1	The interactive global fire module pyrE (v1.0)
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32 Abstract. Fires affect the composition of the atmosphere and Earth's radiation balance 33 by emitting a suite of reactive gases and particles. An interactive fire module in an Earth 34 System Model (ESM) allows us to study the natural and anthropogenic drivers, feedbacks, 35 and interactions of open fires. To do so, we have developed pyrE, the NASA GISS 36 interactive fire emissions module. The pyrE module is driven by environmental variables 37 like flammability and cloud-to-ground lightning, calculated by the GISS ModelE ESM, 38 and parameterized anthropogenic impacts based on population density data. Fire 39 emissions are generated from the flaming phase in pyrE (active fires). Using pyrE, we 40 examine fire occurrence, regional fire suppression, burned area, fire emissions, and how 41 it all affects atmospheric composition. To do so, we evaluate pyrE by comparing it to 42 satellite-based datasets of fire count, burned area, fire emissions, and aerosol optical 43 depth (AOD). We demonstrate pyrE's ability to simulate the daily and seasonal cycles of 44 open fires and resulting emissions. Our results indicate that interactive fire emissions are 45 biased low by 32-42%, depending on emitted species, compared to the GFED4s 46 inventory. The bias in emissions drives underestimation in column densities, which is 47 diluted by natural and anthropogenic emissions sources and production and loss 48 mechanisms. Regionally, the resulting AOD of a simulation with interactive fire 49 emissions is underestimated mostly over Indonesia compared to a simulation with 50 GFED4s emissions and to MODIS AOD. In other parts of the world pyrE's performance 51 in terms of AOD is marginal to a simulation with prescribed fire emissions.

# 52 1 Introduction

53 Open biomass burning (BB), the outdoor combustion of organic material in the 54 form of vegetation, occurs on every continent, with the exception of Antarctica, at a scale 55 observable from space. Open BB is perceived as a natural ecological process that has 56 been modulating the carbon cycle for more than 420 million years [Scott and Glasspool, 57 2006]. However, in practice, BB has been mediated by human activities for more than 58 100,000 years [Bowman et al., 2009, 2011; Archibald et al., 2012]. Bellouin et al. (2008) 59 estimated that at present, only about 20% of fires, compared to preindustrial times, are 60 natural. Andreae (1991) estimated that in the tropics, where about 85% of fire emissions 61 occur [van der Werf et al., 2017], only 10% of fires are natural. In the USA, government 62 records show that about 85% of fires are started by humans [Balch et al., 2017]. Humans 63 affect fires directly through ignition and suppression, and indirectly through man-made 64 changes to land surfaces and climate. According to Hantson et al. (2015), land-use 65 practices are the most important driver of human-fire interactions.

66 BB regimes are often classified based on ecosystem type like boreal, temperate, and tropical forests, savanna and grassland, peat land, and agricultural fires [Ichoku et al., 67 2012]. However, fire characteristics also vary between geographic regions of the same 68 69 ecosystem type; for example, boreal fires in Russia have very different intensity, 70 efficiency, and emissions than boreal fires in Canada [Wooster and Zhang, 2004]. Ichoku 71 et al. (2008) suggested an energy-based classification of open BB indicating fire intensity, 72 similar to hurricanes, using the radiative power of satellite-retrieved fires. Globally, 73 satellite retrievals show that on average about 350 Mha are burned annually [Giglio et al., 74 2013; Chuvieco et al., 2016], about 4% of the global vegetated area [Randerson et al., 75 2012], an area similar to that of India. African fires contribute about 70% to the global 76 total burned area (BA), with about equal contributions from Northern Hemisphere Africa 77 (NHAF, Fig. 1) and Southern Hemisphere Africa (SHAF). The most flammable 78 ecosystem, globally and specifically in Africa, is the savanna [Ichoku et al., 2008; Randerson et al., 2012; Giglio et al., 2013], which in the tropics (23.5° N - 23.5° S) alone 79 is responsible for 62% (1341 TgC a<sup>-1</sup>) of global carbon emissions (2200 TgC a<sup>-1</sup>) [van der 80 81 Werf et al., 2017]. Australian bushfires (grass and shrub) and South American savanna 82 fires are the third and fourth largest regional contributors, with BAs of about 50 Mha and 83 20 Mha annually, respectively. Globally, *Randerson et al.* (2012) estimated an additional 84 contribution of 120 Mha from small fires. The thermal anomalies used to identify those 85 fires, which are mostly associated with agricultural fires, are below the detection limit of 86 satellite-retrieved surface reflectance, and come with large uncertainties. Regionally, 87 small fires can have a significant contribution to BA. By adding the contribution of small 88 fires, burned area increases in Equatorial Asia (EQAS) by 157%, in Central America 89 (CEAM) by 143%, and in Southeast Asia (SEAS) by 90% [Randerson et al., 2012]. This 90 highlights the regional importance of small agricultural fires to regional fire activity. 91 Forest fires, including small fires, contribute about 17 Mha annually to global BA, and 92 are dominant in Temperate North America (TENA), Boreal North America (BONA), 93 Boreal Asia (BOAS) and EQAS.

