A model for marine sedimentary carbonate diagenesis and paleoclimate proxy signal tracking: IMP ~~v0.9~~v1.0

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Abstract. The preservation of calcium carbonate in marine sediments is central to controlling the alkalinity balance of the ocean and hence the ocean-atmosphere partitioning of CO_2. To successfully address carbon cycle-climate dynamics on geologic ($\gg$ 1 kyr) time-scales, Earth system models then require an appropriate representation of the primary controls on $CaCO_3$ preservation. At the same time, marine sedimentary carbonates represent a major archive of Earth history, as they have the potential to preserve how seawater chemistry, and isotopic composition, and even properties of planktic and benthic ecosystems, change with time. However, changes in preservation and even chemical erosion of previously deposited $CaCO_3$, together with the biogenic reworking of upper portions of sediments whereby sediment particles are translocated both locally and non-locally between different depths in the sediments, all act to distort the recorded signal. Numerical models can aid in recovering what the 'true' environmental changes might have been, but only if they appropriately account for these processes.

Building on a classical 1-D reaction-transport framework, we present a new diagenetic model – IMP (Implicit model of Multiple Particles (and diagenesis)) – that simulates biogeochemical transformations in carbonate-hosted proxy signals by allowing for populations of solid carbonate particles to possess different physicochemical characteristics such as isotopic value, solubility, and particle size. The model also utilizes a variable transition matrix to implement different styles of bioturbation. We illustrate the utility of the model for deciphering past environmental changes using several hypothesized transitions of seawater proxies obscured by sediment mixing and chemical erosion. To facilitate the use of IMP, we provide the model in FORTRAN, MATLAB, and Python versions. We described IMP with integration into Earth system models in mind, and present the description of this coupling of IMP with the 'cGENIE.muffin' model in a subsequent paper.

1 Introduction

The removal of carbon and alkalinity through the preservation and burial of carbonate minerals in accumulating marine sediments ~~,~~ plays a central role in the global carbon cycle and hence the regulation of climate over geologic time-scales (e.g., Ridgwell and Zeebe, 2005; Kump et al., 2009). Specifically – burial of $CaCO_3$ is the major long-term sink for atmospheric

CO_2 ($> \sim 10^4$ yr), while chemical erosion of $CaCO_3$ works as a buffer against short-term ($\sim 10^2$ to 10^4 yr) ocean acidification that accompanies CO_2 emissions (e.g., Broecker and Takahashi, 1977; Berner et al., 1983; Archer et al., 1998; Ridgwell and Zeebe, 2005). As such, the dynamics of the calcium carbonate cycle are also important to the stability of the marine environment inhabited by calcifying (and carbonate chemistry sensitive) organisms such as corals (Hönisch et al., 2012) and takes on particular importance in the ~~contest~~ context of the release of CO_2 to the ocean-atmosphere system, both past and present/future (e.g., Archer et al., 1997, 1998; Zeebe and Zachos, 2007; Boudreau et al., 2010; Lord et al., 2016; Penman et al., 2016).

Although calcium carbonate can be produced diagenetically within the sediments (which we do not address in this initial version of the model and will not discuss in any detail in this paper), $CaCO_3$ is predominantly delivered to ocean sediments from calcifying organisms (principally plankton) living in the overlying ocean surface, with a minor contribution from organisms living at or close to the sediment surface itself. Two polymorphs exist – calcite (trigonal), which is precipitated by foraminifera and coccolithophores, and aragonite (orthorhombic), which is precipitated by e.g. modern corals and pteropods. Deep-sea sediments and hence marine archives are generally dominated by the calcitic form (although our model is designed to be sufficiently flexible to consider a mix of polymorphs). The crystal structure of $CaCO_3$ allows for the substitution of a variety of trace elements, which together with measurable isotopic properties of most of these elements, serves as an important archive of paleoceanographic proxies. For example, the δ^{13}C record of $CaCO_3$ has been widely used to constrain C transfers between reservoirs (e.g., Kump and Arthur, 1999), the δ^{18}O record to reconstruct past water temperature and/or global ice volume (e.g., Zachos et al., 2001; Dunkley Jones et al., 2013), the δ^{11}B record for paleo-ocean pH reconstruction (e.g., Gutjahr et al., 2017), and I/Ca ratios to estimate ocean redox state in the past (e.g., Lu et al., 2018). However, reconstruction of paleo-environments using $CaCO_3$-based proxies is complicated by $CaCO_3$ loss via dissolution (chemical erosion) and mixing of $CaCO_3$ particles within sediments by benthic organisms (bioturbation). Both phenomena are ubiquitous and need to be accounted for when one reads proxies in sedimentary carbonates, particularly for events that occur rapidly relative to the sediment accumulation timescale (e.g., Bard et al., 1987; Ridgwell, 2007b; Trauth, 2013).

The effect of bio-mixing on the preservation of proxy signals has been examined analytically and numerically depending on the complexity with which sediment bioturbation is represented (e.g., Berger et al., 1977; Bard et al., 1987; Trauth, 1998, 2013; Hull et al., 2011; Steiner et al., 2016; Kirtland Turner et al., 2017). Most of these studies assume either random mixing or diffusion that follows Fick's law (biodiffusion) for bioturbation. Particle mixing by benthos, however, can be more complex than can be captured by biodiffusion or random mixing, as it depends on animal-specific properties such as burrow geometry and feeding rates and styles (e.g., Meysman et al., 2006; Kristensen et al., 2012). For example, Boudreau and Imboden (1987) suggested, based on their analytical examination of the effect of non-local mixing on distributions of radiotracers, that animal-specific mixing can result in different sediment particle distributions over time than simple biodiffusion. Therefore, specific, more complex animal behaviors and the resulting bio-mixing need to be simulated with a transition matrix method (e.g., Shull, 2001) or a process-based particle-tracking model such as the automaton simulator LABS (Boudreau et al., 2001; Choi et al., 2002; Kanzaki et al., 2019). Specific animal behaviors can be reflected by probabilities in the transition matrix or as automaton rules in LABS. Other (more common) models simply employ a biodiffusion coefficient and consider only bulk properties (e.g., Ridgwell, 2007a,b), simplifying how proxy signals are recorded still further.

Chemical erosion is also known to distort proxy signals ~~(e.g., Keir, 1984; Broecker et al., 1991; Oxburgh, 1998; Barker et al., 2007; Ridg~~ (e.g., Keir, 1984; Broecker et al., 1991; Keir and Michel, 1993; Oxburgh, 1998; Barker et al., 2007; Ridgwell, 2007b; Jennions et al., 2015

60 . Moreover, it has been shown that the extent of signal distortion by chemical erosion is related to the strength of biodiffusion (e.g., Keir, 1984). Generally, however, examination of the effect of chemical erosion on proxy signals has been relatively limited compared to that of bioturbation. Most previous studies have focused on explaining older ^{14}C ages in sedimentary $CaCO_3$ that suffers more significant dissolution ~~(Keir, 1984; Broecker et al., 1991; Oxburgh and Broecker, 1993; Oxburgh, 1998; Barker et al., 2007)~~ (Keir, 1984; Broecker et al., 1991; Keir and Michel, 1993; Oxburgh and Broecker, 1993; Oxburgh, 1998; Barker et al., 2007), and

65 the models used therein cannot be directly applied to other proxies. Only a limited number of studies have quantitatively discussed the effect of dissolution on other proxy signals (e.g., δ^{13}C by Jennions et al., 2015). The reason for this is that published sediment mixing models ~~are generally unable to realize~~ do not generally account for diagenetic reactions (e.g., Trauth, 2013) and even those that enable $CaCO_3$ dissolution are too specific regarding the tracked proxy and style of bioturbation and thus inapplicable to a variety of proxies or to different styles of bioturbation (e.g., Keir, 1984).

70 Caution is particularly warranted in the interpretation of $CaCO_3$-hosted proxy records during episodes of ocean acidification when both chemical erosion (e.g., Zachos et al., 2005) but also changes in benthic ecology and hence bioturbation (e.g., Jennions et al., 2015) are expected, e.g., during hyperthermal events in the early Cenozoic (e.g., Ridgwell, 2007b; Sluijs et al., 2007; McInerney and Wing, 2011). Currently, no model exists that is specifically designed to simulate $CaCO_3$ diagenesis along with different styles of bioturbation, while simultaneously tracking a variety of proxy signals, and hence explicitly tackle

75 complex past geochemical-biological sediment proxy questions.

Here we present the 'Implicit model of Multiple Particles (and diagenesis)' – IMP – that can be used to explore the consequences of chemical erosion and bioturbation on proxy records. IMP is ~~at heart,~~ a reactive-transport model of diagenesis for carbonates, organic matter and refractory detrital materials in marine sediments, along with dissolved oxygen and aqueous CO_2 species in the porewater. Overlaying this ~~,~~ is the ability to track proxy signals in carbonates by representing multiple 'classes' of

80 carbonates particles with different proxy values (for more details see Section 2.1). IMP also has the flexibility of representing various styles of solid phase mixing through the use of different transition matrices. Thus, the model can be used to simulate a wide variety of scenarios of environmental change. Following the presentation of the model framework, we illustrate how the model can be utilized to discern signal distortion caused by chemical erosion and different kinds of bioturbation and better interpret proxy signals for paleo-environments.

85 ## 2 Model description

2.1 Model overview

IMP builds on the reactive-transport framework of Archer (1991) and as such is based on the principals of conservation of carbonate alkalinity and total CO_2 in sediment porewater. However, IMP extends the Archer (1991) model to (i) be explicit about depth-dependent and temporal changes of all considered species, (ii) allow more than one 'class' of $CaCO_3$ particles

90 (see below for the definition of 'class'), and (iii) simulate a variety of mixing styles caused by bioturbation using transition matrices.

The term CaCO$_3$ 'class' refers here to any ensemble of solid CaCO$_3$ particles that (a) record the same proxy value or (b) ~~share~~ possess common, distinct biological and physicochemical characteristics. As an example of the former case ((a) above), if two ensembles of CaCO$_3$ particles have distinctive proxy signals (e.g., different δ^{13}C and/or δ^{18}O values), we refer to these

95 two ensembles as two distinctive CaCO$_3$ classes, even if they belong to the same model species and have exactly the same geochemical properties (i.e. in a 'traditional' reactive-transport framework such as of Archer (1991), this would all just be 'CaCO$_3$'). Similarly ((b) above), if two ensembles of CaCO$_3$ particles belong to different model species (e.g., having distinct sizes and associated dissolution and bio-mixing properties; Keir, 1980; Walter and Morse, 1984, 1985; Bard, 2001; Schmidt et al., 2004), they are referred to as two distinctive CaCO$_3$ classes even when they record the same proxy values (but could

100 ~~now, and~~ not, yet should be distinguished, in a 'traditional' reactive-transport framework). IMP can thus be regarded analogous to the multi-G model of Berner (1980), which separates bulk organic matter into multiple classes of organic compounds with different reactivities. However, the basis upon which we separate bulk CaCO$_3$ into multiple classes of CaCO$_3$ particles is more flexible, as these are not limited to reactivity, but can be any combination of proxy signals as well as biological and physicochemical characteristics. In theory, IMP can simulate the effect of diagenesis and bioturbation on individual CaCO$_3$

105 particles by increasing the total number of CaCO$_3$ classes, though this results in increased computational costs. Our new approach is the first combined diagenetic, bioturbation model to pseudo-explicitly track proxy signals recorded in bulk CaCO$_3$ in the sediment column. This is realized by simulating the depth and time-dependent distribution of more than one CaCO$_3$ class each with distinct proxy signals.

In the following sections, we provide a detailed description of IMP in which the governing equations (Section 2.2), the

110 numerical solutions (Section 2.3), ~~the implementation of transition matrices (Section 2.4),~~ and the simulation of signal tracking (Section ~~2.5~~2.4) are highlighted. The default values of independent parameters (Table 1) and the equations of dependent parameters (Table 2) and thermodynamic parameters (Table 3) are tabulated. The model code for IMP v.1.0 ~~.9~~ is available in Fortran90, MATLAB, and Python (see Code availability).

2.2 Governing equations

115 For solid phase species, IMP considers multiple (n_{cc}) classes of CaCO$_3$ particles, plus a single class of organic matter (OM) with the assumed chemical formula of CH$_2$O, and (a single class of) non-reactive detrital material (referred to as 'clay' hereafter) to act as a 'dilatant' and help determine the final burial velocity. The rate of change with time of the concentrations of these solid species in marine sediments are represented following the classic generalized equations of Boudreau (1997):

$$\frac{\partial(1-\phi)m_\theta}{\partial t} = -\frac{\partial(1-\phi)wm_\theta}{\partial z} - R_\theta - (1-\phi)m_\theta \int_0^{z_{ml}} E_\theta(z,z')dz' + \int_0^{z_{ml}} \{1-\phi(z')\}m_\theta(z')E_\theta(z',z)dz' \tag{1}$$

120 where m_θ (mol cm^{-3}) represents the concentration of solid phase species $\theta \in \{\ell, \text{OM, clay}; \text{here } \ell = 1, 2, \ldots, n_{cc}\}$, ϕ is the porosity, t is the time (yr), $E_\theta(z,z')$ represents the continuous exchange function (cm^{-1} yr^{-1}), which describes transport of

solid species θ from sediment depth z (cm) to any other depth z' (cm) (Section 2.2.2), w is the burial velocity (cm yr^{-1}), z_{ml} is the thickness of the mixed layer (cm), and R_θ (mol cm^{-3} yr^{-1}) represents the net consumption rate of species θ through all biogeochemical reactions. On the right-hand side of Eq. (1) the total change in concentration of the solid species θ is expressed as the change due to advective transport (1st term), biogeochemical reactions (2nd term) and bioturbational transport (3rd and 4th term, note that there is no separate biodiffusion term).

For aqueous species, IMP considers dissolved oxygen (O_2), total dissolved CO_2 species (DIC) and carbonate alkalinity (ALK). The generalized equation for these aqueous species is given by Archer (1991):

$$\frac{\partial \phi c_\sigma}{\partial t} = \frac{\partial}{\partial z}\left(\frac{D_\sigma}{F}\frac{\partial c_\sigma}{\partial z}\right) + R_\sigma \tag{2}$$

where c_σ represents the concentration (mol cm^{-3}), D_σ the diffusion coefficient (cm^2 yr^{-1}) and R_σ the net production rate from all biogeochemical reactions (mol cm^{-3} yr^{-1}) for aqueous species $\sigma \in \{O_2, \mathrm{DIC}, \mathrm{ALK}\}$; and F represents the sediment formation factor (related to the tortuosity; Ullman and Aller, 1982).

2.2.1 Biogeochemical reactions

Following Archer (1991), IMP considers degradation of organic matter and dissolution of $CaCO_3$ ~~,~~ as the main biogeochemical reactions occurring in marine sediments. ~~(In this version of IMP, we omit the role and geochemistry of opal and its dissolved pore-water ph~~ (In this version of IMP, we omit the role and geochemistry of opal and its dissolved pore-water phase, silicic acid, but see e.g. Ridgwell et

The reaction term for organic matter is given by

$$R_{\mathrm{OM}} = (1 - \phi)m_{\mathrm{OM}}k_{\mathrm{OM}} \tag{3}$$

where k_{OM} is the first-order degradation rate constant for organic matter (yr^{-1}). To account for anaerobic degradation of organic matter by SO_4, IMP simulates an anoxic pathway below the dynamically calculated oxygen penetration depth (z_{ox}). Different rate constants for oxic (k_{ox}) and anoxic (k_{anox}) degradation can be adopted:

$$k_{\mathrm{OM}} = \begin{cases} k_{\mathrm{ox}} & (z \leq z_{\mathrm{ox}}) \\ k_{\mathrm{anox}} & (z > z_{\mathrm{ox}}) \end{cases} \tag{4}$$

Following Archer (1991), both rate constants are considered the same for the initial validation of our model in this study. While clearly an oversimplification, it serves as a first approximation of the importance of OM degradation on calcite dissolution and is furthermore a requirement in order to be able to benchmark IMP to the model of Archer (1991). Although other pathways are used to degrade organic matter in marine sediments, such as nitrate and metal oxides, these have been shown to be quantitatively of less importance on a global scale (combined likely < 20%, Archer et al., 2002; Thullner et al., 2009). It is, however, possible to artificially add DIC and ALK fluxes at a given depth, thus simulating the production of ALK and DIC from a pathway that is not explicitly simulated (Supplementary Material) .

The reaction term for any class ℓ of $CaCO_3$ particles is given by:

$$R_\ell = (1-\phi)m_\ell k_{cc,\ell}(1-\Omega_{cc})^{\eta_{cc}} H(1-\Omega_{cc}) \tag{5}$$

where $k_{cc,\ell}$ is the rate constant (yr^{-1}), Ω_{cc} the saturation degree and η_{cc} the reaction order for $CaCO_3$ dissolution and the Heaviside function H guarantees that net $CaCO_3$ precipitation does not occur (Archer, 1991). Note that the model allows assignment of different dissolution rate constants to different classes of $CaCO_3$ particles (e.g., Keir, 1980). For this study, however, unless otherwise described, we assume a dissolution rate of $k_{cc,\ell} = 365.25\ yr^{-1}$ for all classes, a value determined by Archer (1991).

