



- 1 Turbidity maximum zone index: A novel model for remote
- 2 extraction of turbidity maximum zone in different estuaries
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- 18 Abstract. Recognizing and extracting estuarine turbidity maximum zone (TMZ)
- 19 efficiently is important for kinds of terrestrial hydrological process. Although many
- 20 relevant studies of TMZ have been carried out around the world, the method of
- 21 extracting and criteria of describing TMZ vary greatly from different regions and





22 different times. In order to improve the applicability of the fixed threshold in previous 23 studies and develop a novel model extracting TMZ accurately in multi estuaries and 24 different seasons by remote sensing imagery, this study estimated the total suspended 25 solids (TSS) concentrations and chlorophyll a (Chla) concentrations in Pearl River 26 Estuary (PRE), Hanjiang River Estuary (HRE) and Moyangjiang River Estuary (MRE) 27 of Guangdong province, China. The spatial distribution characteristics of both TSS 28 concentrations and Chla concentrations were analyzed subsequently. It was found that 29 there was an almost opposite relationship between TSS concentration and Chla 30 concentration in the three estuaries, especially in PRE. The regions of high (low) TSS 31 concentrations are exactly corresponding to the relative low (high) Chla concentrations. 32 Based on the special feature, an index named turbidity maximum zone index (TMZI), 33 defining as the ratio of the difference and sum of logarithmic transformation of TSS 34 concentrations and Chla concentrations, was firstly proposed. By calculating the values 35 of TMZI in PRE on 20 November 2004 (low-flow season), it was found that the 36 criterion (TMZI > 0.2) could be used to distinguish TMZs of PRE effectively. 37 Compared with the true (false) color imagery and the rudimentary visual interpretation 38 results, the TMZs extraction results by TMZI were mostly consistent with the actual 39 distribution. Moreover, the same criterion was further applied in PRE on 18 October 40 2015. The high accuracy and good consistency across seasons were also found. The 41 west shoal of PRE was the main distribution areas of TMZs. In addition, the good 42 performance in extracting TMZs by this newly proposed index were also found in





43 different estuaries and different times (HRE, 13 August 2008, high-flow season; MRE, 44 on 6 December 2013, low-flow season). Compared to the previous fixed threshold (TSS 45 or turbidity) methods, extracting TMZ by TMZI has a higher accuracy and better 46 applicability. Evidently, this unified TMZI is a potentially optimized method to monitor 47 and extract TMZs of other estuaries in the world by different satellite remote sensing 48 imageries, which can be used to improve the understanding of the spatial and temporal 49 variation of TMZs and estuarial processes on regional and global scales, and the 50 management and sustainable development of regional society and nature environment. 51 Keywords: turbidity maximum zone; turbidity maximum zone index; total suspended 52 solid; chlorophyll a; remote sensing; estuary

# 1 Introduction

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54 Turbidity maximum zone (TMZ) is the dynamic turbid water areas within limits 55 in estuary, where the suspended solids (sediment, matter) concentrations are steadily 56 and significantly higher than landward and seaward (Shen, 1995; Gebhardt et al., 2005; 57 Yu et al., 2014; Li et al., 2019; Wang et al., 2020a). It is a special phenomenon of the 58 progress of suspended sediment movement and migration in estuary throughout the 59 world (Schubel, 1968; Shi et al., 1993; Mitchell et al., 2012; Wang et al., 2020b). The 60 spatial distributions and dynamic change of TMZs not only have a deep and wide 61 impact on the formation and development of estuary morphology, channel, shoal and 62 sandbar (Asp et al., 2018; Azhikodan and Yokoyama, 2019; Li et al., 2019), but also





affect the physics, geochemical and biogeochemical processes of estuarine nature 63 64 environment as well as social production activities significantly (Gebhardt et al., 2005; Jalón-Rojas et al., 2016; Kitheka et al., 2016; Toublanc et al., 2016; Yan et al., 2020). 65 66 It could be found that TMZ has long been a hot topic for scientific inquiries and 67 engineering innovations among a broad spectrum of scholars, government agencies, engineering corporations, and communities (Shen et al., 2001; Shi et al., 2017; Jiang et 68 69 al., 2019; Wang et al. 2020a). 70 Previous works have studied TMZ from various aspects based on different data 71 resources and methods, such as the characteristics and dynamics of total suspended 72 solids (TSS) concentrations in TMZs (Yang et al., 2014; Wan and Wang, 2017; Grasso 73 et al., 2018), the mechanisms and formation of TMZ (Brenon and Hir, 1999; Wai et al., 74 2004; Yu et al., 2014; Toublanc et al., 2016), the location, distribution and change of 75 TMZs across the time (Jiang et al., 2013; Jalón-Rojas et al., 2016; Li et al., 2019; Yan 76 et al., 2020), the interaction with other factors and its long-term trend (Gebhardt et al., 77 2005; Chen et al., 2016; Li et al., 2019). It should be noted that where a TMZ is located 78 in an estuary is a fundamental question and an important task for studying TMZ. It was 79 found that there were two major ways to obtain the locations and distributions of TMZs 80 in current studies (Wang et al., 2020a). One was relatively rough description, such as 81 the locations of TMZ correspond to the front of salinity wedge and moving range of 82 stagnation points, or a distance from coastlines (Feng et al., 2002; Mitchell, 2013; 83 Kitheka et al., 2016; Liu et al., 2016; Toublanc et al., 2016; Gong et al., 2017; Zhang et





al., 2019; Yan et al., 2020). Another was relatively quantitative result, extraction of 84 85 TMZs was conducted by some types of thresholds of TSS concentrations or turbidity criterion (Jiang et al., 2013; Yang and Liu, 2015; Chen et al., 2016; Jalón-Rojas et al., 86 87 2016; Shi et al., 2017; Li et al., 2019). However, the above mentioned fixed threshold 88 method has its latent deficiencies. It is a challenging job to generate precisely TMZs 89 extraction results in different time by a fixed threshold of TSS concentration because 90 TSS concentrations showed remarkable variations in different seasons. Moreover, the 91 threshold values are difficult to be transplanted from local regions to other regions and 92 researches due to lacking of scientific basis. The threshold method and criteria varied 93 greatly in different estuaries, in different regions of a same estuary, in same estuary at 94 different time and by different studies, which showed considerable subjectivity. The 95 results are not comparable (Wang et al., 2020a). 96 TSS concentrations in TMZs and adjacent waters have a significant variation 97 (Uncles et al., 2000; Park et al., 2008; Mitchell, 2013; Wang et al., 2018). And many 98 studies have proved that suspended solids could affect the growth of chlorophyll a (Chla) 99 through absorbing and scattering of the sunlight in water areas (Pozdnyakov et al., 2005; 100 Chen et al., 2015; Montanher et al., 2014; Wang et al., 2017a; Wang et al., 2020b). 101 Therefore, we conclude that there is a relationship between TSS concentrations and 102 Chla concentrations and different characteristics in TMZs and normal water bodies in 103 estuary, which might be used to overcome the drawbacks of previous methods of 104 extracting TMZ, and distinguish and recognize TMZ effectively.

