



1	Simulating Lightning NO _X Production in CMAQv5.2 Using mNLDN, hNLDN,
2	and pNLDN Schemes: Performance Evaluations
3	
4	Daiwen Kang ¹ *, Kristen Foley ¹ , Rohit Mathur ¹ , Shawn Roselle ¹ , Kenneth Pickering ² , and Dale
5	Allen ²
6	
7	¹ Computational Exposure Division, National Exposure Research Laboratory, U.S.
8	Environmental Protection Agency, Research Triangle Park, NC 27711, USA
9 10	² Department of Atmospheric and Oceanic Science, University of Maryland, College Park, MD, USA
11	
12	
13	
14	
15	
16	
17	
18	
19	
20	
21	
22	
23	*Corresponding author: Daiwen Kang, US EPA, 109 T.W. Alexander Drive, Research Triangle Park, NC
24	27711, USA. Tel.: 919-541-4587; fax: 919-541-1379; e-mail: kang.daiwen@epa.gov



25

Geoscientific Model Development

Abstract

26	This study assesses the impact of the lightning NO _X (LNO _X) production schemes in the
27	CMAQ model (Kang et al., 2019) on ground-level air quality as well as aloft atmospheric
28	chemistry through detailed evaluation of model predictions of nitrogen oxides (NO _x) and ozone
29	(O_3) with corresponding observations for the U.S. For ground-level evaluations, hourly O_3 and
30	NOx from the US EPA's AQS monitoring network are used to assess the impact of different
31	LNOx schemes on model prediction of these species in time and space. Vertical evaluations are
32	performed using ozonesonde and P-3B aircraft measurements during the DISCOVER-AQ
33	campaign conducted in the Baltimore/Washington region during July 2011. The impact on wet
34	deposition of nitrate is assessed using measurements from the National Atmospheric Deposition
35	Program's National Trends Network (NADP/NTN). Compared with the base model (without
36	LNOx), the impact of LNOx on surface O ₃ varies from region to region depending on the base
37	model conditions. Overall statistics suggest that for regions where surface O3 mixing ratios are
38	already overestimated, the incorporation of additional NOx from lightning generally increased
39	model overestimation of mean daily maximum 8-hr (DM8HR) O3 by 1-2 ppb. In regions where
40	surface O ₃ is underestimated by the base model, LNOx can significantly reduce the
41	underestimation and bring model predictions close to observations. Analysis of vertical profiles
42	reveals that LNO _x can significantly improve the vertical structure of modeled O ₃ distributions by
43	reducing underestimation aloft, and to a lesser degree decreasing overestimation near the surface.
44	Since the base model underestimates the wet deposition of nitrate in most regions across the
45	modeling domain except the Pacific Coast, the inclusion of LNOx leads to reduction in biases
46	and errors and an increase in correlation coefficients at almost all the NADP/NTN sites. Among
47	the three LNOx schemes described in Kang et al. (2019), the hNLDN scheme, which is
48	implemented using hourly observed lightning flash data from National Lightning Detection
49	Network (NLDN), performs best for the ground-level, vertical profiles, and wet deposition
50	comparisons except that for the accumulated wet deposition of nitrate, the mNLDN scheme (the
51	monthly NLDN-based scheme) performed slightly better. However, when observed lightning
52	flash data are not available, the linear regression-based parameterization scheme, pNLDN,
53	provides an improved estimate for LNOx compared to the base simulation that does not include
54	LNOx.





55 **1. Introduction**

56	The potential importance of NO _X produced by lightning (LNO _X) on regional air quality was
57	recognized more than two decades ago (e.g. Novak and Pierce, 1993), but LNO _X emissions have
58	only been added to regional chemistry and transport models during the last decade (e.g. Allen et
59	al., 2012; Kaynak et al., 2008; Koshak et al., 2014; Smith and Mueller, 2010; Koo et al., 2010)
60	owing in part to the limited understanding of this NO _X source (Schumann and Huntrieser, 2007;
61	Murray, 2016; Pickering et al, 2016). As a result of efforts to reduce anthropogenic NOx
62	emissions in recent decades (Simon et al., 2015; https://gispub.epa.gov/air/trendsreport/2018), it
63	is expected that the relative contribution of LNO_X to the tropospheric NO_X burden and its
64	subsequent impacts on atmospheric chemistry will increase in the United States and other
65	developed countries (Kang and Pickering, 2018). The significant impact of LNO _X on surface air
66	quality was earlier reported by Napelenok et al. (2008), in that low-biases in upper tropospheric
67	NOx in Community Multiscale Air Quality Model (CMAQ) (Byun and Schere, 2006)
68	simulations without LNO _X emissions made it difficult to constrain ground-level NOx emissions
69	using inverse methods and Scanning Imaging Absorption Spectrometer for Atmospheric
70	Cartography (SCIAMACHY) NO2 retrievals (Bovensmann et al., 1999; Sioris et al., 2004;
71	Richter et al., 2005). Appel et al. (2011) and Allen et al. (2012) reported that NO3 ⁻ wet deposition
72	at National Atmospheric Deposition Program (NADP) sites was underestimated by a factor of
73	two when LNOx was not included.
74	LNO _X production and distribution were parameterized initially in global models (e.g.
75	Stockwell et al., 1999; Labrador et al., 2005) relying on the work of Price and Rind (1992) and
76	Price et al. (1997) in that lightning flash frequency was parameterized as a function of the
77	maximum cloud-top-height. Other approaches for LNO _X parameterization include a combination
78	of latent heat release and cloud-top-height (Flatoy and Hov, 1997), convective precipitation rate
79	(e.g. Allen and Pickering, 2002), convective available potential energy (Choi et al., 2005), or
80	convectively induced updraft velocity (Allen et al., 2000; Allen and Pickering, 2002). More
81	recently, Finney et al. (2014, 2016) adopted a lightning parameterization using upward cloud ice
82	flux at 440hPa (based upon definitions of deep convective clouds in the International Satellite
83	Cloud Climatology Project (Rossow et al., 1996)) and implemented it in the United Kingdom
84	Chemistry and Aerosol model (UKCA). With the availability of lightning flash data from the





85	National Lightning Detection Network (NLDN) (Orville et al., 2002), recent LNO _X
86	parameterization schemes started to include the observed lightning flash information to constrain
87	LNO _X in regional Chemical Transport Models (CTMs) (Allen et al., 2012). In Kang et al. (2019),
88	we described the existing LNO_X parameterization scheme that is based on the monthly NLDN
89	(mNLND) lightning flash data, and an updated scheme using hourly NLDN (hNLDN) lightning
90	flash data in the CMAQ lightning module. In addition, we also developed a scheme based on
91	linear and log-linear regression parameters using multiyear NLDN observed lightning flashes
92	and model predicted convective precipitation rate (pNLDN). The preliminary assessment of
93	these schemes based on total column LNO _X suggests that all the schemes provide reasonable
94	LNO _X estimates in time and space, but during summer months, the mNLDN scheme tends to
95	produce the most and the pNLDN scheme the least LNO _X .
96	The first study on the impact of LNO_X on surface air quality using CMAQ was conducted
97	by Allen et al. (2012) and followed by Wang et al. (2013) with different ways for parameterizing
98	LNO _X production and different model configurations. In this study, we present performance
99	evaluations using each of the LNO _X production schemes (mNLDN, hNLDN, pNLDN) described
100	by Kang et al. (2019) to provide estimates of LNO _X in CMAQ. In addition to examination of
101	differences in air quality estimates between these schemes, we compare the model predictions to
102	base model estimates without LNOx and evaluate the estimates from all of the simulations
103	against surface and airborne observations.
104	Section 2 describes the model configuration, simulation scenarios, analysis methodology,

Section 2 describes the model configuration, simulation scenarios, analysis methodology,
 and observational data. Section 3 presents the analysis results and Section 4 presents the
 conclusions.

107

108 2. Methodology

109 2.1 The CMAQ model and simulation configurations

110 The three LNO_x production schemes described in Kang et al (2019) were incorporated

into CMAQ v5.2 (Appel et al. 2017; doi:10.5281/zenodo.1167892). The chemical mechanism

used was CB6 (Yarwood et al., 2010) and the aerosol module was AERO6 (Nolte et al., 2015).





- The meteorological inputs were provided by the Weather Research and Forecasting (WRF) 113 model version 3.8 and the model-ready meteorological input files were created using version 4.2 114 of the meteorology-chemistry interface processor (MCIP; Otte and Pleim, 2010). 115 The modeling domain covers the entire contiguous United States (CONUS) and 116 surrounding portions of northern Mexico and southern Canada, as well as the eastern Pacific and 117 western Atlantic oceans. The model domain consists of 299 north-south grid cells by 459 east-118 west grid cells utilizing 12 km x 12 km horizontal grid spacing, 35 vertical layers with varying 119 thickness extending from the surface to 50 hPa and an approximately 10m midpoint for the 120 lowest (surface) model layer. The simulation time period covers the months from April to 121 September 2011 with a 10-day spin-up period in March. 122 Emission input data were based on the 2011 National Emissions Inventory 123 (https://www.epa.gov/air-emissions-inventories). The raw emission files were processed using 124 version 3.6.5 of the Sparse Matrix Operator Kernel Emissions (SMOKE; 125 https://www.cmascenter.org/smoke/) processor to create gridded speciated hourly model-ready 126 input emission fields for input to CMAQ. Electric generating unit (EGU) emissions were 127 obtained using data from EGUs equipped with a continuous emission monitoring system 128 (CEMS). Plume rise for point and fire sources were calculated in-line for all simulations (Foley 129 et al., 2010). Biogenic emissions were generated in-line in CMAQ using BEIS versions 3.61 130 (Bash et al., 2016). All the simulations employed the bidirectional (bi-di) ammonia flux option 131 for estimating the air-surface exchange of ammonia. 132 133 There are four CMAQ simulation scenarios for this study: 1) simulation without LNO_X (Base), 2) simulation with LNO_X generated by the scheme based on monthly information from 134 the NLDN (mNLDN), 3) simulation with LNO_X generated by scheme based on hourly 135 information from the NLDN (hNLDN), and 4) simulation with LNO_X generated by the scheme 136 137 parameterizing lightning emissions based on modeled convective activity (pNLDN) as described in detail in Kang et al. (2019). All other model inputs, parameters and settings were the same 138 across the four simulations. The vertical distribution algorithm is the same for all the LNO_X 139 schemes as also described in Kang et al. (2019). 140 141 142
- 143