The clay species is assumed to be non-reactive. Hence,

$$R_{clay} = 0 \tag{6}$$

The reaction terms for aqueous species O_2, DIC and ALK are correspondingly given by (Archer, 1991)

$$R_{O_2} = -\gamma_{O_2\text{-}OM}(1-\phi)m_{OM}k_{ox} \tag{7}$$

$$R_{DIC} = R_{OM} + \sum_{\ell=1}^{n_{cc}} R_\ell \tag{8}$$

$$R_{ALK} = (1-\phi)m_{OM}k_{anox} + 2\sum_{\ell=1}^{n_{cc}} R_\ell \tag{9}$$

where $\gamma_{O_2\text{-}OM}$ in Eq. (7) is the mole ratio of oxygen to organic matter consumed upon oxic degradation of organic matter. We assume that the aqueous carbonate system is always at equilibrium, and calculate the partitioning of the aqueous carbonate species (H_2CO_3, HCO_3^- and CO_3^{2-}) based on alkalinity and DIC concentrations in conjunction with the apparent equilibrium dissociation constants adjusted for pressure, salinity and temperature (Tables 2 and 3). Other options to utilize published routines for the calculation of the aqueous carbonate system, mocsy 2.0 (Orr and Epitalon, 2015) and CO2SYS (Lewis and Wallace, 1998; van Heuven et al., 1998; Humphreys et al., 2020), are presented in the Supplementary material.

2.2.2 Bioturbation

Bio-mixing of solid-phase species in the model is simulated by means of a transition matrix. A wide range of bio-mixing styles can be captured by the transition matrix because a transport probability of solid particles from one sediment layer to another can be specified with the value of a cell whose row and column numbers correspond to the two layers between which particles are transported. Thus, the use of the transition matrix facilitates the implementation of user-defined/biology-based particle mixing, whether local or non-local (e.g., Trauth, 1998; Shull, 2001). In this section, we elaborate upon how the bioturbation term in Eq. (1) can be derived from the transition matrix.

The rate at which particles of solid species θ are transported from layer i to layer j, $P_{\theta,ij}$ (yr^{-1}), is given by:

$$P_{\theta,ij} = \frac{N_{\theta,ij}}{\sum_{j=1}^{n_{ml}} N_{\theta,ij}} \frac{1}{\tau} \tag{10}$$

where $N_{\theta,ij}$ is the number of particles of species θ moved from layer i to layer j, n_{ml} is the total number of layers within the bioturbated zone and τ is the time (yr) required for the displacements. Note that $P_{\theta,ij} \times \tau$ represents the particle transport probability and corresponds to components at (i, j) of the transition matrix (Trauth, 1998; Shull, 2001). When bioturbation causes mixing of sediment particles based on the above transport rate, the number of particles of species θ in layer i changes with time according to:

$$\frac{dN_{\theta,i}}{dt} = -N_{\theta,i} \sum_{j=1}^{n_{\mathrm{ml}}} P_{\theta,ij} + \sum_{j=1}^{n_{\mathrm{ml}}} N_{\theta,j} P_{\theta,ji} \tag{11}$$

where $N_{\theta,i}$ is the total number of particles of species θ in layer i (compare Eq. (11) with Eq. (3.117) of Boudreau (1997)).

The concentration of species θ in layer i, $m_{\theta,i}$ (mol cm^{-3}), can be given by (cf., Boudreau, 1997):

$$(1 - \phi_i) m_{\theta,i} \equiv \frac{\alpha_\theta N_{\theta,i}}{A dz_i} \frac{\alpha_\theta N_{\theta,i}}{A \delta z_i} \tag{12}$$

where ϕ_i and $dz_i \, \delta z_i$ are the porosity and the thickness (cm) of layer i, α_θ represents the moles of species θ (mol) included in one particle and A is the ~~considered~~ cross-sectional area in the model (cm^2). Then one can deduce from Eqs. (11) and (12):

$$\frac{d(1 - \phi_i) m_{\theta,i}}{dt} = -(1 - \phi_i) m_{\theta,i} \sum_{j=1}^{n_{\mathrm{ml}}} P_{\theta,ij} + \sum_{j=1}^{n_{\mathrm{ml}}} (1 - \phi_j) \frac{dz_j}{dz_i} \frac{\delta z_j}{\delta z_i} m_{\theta,j} P_{\theta,ji} \tag{13}$$

(compare Eq. (13) with Eq. (3.118) of Boudreau (1997)). Eq. (13) can be simplified with a modified transition matrix for species θ, with components at (i, j) denoted as $K_{\theta,ij}$ and calculated based on the particle transport rate $P_{\theta,ij}$:

$$K_{\theta,ij} = \begin{cases} \delta z_i P_{\theta,ij} / \delta z_j & (i \neq j) \\ -\sum_{j \neq i}^{n_{\mathrm{ml}}} P_{\theta,ij} & (i = j) \end{cases} \tag{14}$$

Using Eq. (14), we can rewrite Eq. (13) as a function of $K_{\theta,ij}$:

$$\frac{d(1 - \phi_i) m_{\theta,i}}{dt} = \sum_{j}^{n_{\mathrm{ml}}} (1 - \phi_j) m_{\theta,j} K_{\theta,ji} \tag{15}$$

Formulation of bioturbation in a continuum system needs a corresponding continuous function. We define a continuous exchange function E_θ (cm^{-1} yr^{-1}) as (cf., Boudreau, 1997):

$$E_\theta(z_i, z_j) \equiv \lim_{dz_j \to 0 \, \delta z_j \to 0} (P_{\theta,ij} / dz \, \delta z_j) \tag{16}$$

where z_i and z_j denote the depths of sediment-layer i and j. With Eq. (16), we can write a continuous form of Eq. (13) in the limits of zero thicknesses for discretized sediment layers:

$$\frac{\partial(1 - \phi) m_\theta}{\partial t} = -(1 - \phi) m_\theta \int_0^{z_{\mathrm{ml}}} E_\theta(z, z') dz' + \int_0^{z_{\mathrm{ml}}} \{1 - \phi(z')\} m_\theta(z') E_\theta(z', z) dz' \tag{17}$$

Here, z' denotes any depth except at z and z_{ml} is the thickness of the mixed layer. Eq. (17) is the same as Eq. (3.121) of Boudreau (1997) and the two bioturbation terms in Eq. (1). Note that Eq. (15) is a finite difference ~~equation~~ version of Eq. (17)

is Eq. (15)and the transition matrix corrected for porosity therein (i.e., $(1 - \phi_i)K_{\theta,ij}$ representing components at (i, ~~formulated~~ ~~with transition matrix , which is used to solve the~~ j)) corresponds to the bioturbational transport part of the Jacobian matrix for species θ, which is used for solving the governing equations (Section 2.3).

Three different transition matrices were created for the present study to illustrate different styles of bio-mixing (Fig. 1): Fickian mixing, homogeneous mixing, and the more mechanistic automaton-based mixing simulated by the particle-tracking bioturbation simulator LABS (e.g., Boudreau et al., 2001; Choi et al., 2002; Kanzaki et al., 2019).

The transition matrix that assumes Fickian diffusion for bioturbation (parameterized with $D_{b,\theta}$, Goldberg and Koide, 1962), can be expressed by:

$$K_{\theta,ij} = \begin{cases} -K_{\theta,ij}(j = i+1) & (i = j = 1) \\ -K_{\theta,ij}(j = i+1) - K_{\theta,ij}(j = i-1) & (1 < i = j < n_{\mathrm{ml}}) \\ -K_{\theta,ij}(j = i-1) & (i = j = n_{\mathrm{ml}}) \\ \{(1-\phi_i)D_{b,\theta,i} + (1-\phi_j)D_{b,\theta,j}\}/\{\delta z_i(1-\phi_i)(\delta z_i + \delta z_j)\} & (2 \le j = i+1 \le n_{\mathrm{ml}} \text{ or } 1 \le j = i-1 \le n_{\mathrm{ml}}-1) \\ 0 & (\text{else}) \end{cases}$$

(18)

where $D_{b,\theta,i}$ represents the biodiffusion coefficient for solid species θ at sediment layer i. As a default biodiffusion parameterization, a depth-independent value of $0.15 \mathrm{~cm}^2 \mathrm{~yr}^{-1}$ is assumed (Emerson, 1985, Table 1). Note that the biodiffusion considered in this study is only intraphase biodiffusion and does not include interphase biodiffusion (e.g., Meysman et al., 2005; Munhoven, 2021). The implementation of interphase biodiffusion requires a different transition matrix.

The transition matrix for homogeneous mixing can be given by:

$$K_{\theta,ij} = \begin{cases} \delta z_i P_{h,\theta}/\delta z_j & (i \ne j \text{ and } 1 \le i, j \le n_{\mathrm{ml}}) \\ -(n_{\mathrm{ml}} - 1)P_{h,\theta} & (1 \le i = j \le n_{\mathrm{ml}}) \\ 0 & (\text{else}) \end{cases}$$

(19)

where $P_{h,\theta}$ (yr^{-1}) is the homogeneous transport rate for solid species θ. A value of $10^{-3} \mathrm{~yr}^{-1}$ is assumed for the default homogeneous mixing (Table 1).

To obtain the mechanistic automaton-based transition matrix, we utilized the eLABS v.0.2 code, the latest release of Lattice-Automaton Bioturbation Simulater (LABS) by Kanzaki et al. (2019), with which a transition matrix can be extracted based on Eqs. (10) and (14). The new features of LABS added by Kanzaki et al. (2019), i.e., 2D porewater flow and diagenesis, were disabled to extract mixing controlled dominantly by benthos biology as in Boudreau et al. (2001). A 200-yr LABS simulation was run with a deposit feeder with a body size of $0.25 \times 0.25 \times 1.65 \mathrm{~cm}^3$, a locomotion speed of $10 \mathrm{~cm~day}^{-1}$ and a maximum ingestion rate of 1 g sediment (g organism)$^{-1}$ day^{-1} in a $0.25 \times 12 \times 15 \mathrm{~cm}^3$ 3D sediment system. Transition matrices were created every 10 model days (cf. Reed et al., 2007) and the averaged transition matrix over 200 model years multipled with a factor of 1/10 was adopted to represent the transition matrix derived from the above LABS simulation. The

factor of 1/10 was introduced above because otherwise the LABS mixing has a relatively high mixing intensity (equivalent to a biodiffusion and also to facilitate the numerical solution of the model with the LABS mixing (see below).

The default transition matrices, corrected for porosity, are shown in Fig. 1. Fickian mixing is a local mixing, allowing translocation of particles only between adjacent sediment layers, resulting in a tridiagonal matrix (Fig. 1a). On the other hand, non-local mixing (homogeneous and LABS mixing) allows transportation of particles between remote layers, and thus is characterized with the spread of non-zero components away from the main diagonal in the transition matrix (Figs. 1b and c). As defined in Eq. (19) the transition matrix for homogeneous mixing has components that systematically change with rows and columns (Fig. 1b) compared to the transition matrix for LABS mixing that has more randomly spread non-continuous values (Fig. 1c). The porosity-corrected transition matrix corresponds to the bioturbational transport part of the Jacobian matrix used for solving the governing equations (Section 2.3). Therefore, the difficulty to achieve a numerical solution of the model differs between chosen mixing styles reflecting corresponding transition matrices: in general, this is the least difficult with Fickian mixing and most difficult with LABS mixing (Fig. 1). Note that the transition matrices for Fickian and homogeneous mixing change with assumed mixed layer depth (related to n_{ml}) and/or parameters that define mixing intensity ($D_{b,\theta}$ and $P_{b,\theta}$) as in Eqs. (18) and (19) (cf. Table 1). Additional LABS simulations, with variations related to deposit feeder behavior and/or modified sediment grid dimensions, are necessary to generate a new LABS-based transition matrix.

2.2.3 Burial velocity/advection

The burial velocity in IMP changes according to the volume change of solid material caused by biogeochemical reactions and ~~non-local mixing~~ bio-mixing because a constant, time-independent porosity profile is assumed (Eq. (~~21~~23)). This section describes how the change in burial rate is calculated in the model.

Multiplying the governing equation (Eq. (1)) with the molar volume V_θ (cm^3 mol^{-1}) for solid species θ leads to:

$$\frac{\partial(1-\phi)V_\theta m_\theta}{\partial t} = -\frac{\partial(1-\phi)wV_\theta m_\theta}{\partial z} - V_\theta R_\theta + V_\theta \left[-(1-\phi)m_\theta \int_0^{z_{ml}} E_\theta(z,z')dz' + \int_0^{z_{ml}} \{1-\phi(z')\}m_\theta(z')E_\theta(z',z)dz' \right] \quad (20)$$

Note that the molar volume V_θ can be obtained from the density, ρ_θ (g cm^{-3}), and the molar mass, M_θ (g mol^{-1}), of species θ as $V_\theta = M_\theta/\rho_\theta$. Summing Eq. (~~18~~20) for all solid-phase species:

$$\frac{\partial(1-\phi)w}{\partial z} = -\sum_\theta V_\theta R_\theta + \sum_\theta V_\theta \left[-(1-\phi)m_\theta \int_0^{z_{ml}} E_\theta(z,z')dz' + \int_0^{z_{ml}} \{1-\phi(z')\}m_\theta(z')E_\theta(z',z)dz' \right] \quad (21)$$

For the derivation of Eq. (~~19~~21), the following relations are enforced:

$$\sum_\theta V_\theta m_\theta = 1 \quad (22)$$

$$\frac{\partial\phi}{\partial t} = 0 \quad (23)$$

Eqs. (~~20~~22) and (~~21~~23) express the constraint that the volume fractions of all solid species sum up to 1 cm^3 cm^{-3} and the assumption of time independency of porosity, respectively. Unless bio-mixing is Fickian (intraphase) biodiffusion with the

same intensity and the same mixed layer depth for all solid species (see below), the burial velocity is calculated based on Eq. (~~19~~21).

260   If bio-mixing of solid species θ is Fickian ~~with a biodiffusion coefficient D_b~~ biodiffusion with a coefficient $D_{b,\theta}$ (cm^2 yr^{-1}), Eq. (~~18~~20) can be expressed as:

$$\frac{\partial(1-\phi)V_\theta m_\theta}{\partial t} = -\frac{\partial(1-\phi)wV_\theta m_\theta}{\partial z} - V_\theta R_\theta + \frac{\partial}{\partial z}\left\{(1-\phi)D_{b,\theta}\frac{\partial V_\theta m_\theta}{\partial z}\right\} \tag{24}$$

Further if bio-mixing of all solid species ~~is Fickian~~ occurs as Fickian biodiffusion with the same mixing intensity (D_b) and depth (z_{ml}), Eqs. (~~20~~22) and (~~21~~23) lead to a simpler burial velocity equation:

265
$$\frac{\partial(1-\phi)w}{\partial z} = -\sum_\theta V_\theta R_\theta \tag{25}$$

Therefore, when the transition matrix is specified to represent ~~biodiffusion (Section 2.4~~intraphase biodiffusion (e.g., Fig. 1a) and the same matrix is applied to all solid species, Eq. (~~23~~25) is used to calculate burial velocity, otherwise Eq. (~~19~~21) is used. In either case, the model generally satisfies Eq. (~~20~~22).

2.3   Initial and boundary conditions and numerical solutions

270   ### 2.3.1   Initial and boundary conditions

At the beginning of the calculation, we must define both initial (e.g. solid and pore-water composition) and boundary conditions, plus the structure of the grid.

    In the default setting of IMP, the calculation domain represents a $z_{tot} = 500$ cm sediment column and is discretized into $N = 100$ ~~irregular grids where the grid size~~ layers whose thickness increases with depth from less than 10^{-2} to more than 10^2

275   cm following a logarithmic function (Table 2). Furthermore, a time-independent exponential porosity profile is imposed (Table 2). One may modify the grid structure and porosity profile by changing the associated parameter values (Table 2) defined in the code (Supplementary material).

    As initial conditions for the sediment grid, the model assumes near vanishingly small concentrations of 10^{-8} mol cm^{-3} for all solid species (carbonate, organic matter, and clay), and ~~adopts ambient~~ the volume deficiency relative to the solid space

280   prescribed by the assumed porosity is filled by the later time-integration (see below). Ambient ocean concentrations at the seawater-sediment interface are adopted as the initial concentrations for all aqueous species at all depths. These initial values, however, do not have an impact on our results as the model is run to steady state before an experiment is started (e.g., a proxy signal change event is simulated).

    The upper boundary conditions at the seawater-sediment interface are given by mass fluxes of simulated solid species and

285   concentrations for simulated aqueous species (Tables 1 and 2). The lower boundary conditions at z_{tot} for all aqueous species are given by zero concentration-gradients. If oxygen is consumed within the simulated sediment column (i.e., $z_{ox} < z_{tot}$), the dynamically calculated oxygen penetration depth marks a lower boundary for oxygen (i.e., $c_{O_2} = 0$ at $z = z_{ox}$). As boundary

conditions can change with model time (e.g. in the proxy signal change experiments) they are specified ~~at the beginning of~~ before each time integration.