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Based on the above analysis, the objectives of this study are to propose a new model with better adaptability and robustness for distinguishing and extracting TMZ in different estuaries and different seasons. To achieve this goal, the TSS concentrations and Chla concentrations in three estuaries (Pearl River Estuary, PRE; Hanjiang River Estuary, HRE; Moyangjiang River Estuary, MRE) were estimated firstly. And the different spatial characteristic of them were further analyzed and compared. Subsequently, the corresponding relationship and special feature between TSS concentration and Chla concentration were fully used to develop a turbidity maximum zone index (TMZI). Finally, this study extracted TMZs in these estuaries at different time by the model (TMZI), and validated and assessed its accuracy. The paper is arranged as follows. The study areas, in situ data, satellite imagery, TSS concentration data were described along with Chla retrieval model and its calibration and validation in Section 2. The spatial analysis of TSS concentration, Chla concentration and the corresponding relationship between them were presented in Section 3.1. The establishment of TMZI and its application and assessment in different estuaries and different times were showed in Sections 3.2-3.5. Finally, the summary and conclusions were given in Section 4.

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# 2 Dataset and methods

#### 2.1 Study areas

124 The study areas include Pearl River Estuary, Hanjiang River Estuary and 125 Moyangjiang River Estuary of Guangdong province, South China (Fig. 1, Fig. 4 and Figs. 7-10). The PRE is located between longitudes 113.45 °114.2 °E and latitudes 126 127 22.25 °22.85 N mainly in the core zone of Guangdong-Hong Kong-Marco Greater Bay 128 Area, Southern Guangdong province; HRE is located between longitudes 116.6 °-117 E 129 and latitudes 23.2 °23.6 N mainly in Shantou city, Eastern Guangdong province; while 130 MRE is located between longitudes 111.9 °112.3 E and latitudes 21.66 °21.8 N mainly 131 in Yangjiang city, Western Guangdong province. Among them, Pearl River has the 132 second large annual runoff and is the third largest river in China. Hanjiang River and 133 Moyangjiang River are the second and third large and famous rivers in Guangdong 134 province (Chen et al., 2011; Wang et al., 2017a; Wang et al., 2018; Wang et al., 2020b). 135 Previous studies reported that the sediment load of them are 7.53•10<sup>7</sup>, 6.93•10<sup>6</sup> and 136 3.27•10<sup>5</sup> ton per year on average (Wang et al., 2017a, b; Wang et al., 2020). It could be found that TMZs often develops in these estuaries and many associated research work 137 138 has been carried out in the regions for a long time, especially in PRE.





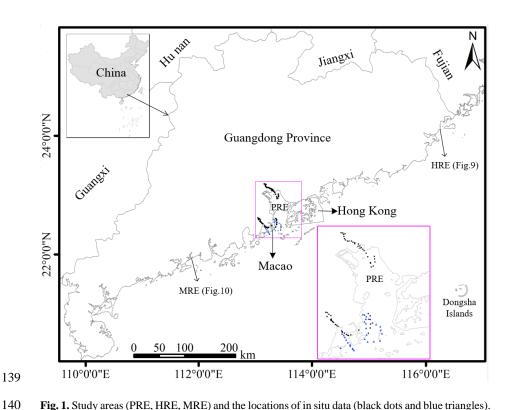


Fig. 1. Study areas (PRE, HRE, MRE) and the locations of in situ data (black dots and blue triangles).

# 2.2 In-situ and satellite data

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The 89 in-situ samples including the reflectance of water surface and Chla concentrations were all collected from PRE, shown in Fig. 1 (black dots and blue triangles). Among them, 60 samples (Fig. 1, black dots; Table 1) were same as our previous work (Chen et al., 2011). This study added newly other 29 samples (Fig. 1, blue triangles; Table 1). The above-mentioned samples were used to recalibrate and validate a Landsat-based Chla concentration retrieval model in this study. Besides, four scenes of Landsat imageries with good quality were used in the study. Two scenes of image from TM and OLI (path/row = 122/44) was captured on 20





November 2004 and 18 October 2015, respectively, covering PRE (Fig. 7a and Fig. 8c).

The other one of them from OLI (path/row = 120/44) was captured on 13 August 2008,

covering HRE (Fig. 9c). The last one from OLI (path/row = 123/45) as well, was

captured on 6 December 2013, covering MRE (Fig. 10c).

#### Table 1

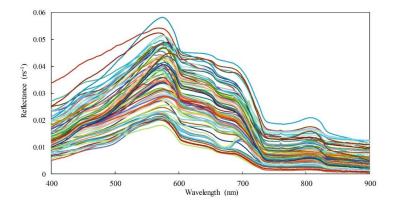
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# 155 Information about the 89 in-situ data.

Date	Samples	Measurements	
Dec 9, 2006	16	Reflectance, Chla	
Dec 21, 2006	12	Reflectance, Chla	Same as
Dec 27, 2007	15	Reflectance, Chla	Chen et al. (2011)
Dec 31, 2007	17	Reflectance, Chla	
Nov 2, 2012	18	Reflectance, Chla	N 1 11 1
Sep 10, 2013	11	Reflectance, Chla	Newly added

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158 Fig. 2. Remote sensing reflectance of surface water of the 89 in situ data.

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## 2.3 Total suspended solids data and chlorophyll a data

This study intends to establish and develop a new model (TMZI) based on TSS concentrations and Chla concentrations, and further extract TMZs in three estuaries of Guangdong province. Therefore, the results of TSS concentrations and Chla concentrations in the study areas should be calculated firstly. For the TSS concentrations data, the results were obtained from our previous works directly (Wang et al., 2017a, b; Wang et al., 2018; Wang et al., 2020). However, the corresponding Chla data needs to be retrieved by Landsat imagery. Consequently, a Landsat-based Chla concentration retrieval model applicable to estuaries of Guangdong province is expected. It can be found that many models have been developed for estimating Chla concentration from different remote sensing data (Gregg and Casey, 2004; Chen et al., 2011; Kim et al., 2016a, b; Attila et al., 2018). Following the feature and form of some typical chlorophyll a retrieval models (Le et al., 2009; Chen et al., 2011; Le et al., 2013; Song et al., 2013), this study recalibrated and validated a three band Landsat-based chlorophyll a model using the 89 in-situ samples (Fig. 3; Equation 1). The model based on Landsat TM and OLI sensors explained about 80% of the Chla concentration variation (Chla: 1.92-92.6 mg/m<sup>3</sup>, N=60, P-value<0.01) and had an acceptable validation accuracy (Chla: 2.33-36.8 mg/m³, RMSE≤3.76 mg/m³, N=29).