144 **2.2** Observations and analysis techniques

145	To assess the impact of LNOx on ground-level air quality, output from the various CMAQ
146	simulations were paired in space and time with observed data from the EPA's Air Quality
147	System (AQS; <u>https://www.epa.gov/aqs</u>) for hourly O3 and NOX. To evaluate the vertical
148	distribution, measurements of trace species from the Deriving Information on Surface Conditions
149	from Column and Vertically Resolved Observations Relevant to Air Quality (DISCOVER-AQ;
150	http://www.nasa.gov/missions/discover-aq) campaign conducted in the Baltimore/Washington
151	region (e.g., Crawford and Pickering, 2014; Anderson et al., 2014; Follette-Cook et al., 2015)
152	were used. During this campaign, the NASA P-3B aircraft measured trace gases including O ₃ ,
153	NO, and NO ₂ . Vertical profiles were obtained over seven locations – Beltsville (Be), Padonia
154	(Pa), Fairhill (Fa), Aldino (Al), Edgewood (Ed), Essex (Es), and Chesapeake Bay (Cb) from
155	approximately 0.3 to 5 km above ground level during P-3B flights over 14 days in July 2011.
156	During this same period, ozonesonde measurements were taken that extended from ground level
157	through the entire model column at two locations (Beltsville, MD and Edgewood, MD shown in
158	Figure 1). Inclusion of LNOx estimates in the CTM simulations also has an important impact on
159	model estimated wet deposition of nitrate. Therefore, assessment was also performed using data
160	from the National Atmospheric Deposition Program's National Trends Network (NADP/NTN,
161	http://ndp.slh.wisc.edu/ntn).
162	Since lightning activity as well as LNOx exhibit distinct spatial variations (Kang and
163	Pickering, 2018), analysis was conducted for the model domain over the contiguous United
164	States, and then for each region as shown in Figure 1. Emphasis is placed on two regions,
165	Southeast (SE) and Rocky Mountains (RM), where lightning activity is more prevalent and
166	LNOx has the greatest impact on model predictions as shown in Results - increasing model bias

166 LNOx has the greatest impact on model predictions as shown in Results - increasing model bia

- in the SE and decreasing bias in the RM. The commonly used statistical metrics, Root Mean
- 168 Square Error (RMSE), Normalized Mean Error (NME), Mean Bias (MB), Normalized Mean
- 169 Bias (NMB), and Correlation Coefficient (R), in the model evaluation field as defined in Kang et
- al. (2005) and Eder et al. (2006) were calculated to assess the basic performance differences
- among all the model cases for their ground-level air quality predictions.
- 172
- 173





174		
175	3.	Results

176 **3.1 Ground-level evaluation for O₃ and NO**x

177 **3.1.1 Statistical performance metrics**

Tables 1 and 2 display the statistical model performance metrics for daily maximum 8-hr 178 (DM8HR) O₃ and daily mean NOx mixing ratios over the domain and each analysis region for 179 all four model cases in July 2011 (Base, mNLDN, hNLDN, and pNLDN). The best performance 180 metrics among the model cases are highlighted in bold. As shown in Table 1, for DM8HR O_3 , 181 the Base simulation has the lowest MB and NMB values over the Domain, while hNLDN 182 produced the smallest RMSE and NME values. mNLDN generated the largest values for both 183 error (RMSE and NME) and biases (MB and NMB), followed by pNLDN. More importantly, all 184 185 model cases with LNOx exhibit slightly higher correlation coefficients than the Base simulation, suggesting the importance of including the contributions of this source for improving the spatial 186 and temporal variability in model predictions. Additionally, the hNLDN simulation exhibited 187 188 higher correlation and lower bias and error relative to the measurements indicating the value of higher temporal resolution lightning activity for representing the associated NO_x emissions and 189 190 their impacts on tropospheric chemistry.

Examining the regional results for DM8HR O₃ in Table 1, the statistical measures indicate 191 192 that in the Northeast (NE), hNLDN outperformed all other model cases with the lowest errors and biases and highest correlation coefficient. In Southeast (SE), the Base performed better with 193 the lowest errors and mean biases, but the correlation coefficient (R) value for hNLDN is slightly 194 higher. Among all the LNOx cases, mNLDN produced the worst statistics in this region. 195 196 Historically, CTMs tend to significantly overestimate surface O₃ in the Southeast US (Lin et al., 2008l Fiore et al., 2009l Brown-Steiner et al., 2015; Canty et al., 2015), and this is speculated to 197 198 be driven in part by an overestimation of anthropogenic NOx emission estimates. Thus, even 199 though lightning is known to impact ambient air quality, including this additional NOx source 200 can worsen model performance in some locations and time periods due to other errors in the modeling system. As noted in Table 1, for SE, the MB values increased by about 1.6 ppb in 201 202 mNLDN and less than 1 ppb in hNLDN and pNLDN. Nevertheless, the correlation coefficients 203 for mNLDN and pNLDN are almost the same with the Base, and hNLDN was slightly higher





204 (0.77 compared to 0.76). These correlations indicate that even though additional NOx increases the mean bias, when it is added correctly in time and space, as with the case of hNLDN, the 205 spatial and temporal correlation are improved. In Upper Midwest (UM), the lowest errors and 206 biases among the model cases are associated with hNLDN, while the worst performance is with 207 mNLDN. In the Lower Midwest (LM), hNLDN performed comparable with the Base, with 208 hNLDN having the highest correlation and lowest mean errors, while the Base has the lowest 209 mean biases. Rocky Mountain (RM) is the only region that shows an underestimation of 210 DM8HR O₃. In this region all the model cases with LNOx outperformed the Base case in all the 211 metrics. Among the three model cases with LNOx, mNLDN produced the lowest MB and NMB 212 values, while hNLDN had the lowest RMSE and NME, and the highest correlation. In the Pacific 213 Coast (PC) region, lightning activity is generally very low compared to other regions (Kang and 214 Pickering, 2018). All model cases with LNOx outperformed the Base case, especially hNLDN 215 which had the lowest mean error and bias and highest correlation among all the cases. 216

Most of the NO_x produced by lightning is distributed in the middle and upper troposphere with only a small portion being distributed close to the surface. As a result, the impact on ground-level NO_x mixing ratios is small. Table 2 shows all the model cases produced similar statistics for the daily mean NO_x mixing ratios at AQS sites across the domain and within all the subregions. Although the changes in model performance are small, the model cases with LNOx exhibit similar or slightly better performance than the Base case.

3.1.2 Time series

Figure 2 presents the timeseries of regional-mean observed and modeled DM8HR O3 for 224 the entire domain and the SE and RM regions during July 2011. Over the domain and in SE, all 225 226 the model cases overestimate the mean DM8HR O₃ mixing ratios on all days with the Base being the closest to the observations. hNLDN is almost the same as the Base with slightly higher 227 228 values on some days. Among all the cases, mNLDN produced the highest values on almost all days through the month, on the order of 1-2 ppb higher than the Base. In contrast, in the RM 229 230 region, the Base significantly underestimates DM8HR O₃ mixing rations on all the days during the month, while all model cases with LNOx improved model predictions relative to 231 232 observations in the region. Among the three model cases with LNOx, mNLDN produced the lowest bias for all the days, closely followed by hNLDN. 233





Figure 3 displays the average daily mean NOx mixing ratios at AQS sites over the same 234 regions as in Figure 2. On most of the days in July 2011, over the domain and in SE, the model 235 cases overestimate NOx values, and on almost half of the days, the overestimation is significant 236 (up to 100%). As noted in Table 2, on average, the overestimation is $\sim 17\%$ over the domain and 237 ~43% in SE. However, in RM, the predicted NOx mixing ratios closely follow the daily 238 observations and on average the modeled and observed magnitude is almost identical (\sim 3% 239 difference). All the model cases, with or without LNOx, produced almost the same mean NOx 240 mixing ratios at the surface. However, the different cases produce different levels of LNOx in the 241 middle and upper troposphere, resulting in differences in O₃ production and transport which 242 impact ground-level O₃ levels. We further explore these features in Section 3.2 which presents 243 evaluation of modeled vertical pollutant distributions. 244

245 **3.1.3 Diurnal variations**

Diurnal plots are used to further examine differences in model evaluation for O3 and 246 NO_x . Figure 4 shows the mean diurnal profiles for hourly O_3 and NO_x over the entire domain, 247 SE, and RM. On a domain mean basis, all model cases overestimate O₃ during the daytime 248 249 hours, while in the SE, the overestimation spans all the hours. In RM, the model cases 250 significantly underestimate O_3 across all the hours except for a few early morning hours, when the model predicted values are very close to the observations. Among all the model cases, as 251 expected, the most prominent differences occurred during the midday hours when the 252 photochemistry is most active. However, the difference between hNLDN (and mNLDN) and the 253 Base is also significant during the night in the RM region, even though the O_3 levels are low. 254 This may be attributed to NOx-related nighttime chemistry in part caused by freshly released NO 255 by cloud-to-ground lightning flashes. The diurnal variations of NOx are similar over the domain 256 and in the regions for all model cases. Appel et al. (2017) reported a significant overestimation of 257 NOx mixing ratios at AQS sites during nighttime hours and underestimation during daytime 258 hours. The bias pattern is identical for all of the LNOx model cases evaluated here (Figure 4). 259

260 **3.1.4 Spatial variations**

Figure 5 shows the impact of the different LNO_x schemes on model performance for DM8HR O₃ at AQS sites. The spatial maps show the difference in absolute MB between the