94 BB can exist when three conditions are met: fuel is available, fuel is combustible, 95 and ignition sources are present [Schoennagel et al., 2004]. The coincidence of these 96 conditions is seasonal, making open BB an inherently seasonal phenomenon. The peak 97 month and duration of fire season are coupled to the seasonal cycle in precipitation, 98 especially in the tropics [Giglio et al., 2006; Hantson et al., 2017]. Precipitation and fire 99 activity are sensitive to natural modes of variability like El Niño Southern Oscillation 100 (ENSO). In particular, the Southern Hemisphere BB activity is strongly coupled to ENSO 101 [Buchholz et al., 2018]. During an El Niño year regional BB emissions can be up to two 102 times higher than their regional average level, due to increased fire activity in tropical 103 rainforests [van der Werf, 2004; Andela and Werf, 2014; Field et al., 2016; Whitburn et 104 al., 2016].

Forest fires are either ignited on purpose, as part of forest management practices [*Ryan et al.*, 2013], ignited by accident, as a by-product of the expansion of urban life to the wildland interface [*Moritz et al.*, 2014; *Fischer et al.*, 2016; *Radeloff et al.*, 2018], or ignited by lightning [*Díaz-Avalos et al.*, 2001]. Thus, fire activity is highly coupled to trends in population density as increased population density at the wildland-urban interface (WUI) increases the probability of fire [*Radeloff et al.*, 2018], while land abandonment leads to shrub encroachment, and fuels fire activity [*Butsic et al.*, 2015].

112 Although BB emissions have high spatiotemporal variability, their impact on 113 atmospheric composition is significant [Crutzen et al., 1979; Seiler and Crutzen, 1980; 114 Crutzen and Andreae, 1990]. BB emissions impact air quality [Johnston et al., 2012, 115 2014, 2016; Bauer et al., 2019], and climate [Ward et al., 2012; Lasslop et al., 2019]. Emitted pollutants include ozone precursors like methane (~49 Tg a<sup>-1</sup>), carbon monoxide 116 117 (~820 Tg a<sup>-1</sup>), and NO<sub>x</sub> (mostly emitted as NO, ~19 Tg a<sup>-1</sup>) [Andreae, 2019]; the latter 118 two are also deleterious for health on their own. In addition to gaseous pollutants, BB 119 emits particulate matter (a total of  $\sim$ 85 Tg a<sup>-1</sup>) like primary emitted black carbon ( $\sim$ 5 Tg a<sup>-1</sup>) and organic carbon (~36 Tg a<sup>-1</sup>), as well as precursors of brown carbon, and 120 121 secondary organic and inorganic aerosols like non-methane volatile organic compounds (NMVOC, ~58 Tg  $a^{-1}$ ), ammonia (~9.9 Tg  $a^{-1}$ ), sulfur dioxide (~6 Tg  $a^{-1}$ ), and NO<sub>x</sub> 122 123 [Andreae, 2019]. Exposure to these pollutants at high concentrations or for a long period 124 of time can compromise the cardiorespiratory system and lead to death [Lelieveld et al.,