2.3.2 Program structure and numerical solution

Solutions for the temporal and spatial evolution of individual solid and aqueous species are obtained by solving the governing equations with the finite difference method (e.g., Hoffman and Chiang, 2000). Figure ~~1~~2 summarizes the structure of the code to solve the governing equations and the calculation at a given time is conducted by the model in the following four main steps.

1. First, organic matter and oxygen concentration profiles are calculated using Eqs. (1) and (2) (for θ = 'OM' and σ = 'O$_2$'). Since both calculations depend on the oxygen penetration depth z_{ox}, they are conducted iteratively ~~(Emerson, 1985; Archer, 1991).~~ by the following steps (cf. Emerson, 1985; Archer, 1991). (a) z_{ox} is calculated based on the O$_2$ profile from the previous iteration or time instance, (b) the OM profile is updated based on the z_{ox} from step a, (c) $(N+1)$ cases of the O$_2$ profile are calculated, each of which assumes that z_{ox} is located in one of N sediment layers or below the model sediment domain with the corresponding boundary conditions (Section 2.3.1) using the aerobic degradation rates calculated from the OM profile obtained in step b, (d) among the $(N+1)$ cases of step c, the O$_2$ profile that is most consistent with the boundary conditions (i.e., $c_{O_2} = 0$ at $z = z_{ox} < z_{tot}$ or $c_{O_2} > 0$ at $z = z_{tot} < z_{ox}$) is adopted with the corresponding z_{ox}, and (e) steps a–d are repeated until z_{ox} in steps a and d are located in the same sediment layer or both below the model sediment domain. After the convergence of the above iteration, anoxic degradation of OM is calculated at $z > z_{ox}$ if $z_{ox} < z_{tot}$.

2. Second, with the obtained oxic and anoxic decomposition of organic matter, concentration profiles of multiple classes of CaCO$_3$, DIC and ALK are solved (Eqs. (1) and (2) for $\theta = \ell$ and σ = 'DIC' and 'ALK') in a fully coupled way (e.g., Steefel and Lasaga, 1994, see below). Concentrations of individual aqueous carbonate species and pH are calculated based on the obtained ALK and DIC profiles assuming charge balance and equilibria for dissociations of carbonic acid and bicarbonate ion (Tables 2 and 3; Archer, 1991).

3. The clay concentration is calculated using Eq. (22) and the concentrations of OM and n_{cc} classes of CaCO$_3$ obtained in steps 1 and 2, following Munhoven (2021). Obtained clay concentration is substituted into Eq. (1) for θ = 'clay' to confirm the satisfaction of the governing equation.

4. Lastly, the reaction and bioturbation terms for solid species are used to update burial velocity using either Eq. (~~19~~21) or (~~23~~25). When the updated burial velocity is significantly different from the previous velocity, iteration is conducted (i.e., calculations of all species are conducted again with the updated burial velocity) until the relative difference becomes negligible ($\leq 10^{-6}$) within the same time step (Fig. ~~1). This procedure generally ensures~~ 2). If the above criterion is not met within 20 iterations (only encountered in a few conditions in lysocline experiments; Section 3.1), the results yielding the minimum relative difference are adopted (still less than a few %). The procedures in steps 3 and 4 ensure that the

volume fractions of solid species sum up to 1 cm^3 cm^{-3} ~~, i.e., (~~(Eq. (~~20) a deviation from 1 cm^3 cm^{-3} is restricted to a~~
~~few %~~22)).

The concentration profiles of individual species are solved based on the difference equations of Eqs. (1) and (2), which are obtained by the finite difference method. The second-order and first-order spatial differential terms are discretized by the second-order central and the first-order upwind differencing schemes, respectively (e.g., Hoffman and Chiang, 2000). The finite difference form of the bioturbation term in Eq. (1) is formulated with a transition matrix (Eq. (15)). The difference equations for all the solid and aqueous species are solved time-implicitly (e.g., Steefel and Lasaga, 1994). For the solution of the difference equations that are non-linear as is the case for the carbonate system (multiple CaCO$_3$ classes, DIC and ALK), ~~Newton's~~ the Newton-Raphson method is utilized (Fig. ~~1) (e.g., Steefel and Lasaga, 1994).~~ 2) where the solution is iteratively updated together with the Jacobian matrix until its relative difference from the previous iteration becomes $\leq 10^{-6}$ (e.g., Steefel and Lasaga, 1994). Note that the porosity-corrected transition matrix corresponds to the bioturbational transport part of the Jacobian matrix (Fig. 1).

2.4 ~~Transition matrices~~

~~Three different transition matrices were created for the present study to illustrate different styles of bio-mixing: Fickian mixing, homogeneous mixing, and the more mechanistic automaton-based mixing simulated by the particle-tracking bioturbation simulator LABS (e.g., Boudreau et al., 2001; Choi et al., 2002; Kanzaki et al., 2019).~~

The ~~transition matrix that assumes Fickian diffusion for bioturbation (parameterized with D_b, Goldberg and Koide, 1962), can be expressed by :~~

$$K_{\theta,ij} = \begin{cases} -K_{\theta,ij}(j=i+1) & (i=j=1) \\ -K_{\theta,ij}(j=i+1) - K_{\theta,ij}(j=i-1) & (1 < i = j < n_{\text{ml}}) \\ -K_{\theta,ij}(j=i-1) & (i=j=n_{\text{ml}}) \\ \{(1-\phi_i)D_{b,i} + (1-\phi_j)D_{b,j}\}/\{dz_i(1-\phi_i)(dz_i + dz_j)\} & (2 \leq j = i+1 = n_{\text{ml}} \text{ or } 1 \leq j = i-1 = n_{\text{ml}}-1) \\ 0 & (\text{else}) \end{cases}$$

~~where $D_{b,i}$ represents the biodiffusion coefficient at sediment layer i. The transition matrix for homogeneous mixing can be given by:~~

$$K_{\theta,ij} = \begin{cases} dz_i P_h/dz_j & (i \neq j \text{ and } 1 \leq i, j \leq n_{\text{ml}}) \\ -(n_{\text{ml}}-1)P_h & (1 \leq i = j \leq n_{\text{ml}}) \\ 0 & (\text{else}) \end{cases}$$

~~where P_h (yr^{-1}) is the homogeneous transport rate of solid particles between sediment layers.~~

~~To obtain the mechanistic automaton-based transition matrix, we ran a 200-yr LABS simulation and created transition matrices every 10 model days (Reed et al., 2007) based on Eqs.~~ (time step taken for the time-integration of the governing

equations can vary between and within simulations, and can be specified by the user (cf. Section 3.1). In the default setting, time step increases with model time from 100 to 10^5 years to reach steady state (e.g., a spin-up phase of simuation prior to imposing a signal change event; Sections 2.4 and 3), and a smaller and fixed time step is taken when simulating a signal change event (5 or 10 years for a 10 or 50 kyr signal change event, respectively) as well as its aftermath (Sections 2.4 and 3).

By default the model monitors and records depth-integrated fluxes of individual rate terms in Eq. (1) or (2) (fluxes caused by amount change in sediment, sediment rain, biogeochemical reactions, advection, bio-mixing, and so on) for each solid/aqueous species, as well as the residual flux as a sum of all the fluxes, which is ideally zero, to confirm the mass balance of the species. The residual fluxes for all the solid and aqueous species are negligible (e.g., $\leq 10^{-6}$ times the rain fluxes) for all the simulations presented in this paper (Section 3).

2.4 Signal tracking

2.4.1 Tracking input signals

Tracking of proxy signals in carbonates is conducted by assigning different numerical values to the simulated $CaCO_3$ classes, and by scaling their input fluxes to reflect the overall change in proxy signal with time. Thus, proxy signal changes are reflected as changes in the boundary conditions (i.e., rain fluxes of different $CaCO_3$ classes) in the model (see Section 2.3). Assignment of proxy signals and fluxes to $CaCO_3$ classes can be realized by three methods (Fig. 3).

In the first method (a 'time-stepping' method) – any change in proxy signal is approximated by a step-function (i.e. a continuously-varying analogue signal is (digitally) discretized; see e.g. dotted curve in the top panel of Fig. 3). Each step is represented by a separate and unique $CaCO_3$ class, characterized by the approximate proxy value (Fig. 3a). For example, if a signal change event is discretized into 10 steps, 10 different $CaCO_3$ classes with unique proxy values are simulated. Any change in a proxy signal during each discretized time interval is thus muted. Accordingly, the accuracy of the proxy signal approximation is increased by increasing the number of steps and thus the number of simulated $CaCO_3$ classes which, however, results in increased computation costs (Supplementary material). As an advantage, one can track any number of proxies, as long as the signal changes of all tracked proxies occur within a simulated event (Supplementary material).

The second method to assign proxy signals (an interpolating method), simulates only the end-member $CaCO_3$ classes each of which possesses a unique combination of the maximum and/or minimum input-signal values. As an example, one proxy can be tracked with two $CaCO_3$ classes, the first possessing the maximum and the second the minimum proxy value. Intermediate values of an input proxy are realized by assigning varying fluxes to the two end-member classes such that the sum of their flux-weighted values

of the two simulated CaCO$_3$ classes matches the results in the input signal value at each time step . Thus(Fig. 3b). Accordingly, the input proxy signal is not just approximated but accurately represented . Another advantage always accurately represented regardless of the resolution of time discretization. As a disadvantage of method 2over method 1 is that, the number of simulated CaCO$_3$ classes increases with the number of proxies to be tracked. In general, 2^{n_p} classes of CaCO$_3$ particles are necessary when tracking n_p proxies because the number of unique combinations of the maximum and/or minimum signal values is increased by a factor of 2 for every additional proxy to be tracked:

$$\underbrace{\overset{\text{proxy 1}}{2} \times \overset{\text{proxy 2}}{2} \times \cdots \times \overset{\text{proxy } n_p}{2}}_{n_p} = 2^{n_p} \tag{26}$$

Nonetheless, the computational demand is lower as compared to method 1 in most cases because we are interested in a limited number of proxies and thus fewer CaCO$_3$ classes are simulated (i.e., 2^{n_p} in method 2 < time steps in method 1)in most cases.

The third method (a direct tracking method), separates bulk CaCO$_3$ into multiple classes based on how the simulated proxies are determined. For example, when the tracked proxy is δ^{13}C which is determined by the ^{13}C/^{12}C ratio (X in Fig. 23), method 3 simulates classes of Ca^{13}CO$_3$ and Ca^{12}CO$_3$ (Y and G, respectively, in Fig. 23c). The rain fluxes of individual classes at a given time step are directly calculated based on the definition of the proxy and the contemporaneous proxy value (see boxes in Fig. 23c). Thus, one can regard method 3 as a derivative of method 2 that defines the end-member CaCO$_3$ classes based on the definition of the tracked proxy. Because the flux calculation must change with the simulated proxy signal, method 3 is not as flexible as methods 1 and 2, but the computational effort can be further reduced because a certain class can be used to define multiple proxies (e.g., tracking 4 proxies with 5 CaCOCa^{12}CO$_3$ classes, Supplementary materialis related to the definition of both ^{14}C age and δ^{13}C). Method 3 has a unique advantage of enabling additional biogeochemical reaction terms for any specific CaCO$_3$ class if necessary. For instance, when tracking ^{14}C age, one needs to account for the radioactive decay of Ca^{14}CO$_3$ and accompanied generation of alkalinity, which can be implemented with method 3. Currently method 3 tracks 4 proxies including ^{14}C age with 5 CaCO$_3$ classes (Supplementary material).

After the signal and flux assignment by any of the three methods, the model is spun up to steady state with only the CaCO$_3$ class(es) with pre-event proxy values being deposited to sediment (Fig. 23). After the spin-up, a proxy-signal change event is simulated by changing the rain fluxes of different CaCO$_3$ classes with different proxy values (i.e., the boundary conditions) with model time (Fig. 23). After the signal change event, the model is run until a new steady state is reached.

Note that the methods and procedures described above can be applied not only to track proxy signals but also any other property of CaCO$_3$ particles such as particle size and deposition time (cf. Section 2.4.2). In this case, methods 2 and 3 track the property in the same way (Section 2.4.2, Supplementary material).

2.4.2 Tracking signals within the sediment

After input signals are reflected in rain fluxes by any of the three methods in Section 2.52.4.1, they are modified within the sediment by bioturbation and chemical erosion. Caution needs to be taken with respect to numerical diffusion, which is inevitably introduced to the difference form of the advection term (1st term on the right-hand side of Eq. (1)) in a finite

difference approach (e.g., Hoffman and Chiang, 2000; Steiner et al., 2016). For an accumulating column of sediment in a fixed grid, numerical diffusion artificially mixes the deposited and buried sediment particles along with their proxy signals, especially at depths where grid cells are relatively coarse (Fig. 34). An alternative is to allow for a partial surface layer and to accrete or remove complete layers depending on the growth or erosion at the surface, such as in Ridgwell (2007b). However, such an approach is impractical if the depth-dependent diagenetic reactions are to be solved rather than just recording historical accumulation (or erosion).

Here, to minimize the effect of numerical diffusion, we read out the proxy signal as a function of time, from just below the mixed layer and before the start of the 'historical' layer (z_{ml}, see arrow in Fig. 34 and Fig. 45). Accordingly, signal values are not plotted against the depth of the sediment domain, but against a sediment stack composed of the sediment layers that were used to record the proxy signal (i.e., at depth z_{ml}) during the course of the simulation. The depth of this sediment stack is called diagnosed depth (z_{diag}, Fig. 45) and can be calculated as:

$$z_{diag} = z_{ml} + \int_t^{t_{tot}} (1 - \phi_{ml})w_{ml}dt \tag{27}$$

where ϕ_{ml} and w_{ml} (cm yr^{-1}) denote the porosity and burial velocity at the mixed layer depth ($z = z_{ml}$) and t_{tot} is the total duration of a simulation (yr). While reading proxy signals at the bottom of the mixed layer is likely effective in most cases (cf. Supplementary material), it is also possible to specify a different depth point to read proxy signals. In such case, the definition of diagnosed depth need be modified by replacing z_{ml}, ϕ_{ml} and w_{ml} in Eq. (27) with the corresponding parameter values at the specified depth.

To convert the signal profiles plotted against diagnosed depth to profiles plotted against model time, an age model is required that can be obtained by tracking model time as a proxy. The application of the three methods explained in Section 2.52.4.1 (i.e. to assign numerical values to multiple classes of CaCO$_3$ particles and calculate their rain fluxes from the input values) is not limited to tracking proxy signals but can also be applied to any other characteristic including the model time at which particles are deposited. In method 1, individual classes of CaCO$_3$ particles are defined based on the time steps discretized from a signal change event (Fig. 23a) and thus already have their own model time to be assigned with. Note, however, that tracking model time with method 1 is computationally more expensive because a larger number of explicit CaCO$_3$ classes is needed to represent the continuously changing model time. When using method 2 or 3 to track model time in addition to paleoceanographic proxies, the number of CaCO$_3$ classes must be doubled (cf. Eq. (26)). For example, when using method 2 one proxy signal can be simulated with two (or a pair of) CaCO$_3$ classes representing the maximum and minimum proxy value. Additionally tracking model-time requires an extra pair of CaCO$_3$ classes, whereas the start and end of model time is assigned to the two pairs, respectively (cf. Eq. (26)). In either method, model time tracked in bulk CaCO$_3$ can be plotted against diagnosed depth, which is the age model of IMP, and can be used to plot the other tracked proxy signals against model time. Examples to obtain and use IMP's age model are provided in Supplementary material.

3 Results and discussion

3.1 Diagenesis

In this section, we highlight diagenetic aspects of the model including comparison with the $CaCO_3$ diagenesis model by Archer (1991).

First, the capability of the model to obtain steady-state and time-dependent sediment profiles of solid and aqueous species is illustrated by showing a spin-up phase and a transient phase between two steady states, respectively, of a simulation. Then, we compare lysoclines estimated by IMP and the diagenesis model of Archer (1991). The lysocline is the ocean depth below which $CaCO_3$ dissolution significantly increases and the depth of the lysocline is an important indicator for determining Earth's carbon cycle response to environmental changes (e.g., sea level change) and associated feedbacks on climate ~~(e.g., Archer and Maier-Reimer, 1994; Ridgwell et al., 2003; Ridgwell and Zeebe, 2005; Greene et al., 2019)~~(e.g., Archer and Maier-Reim . $CaCO_3$ dissolution below the lysocline is caused because the thermodynamic stability of $CaCO_3$ decreases due to increased pressure, but the lysocline is also known to be significantly affected by local rain fluxes of OM and $CaCO_3$, and early diagenesis within sediments (e.g., Archer, 1991). Therefore, simulating the depth of the lysocline is a good test of a $CaCO_3$ diagenesis model. The details of the experiments and results are described in the following subsections.