- 263 cases with lightning NO_x emissions and the Base and is calculated as follows. First, the absolute MB was calculated at each site for each case, e.g. |MB_[Base-Obs]|, then the difference in absolute 264 MB was calculated between model cases, e.g. $|MB_{[hNLDN-Obs]}| - |MB_{[Base-Obs]}|$. The histograms of 265 the differences in absolute MB between model cases in Figure 5 are provided to show the 266 distribution of the change in model performance across space, i.e. the frequency of an 267 improvement in model performance versus a degradation in model performance between cases. 268 As shown in Figure 5, the mNLDN shows increased model bias in the east US and along the 269 California coast, but reduced model bias in the RM. At a majority of the AQS sites, it increases 270 the model bias (only decreases at 26.8% (346) of the sites). The hNLDN also significantly 271 reduces model bias in the RM with a moderate increase in the SE. Overall, in the hNLDN, the 272 mean bias decreased at 61.2% (791) of AQS sites. Similar to mNLDN, increases in mean bias 273 are noted at 29.3% (378) of the AQS sites in the pNLDN simulation. As noted in the histograms, 274 275 the distribution of the model bias in the pNLDN is much narrower than both mNLDN and 276 hNLDN, eliminating the large bias increases in mNLDN and the significant bias decreases in hNLDN. 277
- **3.2 Vertical evaluation for O₃ and NO**_X
- 279

3.2.1 Ozone-sonde observations

A large source of uncertainty in the specification of LNOx is its vertical allocation, which 280 can impact the model's ability to accurately represent the variability in both chemistry and 281 transport. To further assess the impact of the vertical LNOx specification on model results, we 282 compared vertical profiles of simulated model O₃ with extensive ozonesonde measurements 283 available during the study period. Figure 6 presents the vertical profiles for O_3 sonde 284 measurements and paired model estimates of all model cases at Beltsville, MD and Edgewood, 285 MD. At each location, observations from multiple days are available (one or two soundings per 286 day) during the 2011 DISCOVER-AQ campaign in July 2011. The model evaluation was limited 287 to days where the inclusion of LNOx has an obvious impact (the vertical profile lines can be 288 separated) on the model estimates (July 21, 22, 28 and 29 at Beltsville, and July 21, 22, 28, 29, 289 and 30 at Edgewood). We paired the observed data with model estimates in time and space and 290 averaged the model and observed values at each model layer. Only data below 12 km altitude are 291 292 plotted in Figure 6 to exclude possible influence of stratospheric air on O₃. As can be seen in





293 Figure 6, at both locations the Base case underestimates O₃ mixing ratios from around 1 km upwards, but overestimates closer to the surface. When LNOx is included in the simulations, the 294 predicted O₃ mixing ratios increase relative to the Base case starting around 2km, with greater 295 divergence from the Base case at higher altitudes. The two model cases, hNLDN and mNLDN, 296 produced similar O₃ levels until about 6 km, but above that altitude the mNLDN ozone mixing 297 ratios were higher. All the model cases with LNOx performed much better aloft than the Base 298 case. Near the surface, all the model cases overestimated O₃, however hNLDN had smaller bias 299 than the other simulations. This may be attributed to the fact that only hNLDN used the observed 300 lightning flash data directly, and as a result, LNOx was estimated more accurately in time and 301 space. This improvement in model bias at the surface is further investigated in the next section 302 using evaluation against P-3B measurements. 303

304

3.2.2 P-3B measurement

Extensive measurements of lower tropospheric chemical composition distributions over 305 the Northeastern U.S. are available from instruments onboard the P-3B aircraft on 14 days of the 306 DISCOVER-AQ campaign. We utilize measurements from one of the days (28 July 2011) with 307 308 noticeable (the mean vertical profiles of LNO_x cases are separable from that of the base case) 309 lightning impacts, to evaluate the model simulations. Figure 7 shows measured O₃ mixing ratios overlaid on the modeled vertical time-section for 1030 - 1730 UTC. The color-filled circles 310 represent measured O₃ mixing ratios averaged over 60 seconds and the background is the model 311 estimated vertical profiles from the grid cells containing the P-3B flight path for that hour and 312 location. As indicated in the Base case (Figure 7a), the model tends to overestimate O₃ mixing 313 ratios from the surface to about 2 km, but underestimate at altitudes above 2 km. The hNLDN 314 reduced the overestimation below 2km, e.g. fewer grid cells with mixing ratios above 90ppb 315 (shown in red). The other two cases (mNLDN, pNLDN) did not produce the same improvement 316 near the surface. The hNLDN also decreases the underestimation aloft compared to the Base case 317 with O₃ mixing ratios in the 55-65 ppb range (light blue colors), better matching the measured 318 values. This decrease in underestimation aloft is also seen in the mNLDN case, but to a lesser 319 320 degree while the pNLDN case shows only slight improvement aloft over the Base simulation.

To further differentiate the three LNOx model cases, Figures 8-10 show the difference in the time-sections between each of the model cases with LNOx and the Base for NO, NOx, and





O₃ from all the model layers along the P-3B flight path on July 28. As seen in Figure 8, the 323 hNLDN scheme injected most NO above 5 km and small amount near the surface, with the 324 maximum amount injected between 13-14 km. After release into the atmosphere, NO is quickly 325 converted into NO₂ in the presence of O₃, and these collectively result in the NO_x (NO+NO₂) 326 vertical time-section (local production plus transport) shown in the middle panel of Figure 8. 327 NO_x is further mixed down through the time-section and more persistent along the flight path 328 near the surface than is NO. As a result, significant O₃ is produced above 3 km and the maximum 329 O₃ difference appears between 9 and 14 km during the early afternoon hours (from 13:30 to 330 17:30). However, from surface to about 2 km, O_3 is reduced consistently across the entire period, 331 and this is the result of O₃ titration by NO from cloud-to-ground lightning flashes that must have 332 been transported to this layer by storm downdrafts. Since O₃ is significantly underestimated 333 above 3 km and overestimated near the surface by the Base model, the inclusion of LNOx 334 greatly improved the model's performance under both conditions. 335

Comparison of Figure 9 (mNLDN) with Figure 8 (hNLDN) reveals that the time-sections 336 of NO and NOx above 5 km are similar for these two cases, but they are dramatically different 337 near the surface. The near-surface increase in ambient NO noted in the hNLDN is absent in 338 mNLDN, and in fact there are some small decreases in NO, although the reason for this is 339 unclear. The increase in O_3 aloft in the mNLDN case is similar to that seen in the hNLDN case. 340 However, the near-surface reduction in O_3 is almost absent. In the pNLDN case (Figure 10), NO 341 342 mixing ratios are much less than those in hNLDN and mNLDN in the upper layers as a result of less column NO being generated by the linear parameterization. The resulting NOx time-section 343 is also smoothed. The pNLDN time-sections for NO, NOx and O₃ near the surface are similar to 344 the mNLDN case with no change or small decreases compared to the Base case. O₃ mixing ratios 345 increase by more than 30 ppb during the afternoon hours between 10 - 13 km in the pNLDN 346 case, however the increase is not as intense and widespread as the other cases. In summary, the 347 hNLDN scheme produces estimates that are more consistent with measurements at the surface 348 and aloft, compared to the other simulations, reflecting the advantage of using the spatially and 349 temporally-resolved observed lightning flash data. The model performance improvement for 350 simulated O₃ distributions also suggests robustness in the vertical distribution scheme when 351 LNOx is generated at the right time and location. 352





353 To corroborate the above time-section distributions of NO, NO_X, and O_3 in the lightning cases, the lightning NO emissions are traced back on July 28 for each case. It is found that in all 354 cases, the lightning NO was injected about 200 km upwind (north-west) of the flight path. The 355 hNLDN case captured two injections: one occurred during the morning hours (5:00 to 7:00 am) 356 and the other happened during the afternoon hours (after 2:30 pm). Both mNLDN and pNLND 357 captured the afternoon lightning event at the later time (after 3:30 pm for mNLDN and after 4:30 358 pm for pNLDN) with varying intensity, but neither captured the morning lightning event, which 359 explains why the increase of NO and NO_X in the hNLDN case (Figure 8) did not occur in the 360 mNLDN and pNLDN cases (Figures 9 and 10). Also note that the significant increase of NO 361 during the time period from 11:00 to 13:00 occurred about 5 hours after the lightning NO was 362 injected at about 200 km upwind in the hNLDN case. 363

To expand on the evaluation in Figures 7-10 which focused on measurements from July 364 365 28, 2011, we retrieved all the P-3B measurements on days with noticeable lightning impact (July 21, 22, 28, and 29). The 3-D paired observation-model data were grouped together by spiral site 366 and the mean biases (model – observation) were plotted in Figure 11 (a and b) for O₃ and NO, 367 respectively. The boxplots for O₃ in Figure 11a suggests that the Base exhibited larger bias with 368 greater spread (i.e. larger interquartile range) than other model cases incorporating LNOx at most 369 of the locations where aircraft spirals were conducted. At all locations except Aldino, the lowest 370 mean biases in simulated NO and O₃ are noted in the hNLDN simulation. 371

372

373

3.3 Deposition evaluation for nitrate

In addition to contributing to tropospheric O₃ formation, NOx oxidation also leads to gaseous 374 375 nitric acid and particulate-nitrate which are eventually removed from the atmosphere by dry and wet deposition of nitrate (NO_3^-). As a result, inclusion of NOx from lightning also plays an 376 377 important role in nitrogen deposition modeling. To assess the impacts of incorporating LNOx emissions on simulated oxidized nitrogen deposition, we compared model estimated amounts of 378 379 precipitation from NTN network (http://nadp.slh.wisc.edu/ntn/) and wet deposition of NO3⁻ with measurements from the NADP network (http://nadp.slh.wisc.edu/). During summer months in 380 381 2011 (June -August) the WRF model generally reproduces the observed precipitation with a slight underestimate in the east, but the Base model simulation tends to underestimate wet 382





383	deposition of NO_3^- across the domain, with the greatest underestimation in the SE and UM (See
384	Table 3 and Figure 12). All three LNOx simulations increase wet deposition amounts of NO_3^-
385	and decrease model bias in all regions. The bottom panel of Figure 12 shows that the mNLDN
386	simulation resulted in the largest increase over the base model estimates. The NMB is reduced
387	from -35% in the Base to -15% in mNLDN across the domain and from -32% to -2% in the SE.
388	The hNLDN shows very similar model performance to the mNLDN case. In contrast, the wet
389	deposition NO3 ⁻ estimates from the pNLDN case are only slightly higher than the Base case, and
390	as a result the evaluation statistics for pNLND are very similar to the Base statistics. As
391	discussed earlier, the mNLDN tends to produce the most LNOx among the three LNOx schemes,
392	thus it results in the smallest errors in terms of wet deposition of NO3 ⁻ when compared to the
393	Base simulation that significantly underestimated NO3 ⁻ wet deposition. It should be noted that in
394	addition to the LNOx contributions, errors in modeled precipitation amounts and patterns also
395	likely influence the underestimation of NO3 ⁻ wet deposition.