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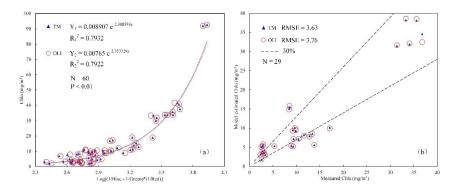
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178 Fig. 3. The calibration (a) and validation (b) results of Chla retrieval models based on 89 in situ data for Landsat sensors.

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$$Chla = a * e^{b*Log[(\frac{1}{R_1} - \frac{1}{R_2})^* \frac{1}{R_3}]}$$
 (1)

Where,  $R_1$ ,  $R_2$  and  $R_3$  represent blue, green, and red band of TM and OLI sensors. 181 182 Parameters a and b corresponding to TM and OLI sensors are 0.008907, 2.308593 and 183 0.00765, 2.353329, respectively. The unit of chlorophyll a concentration is in mg/m<sup>3</sup>.

#### 3 Results and discussion

# 3.1 The spatial characteristic of TSS concentrations and Chla concentrations in estuaries

This study estimated the results of Chla concentrations in each estuary through the developed Chla concentrations retrieval model (Fig. 3). The different spatial distribution characteristics of TSS concentrations and Chla concentrations were analyzed, respectively. Take PRE as an example, it was found that TSS concentrations in low-flow season of PRE (20 November 2004) with a large variation ranging from 1.37 mg/L to more than 200 mg/L (Fig. 4a). Due to the strong interaction between runoff

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and tide, the main region of high TSS concentrations concentrates on west shoal of PRE (Wang et al., 2018), where the TSS concentrations more than 100 mg/L were frequently found. Besides, TSS concentrations in part of east shoal and Neilingding island adjacent waters were also with relative higher value. The other areas of PRE has low TSS concentrations, where the maximum value is general not more than 40 mg/L, especially in Hongkong coastal water bodies (Fig. 4a). Different from TSS concentrations results, the Chla concentrations in PRE are with much lower value (less than 20 mg/m<sup>3</sup> in almost whole PRE) (Fig. 4b). The results was consistent with the findings of Liu et al. (2017) and Huang et al. (2005), which showed Chla concentrations ranged from 0.24 mg/m<sup>3</sup> to 21.5 mg/m<sup>3</sup> in PRE at different time. In addition, it was found that Chla concentrations in PRE showed an almost opposite spatial characteristics compared to TSS concentrations. Except for eastern Lidao district coastal water bodies, the regions of relative high (low) Chla concentrations are exactly the regions of relative low (high) TSS concentrations. These corresponding features are obvious in the four water ways (Humen, Jiaomen, Hongqimen, Hengmen), shoals and channels of PRE (Fig. 4).





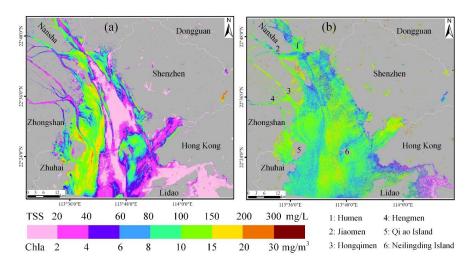


Fig. 4. Estimated TSS concentrations (a) and Chla concentrations (b) in HRE on 20 November 2004.

For further analyzing and assessing the corresponding relationship between TSS concentrations and Chla concentrations in estuary, this study extracted three rows (Fig. 7a; pink lines; rows 1200, 1600, and 1900, columns from 800 to 1300) of TSS concentrations and Chla concentrations values in PRE. The results of row 1600 was shown in Fig. 5(a). The correlation analysis showed remarkable negative correlation of TSS concentrations and Chla concentrations. For the original TSS concentrations and Chla concentrations, the correlation coefficient is -0.6531. While the correlation coefficient reaches about -0.9 for its trend lines (Fig. 5a).

# 3.2 Establishment and application of TMZI

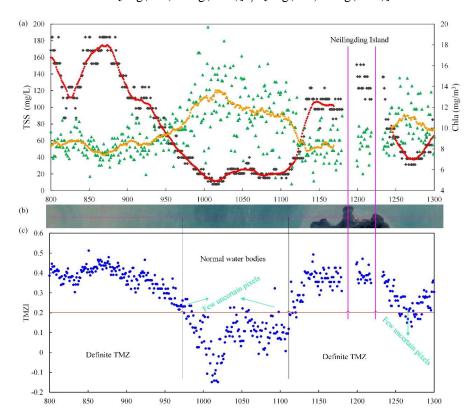
Based on the above analysis and the corresponding features between TSS concentrations and Chla concentrations, we consider that a transform results deriving from the two water color elements may help distinguishing and extracting TMZ better.





Then, this study defined a TMZI as the ratio of the difference and sum of logarithmic transformation of TSS concentrations and Chla concentrations (equation 2) referring to the other remote sensing indexes, such as Normalized Difference Vegetation Index.

$$TMZI = [Log(TSS) - Log(Chla)] / [Log(TSS) + Log(Chla)]$$
 (2)



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Fig. 5. The spatial corresponding relationship between the TSS concentrations (black dots, red trend line) and Chla concentrations (green triangles, orange trend line) of row 1600 (a), the true color imagery (b) and the corresponding values of TMZI (c).

According to the definition and equation, this study calculated the above mentioned corresponding rows TMZI values (Fig. 5c, Fig. 6b and Fig. 6d). Take the results of row 1600 as an example (Fig. 5b and Fig. 5c), the row pixels could be mainly





234 divided into one TMZ (columns from 800 to 975), normal water bodies (columns from 235 975 to 1110) and another TMZ (columns from 1110 to 1300) from left to right. The null data located at columns 1180-1200 and 1220-1235 are Neilingding Island (Fig. 5, Fig. 236 237 7a). Through the comparison to the results of TMZI, it was found that all the values of 238 TMZI corresponding to TMZ pixels are bigger than 0.2 while the values corresponding 239 to normal water bodies pixels are all smaller than 0.2, except for very few blurry pixels 240 (Fig. 5b and Fig. 5c). For the results of rows 1200 and 1900, the similar corresponding 241 characteristics between TMZ and TMZI, and the same criterion were also found (Fig. 6). Hence we can read that TMZI showed a significant feature and had potential to 242 243 develop into a better model for recognizing and extracting estuarine TMZ more 244 effectively.