~~Methodology~~Experimental setup

To illustrate the initial evolution of the model, a spinup experiment was run until a steady-state sediment composition is achieved. For this we assumed Fickian mixing using the default conditions given in Table 1 (Fig. ~~5~~1a). Model output includes depth profiles of density and volume fraction of solid sediment (Figs. ~~5~~6a and c), burial velocity (Fig. ~~5~~6b), concentrations of solid and aqueous species (Figs. ~~5~~6d–k) and rates of biogeochemical reactions (Figs. ~~5~~6l–n) for 5 time ~~steps of~~ instances during the spinup experiment (1, 10, 100 kyr, 1 and ~~10~~3 Myr).

A second experiment illustrates how a change in the boundary conditions affects the temporal evolution of the depth profiles in IMP. This experiment starts from the end of the first spinup experiment and artificially imposes significant carbonate dissolution by changing the water depth from 3.5 to 5.0 km between 10 and 40 kyr (Fig. ~~6~~7). Because of the longer timescale to achieve steady state (see the first experiment), the second experiment run for 50 kyr is in transient states except for the initial steady state at 0 kyr (Fig. ~~6~~7).

Finally, IMP was run to steady state assuming various carbonate rain fluxes (ranging from 6 to 60 μmol cm^{-2} yr^{-1}, in increments of 6 μmol cm^{-2} yr^{-1}), ratios of organic matter to carbonate (0, 0.5, 0.67, 1 and 1.5) and water depths (ranging from 0.24 to 6.00 km, in increments of 0.24 km) (cf. Archer, 1991). These lysocline experiments were performed for both the oxic-only OM degradation model and the oxic-anoxic model (Figs. ~~7 and 8~~8 and 9). To facilitate comparison of our results with Archer (1991) IMP assumes a single class of $CaCO_3$ particles, Fickian mixing for bioturbation and a sediment column depth of 50 cm. All other boundary conditions are as described in Table 1.

One can use the IMP code of any of the three programming languages (i.e., Fortran90, MATLAB or Python) to conduct the simulations presented in this paper. The model code for each language is stored in the respective directory (i.e., 'Fortran', 'MATLAB' and 'Python') and a language-specific readme file provides instructions for how to run the simulations (e.g., \iMP\Fortran\readme_Fortran.txt for the Fortran version). The boundary conditions can be specified with time-invariant values at run time (e.g., the third experiment above; see the readme file for the chosen version of the code), but can also be changed as a function of time (as in the second experiment above). The temporal changes of the boundary conditions must be prescribed in the input files that are stored in a directory 'input' and can be modified by the user (see a readme file therein, \iMP\input\readme_input.txt, for the details). We also provide Python scripts to plot concentrations of solid and aqueous species (e.g., Figs. ~~5~~6–~~8~~9) as well as tracked proxy signals (Section 3.2), stored in a directory 'plot' (see a readme file therein, \iMP\plot\readme_plot.txt, for more details).

Results

In the spin-up to steady-state, spaces for solid sediment defined by assumed porosity $(1 - \phi)$ are initially empty (not filled) because of the low initial concentrations of solid species ($\sum_\theta V_\theta m_\theta \cong 0$; Section 2.3) but get ~~filled with solid species (CaCO$_3$, organic matter and /or clay)~~ soon filled with clay (as a 'dilatant') and OM, and later with CaCO$_3$ as Eq. (~~20~~22) is enforced and steady state is approached ($\sum_\theta V_\theta m_\theta = 1$; Figs. ~~5~~6a and c). In contrast, pore spaces are assumed to be always filled with pore-water and pore-water chemistry achieves ~~the~~ steady state much faster (Figs. ~~5~~6g–k) (e.g., Archer et al., 2002). ~~Changing~~ The steady-state results for bulk phases (Fig. 6) are not affected by changing the number of CaCO$_3$ classes or the time step ~~per one model integration does not change the steady-state results.~~ of each time-integration (cf. Section 2.3.2).

The second experiment demonstrates that once a steady state is achieved, a change in boundary conditions does not generate significant void spaces ($\sum_\theta V_\theta m_\theta \ll 1$) and/or expansions ($\sum_\theta V_\theta m_\theta \gg 1$) in solid sediment (Fig. ~~6~~7c), thus generally satisfying Eq. (~~20~~22). In other words, prescribed spaces for solid sediment by assumed porosity are almost perfectly matched with sums of volumes of all solid-phase species ($\sum_\theta V_\theta m_\theta = 1$; Fig. ~~6~~7c) even when the concentrations of solid species dynamically change with time, leaving a steady state (e.g., Fig. ~~6~~7d). Absence of significant void spaces or expansions in solid sediment ~~has been adopted as~~ provides a convergence diagnostic ~~by the sediment~~ (adapated from one of the convergence diagnostics in the steady-state diagenesis model of ~~Archer et al. (2002), although they considered only steady states. The results of the second experiment thus confirm that the model can also be used for transient calculations.~~ Archer et al. (2002)).

Finally, we compare steady state lysoclines simulated with IMP to results from the CaCO$_3$ diagenesis model of Archer (1991), who showed that the lysocline is sensitive to rain rates of carbonate and organic matter to the seafloor, and in particular to the ratio of these fluxes. The simulated lysocline and carbonate burial rates for the oxic-only OM degradation model are presented in Figs. ~~7~~8a and b. The results for the oxic-anoxic model are shown in Figs. ~~8~~9a and b.

In general, our predicted mixed layer CaCO$_3$ wt% and the CaCO$_3$ burial fluxes match the steady-state estimates by Archer (1991) (compare with his Figs. 5 and 6). For instance, as in Archer (1991) increasing the carbon rain to the sediments for lower OM/CaCO$_3$ rain ratios (i.e., ≤ 0.67) enhances carbonate preservation and causes the lysocline to deepen for both the oxic-only and the oxic-anoxic OM degradation model (Figs. ~~7 and 8~~8 and 9). The only notable difference occurs for the oxic-only OM

505 degradation model under the most extreme carbon rain fluxes (i.e., rain ratio = 1.5; $CaCO_3$ rain > 40 µmol cm^{-2} yr^{-1}). Here, IMP simulates higher $CaCO_3$ preservation than ~~Archer (1991) model~~ the model of Archer (1991) (Fig. ~~7~~8, right panels). This difference can be explained by a burial velocity enhancement caused by high organic matter preservation in the oxic-only model, which is not considered by Archer (1991) (see the lysocline experiment with $V_{OM} = 0$ in the Supplementary material). For the same high OM/$CaCO_3$ rain ratio (1.5) the oxic-anoxic OM degradation model simulates an enhancement in carbonate

510 accumulation rate and a deepening of the lysocline for an increase ~~in~~ of the $CaCO_3$ rain, which is in line with the results of Archer (1991).

3.2 Signal tracking diagenesis

In the following subsections, we illustrate the utility of the model for exploring the combined effects of bioturbation and chemical erosion on the preservation of proxy signals in carbonates. The experiments presented here adopt method 2 for the

515 signal and flux assignment (Fig. ~~2~~3) as it is a more accurate and computationally less expensive approach than method 1 and is more flexible than method 3 (Section ~~2.5~~2.4.1). Equivalent results using methods 1 and 3 are described in the Supplementary material to demonstrate that all methods lead to the same results.

All experiments simulate two paleoceanographic proxies simultaneously, $\delta^{13}C$ and $\delta^{18}O$, and both proxy signals change over the course of the experiments in an idealized fashion. All experiments adopt the oxic-anoxic OM degradation model and,

520 if not stated otherwise, the default conditions in Table 1. Signal values are plotted against diagnosed depth (see Fig. ~~4~~5 and Eq. (~~26~~27)). The same series of experiments in Section 3.2 but with tracking model time in addition to $\delta^{13}C$ and $\delta^{18}O$ are presented in Supplementary material where we illustrate that proxy signal values can be plotted against model-time using the model specific age model (Section ~~2.5~~2.4.2).

3.2.1 Bioturbation

525 *~~Methodology~~Experimental setup*

The effects of three different styles of bioturbation on the recorded proxy signals are considered: (i) Fickian local-mixing with a biodiffusion coefficient of ~~$D_b = 0.15$~~$D_{b,\theta} = 0.15$ cm^2 yr^{-1}, (ii) homogeneous non-local mixing to represent random mixing as simulated by, e.g., TURBO2 (Trauth, 2013), and (iii) process-based non-local mixing simulated by deposit-feeder automata from the LABS model (e.g., Boudreau et al., 2001; Choi et al., 2002; Kanzaki et al., 2019). ~~In some cases, the IMP calculation~~

530 ~~did not converge when adopting the transition matrix created by LABS. This is because the~~ Because the LABS-derived transition matrix contains less continuous and more irregular transport ~~provability~~ probability than the other two styles of bio-mixing ~~.~~ ~~Implementing non-local transport is numerically similar to adding a reaction term (e.g., Van Cappellen and Wang, 1996; Boudreau, 1997, S~~ ~~, and when a reactive-transport model simulates reaction or non-local transport with large irregular, non-continuous values,~~ (Fig. 1), it is susceptible to convergence problems ~~(cf., Boudreau, 1997). To facilitate the calculation, we arbitrarily weakened~~

535 ~~bio-mixing by multiplying the transition matrix from LABS by a factor of 1/10. When this modification did not lead to~~

convergence of the model(cf., Boudreau, 1997, Section 2.2.2). When the convergence was not achieved, model results with the bio-mixing from LABS are ~~now~~ not shown in the following subsections (Sections 3.2.1–3.2.3).

The input proxy values of $\delta^{13}C$ and $\delta^{18}O$ in $CaCO_3$ either experience a step-change over 5 kyr or a 5-kyr duration impulse event, respectively (Fig. ~~9~~10a). Four end-member classes of $CaCO_3$ particles are used for signal tracking (Fig. ~~9~~10c) and simulated proxy signals are recorded just below the sediment mixed layer and plotted against diagnosed depth to minimize the effect of numerical diffusion (Section ~~2.5~~2.4.2). A first set of experiments is conducted with dissolution disabled for all $CaCO_3$ classes ($k_{cc,\ell} = 0$) in order to consider solely the effect of different styles of bioturbation. In a second set of experiments, the default $CaCO_3$ dissolution rate constant is used for all classes.

Results

To visualize signal distortions by comparison, the input signals as a function of time (Fig. ~~9~~10a) are plotted against diagnosed depth in Fig. ~~10~~11, using the age model for the no bioturbation case (Supplementary material). Slight deviations of the recorded signals (pink curves in Figs. ~~10~~11a and b) from the input signals (dotted black lines) in the 'no bioturbation' case can be attributed to numerical diffusion, but are minor compared to signal distortions exhibited by bioturbated sediments (blue, ~~green and orange~~ yellow and green curves). More specifically, dispersion of the recorded signals occurs over a larger depth interval and, for the impulse-event in $\delta^{18}O$, the signal magnitude is significantly reduced with bioturbation (Figs. ~~10~~11a and b). Fickian and homogeneous mixing distorts the input signals similarly (blue and ~~green~~ yellow curves, respectively, which are almost completely superimposed in Figs. ~~10~~11a and b), but LABS mixing results in slightly different signal shifts that extend to ~~shallower depths (orange~~ deeper depths (green curves). This difference may be explained by ~~a net upward transport of sediment particles at depths close to~~ defecating/pushing of particles by deposit-feeder automata resulting in rare occasions where particle displacements propagate to depths even below the mixed layer ~~bottom in the LABS mixing (caused by, e. g., deposit feeder feeding at depths and defecating at shallower depths).~~ (Fig. 1c; e.g., Choi et al., 2002). Note that bio-mixing in LABS can vary with assumed physicochemical and ecological conditions and animal types (e.g., Boudreau et al., 2001; Kanzaki et al., 2019), and thus our results should not be regarded as the exclusive results with a LABS transition matrix (cf. Section ~~2.3~~2.2.2).

Results for the second set of experiments with $CaCO_3$ dissolution enabled are presented in Figs. ~~10~~11d–f. Different modes of bioturbation result in variations in the extent of $CaCO_3$ dissolution (Fig. ~~10~~11f), with no bioturbation leading to the lowest degree of dissolution and efficient homogeneous mixing causing the highest degree of dissolution (Fig. ~~10~~11f). Correspondingly, sediment accumulation rates and thus age models differ between different styles of bioturbation (Supplementary material) and one observes signal change events at shallower depths with a more enhanced dissolution (Figs. ~~10~~11d and e). By enabling dissolution, proxy signals are slightly lost along with $CaCO_3$ particles especially when bio-mixing is not efficient. This can be recognized by a reduction of the magnitude of $\delta^{18}O$ impulse for no bioturbation case by enabling dissolution (slightly smaller peak of pink curve in Fig. ~~10~~11e than in Fig. ~~10~~11b). We examine the dissolution effect in more detail in the next subsection.

3.2.2 Dissolution of carbonates

~~Methodology~~Experimental setup

While evidence for ~~chemical erosion~~ significant dissolution of sedimentary carbonates provides information about ocean chem-
istry (e.g., Oxburgh and Broecker, 1993; Zachos et al., 2005; Panchuk et al., 2008), it also distorts proxy signals recorded in
these carbonates. In this subsection, we examine how and to what extent dissolution distorts proxy signals.

We consider a negative $\delta^{13}C$ excursion over 40 kyr with a relatively rapid onset and recovery of the isotope signal (over 5
kyr). At the same time, a more gradual ramp down and up change of the $\delta^{18}O$ signal over 50 kyrs is simulated (Fig. ~~11~~12a).
The signal shifts for the two proxies are intentionally made decoupled in time and should not be associated with any 'real'
geological event. These signal changes are accompanied by water depth changes from the background depth of 3.5 km to 4.5
and 5.0 km over 5 kyr in order to cause different extents of dissolution (Fig. ~~11~~12c) through destabilizing $CaCO_3$ by increasing
pressure (Millero, 1995). These imposed changes in water depths are not intended to be 'realistic', but rather drive conditions
of enhanced $CaCO_3$ dissolution as might have been caused by environmental changes such as ocean acidification (e.g. see:
Ridgwell (2007b)), but without the additional interpretative complications of actually changing the ocean chemistry at the
sediment surface in the model. (Note that it is ~~perfectly~~ also possible to drive IMP with changing upper geochemical boundary
conditions to explicitly simulate e.g. ocean acidification.) The water depth and related dissolution changes are assumed to be
synchronous with the proxy signal changes (Figs. ~~11~~12a and c).

Signal tracking is conducted by simulating the same four classes of $CaCO_3$ as in the previous subsection (Fig. ~~11~~12d; cf.
Fig. ~~9c) .~~10c) with enabling Fickian or homogeneous bio-mixing (Figs. 1a and b), or without bioturbation. An additional set
of experiments was run without changing the water depth as a 'no dissolution' control (dotted line in Fig. ~~11~~12c). Simulated
signals against sediment depth (Fig. ~~12~~13) are compared with input signals (dotted black curves in Fig. ~~12~~13) which are
obtained from their temporal changes (Fig. ~~11~~12a) and the age model for the no bioturbation case (cf. Supplementary material)
as in the previous subsection.

Results

When dissolution is ~~imposed~~ intensified by changing the water depth from 3.5 to 4.5 km (experiment #1, solid line in Fig.
~~11~~12c), the total amount of $CaCO_3$ is reduced from ∼90 to ∼50 wt% for all cases with and without bioturbation (Fig. ~~12~~13f).
As described in Section 3.2.1, dissolution is enhanced by bio-mixing and correspondingly signal change events are observed
at different depths between different modes of bioturbation (Figs. ~~12~~13d–f; cf. Supplementary material). Apparent durations
of the signal change events become shorter compared to the control experiment (Figs. ~~12~~13a–c) because less sediment accu-
mulates during the events with a more enhanced dissolution (Figs. ~~12~~13c and f). However, because imposed dissolution is still
moderate (Fig. ~~12~~13f) and relatively long-term signal change events are considered (e.g., compare Fig. ~~11~~12a with Fig. ~~9~~10a),
no significant reduction of the magnitude of signal peaks is observed in experiment #1.

Further increasing dissolution rate by changing the water depth to 5.0 km during the isotope excursion (experiment #2,
dashed line in Fig. ~~11~~12c) causes $CaCO_3$ to completely disappear for all cases with and without bioturbation (Fig. ~~12~~13i). Note

that a concentration of absolute zero is not allowed for solid species in the model. Simulated concentrations are truncated at a threshold of 10^{-300} mol cm^{-3}. As for dissolution experiment #1 (Fig. ~~12~~13f), different styles of bioturbation cause different CaCO$_3$ dissolution rates (Fig. ~~12~~13i). Under this more intense dissolution scenario, simulated proxy signals are considerably distorted and reduced for all styles of bioturbation (Figs. ~~12~~13g and h). Simulated _excursions of_ proxy signals are ~~considerably shorter in apparent duration~~ _observed for considerably shorter apparent duration or sediment depth iterval_ as described in the above paragraph.