396

397 4. Conclusions

A detailed evaluation of lightning NOx emission estimation parameterizations available
 in the CMAQ modeling system was performed through comparisons of model simulation
 results with surface and aloft air quality measurements.

Our analysis indicates that incorporation of LNOx emissions enhanced O₃ production in 401 the middle and upper troposphere, where O₃ mixing ratios were often significantly 402 403 underestimated without the representation of LNOx. Though the impact on surface O₃ varies from region to region and is also dependent on the accuracy of the NOx emissions from other 404 sources, the inclusion of LNOx, when it is injected at the appropriate time and location, can 405 improve the model estimates. In regions where the base model estimates of O3 were biased 406 high, the inclusion of LNOx further increased the model bias; and a systematic increase is 407 408 noted in the correlation with measurements, suggesting that emissions from other sources likely drive the overestimation. Identifying how errors in emissions inputs from different 409 sources interact with errors in meteorological modeling of mixing and transport, remains a 410 challenging but critical task. Likewise, all the LNOx schemes also enhanced the accumulated 411





412	wet deposition of NO3 ⁻ , that was significantly underestimated by the base model without
413	LNOx throughout the modeling domain except the Pacific Coast.
414	Uncertainty remains in modeling the magnitude and spatial, temporal and vertical
415	distribution of lightning produced NOx. LNOx schemes are built on numerous assumptions
416	and all current schemes also depend on the skill of the upstream meteorological models in
417	describing convective activity. Nevertheless, these schemes reflect our best understanding
418	and knowledge at the time when the schemes were implemented. The use of hourly
419	information on lightning activity yielded LNOx emissions that generally improved model
420	performance for ambient O3 and NOx as well as oxidized nitrogen wet deposition amounts.
421	As more high-quality data from both ground and satellite measurements become available,
422	the performance of the LNOx schemes will continue to improve.

423

424 Code and data availability

CMAQ model documentation and released versions of the source code, including all model
code used in his study, are available at https://www.epa.gov/cmaq. The data processing and
analysis scripts are available upon request. The WRF model is available for download through
the WRF website (http://www.wrf-model.org/index.php).

429 The raw lightning flash observation data used are not available to the public but can be

430 purchased through Vaisala Inc. (https://www.vaisala.com/en/products/systems/lightning-

431 detection). The immediate data behind the tables and figures are available from

https://zenodo.org/record/2621096 (Kang and Foley, 2019). Additional input/output data for
 CMAQ model utilized for this analysis are available upon request as well.

434 435

436 Disclaimer: The views expressed in this paper are those of the authors and do not necessarily437 represent the views or policies of the U.S. EPA.

438

439 Author Contribution

440 Daiwen Kang: data collection, algorithm design, model simulation, analysis, and manuscript
 441 writing.

442 Kristen Foley: data analysis and manuscript writing.

- 443 **Rohit Mathur**: manuscript editing.
- 444 Shawn Roselle: manuscript editing.





- 445 Kenneth Pickering: manuscript editing.
- 446 **Dale Allen:** manuscript editing.
- 447
- 448 Acknowledgement:
- The authors thank Brian Eder, Golam Sarwar, and Janet Burke (U.S. /EPA) for their
 constructive comments and suggestions during the internal review process.
- 451

453	Allen, D., Pickering, K., Stenchikov, G., Thompson, A., and Kondo, Y.: A three-dimensional
454	total odd nitrogen (NOy) simulation during SONEX using a stretched-grid chemical
455	transport model, J. Geophys. Res., 105, 3851–3876, doi:10.1029/1999JD901029, 2000.
456	Allen, D. J. and Pickering, K. E.: Evaluation of lightning flash rate parameterizations for use in a
457	global chemical transport model, J. Geophys. Res., 107, 4711,
458	doi:10.1029/2002JD002066, 2002.
459	Allen, D. J., Pickering, K. E., Pinder, R. W., Henderson, B. H., Appel, K. W., and Prados, A.:
460	Impact of lightning-NO on eastern United States photochemistry during the summer of
461	2006 as determined using the CMAQ model, Atmos. Chem. Phys., 12, 1737–1758,
462	doi:10.5194/acp-12-1737-2012, 2012.
463	Anderson, D. C., Loughner, C. P., Diskin, G., Weinheimer, A., Canty, T. P., Salawitch, R. J,
464	Worden, H. M., Fried, A., Mikoviny, T., Wisthaler, A., and Dickerson, R. R.: Measured
465	and modeled CO and NOy in DISCOVER-AQ: An evaluation of emissions and
466	chemistry over the eastern US, Atmos. Environ., 96, 78-87,
467	doi:10.1016/j.atmosenv.2014.07.004, 2014.
468 469 470 471 472 473	 Appel, K. W., Napelenok, S. L., Foley, K. M., Pye, H. O., Hogrefe, C., Luecken, D. J., Bash, J. O., Roselle, S. J., Pleim, J. E., Foroutan, H., Hutzell1, W. D., Pouliot, G. O., Sarwar, G., Fahey, K. M., Gantt, G., Gilliam, R. C., Heath, N. K., Kang, D., Mathur, R., Schwede, D. B., Spero, T. L., Wong, D. C., and Young, J. O.: Description and evaluation of the Community Multiscale Air Quality (CMAQ) modeling system version 5.1, Geosci. Model Dev., 10, 1703–1732, doi:10.5194/gmd-10-1703-2017, 2017.
474	Appel, K. W., Foley, K. M., Bash, J. O., Pinder, R. W., Dennis, R. L., Allen, D. J., and
475	Pickering, K.: A multi-resolution assessment of the Community Multiscale Air Quality
476	(CMAQ) model v4.7 wet deposition estimates for 2002-2006, Geosci. Model Dev., 4,
477	357–371, doi:10.5194/gmd-4-357-2011, 2011.





478	Bash, J. O., Baker, K. R., and Beaver, M. R.: Evaluation of improved land use and canopy
479	representation in BEIS v3.61 with biogenic VOC measurements in California, Geosci.
480	Model Dev., 9, 2191–2207, doi:10.5194/gmd-9-2191-2016, 2016.
481	Bovensmann, H., Burrows, J. P., Buchwitz, M., Frerick, J., No [°] el, S., Rozanov, V. V., Chance,
482	K. V., and Goede, A. P. H.: SCIAMACHY: Mission Objectives and Measurement
483	Modes, J. Atmos. Sci., 56, 127–150, 1999.c
484 485 486 487	Brown-Steiner, B., Hess, P. G., and Lin, M. Y.: On the capabilities and limitations of GCCM simulations of summertime regional air quality: A diagnostic analysis of ozone and temperature simulations in the US using CESM CAM-Chem, Atmos. Environ., 101, 134–148, doi:10.1016/j.atmosenv.2014.11.001, 2015
488	Byun, D. W. and Schere, K. L.: Rewiew of the governing equations, computational algorithms,
489	and other components of the Models-3 Community Multiscale Air Quality (CMAQ)
490	modeling system, Appl. Mech. Rev., 59, 51-77, 2006.
491 492 493 494	 Canty, T. P., Hembeck, L., Vinciguerra, T. P., Anderson, D. C., Goldberg, D. L., Carpenter, S. F., Allen, D. J., Loughner, C. P., Salawitch, R. J., and Dickerson, R. R.: Ozone and NOx chemistry in the eastern US: evaluation of CMAQ/CB05 with satellite (OMI) data, Atmos. Chem. Phys., 15, 10965–10982, doi:10.5194/acp-15-10965-2015, 2015.
495	Choi, Y., Wang, Y., Zeng, T., Martin, R. V., Kurosu, T. P., and Chance, K.: Evidence of
496	lightning NOx and convective transport of pollutants in satellite observations over North
497	America, Geophys. Res. Lett., 32, L02805, doi:10.1029/2004GL021436, 2005.
498 499	Crawford, J. H. and Pickering, K. E.: DISCOVER-AQ: Advancing strategies for air quality observations for the next decade, EM, A&WMA, September, 2014.
500	Eder, B. K., Kang, D., Mathur, R., Yu, S., and Schere, K.: An operational evaluation of the Eta-
501	CMAQ air quality forecast model, Atmos. Environ., 40, 4894-4905, 2006.
502	Finney, D. L., Doherty, R. M., Wild, O., Huntrieser, H., Pumphrey, H. C., and Blyth, A. M.:
503	Using cloud ice flux to parametrize large-scale lightning, Atmos. Chem. Phys., 14,
504	12665–12682, doi:10.5194/acp-14-12665-2014, 2014.
505	Finney, D. L., Doherty, R. M., Wild, O., and Abraham, N. L.: The impact of lightning on
506	tropospheric ozone chemistry using a new global lightning parameterization, Atmos.
507	Chem. Phys., 16, 7507–7522, doi:10.5194/acp-16-7507-2016, 2016.
508 509	Flatøy, F. and Hov, O.: NOx from lightning and the calculated chemical composition of the free troposphere, J. Geophys. Res., 102, 21 373–21 381, 1997.
510 511 512 513 514	 Fiore, A. M., Dentener, F. J., Wild, O., Cuvelier, C., Schultz, M. G., Hess, P., Textor, C., Schulz, M., Doherty, R. M., Horowitz, L. W., MacKenzie, I. A., Sanderson, M. G., Shindell, D. T., Stevenson, D. S., Szopa, S., Van Dingenen, R., Zeng, G., Atherton, C., Bergmann, D., Bey, I., Carmichael, G., Collins, W. J., Duncan, B. N., Faluvegi, G., Folberth, G., Gauss, M., Gong, S., Hauglustaine, D., Holloway, T., Isaksen, I. S. A., Jacob, D. J., Jonson, J. E., 17