125 2015]. These pollutants, along with BB-emitted greenhouse gases (GHGs) like carbon dioxide (CO<sub>2</sub>; ~13,900 Tg a<sup>-1</sup>) and nitrous oxide (N<sub>2</sub>O; ~1.38 Tg a<sup>-1</sup>), interact with 126 127 radiation, directly and indirectly. Fires are a net source of carbon dioxide only where 128 vegetation regrowth is inhibited, i.e. in deforested areas; otherwise BB is not viewed as a 129 source of CO<sub>2</sub> but as "fast respiration" [van der Werf et al., 2017]. Absorbing black and 130 brown carbon [Lack et al., 2012; Lack and Langridge, 2013; Laskin et al., 2015], and 131 reflecting primary and secondary organic and inorganic aerosols interact with solar 132 radiation directly by scattering and absorbing radiation, and indirectly by modifying 133 clouds. The radiative properties of particles and their hygroscopicity are also influenced 134 by their mixing state [Bauer and Menon, 2012]. For example, when black carbon (BC) is 135 coated it becomes even more absorbing per unit mass [Bond and Bergstrom, 2006]. There 136 is evidence that smoke plumes can suppress or invigorate precipitation [Feingold et al., 137 2001; Andreae et al., 2004; Tosca et al., 2015]. Aerosols impact cloud height and cover 138 by modifying the heat profile of the atmosphere and increasing the number of cloud 139 condensation nuclei. There are large uncertainties associated with aerosols' impact on 140 climate. Modeling studies suggest that the aerosol effects from BB emissions overrides 141 the BB-GHG effect to a net negative radiative forcing [Mao et al., 2013], with the 142 indirect effect of clouds dominating the forcing [Ward et al., 2012]. The present day BB forcing is estimated at -0.5-(-0.1)±0.05 Wm<sup>-2</sup> [Ward et al., 2012; Mao et al., 2013; Jiang 143 144 et al., 2016; Landry and Matthews, 2016; Lasslop et al., 2019].

145 The quantification of speciated BB emissions is challenging due to the fact that no 146 one fire is the same as another [Ito and Penner, 2005]. The composition of the resulting 147 smoke plume depends on the fuel type, burning conditions (i.e. flaming or smoldering), 148 fuel consumption, and on background chemistry. More complete combustion has a higher 149 fraction of oxidized species (e.g.  $CO_2$  and  $NO_x$ ) while smoldering fires release more 150 reduced species (e.g. CO, NH<sub>3</sub>, NMVOCs). Globally, most fire emissions occur during 151 the active phase of the fire, with peat fires as the main exception [Andreae, 2019]. Thus, 152 emissions in different regions contribute different amounts of pollutants; Indonesia, for 153 example, is responsible for 8% of global carbon BB emissions, but 23% of methane BB 154 emissions [van der Werf et al., 2017]. Emissions are sensitive to season and region. Even 155 within one region, like a boreal forest, emissions from crown fires differ from those from

156 ground fires. The amount of fuel consumed by a fire is highly variable and depends on 157 fuel load, density, moisture, vegetation type, and on environmental factors such as wind 158 speed, soil moisture and soil composition. Additional challenges relate to external forcing 159 like insect herbivority, mammal grazing, and manmade land fragmentation and 160 deforestation [Schultz et al., 2008]. The quantification of BB emissions has an even 161 bigger importance during preindustrial times, where fire emission are identified as the 162 largest source of uncertainty for aerosol loading in Earth system models [Hamilton et al., 2018]. BB emissions are a key quantity needed for quantifying the unperturbed-from-163 164 humans background conditions of the atmosphere [Carslaw et al., 2013].