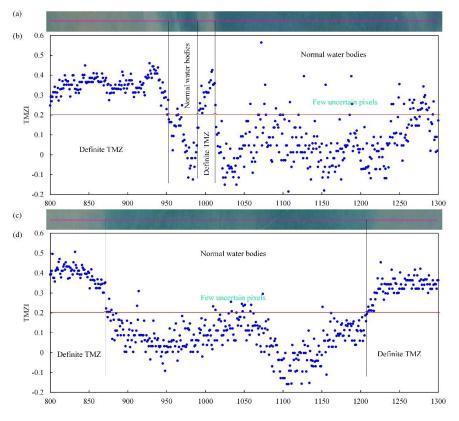


Fig. 6. The true color imagery and the corresponding values of TMZI of rows 1200 (a, b) and 1900

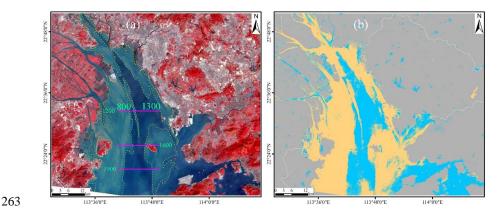
(c, d).

After that, TMZI of the whole Landsat TM imagery was calculated and TMZs in PRE were extracted. Fig. 7(b) showed the spatial distribution results of TMZ in PRE on 20 November 2004 (regions with mango colors). It was found that TMZ is widely distributed throughout PRE, accounting for more than half of water areas in the imagery. Among them, a main TMZ was located within an average distance of 11 km from Panyu, Nansha, Zhongshang and Zhuhai coasts, which is roughly corresponding to the west shoal in PRE. In western Dongguan and Shenzhen coastal water bodies, an approximately rectangular TMZ was developed (within a distance of about 5 km from





coastline), where the East Shoal frequently appears in PRE (Wang et al., 2018). Besides, a third main TMZ in PRE located from surrounding Neilingding Island to western Hong Kong water bodies was been found, although TSS concentrations in the TMZ were lower than that of the former TMZs (Fig.4 a and Fig.7 b). Compared to the preliminary diagrams in our previous works (Fig.7 a) (Wang et al., 2020a, b), it was found that the extracted TMZ results by TMZI in this study got a better accuracy and more natural, which indicates a more effective way to recognize TMZs in estuaries (Figs. 6-7).



**Fig. 7.** False color imagery (USGS 1982; NASA 2001), rough spatial distributions (yellow dashed frames) of TMZ (Wang et al., 2020a, b) (a), and the extracted TMZ results (regions with mango colors) by TMZI (b) in PRE on 20 November 2004 (low-flow season)..

# 3.3 Validation of the accuracy of TMZI in different seasons

Due to the complexity of hydrodynamic environments, estuarine factors and water color elements showed great variations in different seasons, even in the same estuary at





- 270 the different time of the day. Therefore, this study further validated the accuracy of
- 271 TMZI of extracting TMZ in PRE in the high-flow season (18 October 2015).

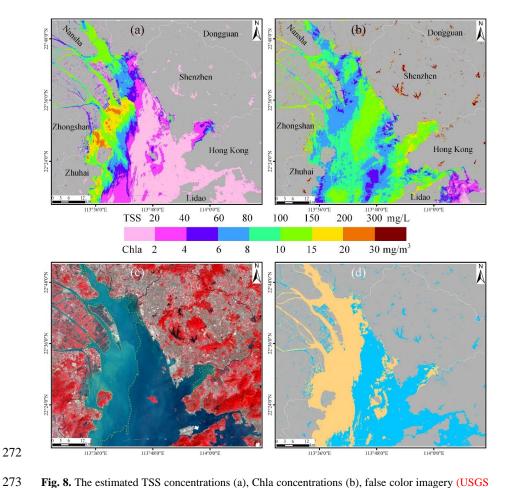


Fig. 8. The estimated TSS concentrations (a), Chla concentrations (b), false color imagery (USGS

274 1982; NASA 2001), rough spatial distributions (yellow dashed frames) of TMZ (Wang et al., 2020b)

(c), and extracting TMZ results (regions with mango colors) by TMZI (d) in PRE on 18 October

276 2015 (high-flow season).

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Fig. 8(a) and Fig. 8(b) showed the retrieved TSS concentrations and Chla concentrations results in high-flow season of PRE. It is clearly that the results in

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different sesaon had big difference (Fig. 4 and Fig. 8). On 18 October 2015, TSS concentrations in PRE had wider variables, ranging from 2.23 to 286.6 mg/L. However, the water bodies with high TSS concentrations (more than 80 mg/L) were mainly in the outlets of four waterways (Humen, Jiaomen, Hongqimen and Hengmen). The other areas of PRE had much lower TSS concentrations, where the TSS concentrations were less than 20 mg/L generally (Fig. 8a). Similar to the corresponding features between TSS concentrations and Chla concentrations in low-flow season, the almost opposite spatial characteristics still existed in high-flow season. For regions with relative high (low) Chla concentrations where showed relative low (high) TSS concentrations (Fig. 8a and b). But it shoul be noted that eastern Lidao district coastal water bodies were an exception, which same to the results in low-flow season (Fig. 4). Both TSS concentrations and Chla concentrations in the zone were relative low (Fig. 4 and Fig. **8**). Based on the resluts of TSS concentrations and Chla concentrations of PRE on 18 October 2015, the study calculated TMZI and extracted TMZs of PRE in high-flow season (Fig. 8d; regions with mango colors). In comparison with the rough diagram results by the original imagery directly (Fig.8 c) (Wang et al., 2020b), the newly extracted TMZs in this study showed a higher accuracy and agreed better with the reality. It was found that there remained only one main TMZ along the west coast of PRE (Fig. 8d), which similar to one of the main TMZs in low-flow season 2004 (Fig. 7b). However, there still existed obvious difference at different seasons, such as TMZs





300 in Hongqimen and Hengmen waterways and eastern Zhuhai coasts (Fig. 7b and Fig. 301 8d). The other TMZs in high-flow season 2015 were mainly located in the sourronding 302 Dachanwan Wharf of Shenzhen and Neilingding Island. The distributions were obvious 303 less than them in low-flow season 2004 (Fig. 7b). Besides, two relative small isolated 304 TMZs could be found at western two artificial islands of the Hong Kong-Zhuhai-Macao 305 Bridge (Fig. 8d), respectively, which may imply the associated influence of human 306 activities. 307 According to the analysis of results in PRE on 18 October 2015, it indicated that the TMZI and the criterion (TMZI > 0.2) also worked well in extracting esturaine TMZ 308 309 in different seasons by Landsat OLI imagery. 3.4 Assessment of the applicability of TMZI in different estuaries 310 311 In order to further assess the applicability of TMZI in different estuaries, the 312 corresponding TMZs results in HRE and MRE were also calculated and validated, 313 similar to PRE.