515 516 517 518	Kaminski, J. W., Keating, T. J., Lupu, A., Marmer, E., Montanaro, V., Park, R. J., Pitari, G., Pringle, K. J., Pyle, J. A., Schroeder, S., Vivanco, M. G., Wind, P., Wojcik, G., Wu, S., and Zuber, A.: Multimodel estimates of intercontinental sourcereceptor relationships for ozone pollution, J. Geophys. Res., 114, D04301, doi:10.1029/2008jd010816, 2009.
519	Foley, K. M., Roselle, S. J., Appel, K. W., Bhave, P. V., Pleim, J. E., Otte, T. L., Mathur, R.,
520	Sarwar, G., Young, J. O., Gilliam, R. C., Nolte, C. G., Kelly, J. T., Gilliland, A. B., and
521	Bash, J. O.: Incremental testing of the Community Multiscale Air Quality (CMAQ)
522	modeling system version 4.7, Geosci. Model Dev., 3, 205–226, doi:10.5194/gmd-3-205-
523	2010, 2010.
524	Follette-Cook, M. B., Pickering, K. E., Crawford, J. H., Duncan, B. N., Loughner, C. P., Diskin,
525	G. S., Fried, A., and Weinheimer, A. J.: Spatial and temporal variability of trance gas
526	columns derived from WRF/Chem regional model output: Planning for geostationary
527	observations of atmospheric composition, Atmos. Environ., 118, 28-44,
528	doi:10.1016/j.atmosenv.2015.07.024, 2015.
529	Huntrieser, H., Schlager, H., Lichtenstern, M., Stock, P., Hamburger, T., Holler, H., Schmidt, K.,
530	Betz, H. D., Ulanovsky, A., and Ravegnani, F.: Mesoscale convective systems observed
531	during AMMA and their impact on the NOx and O3 budget overWest Africa. Atmos
532	Chem Phys., 11(6):2503–2536. doi:10.5194/acp-11-2503-2011, 2011.
533	Kang, D., Eder, B. K., Stein, A. F., Grell, G. A., Peckham, S. E., and Mchenry, J.: The New
534	England air quality forecasting pilot program: development of an evaluation protocol and
535	performance benchmark, J. Air & Waste Manage. Assoc., 55, 1782-1796, 2005.
536	Kang, D., and Foley, K.: Simulating Lightning NOX Production in CMAQv5.2: Performance
537	Evauations, data set, <u>https://doi.org/10.5281/zenodo.2621096</u> , 2019.
538	Kang, D., Heath, N., Foley, K., Bash, J., Roselle, S., and Mathur, R.: On the relationship
539	between observed NLDN lightning strikes and modeled convective precipitation rates:
540	parameterization of lightning NOx production in CMAQ, Air Pollution Modeling and its
541	Application XXV, Chapter 65, ISBN 978-3-319-57644-2, doi: 10.1007/978-3-319-
542	57645-9, 2018.
543 544 545 546 547 548	 Kang, D., Heath, N., Wong, D., Pleim, J., Roselle, S. J., Foley, K. M., and Mathur, R.: Lightning NO_X Production in CMAQ: Part I – Using hourly NLDN Lightning Strike Data, Presented at 15th Annual CMAS Models-3 Users' Conference, 24–26 October 2016, UNC-Chapel Hill, available at: https://www.cmascenter.org/conference/2016/slides/kang_lightning_nox_2016.pptx, 2016.
549	Kang, D., Pickering, K. E., Allen, D. J., Foley, K. M., Wong, D., Mathur, R., and Roselle, S. J.:
550	Simulating Lightning NOx Production in CMAQ: Evolution of Scientific Updates,
551	Geosci. Model Dev. Disc., doi:10.5194/gmd-2019-33, 2019.





552	Kang, D. and Pickering, K. E.: Lightning NOX emissions and the Implications for Surface Air
553	Quality over the Contiguous United States, EM, A&WMA, November, 2018.
554	Kaynak, B., Hu, Y., Martin, R. V., Russell, A. G., Choi, Y., and Wang, Y.: The effect of
555	lightning NOx production on surface ozone in the continental United States. Atmos Chem
556	Phys. 8(17):5151–5159. doi:10.5194/acp-8-5151-2008, 2008.
557	 Koo, B., Chien, C. J., Tonnesen, G., Morris, R., Johnson, J., Sakulyanontvittaya T.,
558	Piyachaturawat, P., and Yarwood, G.: Natural emissions for regional modeling of
559	background ozone and particulate matter and impacts on emissions control strategies.
560	Atmos Environ.,44(19):2372–2382. doi:10.1016/j.atmosenv.2010.02.041, 2010.
561	Koshak, W., Peterson, H., Biazar, A., Khan, M., and Wang, L.: The NASA Lightning Nitrogen
562	Oxides Model (LNOM): Application to air quality modeling, Atmos. Res.,
563	doi:10.1016/j.atmosres.2012.12.015, 2014.
564 565 566	Labrador, L. J., von Kuhlmann, R., and Lawrence, M. G.: The effects of lightning-produced NOx and its vertical distribution on atmospheric chemistry: sensitivity simulations with MATCHMPIC, Atmos. Chem. Phys., 5, 1815–1834, 2005,
567	Lin, J., Youn, D., Liang, X., and Wuebbles, D.: Global model simulation of summertime U.S.
568	ozone diurnal cycle and its sensitivity to PBL mixing, spatial resolution, and emissions,
569	Atmos. Environ., 42, 8470–8483, doi:10.1016/j.atmosenv.2008.08.012, 2008.
570 571	Murray, L. T.: Lightning NO _x and Impacts on Air Quality, <i>Curr Pollution Rep.</i> , doi: 10.1007/s40726-016-0031-7, 2016.
572	Nolte, C. G., Appel, K. W., Kelly, J. T., Bhave, P. V., Fahey, K. M., Collett Jr., J. L., Zhang, L.,
573	and Young, J. O.: Evaluation of the Community Multiscale Air Quality (CMAQ) model
574	v5.0 against size-resolved measurements of inorganic particle composition across sites in
575	North America, Geosci. Model Dev., 8, 2877–2892, doi:10.5194/gmd-8-2877-2015,
576	2015.
577	Napelenok, S. L., Pinder, R. W., Gilliland, A. B., and Martin, R. V.: A method for evaluating
578	spatially-resolved NOx emissions using Kalman filter inversion, direct sensitivities, and
579	spacebased NO ₂ observations, Atmos. Chem. Phys., 8, 5603–5614, doi:10.5194/acp-8-
580	5603-2008, 2008.
581 582	Novak, J. H. and Pierce, T. E.: Natural emissions of oxidant precursors, Water Air Soil Poll., 67, 57-77, 1993.
583	Orville, R. E., Huffines, G. R., Burrows, W. R., Holle, R. L., and Cummins, K. L.: The North
584	American Lightning Detection Network (NALDN) – first results: 1998-2000, Mon. Wea.
585	Rev., 130, 2098–2109, 2002.
586 587 588	Otte, T. L. and Pleim, J. E.: The Meteorology-Chemistry Interface Processor (MCIP) for the CMAQ modeling system: updates through MCIPv3.4.1, Geosci. Model Dev., 3, 243–256, doi:10.5194/gmd-3-243-2010, 2010.





589	Pickering, K. E., Bucsela, E., Allen, D., Ring, A., Holzworth, R., and Krotkov, N.: Estimates of
590	lightning NOx production based on OMI NO2 observations over the Gulf of Mexico, J.
591	Geophys. Res. Atmos., 121, 8668–8691, doi:10.1002/2015JD024179, 2016.
592	Price, C., Penner, J., and Prather, M.: NOx from lightning. 2. Constraints from the global
593	atmospheric electric circuit, J. Geophys. Res., 102, 5943–5951, doi:10.1029/96JD02551, 1997.
594 595	Price, C. and Rind, D.: A simple lightning parameterization for calculating global lightning distributions, J. Geophys. Res., 97, 9919–9933, doi:10.1029/92JD00719, 1992.
596	Richter, A., Burrows, J. P., N [°] uß, H., Granier, C., and Niemeier, U.: Increase in tropospheric
597	nitrogen dioxide over China observed from space, Nature, 437, 129–132,
598	doi:10.1038/nature04092, 2005.
599	Rossow, W. B., Walker, A. W., Beuschel, D. E., and Roiter, M. D.: International Satellite Cloud
600	Climatology Project (ISCCP) documentation of new cloud data sets, Tech. Rep. January,
601	World Meteorological Organisation, WMO/TD 737, Geneva, 1996.
602 603	Schumann, U. and Huntrieser, H.: The global lightning-induced nitrogen oxides source, Atmos. Chem. Phys., 7, 3823-3907, doi:10.5194/acp-7-3823-2007, 2007.
604 605	Sioris, C. E., Kurosu, T. P., Martin, R. V., and Chance, K.: Stratospheric and tropospheric NO2 observed by SCIAMACHY: first results, Adv. Space Res., 34, 780–785, 2004.
606	Stockwell, D. Z., Giannakopoulos, C., Plantevin, P. H., Carver, G. D., Chipperfield, M. P., Law,
607	K. S., Pyle, J. A., Shallcross, D. E., and Wang, K. Y.: Modelling NOx from lightning and
608	its impact on global chemical fields, Atmos. Environ., 33, 4477–4493, 1999.
609	Smith, S. N., and Mueller, S. F.: Modeling natural emissions in the Community Multiscale Air
610	Quality (CMAQ) Model-I: building an emissions data base. Atmos Chem Phys.,
611	10(10):4931–4952. doi:10.5194/acp-10-4931-2010, 2010.
612	Simon, H., Reff, A., Wells, B., Xing, J., and Frank, N.: Ozone trends across the United States
613	over a period of decreasing NO _X and VOC emissions. Environ. Sci. Technol., 49, 186-
614	195, 2015.
658	Wang, L., Newchurch, M. J., Pour-Biazar, A., Kuang, S., Khan, M., Liu, X., Koshak, W., and
659	Chance, K.: Estimating the influence of lightning on upper tropospheric ozone using
660	NLDN lightning data and CMAQ model, Atmos. Environ., 67, 219–228, 2013.
661	Yarwood, G., Whitten, G. Z., Jung, J., Heo, G., and Allen, D. T.: Final Report: Development,
662	Evaluation and Testing of Version 6 of the Carbon Bond Chemical Mechanism (CB6),
663	available at:
664	<u>https://www.tceq.texas.gov/assets/public/implementation/air/am/contracts/reports/pm/582</u>
665	0784005FY1026-20100922-environ-cb6.pdf, 2010.