165 Traditionally, fires are included in climate models using emission inventories [Lamarque et al., 2010; van der Werf et al., 2010, 2017; van Marle et al., 2017]. Some 166 167 models have the ability to simulate BB emissions interactively with a varying level of 168 complexity [Thonicke et al., 2001; Arora and Boer, 2005; Pechony and Shindell, 2009; Li 169 et al., 2012; Lasslop et al., 2014; Hantson et al., 2016; Mangeon et al., 2016; Rabin et al., 170 2017; Zou et al., 2019]. On the one end of the spectrum, there are statistically-based 171 models, and on the other end there are detailed empirical and physical process-based 172 models. Statistical models are skilled at making predictions based on present-day 173 relationships between climate and fire (their training data). Process-based models 174 encapsulate the complex feedbacks within the climate system at various levels. They 175 combine physical processes such as fuel condition, cloud-to-ground lightning ignitions, 176 and wind-driven fire expansion. The most sophisticated models are coupled to dynamic 177 global vegetation models and directly connect fire-Earth system interactions through fuel 178 consumption (e.g. LPJ-GUESS-GlobFIRM, LPJ-GUESS-SIMFIRE-BLAZE (Smith et al., 179 2001, 2014; Lindeskog et al., 2013), and MC-Fire (Bachelet et al., 2015; Sheehan et al., 180 2015)). Some models also include simplified empirical relationships of anthropogenic 181 ignition and suppression, which, at present, are not understood in a dynamic process level. 182 State-of-the-art process-based fire models are well equipped to study the feedbacks 183 between the climate system and fires [Hantson et al., 2016]. However, there is indication 184 that they lack accurate predictive capabilities, as they only partly capture trends in present 185 day observations. For example, satellite products show a global decrease in burned area from about 500 Mha a<sup>-1</sup> in 1997 to 400 Mha a<sup>-1</sup> in 2013, a trend which fire models do not 186

187 capture [Andela et al., 2017]. This trend is mostly driven by land fragmentation and 188 grazing practices over African savanna, highlighting the challenge of fire models to 189 account for the combined changes in climate, vegetation and socio-economic drivers 190 [Forkel et al., 2019]. Though less accurate than observational datasets, when trying to 191 simulate individual fire events, fire models provide the unique advantage of linking the 192 atmosphere, biosphere and hydrosphere in a consistent way, a crucial step when studying 193 Earth System interactions. They are also able to predict fire during climate periods for 194 which we have no observational data available (e.g. preindustrial and future).

195 In this paper we present a new global fire module, pyrE, based on an improved 196 scheme of [Pechony and Shindell, 2009, 2010] with new capabilities. The pyrE module is 197 process-based, as it includes the two basic parameters of fuel availability and 198 combustibility, which are used to calculate active fires. It utilizes empirical relationships 199 with population density to account for the anthropogenic impact on fire ignition and 200 suppression. However, unlike most fire models where fire suppression is applied 201 uniformly across all regions [Rabin et al., 2017], in pyrE fire suppression depends both 202 on population density and region. Additionally, pyrE uses active fires to derive emissions 203 in contrast to other fire models that use BA. The fire module is part of the NASA GISS 204 ModelE Earth System model, ModelE2.1 (an updated version based on Schmidt et al. 205 (2014)), and is described below.

#### 206 2 Model description

207 pyrE, from the Greek word for fire (pyr,  $\pi \nu \rho$ ), is a global fire module within GISS 208 ModelE. It incorporates the active fire parameterization of Pechony and Shindell (2009, 209 2010), with the addition of fire spread and BA, following the Community Land Model's 210 (CLM) approach [Li et al., 2012]. The module is a collection of physical processes like 211 flammability, natural ignition, fire spread, and fire emissions, and empirical processes 212 that include accidental ignition and suppression (Fig. 2). The climate model input 213 required, includes surface temperature, surface relative humidity (RH), precipitation, 214 surface wind speed, vegetation density and type, cloud-to-ground lightning frequency and population density. Like many fire modules it lacks explicit intentional ignition (e.g. crop, 215 216 deforestation) and peat fires.

217 2.1 Flammability

- Flammability is a parameter that indicates conditions favorable for fire occurrence [*Pechony and Shindell*, 2009, 2010]. It is a unit-less number that ranges between zero and one, and is calculated using vapor pressure deficit (*VPD*), monthly-accumulated precipitation, and vegetation density (*VD*).
- 222 *VPD*, an indicator of drought [*Seager et al.*, 2015; *Williams et al.*, 2015], is 223 calculated via the Goff-Gratch equation [*Goff and Gratch*, 1946; *Goff*, 1957] using the 224 saturation vapor pressure ( $e_s$ ) and surface relative humidity (*RH*):

225 
$$VPD = e_s \left(1 - \frac{RH}{100}\right) (1)$$

226 Where  $e_{st} = 1013.245 \ [mb]$  is the saturation vapor pressure at the boiling point 227 of water and  $e_s = e_{st} 10^{Z(T)}$  depends on temperature (*T*):

228 
$$Z(T) = a\left(\frac{T_s}{T} - 1\right) + b \cdot \log\left(\frac{T_s}{T}\right) + c\left(10^{d\left(1 - \frac{T_s}{T}\right)} - 1\right) + f\left(10^{h\left(\frac{T_s}{T} - 1\right)} - 1\right)(2)$$

229 With the coefficients:  $a = -7.90298; b = 5.02808; c = -1.3816 \cdot 10^{-7}; d =$ 230 11.344;  $f = 8.1328 \cdot 10^{-3}; h = -3.49149$  [*Goff and Gratch*, 1946], and  $T_s =$ 231 373.16 [°K] (water boiling point temperature).