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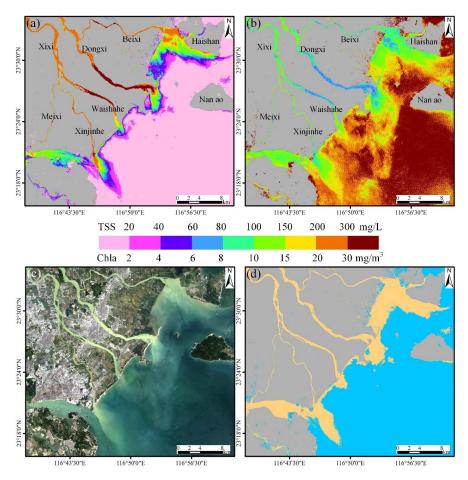


Fig. 9. The estimated TSS concentrations (a), Chla concentrations (b), true color imagery (USGS

1982; NASA 2001) (c), and extracted TMZ results (regions with mango colors) (d) in HRE on 13 August 2008 (high-flow season).

Fig. 9(a) and Fig. 9(b) showed the results of TSS concentrations and Chla concentrations in HRE on 13 August 2008. It is clear that the TSS concentrations in downstream and estuary of HRE are much higher than outer shelf area, especially in the downstream of Dongxi River and Xinjinhe River waterways of Hanjiang River, with a mean value of even more than 300 mg/L (Fig. 9 a). TSS concentrations in the offshore

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area (South China Sea) were less than 20 mg/L frequently. Hence, a significant decreasing trend of TSS concentration could be found from the northwest to southeast in HRE (Fig. 9 a). Besides, the Chla concentrations in HRE showed an opposite spatial distributions characteristics, which similar to the findings in PRE (Fig. 4 and Fig. 8). In general, relative low Chla concentrations were mainly found in downstream and estuary while outer shelf area with high values (Fig. 9 b). Chla concentrations in HRE ranged from 4.1 to 37.3 mg/m<sup>3</sup> (Fig. 9 b), which were a little higher than that of PRE (Fig. 4 and Fig. 8). The results of extracted TMZs in HRE were showed in Fig. 9(d) (regions with mango colors). We found that the TMZs distributed in all downstream and estuary of Hangjiang River. They could be devided into four main TMZs based on different waterways (Beixi, Dongxi, Waishahe, Xinjinhe and Meixi waterways) of Hanjing River. The maximum TMZ was located within an average distance of 3 km from Beixi estuary, western Haishan coasts and the coastlines between Beixi and Dongxi esturay. From Meixi estuary to Xinjinhe estuary, the second large TMZ of HRE was distributed. The region of the main TMZ of Xinjinhe estuary looks knife-shaped, which was mainly caused by the runoff of Xinjinhe waterway and the flow guiding line connected to Longhu District, Shantou City (Fig. 9 d). The other two relative smaller TMZs were distributed in Dongxi estuary and Waishahe estuary, respectively. The results indicated that the TMZs distribution in HRE mainly connected to tide, runoff, estuarine





343 topography and human activity. We found that TMZI has a high applicability in a 344 different eatuary beside PRE. 345 While in HRE, the region of high TSS concentrations were mainly distributed in 346 an average distance of 1.2 km from Yangjiang coastlines, especially in eastern 347 Hailingdati dike water bodies, with a mean value of more than 150 mg/L (Fig. 10a). 348 The outer shelf area has much lower TSS concentrations, where the TSS concentrations were less than 35 mg/L generally. It was also found that Chla concentrations in most 349 350 region of MRE were more than 4 mg/m<sup>3</sup>, except for southwestern Dongping town 351 coastal water bodies where Chla concentrations main ranged from 2 to 4 mg/m<sup>3</sup>. Chla 352 concentrations in Moyangjiang River downstream, Fuchang town coast and outside of 353 Shouchanghe River estuary have relative high values where Chla concentrations more 354 than 8 mg/m<sup>3</sup> were often found (Fig. 10b). Compared to PRE and HRE, the 355 corresponding relationship between TSS concentrations and Chla concentrations in 356 MRE was a little weak. However, there still existed a trend that high (low) TSS 357 concentrations water bodies with relative low (high) Chla concentrations (Fig. 10a and 358 Fig. 10b).



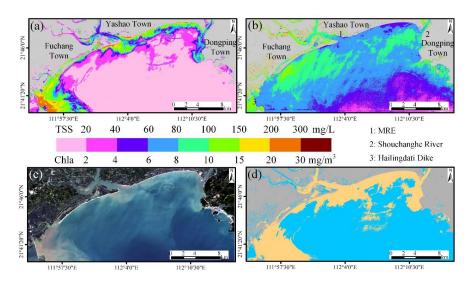


Fig. 10. The estimated TSS concentrations (a), Chla concentrations (b), true color imagery (USGS

1982; NASA 2001) (c), and extracted TMZ results (regions with mango colors) (d) in MRE on 6 December 2013 (Low-flow season).

Fig. 10(c) and Fig. 10(d) showed the ture color imagery of MRE and the extracted TMZs results (regions with mango colors). It was found that there were two main TMZs in MRE on 6 December 2013. The first TMZ mainly distributed from inside and outside of Moyangjiang River estuary to Shouchanghe River estuary, with a distance of about 1.8 km from coastlines (Fig. 10 d). The TMZ distribution in this region mainly attribute to interaction of tide and runoff. Another mian TMZ was in the regions with a distance of 4 km from Hailingdati dike, which mainly caused by the obstruction agains ocean currents of Hailingdati dike (Fig. 10 d). In addition, it was clear that several small long narrow TMZs were also accuracy extreacted through TMZI and the same criterion as that in PRE and HRE. All the results in the three estuaries showed that extracting TMZ

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based on TMZI and the criterion (TMZI > 0.2) has a better applicability in multi eatuaries.