It is noted that the carbon and alkalinity fluxes from dissolved CaCO$_3$ in sediments under any destabilization can vary with the mode of bioturbation (Figs. 11 and 13). This incidates the potential role of benthic ecosystems to determine the feedback of sedimentary CaCO$_3$ to a climate perturbation (e.g., Ridgwell, 2007b; Jennions et al., 2015).

3.2.3 Species-specific mixing/dissolution

~~Methodology~~Experimental setup

It has been suggested that carbonates of different sizes can be differently bioturbated and dissolved in marine sediments (e.g., Broecker et al., 1991; Bard, 2001; Barker et al., 2007). IMP is well-suited for examining the effect of differential mixing and/or dissolution rate among CaCO$_3$ size classes on the signal distortion.

Here we consider eight CaCO$_3$ classes, consisting of two sets of the same four CaCO$_3$ classes as in the previous subsections _(Table 4)_. We assign two distinctive sizes to these two sets (Figs. ~~13~~14c and d). CaCO$_3$ particles in the first set are assumed to be of 'fine' grain size, and are consequently bioturbated _(by Fickian and in a second experiment by homogeneous mixing)_ to deeper depths (20 cm; cf., Bard, 2001) ~~.~~_with the correspondingly modified transition matrices (Eqs. (18) and (19); Section 2.2.2)._ They are also dissolved at a faster rate by adopting a dissolution rate constant increased by a factor of 10 (cf., Keir, 1980) (class #1–4 in Fig. ~~13~~14 and Table 4). CaCO$_3$ particles in the second set are of 'coarse' grain size and adopt the default particle characteristics (Table 1, class #5–8 in Fig. ~~13) .~~_14 and Table 4) and transition matrices (Figs. 1a and b)._ The total mass flux and isotope signal input are the same as in Section 3.2.2 and the water depth remains unaltered at 3.5 km. In concert with the δ^{18}O decrease, the coarse species becomes more dominant over the fine species ~~(Fig. 13; cf., Schmidt et al., 2004)~~ _(the rain fraction of coarse species increases from 50 to 90 %, Fig. 14c; cf., Schmidt et al., 2004)._

Results

The differences in dissolution and mixing properties of fine and coarse CaCO$_3$ species have a prominent effect on their relative preservation (Fig. ~~14~~15c). In general, the coarse species shows higher preservation due to its lower dissolution rate. The more efficient the adopted mixing mode ~~,~~_(e.g., homogeneous mixing), the better_ the ~~more the~~ coarse species is preserved and the more obscured is the preservation of the imposed CaCO$_3$ input flux changes. Correspondingly accumulation ~~rate differs between~~ _rates are different for_ fine and coarse CaCO$_3$ species and thus excursions of proxy signals as well as peaks in coarse vs. fine species abundance are offset between the two species by $\sim$10 cm (compare solid and dotted curves in Fig. ~~14~~15). Observed apparent offsets of peaks in proxy signals and species abundance can be mostly removed by applying individual age

models to the two species, although the reduction of the magnitude in abundance shifts cannot be recovered (Supplementary material).

Although the above experiment is not designed to simulate any specific surface-environment change event in the past, signal offsets among $CaCO_3$ species have been observed in e.g. hyperthermal events (e.g., Kirtland Turner et al., 2017). Application of IMP to such events can be useful, as it might lead to an insight into population shifts among calcifiers associated with enviromental changes in the past (cf. Figs. 14 and 15c).

3.3 Proxy signals in an extended enviromental parameter space

The complexity of IMP furthermore allows for hypothesis testing that has not been possible with traditional diagenetic models. For instance, changes in the rain the fraction of fine vs. coarse species in the signal tracking experiment in Section 3.2.3 affected proxy signals of both species differently. However, in traditional 1D diagenetic models such an environmental variable is not explicitly considered. This section reiterates the utility of IMP to interpret proxy signals in a parameter space that is not accessible when considering only bulk $CaCO_3$. Here we focus on ^{14}C age as another example proxy.

In equatorial Pacific sediments carbonate ^{14}C ages have been observed to get older with decreasing $CaCO_3$ wt% (circles in Fig. 16), a counter-intuitive trend if $CaCO_3$ is dissolved homogeneously, as dissolution should shift the distribution towards younger $CaCO_3$ particles. Although Broecker et al. (1991) demonstrated with an idealized sediment box model that interface dissolution ($CaCO_3$ dissolution completed before bio-mixing and burial) can reproduce the observation, the mechanism does not allow $CaCO_3$ dissolution to continue within the sediment column and thus cannot be implemented by 1D reactive-transport models, which usually assume homogeneous dissolution (cf. Keir, 1984; Broecker et al., 1991; Keir and Michel, 1993). However, it is not known whether homogeneous dissolution can lead to a different ^{14}C age vs. $CaCO_3$ wt% relationship in a more complicated and thus realistic parameter space, especially where distinct $CaCO_3$ size classes are explicitly accounted for. Here we simulate steady-state ^{14}C age in the mixed layer for the coarse and fine species considered in Section 3.2.3.

Experimental setup

To track radiocarbon age, the direct tracking method (method 3 in Section 2.4.1) is utilized. The method simulates 5 $CaCO_3$ classes corresponding to 5 isotopologues ($Ca^{12}C^{16}O_3$, $Ca^{12}C^{18}O^{16}O_2$, $Ca^{13}C^{16}O_3$, $Ca^{13}C^{18}O^{16}O_2$ and $Ca^{14}CO_3$) to track 4 associated isotopic signals ($\delta^{13}C$, $\delta^{18}O$, Δ_{47}, and ^{14}C age) recorded in $CaCO_3$ particles that are of the same size (see Supplementary material for the details). Because we further track 'size' of $CaCO_3$ particles by simulating 2 distinct $CaCO_3$ species of 'fine' and 'coarse' sizes, two sets of the above 5 classes, i.e., 10 classes in total, are necessary (cf. Eq. (26); Table 5). The first set of 5 classes (class #1–5) possesses the dissolution and bio-mixing properties for the fine species defined in Section 3.2.3, while the second set (class #6–10) represents the coarse species (Table 5).

Steady-state simulations were run with the above 10 $CaCO_3$ classes adopting Fickian mixing, for three water depths (3.7, 3.9 and 4.1 km), three total sediment fluxes (12 (default), 6 and 3 μmol total $CaCO_3$ cm^{-2} yr^{-1} with the fixed default OM/$CaCO_3$ and clay/$CaCO_3$ rain ratios; Tables 1 and 2) and for different rain fractions of the fine $CaCO_3$ species (10, 50, 90 and 99 %).

The rain fluxes of individual $CaCO_3$ classes are calculated from the total $CaCO_3$ rain, the rain fraction for fine species, and assuming that $\delta^{13}C = \delta^{18}O = \Delta_{47} = 0\,‰$ and the $^{14}C/^{12}C$ ratio is 1.2×10^{-12} (Aloisi et al., 2004) (cf. Supplementary material). The mixed layer depth (and thus the transition matrix) and the dissolution rate constant are defined differently between the fine (class #1–5) and coarse (class #6–10) species (compare Section 3.2.3). All the other parameters were set at the default values (Table 1).

Results

Because dissolution and transport is fully coupled in IMP (i.e., dissolution is 'homogeneous'), a decrease in $CaCO_3$ wt% caused by increasing water depth generally leads to a younger radiocarbon age (e.g., see Fig. 16c where ^{14}C ages are highest with blue curves and lowest with green curves). However, when the decrease of $CaCO_3$ concentration is caused not by increasing the water depth but by increasing the rain fraction of the fine species that dissolves faster (trajectories depicted with curves in Fig. 16), the trend of ^{14}C age vs. $CaCO_3$ wt% differs. The trend for the coarse species is especially counter-intuitive, where an older ^{14}C age is observed for lower $CaCO_3$ wt% (Fig. 16a). The opposite trend is recognized for the fine species (Fig. 16b). Bulk $CaCO_3$ shows a combination of the above two contrasting aging trends, and whether bulk ^{14}C age increases or decreases with bulk $CaCO_3$ wt% depends on the contribution of fine vs. coarse species (Fig. 16c). The magnitude of the aging effect (whether by changes in the rain fraction of the fine species or the water depth) can be amplified when the total sediment rain is decreased because both $CaCO_3$ species are buried at a slower rate (dashed and dotted curves in Fig. 16).

Note that it is not our intention to perfectly reproduce the observations with the parameterization adopted in this experiment, given that a large number of parameters would need to be constrained and/or modified (e.g. Keir, 1980; Walter and Morse, 1984, 1985; Bard . Nonetheless, the ^{14}C age sensitivity to the rain fraction of fine species shown above illustrates the utility of the model to interpret proxy singals in an extended and more realistic enviornmental parameter space.

4   Conclusions and summary

Our new Implicit model of Multiple Particles (diagenesis) – IMP – is capable of tracking proxy signals by implicitly simulating reactive transport of multiple solid carbonate particles, along with calculations of organic matter, refractory detrital materials, and aqueous oxygen and dissolved CO_2 species. The model also realizes simulations of different kinds of bioturbation by adopting different transition matrices. As shown with illustrative experiments, signal distortion can vary with the style of bioturbation, intensity of chemical erosion and distributions of $CaCO_3$ species with different dissolution/mixing characteristics. Such complexity needs to be carefully evaluated when reading proxies in marine sedimentary carbonates for reconstruction of past environmental changes.

Future developments of the model include coupling with Earth system models, which will provide synthetic sedimentary records that are process-based and can be directly compared with geological records. Coupling the model with an efficient Earth system model such as 'cGENIE' (Ridgwell and Hargreaves, 2007; Ridgwell, 2007b) is particularly promising as it may

695 allow iterative runs to predict environment changes that minimizes the difference between synthetic and observed sedimentary records (e.g., Kirtland Turner and Ridgwell, 2013).

Code availability. The IMP source codes are available on GitHub (https://github.com/imuds/iMP) under the MIT License. The specific version used in this paper is tagged as 'v1.0' and has been assigned a DOI (https://doi.org/10.5281/zenodo.5213875). A readme file on the web provides the instructions for executing the simulations.

700 **Appendix A: Supplementary material**

Supplementary material related to this article can be found online at https://doi.org/xxxxxx.

Author contributions. YK designed and implemented the model in Fortran90 with contributions from the other authors. DH and YK converted the Fortran90 version to MATLAB and Python versions, respectively. YK designed the simulations with contributions from the other authors. All authors contributed to the writing of the paper.

705 *Competing interests.* The authors declare no competing interests.