The precipitation dependence of flammability is in the form of an inverse exponential (Following [*Keetch and Byram*, 1968]):

234

$$f(R) = \exp(-c_R R) (3)$$

Where *R* is the surface rain rate in mm per day and  $c_R = 2 [day/mm]$  is an empirical constant [*Pechony and Shindell*, 2009].

Vegetation density (*VD*) is taken as the normalized leaf area index (LAI) in the land fraction of a grid cell, varying between 0 for no vegetation and 1 for dense vegetation.

We modified the original calculation proposed by [*Pechony and Shindell*, 2009] by calculating flammability only for the fraction of the model's grid cell that is not burned from previous fires. The flammability *F* at a time step *t* in a grid cell (*i*, *j*) is:

243 
$$F(t) = 10^{Z(T(t)_{i,j})} \left(1 - \frac{RH(t)_{i,j}}{100}\right) VD(t)_{i,j} \left(1 - \frac{BA(t)_{i,j}}{LA_{i,j}}\right) \exp\left(-c_R R(t)_{i,j}\right) (4)$$

244 Where  $LA_{i,j}$  is the total land area (LA) in the grid cell (i, j).

245 **2.2 Ignition** 

246 Natural and anthropogenic ignition varies in space and time, and is necessary for 247 the calculation of active fires. If ignition is zero, the resulting number of active fires will 248 be zero, independent of flammability. Natural ignition is in the form of cloud-to-ground 249 lightning frequency, which is interactively calculated in ModelE2.1 [Price and Rind, 250 1992, 1993]. The parameterization of anthropogenic ignition follows Venevsky et al. 251 (2002) and is based on the assumption that in sparsely populated regions people interact 252 more with the natural environment, thus increasing the potential for ignition. The 253 parameterization uses population density data and empirical scaling factors, as described 254 by Pechony and Shindell (2009), and does not include intentional ignition. The number of anthropogenic accidental ignitions per km<sup>2</sup> per month is: 255

256

 $I_A = k(PD)PD\alpha (5)$ 

257 Where PD is the population density;  $k(PD) = 6.8PD^{-0.6}$  represents the varying 258 anthropogenic ignition potentials as a function of population density;  $\alpha = 0.03$  is the 259 number of potential ignitions per person per month. Coefficients are taken following 260 *Pechony and Shindell* (2009) and *Mangeon et al.* (2016) which utilized correlation 261 calculations done by *Venevsky et al.* (2002).

#### 262 **2.3 Suppression**

263 A first-order approximation of the impact of population density on explicit fire 264 suppression was proposed by Pechony and Shindell (2009). According to that 265 parameterization, more fires are suppressed in densely populated areas compared to 266 sparsely populated areas, regardless of ignition source. Specifically, suppression varies 267 from 5% to 95% of fires. However, fire management is a region-specific practice, which 268 depends on cultural norms and economic capabilities. For example, fire suppression in 269 the United States of America (USA) is a common practice (Parisien and Moritz, 2009; 270 Marlon et al., 2012) while active fire suppression in most parts of Africa is not 271 commonly practiced. Most fire suppression in Africa is an indirect byproduct of changes 272 in land surface properties through grazing and fragmentation (Archibald, 2016). Hence, 273 we modified the simplistic approach suggested by Pechony and Shindell (2009), guided 274 by the results presented in Sect. 5.1.1 to better match with observed fire activity at 275 specific regions. Our initial analysis showed that with the original *Pechony and Shindell* 276 (2009) suppression scheme fire activity is overestimated in the TENA and MIDE regions 277 while being underestimated in NHAF and SHAF. Following these initial results a series

of sensitivity simulations were conducted with varying values of suppression coefficients.