666





Region	Case	Record	OBS (ppb)	MOD (ppb)	RMSE (ppb)	NME (%)	MB (ppb)	NMB (%)	R
	Base	36242	48.21	52.04	12.6	19.2	3.8	8.0	0.69
Domain	mNLDN	36242	48.21	53.40	12.9	19.8	5.2	10.8	0.70
Domain	hNLDN	36242	48.21	52.21	11.9	18.4	4.0	8.3	0.72
	pNLDN	36242	48.21	52.52	12.7	19.5	4.3	8.9	0.70
	Base	5512	50.97	55.08	13.0	17.8	4.1	8.1	0.74
NE	mNLDN	5512	50.97	55.77	13.4	18.5	4.8	9.4	0.74
INE	hNLDN	5512	50.97	54.23	11.9	16.7	3.3	6.4	0.75
	pNLDN	5512	50.97	55.32	13.1	18.0	4.4	8.5	0.74
	Base	7061	44.55	51.71	12.6	21.0	7.2	16.1	0.76
65	mNLDN	7061	44.55	53.33	13.6	236	8.8	19.7	0.76
SE	hNLDN	7061	44.55	52.30	12.6	21.7	7.8	17.4	0.77
	pNLDN	7061	44.55	52.39	13.0	22.0	7.8	17.6	0.76
	Base	8072	51.60	58.99	13.6	18.8	7.4	14.3	0.64
	mNLDN	8072	51.60	60.14	14.4	20.5	8.5	16.6	0.64
UM	hNLDN	8072	51.60	58.35	12.8	18.0	6.8	13.1	0.64
	pNLDN	8072	51.60	59.42	13.9	19.4	7.8	15.1	0.64
	Base	3609	42.15	46.21	12.4	21.5	4.1	9.6	0.73
	mNLDN	3609	42.15	47.93	12.9	22.3	5.8	13.7	0.74
LM	hNLDN	3609	42.15	47.12	12.3	21.3	5.0	11.8	0.76
	pNLDN	3609	42.15	46.93	12.6	21.8	4.8	11.3	0.74
	Base	6256	52.52	48.13	11.3	17.0	-4.4	-8.4	0.52
514	mNLDN	6256	52.52	50.93	10.2	14.7	-1.6	-3.0	0.56
RM	hNLDN	6256	52.52	50.35	9.9	14.4	-2.2	-4.1	0.57
	pNLDN	6256	52.52	48.93	10.9	16.2	-3.6	-6.9	0.53
	Base	5570	44.72	47.58	11.7	20.1	2.9	6.4	0.80
	mNLDN	5570	44.72	47.73	11.6	20.0	3.0	6.7	0.80
PC	hNLDN	5570	44.72	46.65	11.3	19.5	1.9	4.3	0.81
	pNLDN	5570	44.72	47.62	11.6	20.0	2.9	6.5	0.80

Table 1. Statistics of DM8HR O_3 for all model cases over the domain and analysis regions in July 2011. The best performance metrics among the model cases are highlighted in bold.





Region	Case	Record	OBS (ppb)	MOD (ppb)	RMSE (ppb)	NME (%)	MB (ppb)	NMB (%)	R
Domain	Base	6912	7.58	8.88	8.7	62.6	1.3	17.1	0.54
	mNLDN	6912	7.58	8.87	8.7	62.5	1.3	17.1	0.54
	hNLDN	6912	7.58	8.92	8.7	62.7	1.3	17.7	0.55
	pNLDN	6912	7.58	8.87	8.7	62.5	1.3	17.1	0.54
	Base	989	10.48	9.72	7.0	46.0	-0.8	-7.3	0.55
	mNLDN	989	10.48	9.71	7.0	46.0	-0.8	-7.3	0.55
NE	hNLDN	989	10.48	9.77	7.1	46.1	-0.7	-6.8	0.55
	pNLDN	989	10.48	9.72	7.0	46.0	-0.8	-7.3	0.55
	Base	645	6.44	9.18	7.2	75.3	2.7	42.6	0.34
	mNLDN	645	6.44	9.17	7.2	75.1	2.7	42.4	0.34
SE	hNLDN	645	6.44	9.18	7.2	75.3	2.7	42.6	0.34
	pNLDN	645	6.44	9.17	7.2	75.2	2.7	42.5	0.34
	Base	542	11.42	18.09	18.7	82.7	6.7	58.4	0.58
	mNLDN	542	11.42	18.10	18.7	82.8	6.7	58.5	0.58
UM	hNLDN	542	11.42	18.22	18.9	83.6	6.8	59.5	0.58
	pNLDN	542	11.42	18.09	18.7	82.7	6.7	58.4	0.58
	Base	1240	6.11	8.32	6.0	61.2	2.2	36.1	0.68
	mNLDN	1240	6.11	8.30	6.0	61.1	2.2	35.9	0.68
LM	hNLDN	1240	6.11	8.33	6.0	61.3	2.2	36.3	0.68
	pNLDN	1240	6.11	8.31	6.0	61.2	2.2	36.0	0.68
	Base	1370	3.90	4.00	3.7	60.0	0.1	2.4	0.58
	mNLDN	1370	3.90	4.01	3.7	59.9	0.1	2.6	0.58
RM	hNLDN	1370	3.90	4.02	3.7	60.0	0.1	3.3	0.58
	pNLDN	1370	3.90	4.00	3.7	60.0	0.1	2.4	0.58
	Base	2056	8.61	9.52	9.1	62.8	0.9	10.6	0.48
	mNLDN	2056	8.61	9.52	9.1	62.8	0.9	10.6	0.48
PC	hNLDN	2056	8.61	9.59	9.1	62.9	1.0	11.4	0.48
	pNLDN	2056	8.61	9.52	9.1	62.8	0.9	10.6	0.48

Table 2. Statistics of daily mean NO_x for all model cases over the domain and analysis regions July 2011. The best performance metrics among the model cases are highlighted in bold.





Table 3. Statistics of June-August 2011 accumulated precipitation (cm) and wet deposition of nitrate (NO_3) for all model cases over the domain. The best performance metrics among the model cases are highlighted in bold.

Region	Case	Record	OBS (cm, kg/ha)	MOD (cm, kg/ha)	RMSE (cm, kg/ha)	NME (%)	MB (cm, kg/ha)	NMB (%)	R
Domain	precip	196	24.8	23.9	7.5	23	-0.9	-4	0.87
	Base	196	2.34	1.52	1.1	38	-0.8	-35	0.84
	mNLDN	196	2.34	1.98	0.8	26	-0.4	-15	0.86
	hNLDN	196	2.34	1.95	0.8	26	-0.4	-17	0.86
	pNLDN	196	2.34	1.68	1.0	33	-0.7	-28	0.85
	precip	31	38.6	35.9	9.5	19	-2.7	-7	0.79
	Base	31	2.96	2.32	1.1	29	-0.6	-23	0.70
NE	mNLDN	31	2.96	2.71	0.9	24	-0.3	-8	0.76
	hNLDN	31	2.96	2.74	0.9	24	-0.2	-7	0.74
	pNLDN	31	2.96	2.48	1.0	27	-0.5	-16	0.73
	precip	39	36.1	31.7	9.4	21	-4.3	-12	0.80
	Base	39	3.05	2.09	1.2	35	-1.0	-32	0.51
SE	mNLDN	39	3.05	2.97	0.8	21	-0.1	-2	0.56
	hNLDN	39	3.05	2.82	0.9	23	-0.2	-8	0.53
	pNLDN	39	3.05	2.43	1.0	27	-0.6	-20	0.54
	precip	45	28.8	26.1	6.8	20	-2.7	-9	0.51
	Base	45	3.17	1.98	1.4	38	-1.2	-38	0.73
UM	mNLDN	45	3.17	2.51	0.9	24	-0.7	-21	0.77
	hNLDN	45	3.17	2.48	0.9	25	-0.7	-22	0.77
	pNLDN	45	3.17	2.15	1.2	33	-1.0	-32	0.76
	precip	12	12.3	10.4	4.1	29	-2.0	-16	0.90
	Base	12	1.44	0.85	0.7	41	-0.6	-41	0.90
LM	mNLDN	12	1.44	1.16	0.6	33	-0.3	-19	0.88
	hNLDN	12	1.44	1.13	0.6	32	-0.3	-21	0.89
	pNLDN	12	1.44	0.93	0.7	36	-0.5	-35	0.88
	precip	50	13.7	18.2	6.9	39	4.4	32	0.91
	Base	50	1.63	0.8	1.0	51	-0.8	-51	0.90
RM	mNLDN	50	1.63	1.1	0.7	34	-0.5	-32	0.91
	hNLDN	50	1.63	1.12	0.7	33	-0.5	-31	0.90
	pNLDN	50	1.63	0.86	1.0	48	-0.8	-47	0.91
	precip	19	7.01	6.53	2.4	29	-0.48	-6.8	0.84
	Base	19	0.31	0.31	0.18	44	0.00	-1.0	0.88
PC	mNLDN	19	0.31	0.33	0.19	48	0.01	3.9	0.89
	hNLDN	19	0.31	0.33	0.20	50	0.02	6.6	0.89
	pNLDN	19	0.31	0.31	0.18	44	0.00	-0.3	0.88