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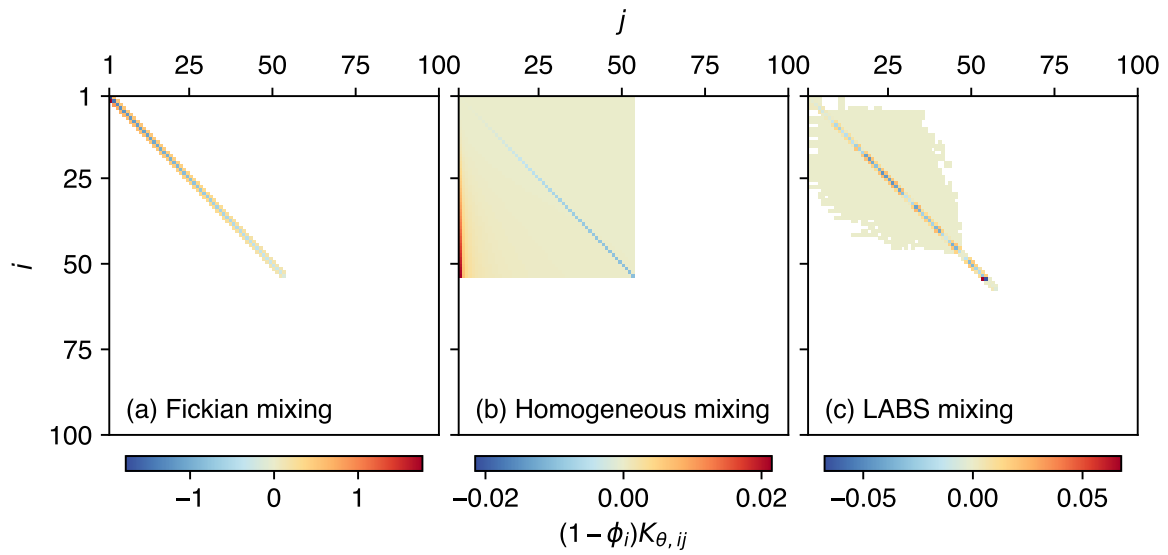
Figure 1. Transition matrices corrected for porosity, representing three different bio-mixing styles: (a) Fickian intraphase biodiffusion, (b) homogeneous mixing, and (c) automaton-based mixing by a particle-tracking bioturbation simulator LABS.
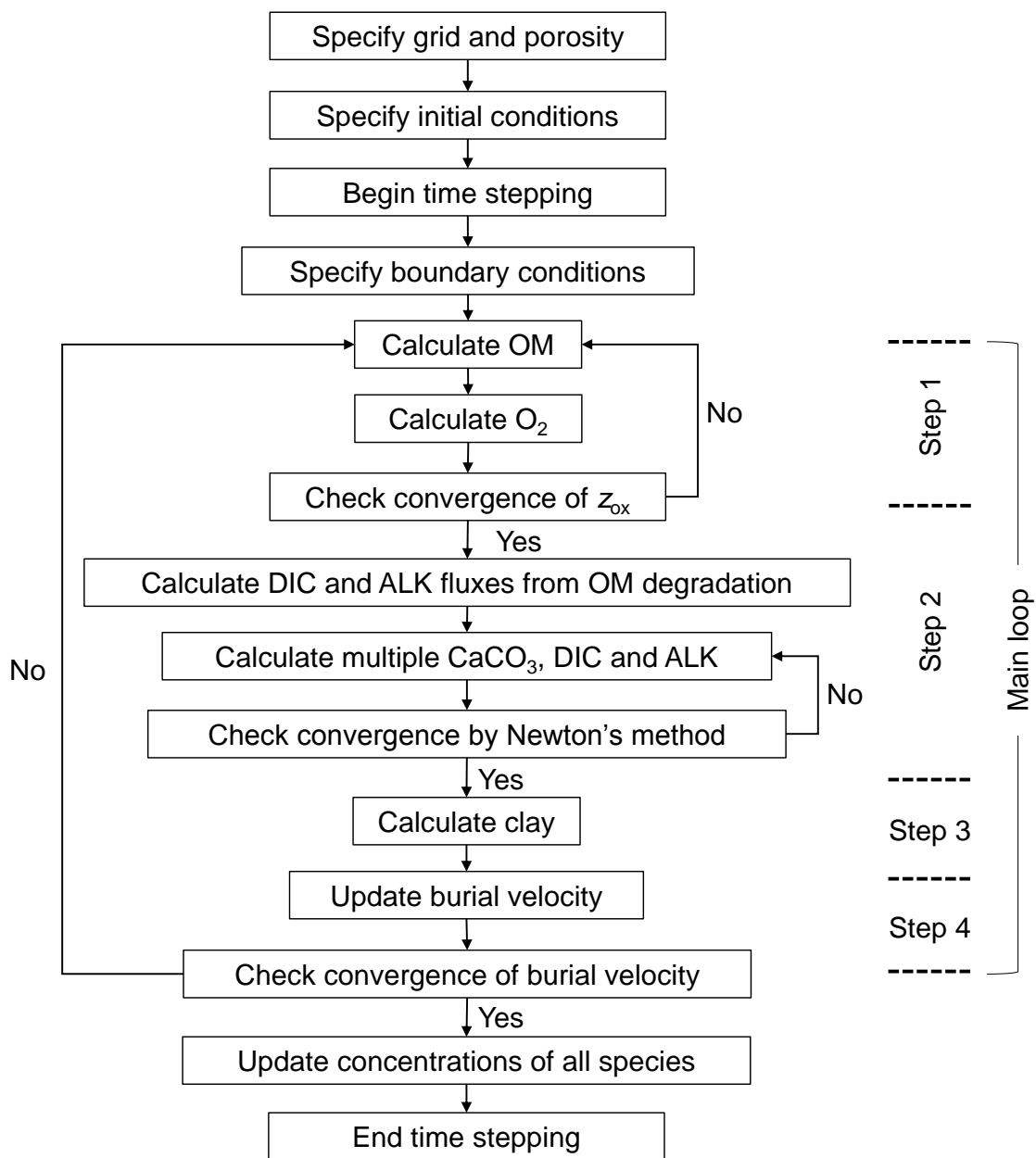
Figure 2. Program structure for reactive transport modeling of diagenesis.
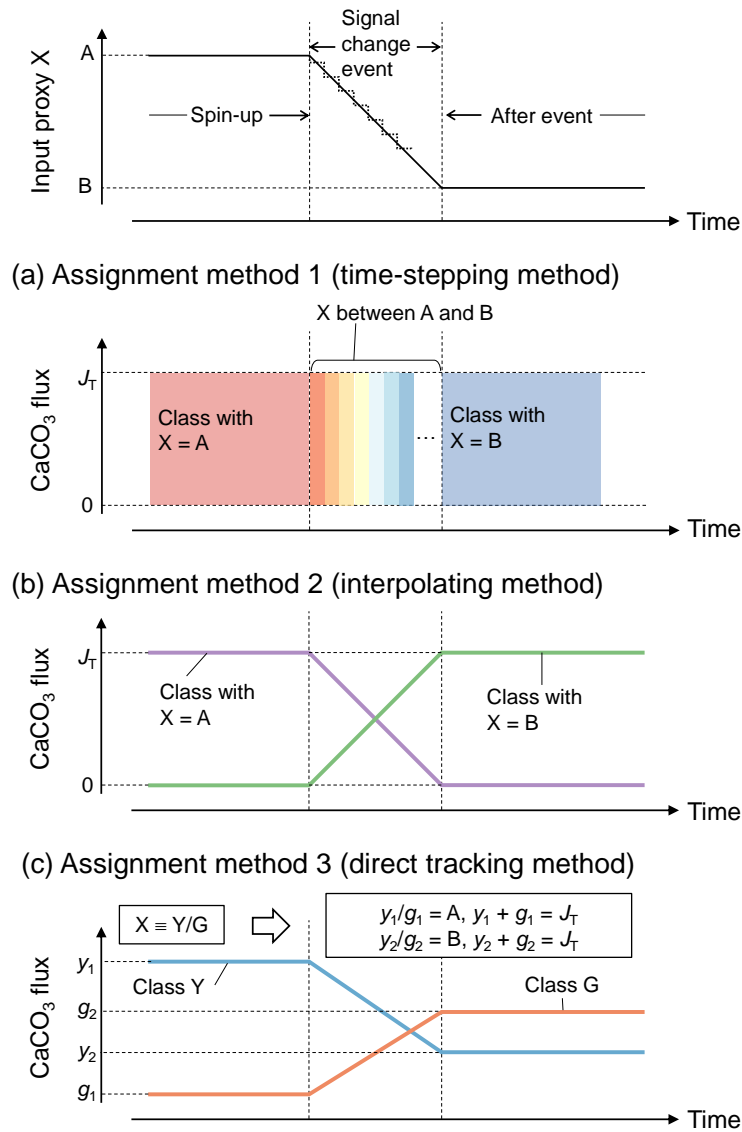
Figure 3. Schematic of signal tracking simulation. Input proxy signal X (solid line in the uppermost panel) is reflected in rain fluxes of multiple classes of CaCO₃ particles using three different methods (a–c). Method 1 (a) approximates input proxy signal by a step function (dotted line in the uppermost panel) and uses different classes of CaCO₃ with separate and unique proxy values at individual time steps. The rain flux of each CaCO₃ class can take either 0 or the total rain flux value J_T. Method 2 (b) uses CaCO₃ classes with the maximum and minimum values of proxy (A and B) and rain fluxes of these CaCO₃ classes are changed so that flux-weighted sums of proxy values of CaCO₃ classes become the same as the input proxy values. Method 3 (c) separates bulk CaCO₃ into CaCO₃ classes that define the proxy signal (classes Y and G), and rain fluxes of these CaCO₃ classes are calculated based on the proxy signal values (see boxes). See Section ~~2.5~~2.4.1 for more details.
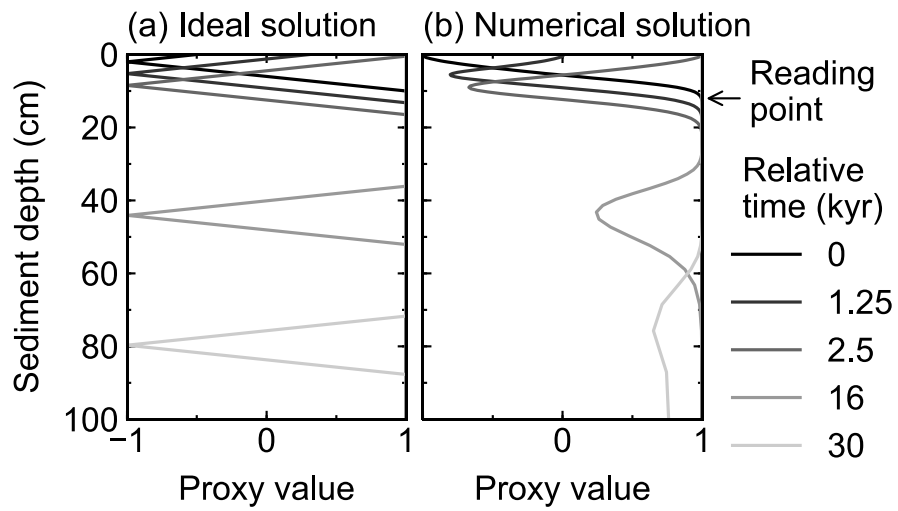
Figure 4. Comparison of ideal (a) and numerical (b) solutions for burial advection of proxy signal. To minimize the effect of numerical diffusion in numerical solution, signal values are read at just below the mixed layer as denoted by an arrow.
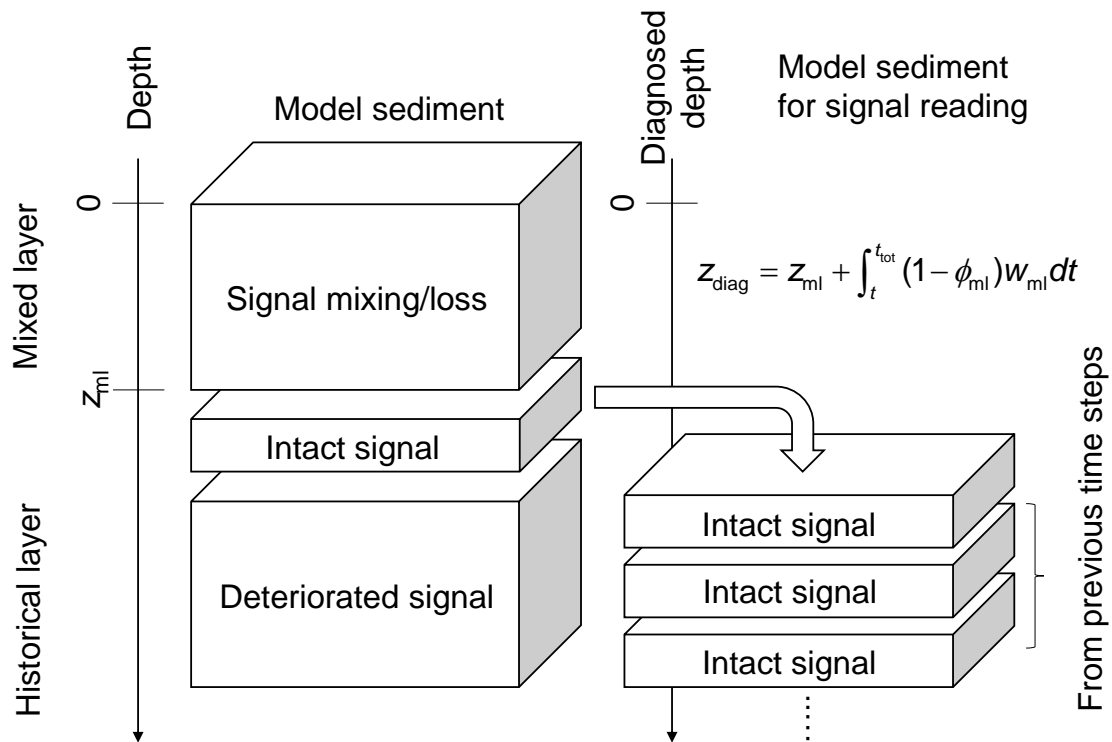
Figure 5. Schematic of sediment column for signal tracking. Left side of diagram shows the sediment calculation domain that can be divided into mixed and historical layers. Signals are bio-mixed or lost by dissolution in the mixed layer and deteriorated at deep depths in the historical layer by numerical diffusion. Right side of diagram shows the sediment column for signal tracking which is composed of sediment layers that used to be located just below the mixed layer in the calculation domain and preserve proxy signals relatively well. Sediment depth in the latter system is denoted as 'diagnosed depth' which can be calculated by the equation in the diagram or Eq. (26).
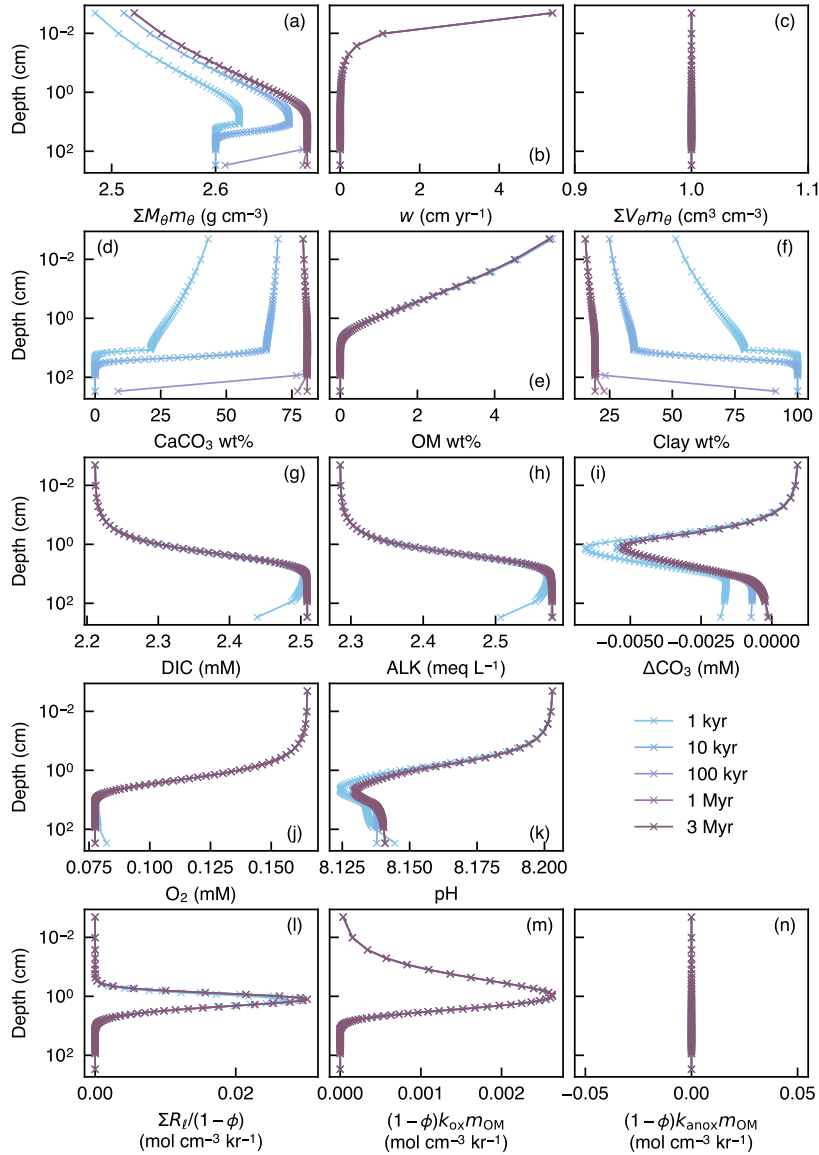
Figure 6. Depth profiles of density (a) and volume fraction (c) of solid sediment, burial velocity (b), weight fractions of bulk $CaCO_3$ (d), organic matter (e) and non-reactive detrital materials (f) in solid sediment, porewater concentrations of total dissolved CO_2 species (g), carbonate alkalinity (h) and oxygen (j), deviation of porewater carbonate concentration from that in equilibrium with $CaCO_3$ (i), porewater pH (k), dissolution rate of $CaCO_3$ (l) and decomposition rate of organic matter in the oxic (m) and anoxic (n) zone of sediment, as a function of time. The boundary conditions of the model are parameterized with the default parameter values (Table 1). The calculations assume 4 classes of $CaCO_3$ particles and Fickian mixing for bioturbation. Illustrated is the temporal evolution of the depth profiles from initial conditions (Section 2.3) to a steady state.
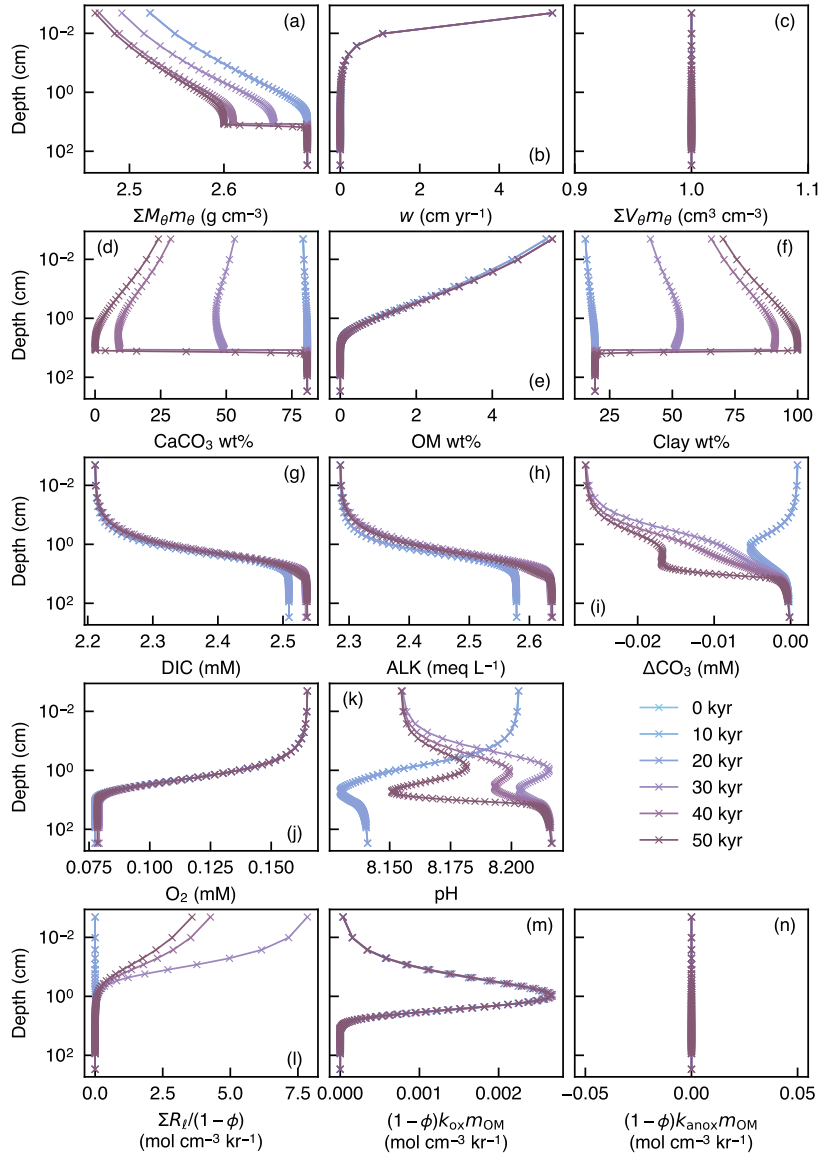
Figure 7. Depth profiles of density (a) and volume fraction (c) of solid sediment, burial velocity (b), weight fractions of bulk $CaCO_3$ (d), organic matter (e) and non-reactive detrital materials (f) in solid sediment, porewater concentrations of total dissolved CO_2 species (g), carbonate alkalinity (h) and oxygen (j), deviation of porewater carbonate concentration from that in equilibrium with $CaCO_3$ (i), porewater pH (k), dissolution rate of $CaCO_3$ (l) and decomposition rate of organic matter in the oxic (m) and anoxic (n) zone of sediment, as a function of time. The boundary conditions of the model change with time as in dissolution experiment #2 (Section 3.2.2, Fig. ~~11~~12). The calculations assume 4 classes of $CaCO_3$ particles and Fickian mixing for bioturbation. Illustrated are the temporal evolutions of the depth profiles which are initially at steady state at 3.5 km of water depth but perturbed by water depth change to 5.0 km between 10 and 50 kyr.
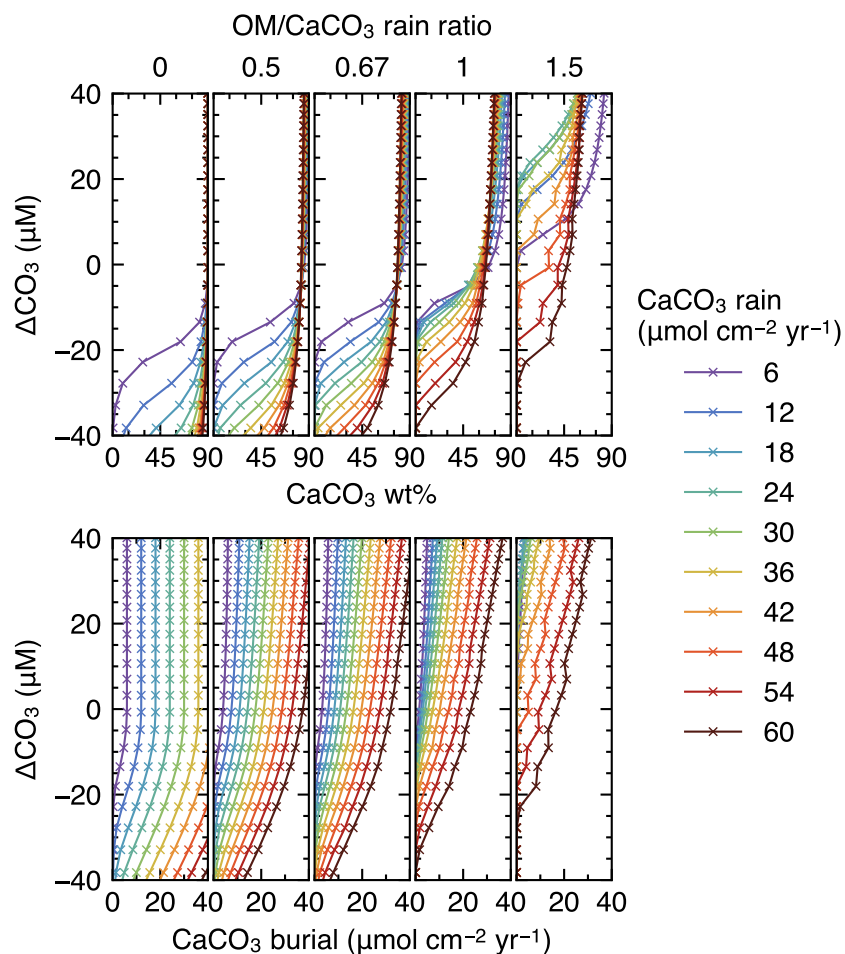
Figure 8. Estimated CaCO₃ weight fractions in mixed layer (a) and burial fluxes (b) as functions of CaCO₃ saturation degree and rain fluxes, with enabling only oxic degradation of organic matter. Saturation degree is measured by the difference of carbonate ion concentration at the seawater-sediment interface from that at calcite saturation, ΔCO_3. The results shown are from the model with a shallower sediment depth (50 cm) and single class of CaCO₃ particles.
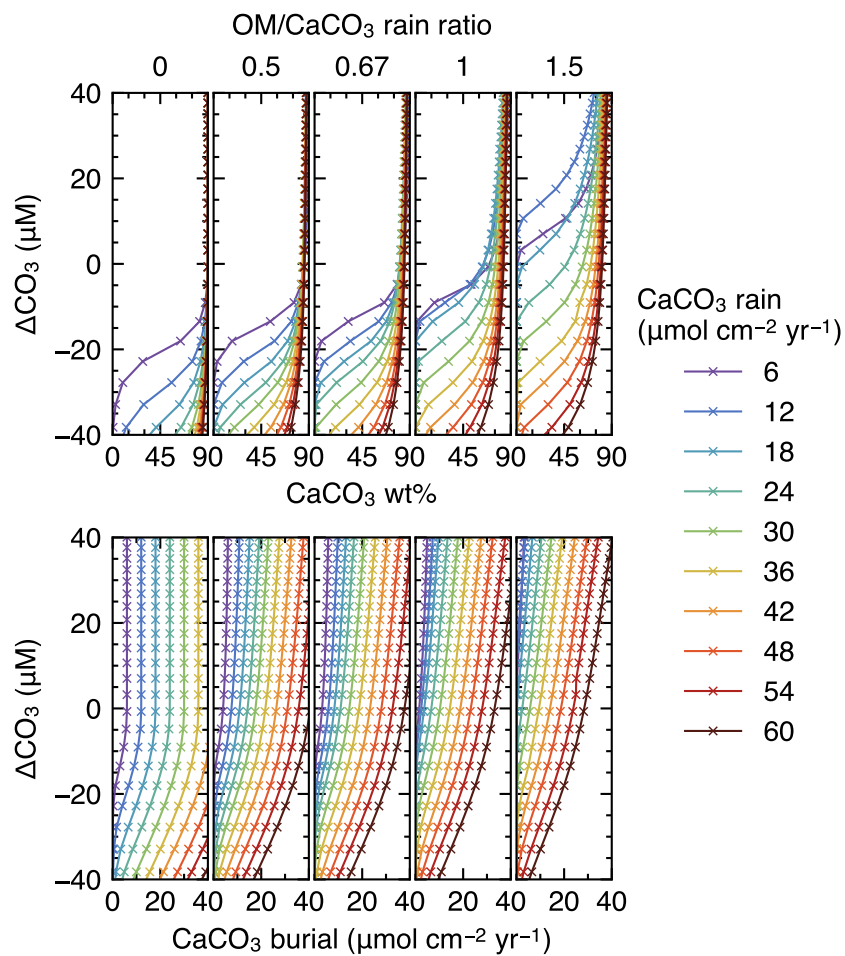
Figure 9. As for Fig. 78, except enabling both oxic and anoxic degradation of organic matter.
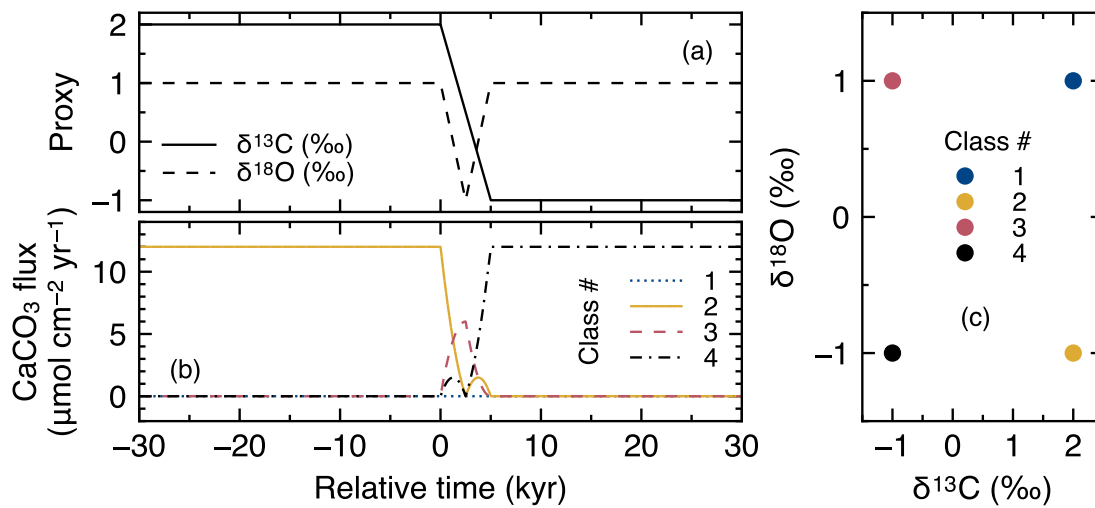
Figure 10. Timelines of proxy inputs (a) and rain fluxes of individual classes of CaCO$_3$ particles (b) with different proxy values (c) in simulations examining signal distortion by bioturbation.
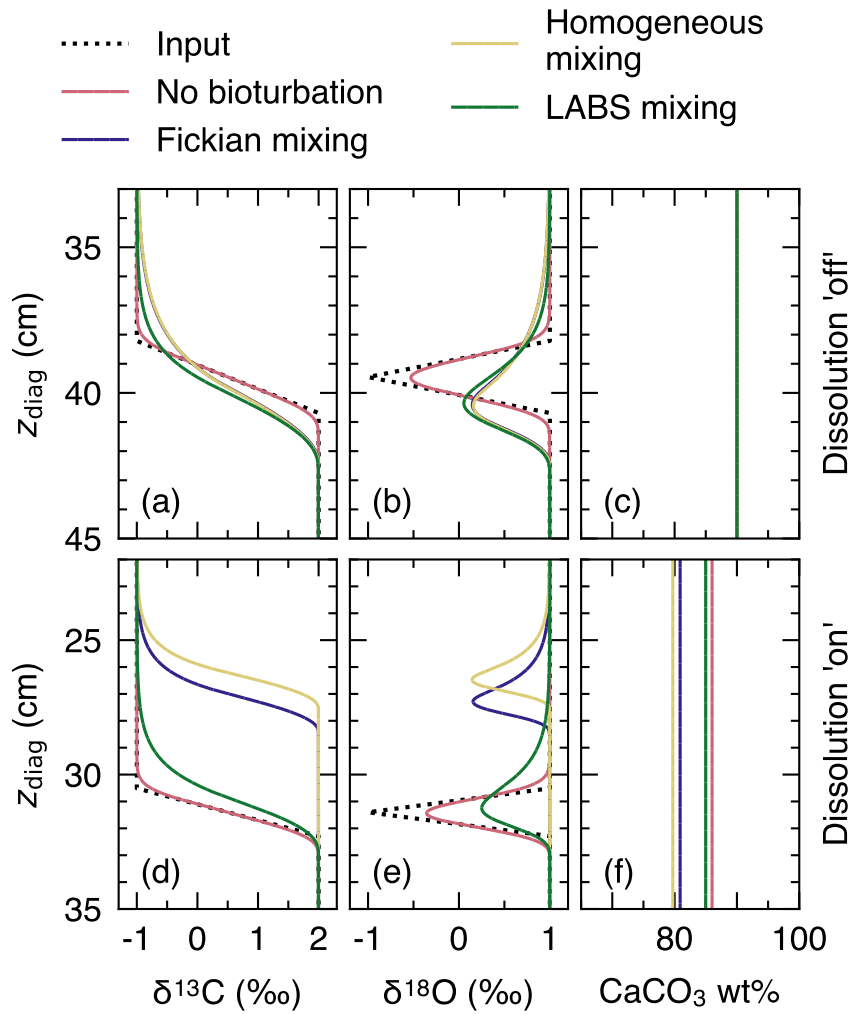
Figure 11. Proxy signals (a, b, d and e) and weight fraction of bulk $CaCO_3$ in solid sediment (c and f) tracked by 4 classes of $CaCO_3$ particles plotted against diagnosed depth in simulations examining signal distortion by bioturbation. In a–c, dissolution rate constants of all $CaCO_3$ classes are fixed at zero, while in d–f, at the default value (Table 1).
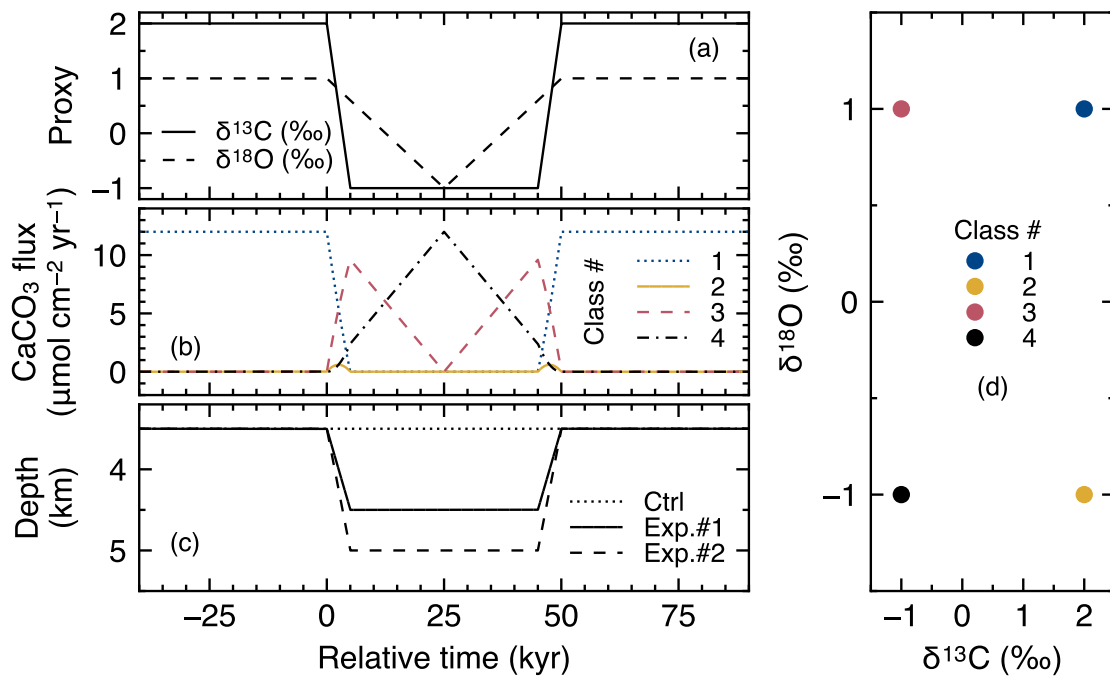
Figure 12. Timelines of proxy inputs (a), rain fluxes of individual classes of CaCO$_3$ particles (b) with different proxy values (d) and water depth changes (c) in simulations examining signal distortion by CaCO$_3$ dissolution. Two different water depth changes are considered, denoted as dissolution experiments #1 and 2 (c). One set of experiments was conducted without changing the water depth for comparison (dotted line in c).
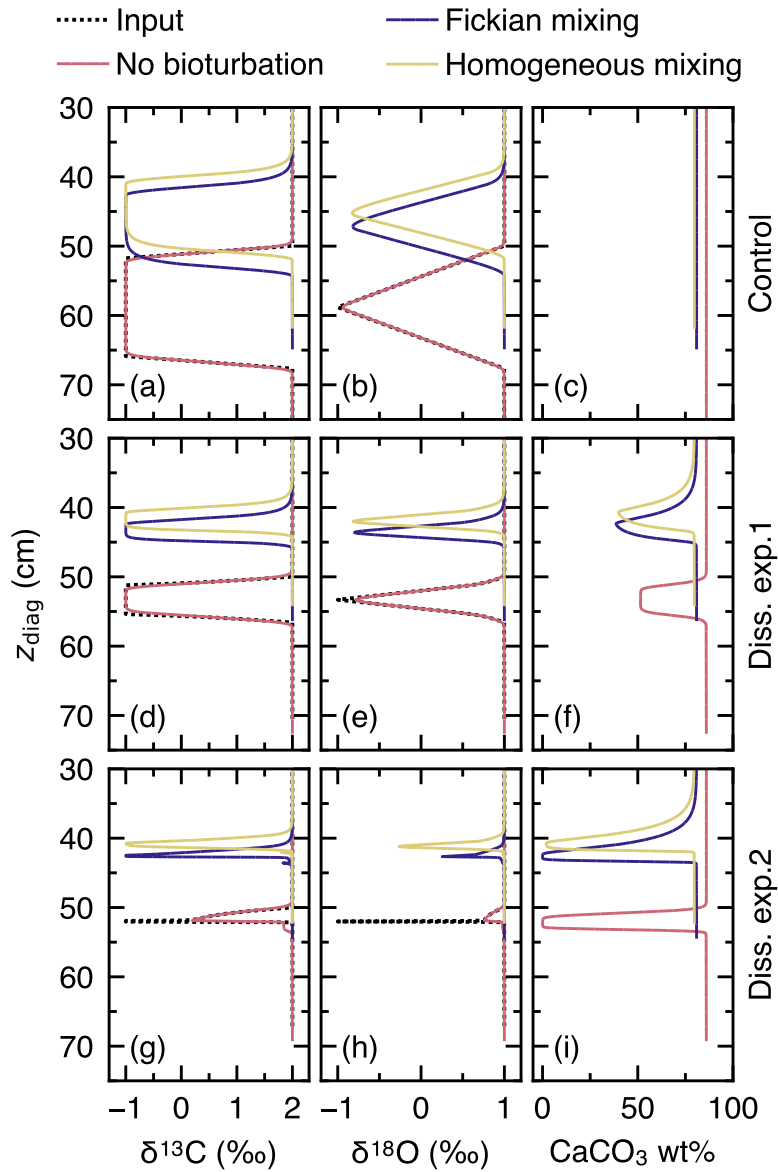
Figure 13. Proxy signals (a, b, d, e, g and h) and weight fraction of bulk CaCO₃ in solid sediment (c, f and i) tracked by 4 classes of CaCO₃ particles plotted against diagnosed depth in simulations examining signal distortion by CaCO₃ dissolution. Two different water depth changes are considered, denoted as dissolution experiments #1 and 2, and compared to the case without water depth change, denoted as control. See Fig. 11c for the assumed water depth changes.
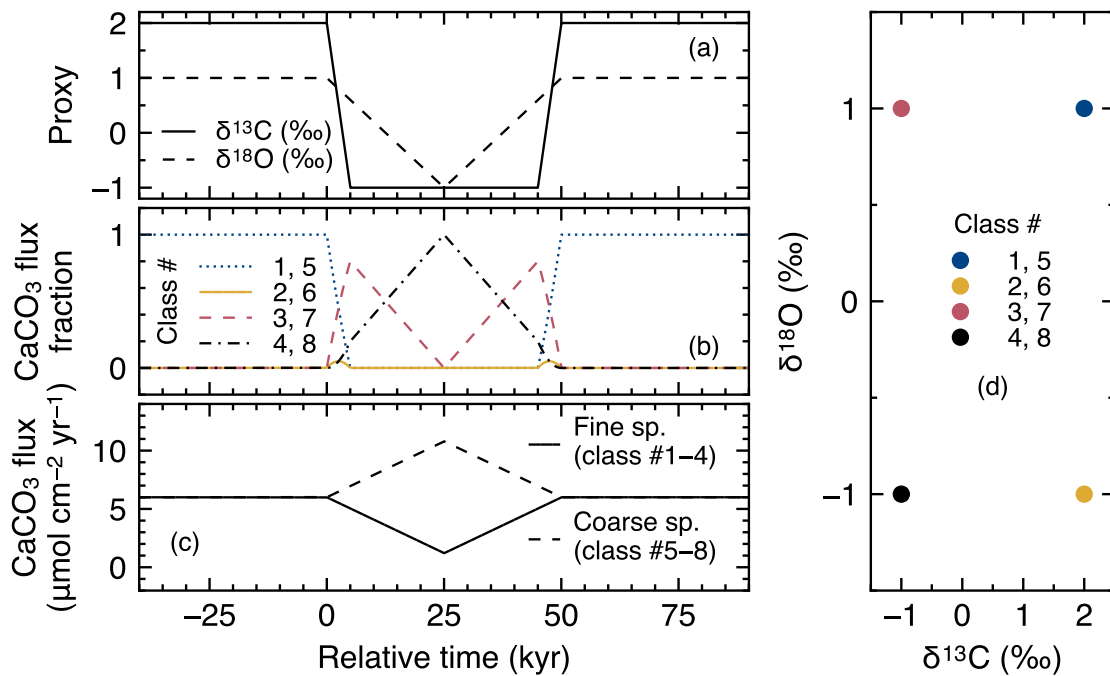
Figure 14. Timelines of proxy inputs (a), normalized rain fluxes of individual classes of $CaCO_3$ particles (b) with different proxy values (d) and total rain fluxes of fine- and coarse-sized $CaCO_3$ species (c) in simulations examining effect of species-specific mixing/dissolution properties. In b, rain fluxes of individual classes of fine and coarse $CaCO_3$ species are normalized against the total rain fluxes of fine and coarse $CaCO_3$ species, respectively, in c.