279 The final values were chosen in a heuristic manner that improved the simulations yet did

280 not over-fit them to the observations, similarly to Pechony and Shindell (2009) and other

281 fire parameterization, due to the lack of appropriate global data.

We use the complement of the fraction of suppressed fires that is the fraction of non-suppressed fires,  $f_{NS}$ :

284 
$$f_{NS} = \begin{cases} 0.2 \exp(-0.05PD), \ USA \ and \ MIDE \\ 1, \ Africa \\ 0.05 + 0.9 \exp(-0.05PD), \ Elsewhere \end{cases}$$
(6)

**285 2.4 Active fires** 

Active fires are a key metric used to drive burned area and fire emissions in pyrE. The number of fires in a time step per km<sup>2</sup> is calculated as the product of flammability, sum of natural and anthropogenic ignition, and suppression [*Pechony and Shindell*, 2009] (Fig. 2):

290

$$N_{fire}(t)_{i,j} = F(t)_{i,j} \cdot \left( I_N(t)_{i,j} + I_A(t)_{i,j} \right) \cdot f_{NS}(t)_{i,j}$$
(7)

**291 2.5 Burned area (BA)** 

We adopted the process-based approach of *Li et al.* (2012) to calculate fire spread and burned area. The burned area in grid cell (i, j) at a model time step *t* is the product of active fires and the weighted average over plant functional types (PFTs) of the area burned by one fire:

296

$$BA_{i,j} = N_{fire}(t)_{i,j} \cdot \sum_{\nu} a_{i,j,\nu} \cdot f_{i,j,\nu}$$
(8)

Where  $f_{i,j,v}$  is the fractional area covered by plant functional type v, and the burned area of a single fire  $a_{i,j,v}$  is assumed to have an elliptical shape (Fig. 3). Wind speed, surface relative humidity, and vegetation type control the eccentricity of the ellipsoid that represents the burned area of a single fire (based on *van Wagner* (1969)):

301 
$$a_{i,j,v} = \frac{\pi ROS^2 \tau^3}{4LB} \left(1 + \frac{1}{HB}\right)^2 (9)$$

Where *ROS* is the rate of fire spread, *LB* is the length-to-breadth ratio, and *HB* is the head-to-breadth ratio. The stronger the wind, the more eccentric the ellipse, i.e. the bigger the length-to-breadth ratio:

305 
$$LB = 1 + 10 \cdot (1 - \exp(-0.06W)) (10)$$

306 Where W is the surface wind speed in m  $s^{-1}$ .

307 Strong winds also increase the head to back ratio; the ratio of the downwind 308 spread compared to the upwind spread:

309 
$$HB = \frac{LB + \sqrt{LB^2 - 1}}{LB - \sqrt{LB^2 - 1}} (11)$$

310 The rate of spread (ROS) of a fire is a function of vegetation type, wind speed, and atmospheric and soil moisture: 311

312 
$$ROS = ROS_{max} \cdot gW \cdot f_{RH} \cdot f_{\theta}$$
 (12)

ROS<sub>max</sub> is the maximum fire spread rate. Following Li et al. (2012), we set it to 313  $0.2 \text{ m s}^{-1}$  for grasses,  $0.17 \text{ m s}^{-1}$  for shrubs,  $0.15 \text{ m s}^{-1}$  for needle leaf trees, and  $0.11 \text{ m s}^{-1}$ 314 for other trees. Li et al. (2012) estimated the fire spread coefficients to be on the lower 315 316 range of observed ROS, but are yet higher than the global value of 0.13 m s<sup>-1</sup> suggested 317 by Arora and Boer (2005).

by:

319 
$$gW = \frac{2L_B}{1 + \frac{1}{H_B}} g0 (13)$$

320 Where 
$$g0 = \frac{1 + HB_{max}^{-1}}{2LB_{max}} \approx 0.05$$

 $f_{RH}$ ,  $f_{\theta}$  are the dependencies of fire spread on RH and root zone soil moisture: 321

322 
$$f_{RH} = \begin{cases} 1 & RH \le RH_{low} \\ \frac{RH_{up} - RH}{RH_{up} - RH_{low}} & RH_{low} < RH < RH_{up} (14) \\ 0 & RH \ge RH_{up} \end{cases}$$

Following *Li et al.* (2012), we set  $RH_{low} = 30\%$ ,  $RH_{up} = 70\%$  and  $f_{\theta} = 0.5$ 323 324 as ModelE2.1 does not simulate prognostic root zone soil moisture.