## 3.5 TMZI compared to fixed threshold criterion of previous studies

As the above mentioned, previous studies extracts TMZ mainly based on threshold of TSS concentrations or turbidity. For example, Jalón-Rojas et al. (2016) used thresholds of 500 mg/L (300 NTU) and 1000 mg/L (600 NTU) to define moderatelyconcentrated TMZ and highly-concentrated TMZ in France Loire Estuary; Jiang et al. (2013) and Li et al. (2019) defined TMZ as the areas with TSS values larger than 700 mg/L in Yangtze Estuary and Hangzhou Bay. While for the TMZ in PRE, it was found that TSS values in studies of Shi et al. (2017) and Wai et al. (2004) were more than 89.4 mg/L and about 150 mg/L, respectively. Based on the two criteria (TMZ: TSS > 89.4 mg/L or TSS > 150 mg/L), this study calculated and extracted TMZs in PRE. The results were showed in Fig. 11(c-f), regions with cyan or yellow colors. It was easy to find that the extracted TMZs results in PRE based on the criterion of Shi et al. (2017) were better that of Wai et al. (2004), no matter on 20 November 2004 (Fig. 11c vs. Fig. 11e, low-flow season) or 18 October 2015 (Fig. 11d vs. Fig. 11f, highflow season). The main reason might be that the time of data source in Shi et al. (2017) was more close to our study than that in the study of Wai et al. (2004), which causes that the criterion of Shi et al. (2017) was more suitable to this study than that of Wai et al. (2004).

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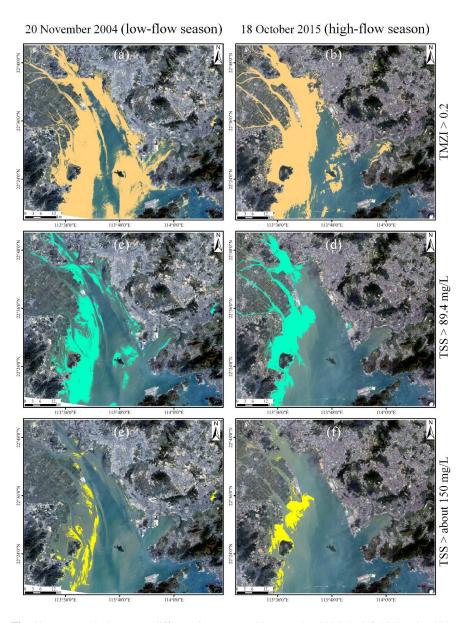
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**Fig. 11.** TMZ results in PRE at different time (a, c, e: 20 November 2004; b, d, f: 18 October 2015; (USGS 1982; NASA 2001)) based on TMZI method by this study (a, b, regions with mango color; same as Fig. 7b and Fig. 8d), the criterion by Shi et al. (2017) (c, d, regions with cyan color), and the criterion by Wai et al. (2004) (e, f, regions with yellow color), respectively.

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Besides, it was also found that a relative good result was obtained in the west shoal of PRE on 20 November 2004 by the criterion of Shi et al. (2017) (Fig. 11c). The extracted TMZs were almost consistent with the reality compared to the true color imagery and our rudimentary visual interpretion results (Wang et al., 2020a, b). However, the accuracy in the east shoal and surrounding Neilingding Island of PRE was not as high as that in the west shoal, where much obvious distributions of TMZs were not recognized effectively (Fig. 11c). What is worse, it was found that the same criterion did not work well in the west shoal of PRE at a different time (Fig. 11c vs. Fig. 11d). Almost a third of the distributions of TMZs in the west shoal of PRE on highflow season was not distinguished and extracted (Fig. 11d). The results based on the criteria of previous studies indicated that fixed thresholds have a distinct disadvantage when extracting TMZ in different times or estuaries. Based on the evaluation and analysis of all the above results (Figs. 7-11), we can find that the TMZI could be widely and effectively applied to the accurate extraction of estuarine TMZ, regardless of the significant variations of hydrodynamic environments, TSS concentrations, Chla concentrations in different estuaries and seasons. Compared to the previous studies and the results from fixed thresholds, we conclude that TMZI has the great potential to develop into a unified model for distinguishing and extracting TMZ effectively and accurately in many other estuaries of the world (Figs. 7-11).

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# 4 Summary and Conclusions

This study established and developed a novel model (turbidity maximum zone index) based on TSS concentration and Chla concentration for distinguishing estuarine turbidity maximum zone from Landsat imageries. It was found that both TSS concentration and Chla concentration showed significant variations and different characteristics in PRE, HRE and MRE in different times (Fig. 4 and Figs. 8-10). However, we found that there still exists a corresponding relationship between TSS concentration and Chla concentration in the three estuaries of Guangdong province. Chla concentrations and TSS concentrations in this study showed an almost opposite spatial distributions characteristics, where relative high (low) Chla concentrations are exactly corresponding to the relative low (high) TSS concentrations (Figs. 4-5 and Figs. 8-10). Therefore, the turbidity maximum zone index (TMZI) was defined and designed as the ratio of the difference and sum of logarithmic transformation of TSS concentrations and Chla concentrations in this study. Compared with the true (false) color imagery or visual interpretation results, it was found that the extracted TMZs results by TMZI were consistent with the reality (Figs. 7-10). Besides, it should be noted that the criterion used for extracting TMZs in different estuaries and seasons was exactly the same (TMZI > 0.2) and got a reasonable accuracy and better performance compared to the previous fixed TSS concentration or turbidity threshold (Fig. 11), which showed that TMZI has a higher adaptability and robustness.





438 The results indicated that there is great potential for optimizing the TMZI to 439 distinguish and extract TMZs from multi-source satellite remote sensing, such as Sentinel, Aqua & Terra-MODIS and SeaWiFS, which also provides great help in 440 441 establishing and developing a unified criterion for extracting TMZs effectively in 442 different estuaries and different times throughout the world. Code and data availability 443 444 All the Landsat remote sensing imageries are 445 https://glovis.usgs.gov/ (USGS 1982; NASA 2001). **Author Contribution** 446 447 The individual contributions and responsibilities of the authors are listed as follows: Chongyang Wang and Li Wang designed the research and wrote the paper; 448 449 Chenghu Zhou and Dan Li guided the research process; Danni Wang, Qiong Zheng, 450 Hao Jiang and Yangxiaoyue Liu collected and analyzed the data; Shuisen Chen, Ji Yang, 451 Xia Zhou and Yong Li revised the manuscript, provided some comments and helped 452 edit the manuscript. All authors have read and agreed to the published version of the 453 manuscript. **Competing interests** 454

The authors declare that they have no conflict of interest.