Figure 15. Proxy signals (a and b) and weight fraction of bulk CaCO$_3$ in solid sediment (c) for fine and coarse CaCO$_3$ species (solid and dotted curves, respectively) tracked by 8 classes of CaCO$_3$ particles in simulations examining effect of species-specific mixing/dissolution properties.
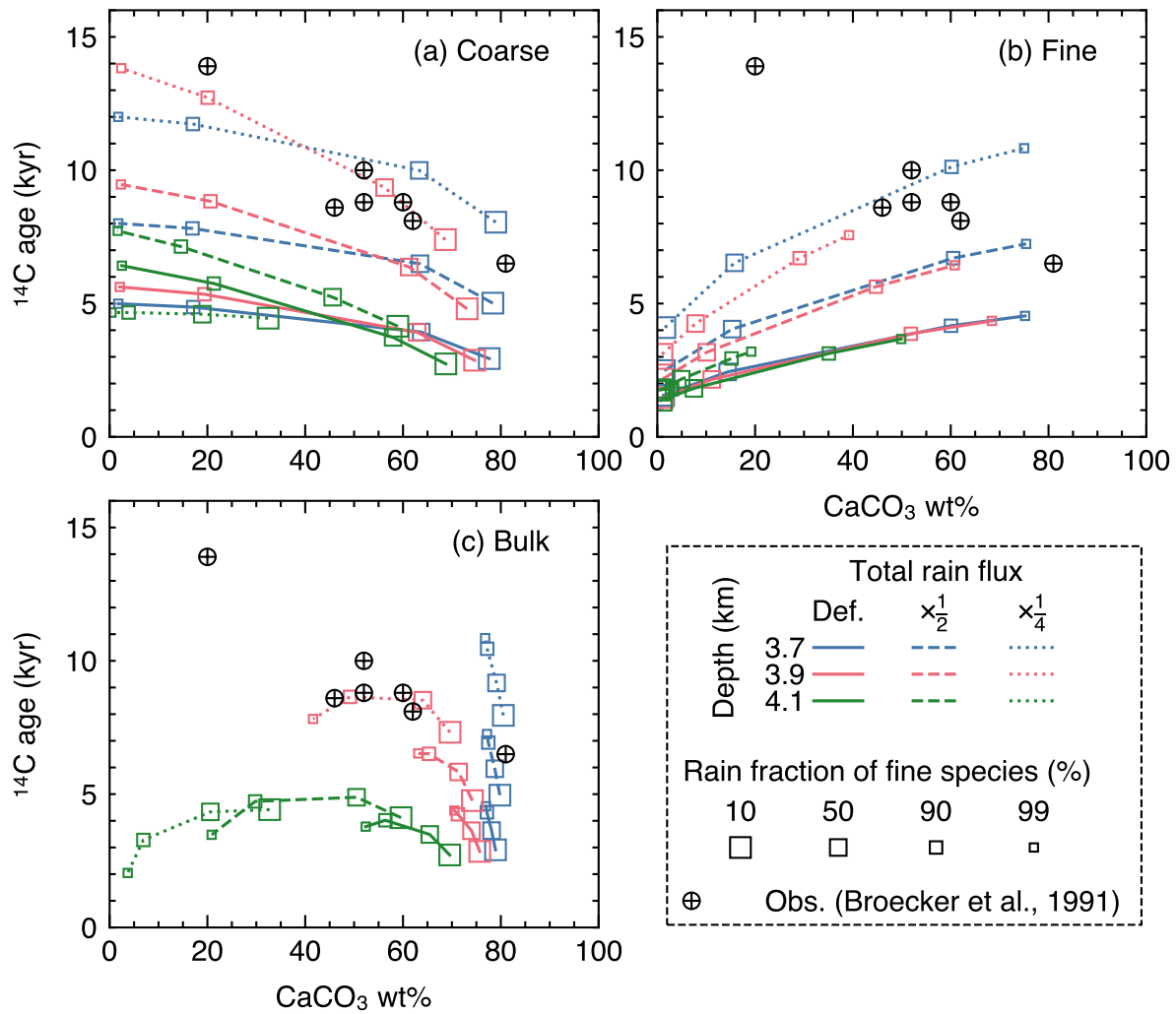
Figure 16. ^{14}C age plotted against CaCO$_3$ wt% in the mixed layer for (a) coarse and (b) fine CaCO$_3$ species, and (c) bulk CaCO$_3$. The values at 12 cm sediment depth are assumed to represent those in the mixed layer.