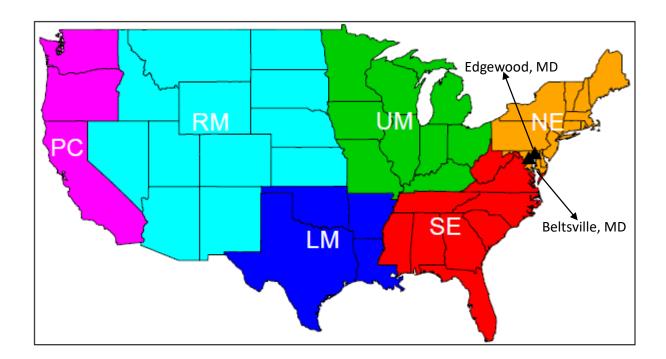


Figure 1. Analysis regions and ozonesonde locations during the 2011 DISCOVER-AQ field study.





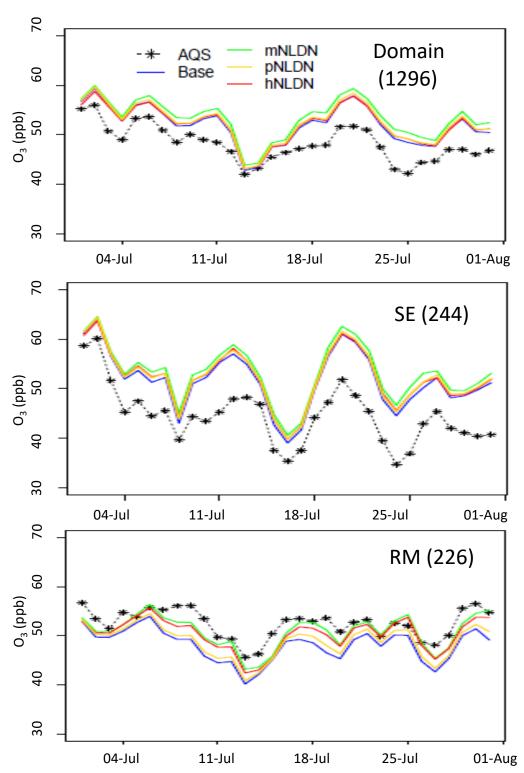


Figure 2. Timeseries of regional-mean daily maximum 8-hr O_3 comparing observations (AQS) and CMAQ model predictions using the LNO_x schemes to Base simulation for the domain (a), SE (b), and RM (c) in July, 2011. The numbers in the parentheses following the region names are the number of AQS sites.





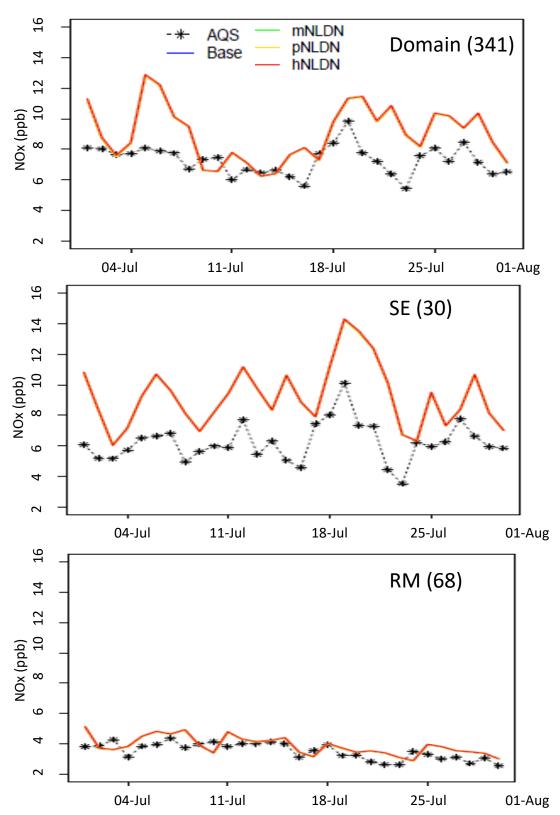


Figure 3. Timeseries of daily mean NO_x over the domain (a), SE (b), and RM (c) in July, 2011. The numbers in the parentheses following the region names are the number of AQS sites.





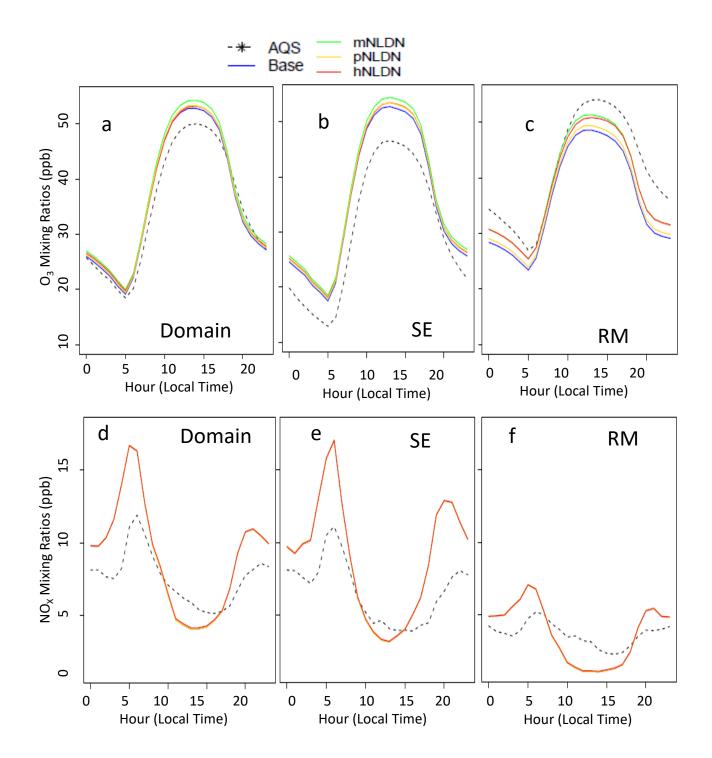


Figure 4. Diurnal profiles for hourly O_3 and NO_x over the domain (a,d), SE (b,e), and RM (c,f) in July, 2011.





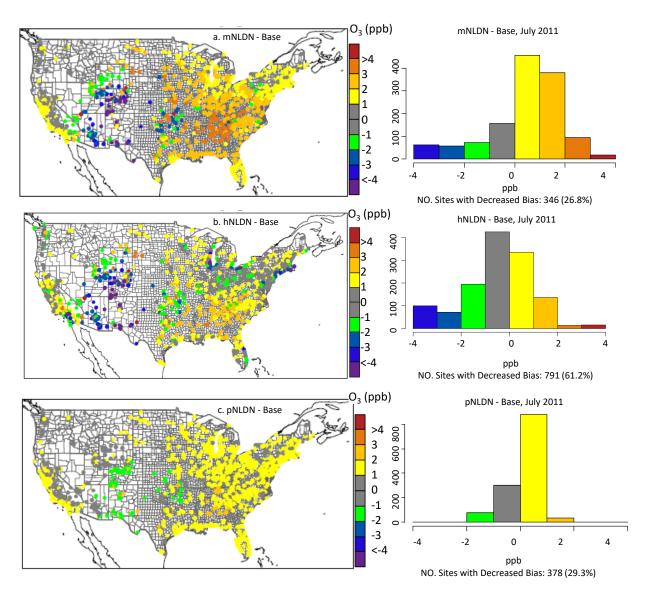


Figure 5. Spatial maps of the mean bias of DM8HR O_3 (model – observation) differences between model case with LNO_x and the Base as well as the corresponding histograms indicating the number of sites with decreased mean bias for each pair of model cases in July, 2011.





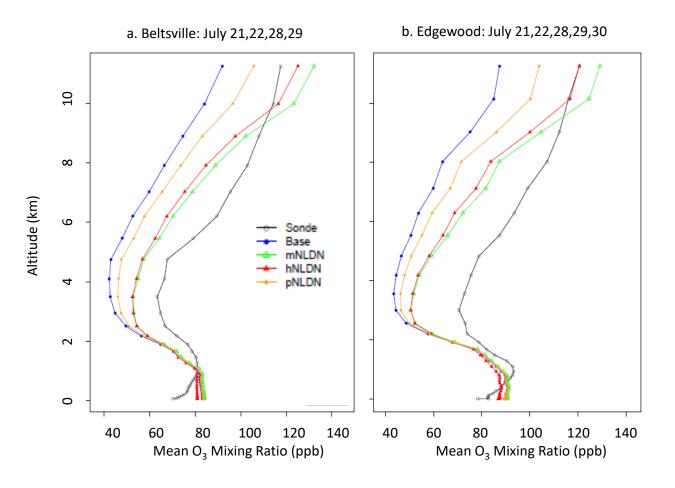


Figure 6. Vertical profiles of O_3 mixing ratios from ozonesonde measurements and model simulations at Beltsville, MD (a) and Edgewood, MD (b) on the days when lightning NO produced significant impact on O_3 during the Discover-AQ field study in July, 2011.