#### 325 2.6 Emissions

326 Trace gas and aerosol emissions are generated during the active phase of the fire and are calculated as the product of the simulated active fires and emission factors 327  $(EF_{s,v})$  and are a function of PFT (denoted by v) and chemical specie (denoted by s). The 328 use of active fires to derive emissions is driven by the extremely rudimentary 329 representation of the terrestrial biosphere in ModelE, under which interactive fuel 330 331 consumption cannot be calculated. The emissions per grid cell (i, j) of specie s at a 332 model time step *t* are calculated by:

333  $E_{i,j,s}(t) = N_{fire}(t)_{i,j} \cdot \sum_{v} EF_{s,v} \cdot f_{i,j,v}$ (15)

Where  $E_{i,j,s}(t)$  is the emissions flux rate in kg m<sup>-2</sup> s<sup>-1</sup>,  $N_{fire}(t)_{i,j}$  are the number of active fires,  $EF_{s,v}$  are the offline emission factors, and  $f_v$  is the fractional area of that PFT in the grid cell.

337 Emission factors describe the PFT-specific speciated mass (in kg) of the smoke, 338 normalized per fire (Table 1). Emission factors were calculated offline using ModelE2.1 339 PFTs, annual mean global MODIS Terra fire count, and GFED4s emissions from the 340 period of 2003-2009. Our technique, known as multivariate curve fitting, matched the 341 emissions within the PFT fraction of the grid cell with the respective fire count. We 342 correlated a time series of GFED4s emissions with a time series of MODIS fire count for 343 each modeled PFT in a grid cell. Our settings included statistical (Poisson) weighting of 344 the GFED4s emissions (1 over emissions) and a uniform initial estimate of 100,000 kg m<sup>-</sup> <sup>2</sup> s<sup>-1</sup> per fire per PFT. This calculation resulted with a specific emission factors per PFT 345 346 (Table 1).

## 347 2.7 Implementation within ModelE

348 ModelE2.1 can be used with either GFED4s prescribed fire emissions or 349 interactive pyrE emissions. The pyrE module generates emissions at every model time 350 step with ESM-simulated climate as a driver. Flammability is calculated only in the 351 fraction of grid cells with natural vegetation. It is driven by the simulated surface RH, 352 surface temperature, monthly accumulated precipitation, and LAI. LAI is calculated by 353 Ent [Kim et al., 2015], the Terrestrial Biosphere Model component of ModelE2.1, and is 354 currently derived from 2005 MODIS LAI data [Tian et al., 2002a, 2002b]. Cloud-to-355 ground lightning, calculated by ModelE2.1, is used as the natural ignition source. Most 356 ESMs have low skill in reproducing flash rate distributions [Murray, 2016], and the GISS 357 model is no exception. A qualitative comparison with the World Wide Lightning 358 Location Network (WWLN) (not presented here) showed that modeled cloud-to-ground 359 lightning, which makes up only about 30% of total lightning, is biased high in ModelE2.1. 360 We decided to use a simple scaling factor of 0.1 in the calculation of natural ignition to 361 better match observed flash rates, as improving the lightning parameterization is beyond 362 the scope of this study.

363 All fire-related parameters like flammability, active fires, burned area, and fire 364 emissions are recalculated in every model time step (30 min) with memory only of the 365 burned area in the previous time step. We could not extend the "fire memory" past the 366 previous time step due to limitations related to ModelE's terrestrial biosphere module. 367 However this is a reasonable application, given that the climate inputs we use for fire 368 calculations such as monthly accumulated precipitation, surface RH and temperature 369 don't change significantly between each time step. The fire module's impact on the Earth 370 system is currently only through interactive emissions. Albedo, carbon stocks and LAI 371 are not modified by pyrE.