Table 1. Values of independent parameters and boundary conditions.

Parameter [units]	Symbol[a]	Value[b]	Ref./note[c]
Independent parameters			
Biodiffusion coefficient [$cm^2 \ yr^{-1}$]	~~D_b~~ $D_{b,\theta}$	0.15	1
Density of $CaCO_3$ [$g \ cm^{-3}$]	ρ_ℓ	2.71	2
Density of clay [$g \ cm^{-3}$]	ρ_{clay}	2.60	2
Density of OM[d] [$g \ cm^{-3}$]	ρ_{OM}	1.2	3
Homogeneous transport rate of sediment particles [yr^{-1}]	~~P_h~~ $P_{h,\theta}$	0.001	4
Mixed layer thickness [cm]	z_{ml}	12	5
Molar mass of $CaCO_3$ [$g \ mol^{-1}$]	M_ℓ	100	2
Molar mass of clay [$g \ mol^{-1}$]	M_{clay}	258.16	2
Molar mass of OM [$g \ mol^{-1}$]	M_{OM}	30	6
Mole ratio of O_2 to OM consumed by oxic degradation of OM [dimensionless]	$\gamma_{O_2\text{-}OM}$	1.3	5
Number of sediment ~~grids~~ grid points [dimensionless]	N	100	4
OM/$CaCO_3$ rain ratio [dimensionless]	r	0.7	5
Rate constant for $CaCO_3$ dissolution [yr^{-1}]	$k_{cc,\ell}$	365.25	5
Rate constant for oxic degradation of organic matter [yr^{-1}]	k_{oxic}	0.06	1
Rate constant for anoxic degradation of organic matter [yr^{-1}]	k_{anoxic}	0.06	1
Reaction order for calcite dissolution [dimensionless]	η_{cc}	4.5	5
Boundary conditions			
Carbonate alkalinity at seawater-sediment interface [mM]		2.285	4
Oxygen concentration at seawater-sediment interface [mM]		0.165	4
Salinity [‰]	S	35	4
Temperature [°C]	T_C	2	7
Total $CaCO_3$ rain flux [$\mu mol \ cm^{-2} \ yr^{-1}$]	J_{CaCO_3}	12	5
Total concentration of aqueous CO_2 species at seawater-sediment interface [mM]		2.211	4
Total sediment depth [cm]	z_{tot}	500	4
Water depth [km]	L	3.5	8

[a] Given if defined in main text or used in equations in Tables 2 and 3.

[b] Default values are given, which are used unless otherwise described.

[c] [1] Emerson (1985). [2] From Robie and Hemingway (1995), assuming kaolinite ($Al_2Si_2O_5(OH)_4$) and calcite as representative clay and $CaCO_3$ phases, respectively. [3] A value close to the lower limit of the range (1.14–$1.68 \ g \ cm^{-3}$) reported by Mayer et al. (2004) is adopted (cf. Meyers, 2007). [4] Assumed. [5] Archer (1991). [6] Calculated assuming the chemical formula of OM as CH_2O. [7] Boudreau (1996). [8] Assumed, close to calcite saturation horizon and above calcite compensation depth in the modern oceans (e.g., Emerson and Archer, 1990; Oxburgh and Broecker, 1993).

[d] OM denotes organic matter.

Table 2. Dependent parameters and their equations.

Parameter [units]	Symbol[a]	Equation[b]	Ref./note[c]
Absolute temperature [K]	T	$T = T_C + 273.15$	
Concentration of aqueous CO_2 [mol cm^{-3}]		$c_{ALK}/(K_1/[H^+] + 2K_1 K_2/[H^+]^2)$	1
Concentration of aqueous species σ [mol cm^{-3}]	c_σ	Eq. (2)	2
Concentration of bicarbonate ion [mol cm^{-3}]		$c_{ALK}/(1 + 2K_2/[H^+])$	1
Concentration of carbonate ion [mol cm^{-3}]	$c_{CO_3^{2-}}$	$c_{CO_3^{2-}} = c_{ALK}/([H^+]/K_2 + 2)$	1
Concentration of H$^+$ [mol kg^{-1}]	[H$^+$]	$[H^+] = [-K_1(1 - c_{DIC}/c_{ALK}) + \{K_1^2(1 - c_{DIC}/c_{ALK})^2 \\ - 4K_1 K_2(1 - 2c_{DIC}/c_{ALK})\}^{0.5}]/2$	1
Concentration of solid species θ [mol cm^{-3}]	m_θ	Eq. (1)	2
Detrital rain flux [μg cm^{-2} yr^{-1}]		$(1/9) J_{CaCO_3} M_\ell$	1
Diffusion coefficient for ALK [cm^2 yr^{-1}]	D_{ALK}	$D_{ALK} = 151.69 + 7.93 T_C$	3
Diffusion coefficient for DIC [cm^2 yr^{-1}]	D_{DIC}	$D_{DIC} = 151.69 + 7.93 T_C$	3
Diffusion coefficient for dissolved O$_2$ [cm^2 yr^{-1}]	D_{O_2}	$D_{O_2} = 348.62 + 14.09 T_C$	3
Formation factor [dimensionless]	F	$F = \phi^{-3}$	4
Molar volume [cm^3 mol^{-1}]	V_θ	$V_\theta = M_\theta/\rho_\theta$	2
OM rain flux [μmol cm^{-2} yr^{-1}]		$r J_{CaCO_3}$	1
Porosity [dimensionless]	ϕ	$\phi = 0.1932 \exp(-z/3) + 0.8068$	5
Pressure [bar]	p	$p = 100L$	6
Saturation degree of calcite [dimensionless]	Ω_{cc}	$\Omega_{cc} = c_{CO_3^{2-}} \times 10^{-3} \times 10.3 \times 10^{-3}/K_{cc}$	1,7
Sediment depth [cm]	z	$z = z_{tot} \times \ln\{(\beta + \zeta^2)/(\beta - \zeta^2)\}/\ln\{(\beta+1)/(\beta-1)\}$	8

[a] Given if defined in main text or used in equations in Tables 2 and 3.

[b] Parameter values are calculated based on the listed equations unless otherwise described.

[c] [1] Archer (1991). [2] Section 2. [3] Hülse et al. (2018). [4] Ullman and Aller (1982). [5] Archer (1996). No porosity dependence on CaCO$_3$ is assumed. [6] Approximate relation, cf., Saunders and Fofonoff (1976). [7] Dissolved calcium concentration is assumed to be constant at 10.3 mM. [8] Modified after Eq. (9-32) of Hoffman and Chiang (2000, Ch. 9), where ζ denotes the normalized regular grid and $\beta = 5 \times 10^{-11} + 1$.

Table 3. Thermodynamic parameters.

Parameter [units]	Symbol[a]	Equation	Ref./note[b]
Equilibrium constant for carbonic acid dissociation [mol kg^{-1}]	K_1	$-\log K_1 = -126.34048 + 6320.813/T + 19.568224 \times \ln T$ $+ 13.4191 \times S^{0.5} + 0.0331 \times S - 5.33 \times 10^{-5} \times S^2$ $+ (-530.1228 \times S^{0.5} - 6.103 \times S)/T - 2.06950 \times S^{0.5} \times \ln T$ $- \{-(-25.50 + 0.1271 \times T_C) \times p + 0.5 \times (-3.08 \times 10^{-3}$ $+ 0.0877 \times 10^{-3} \times T_C) \times p^2\}/83.131/T/\ln 10$	1
Equilibrium constant for bicarbonate dissociation [mol kg^{-1}]	K_2	$-\log K_2 = -90.18333 + 5143.692/T + 14.613358 \times \ln T$ $+ 21.0894 \times S^{0.5} + 0.1248 \times S - 0.0003687 \times S^2$ $+ (-772.483 \times S^{0.5} - 20.051 \times S)/T - 3.32254 \times S^{0.5} \times \ln T$ $- \{-(-15.82 - 0.0219 \times T_C) \times p + 0.5 \times (1.13 \times 10^{-3}$ $- 0.1475 \times 10^{-3} \times T_C) \times p^2\}/83.131/T/\ln 10$	1
Solubility product of calcite [mol^2 kg^{-2}]	K_{cc}	$-\log K_{cc} = -171.9065 - 0.077993 \times T + 2839.319/T + 71.595 \times \log T$ $+ (-0.77712 + 0.0028426 \times T + 178.34/T) \times S^{0.5} - 0.07711 \times S$ $+ 0.0041249 \times S^{1.5} - \{-(-48.76 + 0.5304 \times T_C) \times p$ $+ 0.5 \times (-11.76 \times 10^{-3} + 0.3692 \times 10^{-3} \times T_C) \times p^2\}/83.131/T/\ln 10$	2

[a] Given if defined in main text or used in equations in Tables 2 and 3.

[b] [1] Millero (1995), Millero et al. (2006). [2] Mucci (1983), Millero (1995)

Table 4. Properties of CaCO$_3$ classes for simulations in Section 3.2.3.

Property [a]	CaCO$_3$ class #							
	1	2	3	4	5	6	7	8
δ^{13}C (‰)	2	2	−1	−1	2	2	−1	−1
δ^{18}O (‰)	1	−1	1	−1	1	−1	1	−1
Size [a]	Fine	Fine	Fine	Fine	Coarse	Coarse	Coarse	Coarse

[a] Coarse classes have the default values for dissolution rate constant and bio-mixing parameters in Table 1. Fine classes have 10 times higher dissolution rate constant and 20 cm mixed layer depth, otherwise the same parameter values as coarse classes.

Table 5. Properties of CaCO$_3$ classes for simulations in Section 3.3.

Property	CaCO$_3$ class #									
	1	2	3	4	5	6	7	8	9	1
Comp. (Ca-) [a]	^{12}C^{16}O$_3$	^{12}C^{18}O^{16}O$_2$	^{13}C^{16}O$_3$	^{13}C^{18}O^{16}O$_2$	^{14}CO$_3$	^{12}C^{16}O$_3$	^{12}C^{18}O^{16}O$_2$	^{13}C^{16}O$_3$	^{13}C^{18}O^{16}O$_2$	^{14}C
Size [b]	Fine	Fine	Fine	Fine	Fine	Coarse	Coarse	Coarse	Coarse	Co

[a] Isotopologue composition of each CaCO$_3$ class, denoted without Ca.

[b] Coarse classes have the default values for dissolution rate constant and bio-mixing parameters in Table 1. Fine classes have 10 times higher dissolution rate constant and 20 cm mixed layer depth, otherwise the same parameter values as coarse classes.