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## Acknowledgements

458 (41801364), Natural Science Foundation of Guangdong Province (2021A1515012579), 459 Key Special Project for Introduced Talents Team of Southern Marine Science and 460 Engineering Guangdong Laboratory (Guangzhou) (GML2019ZD0301), Scientific 461 Research Project approved by Department of Education of Guangdong Province (2019KONCX209), Guangdong Innovative and Entrepreneurial Research Team 462 Program (2016ZT06D336) and GDAS' Project of Science and Technology 463 464 Development (2020GDASYL-20200104006, 2020GDASYL-20200302001, 465 2019GDASYL-0503001, 2019GDASYL-0301001 and 2019GDASYL-0501001). We 466 would also like to thank USGS for providing the Landsat remote sensing imageries. 467 References Asp, N.E., Gomes, V., Schettini, C.A.F., Filho, P.W.S., Siegle, E., Ogston, A.s., 468 469 Nittrouer, C.A., Silva, J.N.S., Nascimento, W.R., Jr, Souza, S.R., Pereira, L.C.C., 470 Queiroz, M.C., 2018. Sediment dynamics of a tropical tide-dominated estuary: 471 Turbidity maximum, mangroves and the role of the Amazon River sediment load. 472 Estuarine, Coastal and Shelf Science. 473 Attila, J., Kauppila, P., Kallio, K.Y., Alasalmi, H., Keto, V., Bruun, E., Koponen, S., 474 2018. Applicability of Earth Observation chlorophyll-a data in assessment of

This work was funded jointly by National Natural Science Foundation of China





475	water status via MERIS-With implications for the use of OLCI sensor. Remote
476	Sensing of Environment. 212, 273-287.
477	Azhikodan, G., Yokoyama, K., 2019. Seasonal morphodynamic evolution in a
478	meandering channel of a macrotidal estuary. Science of the Total Environment.
479	684, 281-295.
480	Brenon, I., Hir, P.L., 1999. Modelling the Turbidity Maximum in the Seine Estuary
481	(France): Identification of Formation Processes. Estuarine, Coastal and Shelf
482	Science. 49, 525-544.
483	Chen, S., Fang, L., Li, H., Chen, W., Huang, W., 2011. Evaluation of a three-band
484	model for estimating chlorophyll-a concentration in tidal reaches of the Pearl River
485	Estuary, China. ISPRS Journal of Photogrammetry and Remote Sensing. 68, 356-
486	364.
487	Chen, S., Han, L., Chen, X., Li, D., Sun, L., Li, Y., 2015. Estimating wide range Total
488	Suspended Solids concentrations from MODIS 250-m imageries: An improved
489	method. ISPRS Journal of Photogrammetry and Remote Sensing. 99, 58-69.
490	Chen, X., Shen, Z., Yang, Y., 2016. Response of the turbidity maximum zone in the
491	Yangtze River Estuary due to human activities during the dry season.
492	Environmental Science and Pollution Research. 11, 1-16.
493	Feng, H., Cochran, J.K., Hirschberg, D.J., 2002. Transport and sources of metal
494	contaminants over the course of tidal cycle in the turbidity maximum zone of the
495	Hudson River estuary. Water Research. 36, 733-743.





496	Gebhardt, A.C., Schoster, F., Gaye-Haake, B., Beeskow, B., Rachold, V., Unger, D.,
497	Ittekkot, V., 2005. The turbidity maximum zone of the Yenisei River (Siberia) and
498	its impact on organic and inorganic proxies. Estuarine, Coastal and Shelf Science.
499	65, 61-73.
500	Gong, S., Gao, A., Lin, J., Zhu, X., Zhang, Y., Hou, Y., 2017. Temporal-spatial
501	distribution and its influencing factors of suspended particulate matters in
502	Minjiang lower reaches and estuary. Journal of Earth Sciences and Evironment.
503	39(6), 826-836.
504	Grasso, F., Verney, R., Hir, P.L., Thouvenin, B., Schulz, E., Kervella, Y., Fard, I.K.P.,
505	Lemoine, JP., Dumas, F., Garnie, V., 2018. Suspended Sediment Dynamics in the
506	Macrotidal Seine Estuary (France) - Part 1: Numerical Modeling of Turbidity
507	Maximum Dynamics. Journal of Geophysical Research: Oceans. 123, 558-577.
508	Gregg, W.W., Casey, N.W., 2004. Global and regional evaluation of the SeaWiFS
509	chlorophyll data set. Remote Sensing of Environment. 93, 463-479.
510	Huang, B., Hong, H., Ke, L., Cao, Z., 2005. Size-fractionated phytoplankton biomass
511	and productivity in the Zhujiang River Estuary in China. Acta Oceanologica Sinica.
512	27, 180-186.
513	Jalón-Rojas, I., Schmidt, S., Sottolichio, A., Bertier, C., 2016. Tracking the turbidity
514	maximum zone in the Loire Estuary (France) based on a long-term, high-
515	resolution and high-frequency monitoring network. Continental Shelf Research.
516	117, 1-11.





517	Jiang, J., He, Q., Zhu, L., Lin, J., 2019. Analysis of hydrodynamic features of the North
518	Passage in the turbidity maximum, Changjinag Estuary. Haiyang Xuebo. 41(1),
519	11-20.
520	Jiang, X., Lu, B., He, Y., 2013. Response of the turbidity maximum zone to fluctuations
521	in sediment discharge from river to estuary in the Changjiang Estuary (China).
522	Estuarine, Coastal and Shelf Science. 131, 24-30.
523	Kim, H.H., Ko, B.C., Nam, J.Y., 2016a. Predicting chlorophyll-a using Landsat 8 OLI
524	sensor data and the non-linear RANSAC method -a case study of Nakdong River,
525	South Korea. International Journal of Remote Sensing. 37, 3255-3271.
526	Kim, W., Moon, JE., Park, YJ., Ishizaka, J., 2016b. Evaluation of chlorophyll
527	retrievals from Geostationary Ocean Color Imager (GOCI) for the North-East
528	Asian region. Remote Sensing of Environment. 184, 482-495.
529	Kitheka, J.U., Mavuti, K.M., Nthenge, P., Obiero, M., 2016. The turbidity maximum
530	zone in a shallow, well-flushed Sabaki estuary in Kenya. Journal of Sea Research.
531	110, 17-28.
532	Le, C., Hu, C., Cannizzaro, J., English, D., Muller-Karger, F., Lee, Z., 2013. Evaluation
533	of chlorophyll-a remote sensing algorithms for an optically complex estuary.
534	Remote Sensing of Environment. 129, 75-89.
535	Le, C., Li, Y., Zha, Y., Sun, D., Huang, C., Lu, H., 2009. A four-band semi-analytical
536	model for estimating chlorophyll a in highly turbid lakes: The case of Taihu Lake,
537	China. Remote Sensing of Environment. 113, 1175-1182.





