:
 10.1111/j.1523-1739.2005.00299.x, 2006
- 687
- Rondinini, C., M. D. Marco, F. Chiozza, G. Santulli, D. Baisero, P. Visconti, M. Hoffmann, J.
- 689 Schipper, S. N. Stuart, M. F.Tognelli, G. Amori, A. Falcucci, L. Maiorano, and L. Boitani.: Global
- habitat suitability models of terrestrial mammals, PHILOS T R SOC B., 366, 1578, 2633-41, DOI:
- 691 10.1098/rstb.2011.0113, 2011

692	
693	Stoms, D. M., F. W. Davis, and C. B. Cogan. : Sensitivity of wildlife habitat models to uncertainties
694	in GIS data, PHOTOGRAMM ENG REM S., 58, 843- 850, 1992.
695	
696	Strimas-Mackey, M., E. Miller, W. Hochachka.: auk: eBird Data Extraction and Processing with
697	AWK. R package version 0.3.0, https://cornelllabofornithology.github.io/auk/, 2018
698	
699	Sullivan, L., B., C. L.Wood, M. J. Iliff, R. E. Bonney, D. Fink, and S. Kelling.: eBird: A citizen-
700	based bird observation network in the biological sciences, BIOL CONSERV., 142, 10, 2009
701	
702	The IUCN Red List of Threatened Species. Version 2020-2.: IUCN, 2020
703	
704	Tracewski, L., S. H. Butchart, M. Di Marco, G. F. Ficetola, C. Rondinini, A. Symes, H. Wheatley,
705	A. E. Beresford, and G. M.Buchanan (2016).: Toward quantification of the impact of 21 st -century
706	deforestation on the extinction risk of terrestrial vertebrates, CONSERV BIOL., 30, 5, 2016
707	
708	Wilcox, R. R., Introduction to robust estimation and hypothesis testing.: 4th edition, Elsevier,713
709	Waltham, Massachusetts, USA, 2017
710	
711	
/ 11	