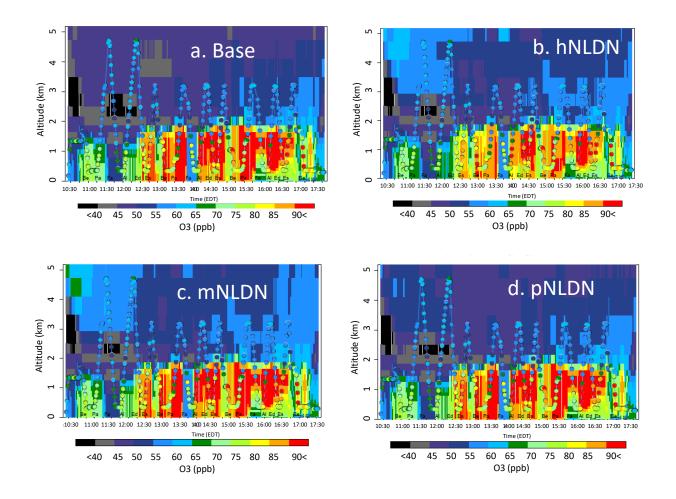


Figure 7. Overlay of P3B observed O_3 (1 minute mean values) over the corresponding vertical cross sections of simulated values extracted at the flying locations on July 28, 2018, (a) Base, (b) hNLDN,(c) mNLDN, and (d) pNLDN. The letters marked at the bottom of the plots are P3B spiral sites, Be: Beltsville, Pa: Padonia, Fa: Fairhill, Al: Aldino, Ed: Edgewood, Es: Essex.





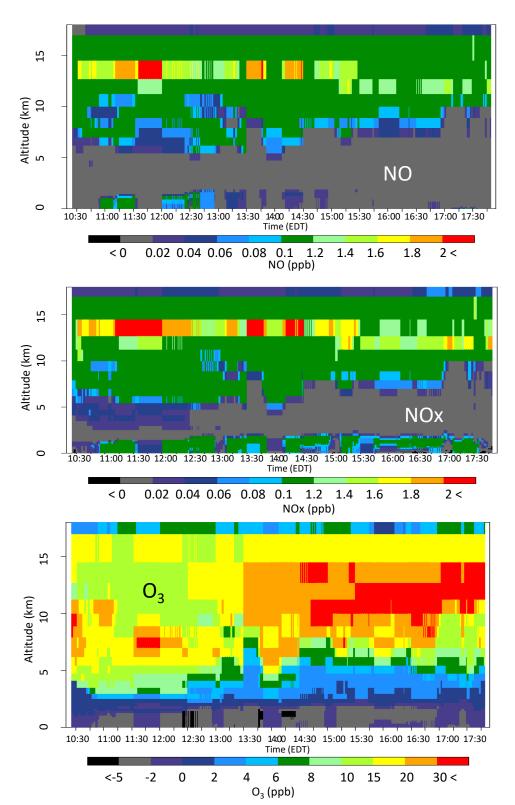


Figure 8. The vertical-time difference between hNLDN and Base during the P3B flight period on July 28, 2011 for (a) NO, (b) NO_x , and (c) O_3 .





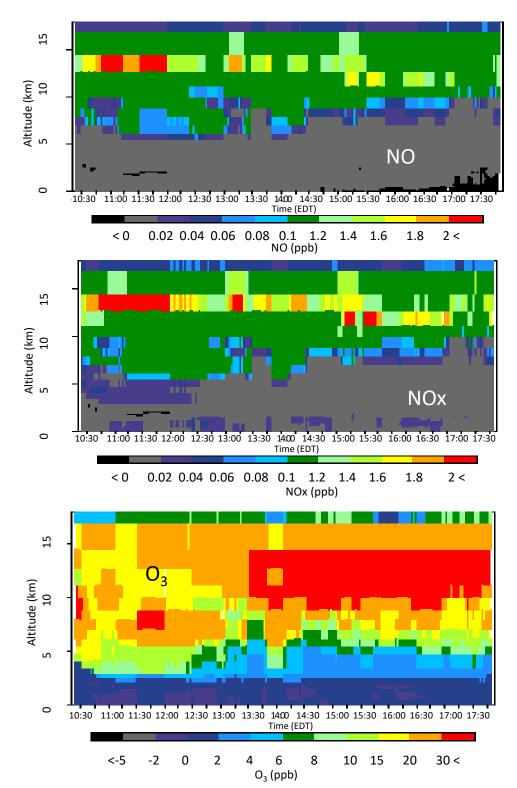


Figure 9. The vertical-time difference between mNLDN and Base during the P3B flight period on July 28, 2011 for (a) NO, (b) NO_x , and (c) O_3 .





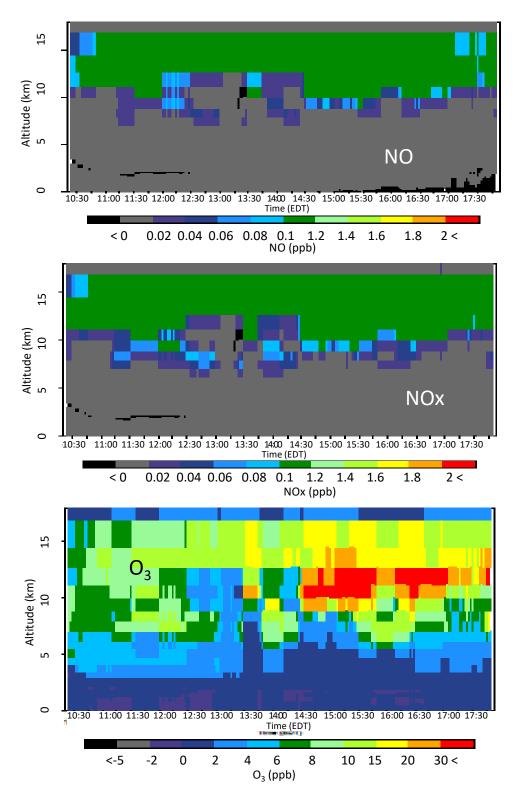


Figure 10. The vertical-time difference between pNLDN and Base during the P3B flight period on July 28, 2011 for (a) NO, (b) NO_x , and (c) O_3 .





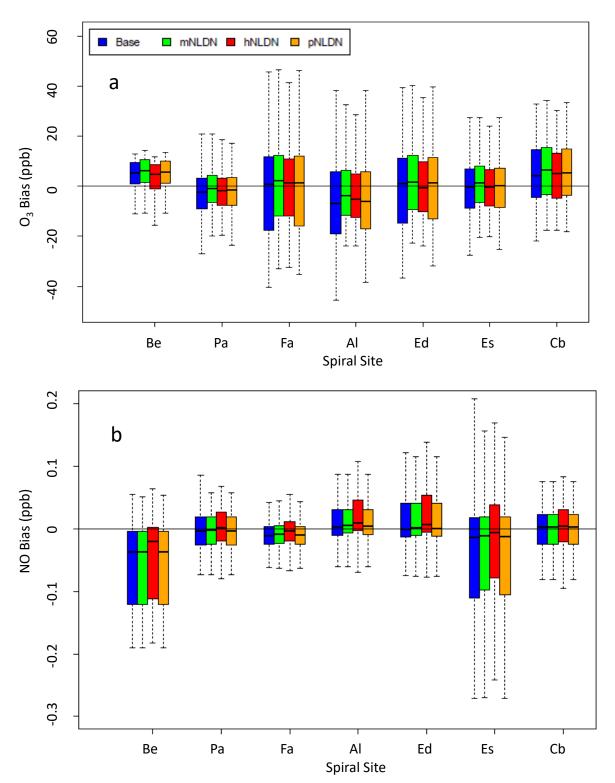


Figure 11. Bias (model – observation) distributions of O_3 (a) and NO (b) at each P3B spiral site on July 21, 22, 28, and 29, 2011. Be: Beltsville, Pa: Padonia, Fa: Fairhill, Al: Aldino, Ed: Edgewood, Es: Essex, Cb: Chesapeake Bay.





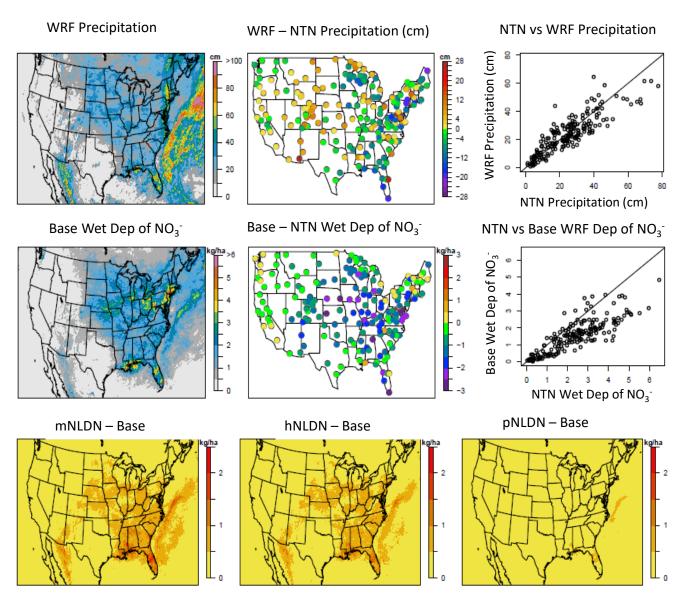


Figure 12. The top row shows precipitation estimates from WRF (left), the bias in the WRF predicted precipitation at NTN locations (middle), and the corresponding scatter plots (right). The middle row shows wet deposition (Dep) of nitrate estimates from the Base simulation (left), the bias in the Base model estimates of wet deposition of NO_3^- at NADP/NTN locations (middle), and the corresponding scatter plots (right). The bottom row shows the difference in the LNO_x sensitivity simulations and the Base case estimates of wet deposition of NO_3^- : mNLDN – Base (left); hNLDN – Base (middle), and pNLDN – Base (right). All maps are based on accumulated values (precipitation or wet deposition) during June – August 2011. Precipitation totals are in cm and wet deposition totals are in kg/ha.