538	Li, L., Ye, T., Wang, X., Xia, Y., 2019. Tracking the multidecadal variability of the
539	surface turbidity maximum zone in Hangzhou Bay, China. International Journal of
540	Remote Sensing. 1-22.
541	Liu, H., Huang, L., Tan, Y., Ke, Z., Liu, J., Zhao, C., Wang, J., 2017. Seasonal variations
542	of chlorophyll a and primary production and their influencing factors in the Pearl
543	River Estuary. Journal of Tropical Oceanography. 36, 81-91.
544	Liu, R., Wang, Y., Gao, J., Wu, Z., Guan, W., 2016. Turbidity maximum formation and
545	its seasonal variations in the Zhujiang (Pearl River) Estuary, southern China. Acta
546	Oceanologica Sinica. 35, 22-31.
547	Mitchell, S., 2013. Turbidity maxima in four macrotidal estuaries. Ocean & Coastal
548	Management. 79, 62-69.
549	Mitchell, S., Akesson, L., Uncles, R., 2012. Observations of turbidity in the Thames
550	Estuary, United Kingdom. Water and Environment Journal. 26, 511-520.
551	Montanher, O., Novo, E., Barbosa, C., Renno, C., Silva, T., 2014. Empirical models for
552	estimating the suspended sediment concentration in Amazonian white water rivers
553	using Landsat 5/TM. International Journal of Applied Earth Observation and
554	Geoinformation, 29, 67-77.
555	Park, K., Wang, H.V., Kim, SC., Oh, JH., 2008. A Model Study of the Estuarine
556	Turbidity Maximum along the Main Channel of the Upper Chesapeake Bay.
557	Estuaries and Coasts. 31, 115-133.





558 Pozdnyakov, D., Shuchman, R., Korosov, A., Hatt, C., 2005. Operational algorithm for 559 the retrieval of water quality in the Great Lakes. Remote Sensing of Environment. 97, 352-370. 560 561 Schubel, J., 1968. Turbidity maximum of the northern chesapeake bay. SCIENCE. 161, 1013-1015. 562 Shen, H., 1995. New understanding on the study of the maximum turbidity zone in 563 estuaries of China. Advence in Earth Sciences. 10, 210-212. 564 Shen, H., He, S., Mao, Z., Li, J., 2001. On the turbidity maximum in the Chinese 565 566 estuaries. Journal of Sediment Research. 1, 23-29. 567 Shi, W., Shen, H., Li, J., 1993. Review on the formation of estuarine turbidity maximum. 568 Advence in Earth Sciences. 8, 8-13. 569 Shi, Z., Xu, J., Huang, X., Zhang, X., Jiang, Z., Ye, F., Liang, X., 2017. Relationship 570 between nutrients and plankton biomass in the turbidity maximum zone of the 571 Pearl River Estuary. Journal of Environmental Sciences. 57, 72-84. 572 Song, K., Li, L., Tedesco, L.P., Li, S., Duan, H., Liu, D., Hall, B.E., Du, J., Li, Z., Shi, 573 K., Zhao, Y., 2013. Remote estimation of chlorophyll-a in turbid inland waters: 574 Three-band model versus GA-PLS model. Remote Sensing of Environment. 136, 342-357. 575 576 Toublanc, F., Brenon, I., Coulombier, T., 2016. Formation and structure of the turbidity 577 maximum in the macrotidal Charente estuary (France)\_ Influence of fluvial and 578 tidal forcing. Estuarine, Coastal and Shelf Science. 169, 1-14.





579	Uncles, R.J., Bloomer, N.J., Frickers, P.E., Griffiths, M.L., Harris, C., Howland, R.J.M.,
580	Morris, A.W., Plummer, D.H., Tappin, A.D., 2000. Seasonal variability of salinity,
581	temperature, turbidity and suspended chlorophyll in the Tweed Estuary. The
582	Science of the Total Environment. 251/252, 115-124.
583	Wai, O.W.H., Wang, C.H., Li, Y.S., Li, X.D., 2004. The formation mechanisms of
584	turbidity maximum in the Pearl River estuary, China. Marine Pollution Bulletin.
585	48, 441-448.
586	Wan, Y., Wang, L., 2017. Numerical investigation of the factors influencing the vertical
587	profiles of current, salinity, and SSC with in a turbidity maximum zone.
588	International Journal of Sediment Research. 32, 20-33.
589	Wang, C., Chen, S., Li, D., Wang, D., liu, W., Yang, J., 2017a. A Landsat-based model
590	for retrieving total suspended solids concentration of estuaries and coasts in China.
591	Geoscientific Model Development. 10, 4347-4365.
592	Wang, C., Chen, S., Yang, J., Li, Y., Zhou, X., Li, D., Wang, D., 2020. Monitoring total
593	suspended solids concentrations in estuaries based on remote sensing. Beijing:
594	China Water & Power Press.
595	Wang, C., Li, D., Wang, D., Chen, S., 2017b. Detecting the Temporal and Spatial
596	Changes of Suspended Sediment Concentration in Hanjiang River Estuary During
597	the Past 30 Years Using Landsat Imageries. Research Journal of Environmental
598	Science. 11, 143-155.





599	Wang, C., Li, W., Chen, S., Li, D., Wang, D., Liu, J., 2018. The spatial and temporal
600	variation of total suspended solid concentration in Pearl River Estuary during
601	1987–2015 based on remote sensing. Science of the Total Environment. 618, 1125-
602	1138.
603	Wang, C., Wang, D., Yang, J., Fu, S., Li, D., 2020b. Suspended Sediment within
604	Estuaries and along Coasts: A Review of Spatial and Temporal Variations based
605	on Remote Sensing. Journal of Coastal Research. doi:10.2112/JCOASTRES-D-
606	19-00164.1.
607	Wang, C., Zhou, C., Chen, S., Xie, Y., Li, D., Yang, J., Zhou, X., Li, Y., Wang, D., Liu,
608	Y., 2020a. Retrospect and perspective of the estuarine turbidity maximum zone
609	researches. Chinese Science Bulletin. doi: 10.1360/TB-2020-0938.
610	Yan, D., Song, D., Bao, X., 2020. Spring-neap tidal variation and mechanism analysis
611	of the maximum turbidity in the Pearl River Estuary during flood season. Journal
612	of Tropical Oceanography. 39, 20-35.
613	Yang, J., Liu, W., 2015. Characteristics of the maximum turbidity zone in the
614	lingdingyang-Pearl river estuary during the flood season in the recent 30 years.
615	Pearl River Water Transport. 16, 58-62.
616	Yang, Y., Li, Y., Sun, Z., Fan, Y., 2014. Suspended sediment load in the turbidity
617	maximum zone at the Yangtze River Estuary: The trends and causes. Journal of
618	Geographical Sciences. 24, 129-142.





619	Yu, Q., Wang, Y., Gao, J., Gao, S., Flemming, B., 2014. Turbidity maximum formation
620	in a well-mixed macrotidal estuary: The role of tidal pumping. Journal of
621	Geophysical Research: Oceans. 119, 7705-7724.
622	Zhang, X., Chen, X., Dou, X., Zhao, X., Xia, W., Jiao, J., Xu, H., 2019. Study on
623	formation mechanism of turbidity maximum zone and numerical simulations in
624	the macro tidal estuaries. Advances in Water Science. 30, 84-92.