372 The modeling approach presented in this paper provides a good reproduction of 373 the seasonality compared to satellite retrievals (see Results section). However, the 374 simulated magnitude of active fires and burned area was too small compared to satellite 375 retrievals and required the use of a scaling factor, a common practice among other fire 376 models [Pfeifer et al., 2013; Hantson et al., 2016; Mangeon et al., 2016; Zou et al., 377 2019]. To calibrate the global modeled active fires to MODIS retrievals, we used a global 378 scaling factor of 30 for all active fires. A similar approach was taken by Pechony and 379 Shindell (2009). We scaled burned area by a factor of 250 to reach the magnitude of 380 GFED4s. Nevertheless, even with this large correction factor, burned area, which 381 accounts for a small fraction of the grid cell that is able to burn, has a very minor impact 382 on fire activity and fire emissions as its only impact to fire activity is through 383 flammability.

## **384 3 Model configuration**

We used ModelE2.1 with a spatial resolution of 2° in latitude by 2.5° in longitude, 40 vertical layers and a model top at 0.1 hPa. The vegetation component of ModelE2.1 is the Ent Terrestrial Biosphere Model (Ent TBM), which is coupled with the land use/land cover data in the model [*Kim et al.*, 2015]. Ent prescribes leaf area index (LAI) for 14 plant functional types (presented in Table 1) derived from MODIS 2005 data (cover and biome types [*Friedl et al.*, 2010]; LAI [*Tian et al.*, 2002a, 2002b]), historical crop cover [*Pongratz et al.*, 2008], and vegetation heights from [*Simard et al.*, 2011].

In this study we show results from runs of ModelE2.1 coupled to the aerosol
 microphysical scheme MATRIX (Multiconfiguration Aerosol TRacker of mIXing state)

394 [*Bauer et al.*, 2008]. MATRIX simulates aerosol formation, condensation and 395 coagulation, calculates the size distribution of aerosols and tracks their mixing state. Sea 396 salt, dust, and dimethyl sulfide (DMS) emissions were calculated interactively, driven by 397 the simulated climate, while other natural and anthropogenic fluxes, except for fires, were 398 prescribed from the CEDS (Community Emissions Data System) inventory [*Hoesly et al.*, 399 2018].

In the following, we will present a simulation with pyrE turned on, generating interactive fire emissions, and a simulation with pyrE turned off, using prescribed 2005 climatological (interpolated 2000-2010) GFED4s emissions instead. Also, we will discuss sensitivity studies using two simulations where pyrE generates interactive fire emissions but suppression is changed from a global parameterization to a regional one. Prescribed climatological monthly varying mean (1996-2004) sea surface temperature and sea ice thickness and extent were used as boundary conditions [*Rayner et al.*, 2003].

407 **4 Dataset** 

408 Most of the data below are based on a composite of level 3 Aqua and Terra 409 Moderate-resolution Imaging Spectro-radiometer (MODIS) Collection 5.1 data [Giglio et 410 al., 2003b; Giglio, 2013], unless otherwise stated. Aqua and Terra are sun-synchronous, 411 near-polar orbiting satellites with a global continuous record of more than 15 years; Aqua 412 was launched in May 2002 and Terra in December 1999. Aqua's overpass time is 413 1:30AM and 1:30PM local, and Terra's overpass time 10:30AM and 10:30PM local, and 414 their period is between one to two days. All reference data used in this study are 415 interpolated and re-gridded to the resolution of ModelE2.1.

## 416 **4.1 Population density**

Gridded population density (PD) that drives both anthropogenic ignition and fire
suppression is based on historical data for years prior to 2010 [*Klein Goldewijk et al.*,
2010]. PD has a time resolution of 10 years and is interpolated in between.

# 420 **4.2 Fire count**

To detect fires, MODIS uses brightness temperatures (thermal anomaly) derived from two channels [*Justice et al.*, 2002; *Giglio et al.*, 2006]. In our study we used the monthly cloud-corrected fire count (CloudCorrFirePix) climate model grid data (MYD14CMH, MOD14CMH). One single fire might include multiple fire pixels. The 425 spatial resolution of the data is  $0.5^{\circ}$ . Static, persistent hot spots are excluded from this 426 product [Giglio, 2013]. Because of its non-uniform spatial and temporal sampling, raw 427 MODIS data are biased high at high latitudes [Giglio et al., 2003a, 2006]. The product 428 we used is corrected for the multiple satellite overpasses, the missing data, and variable 429 cloud cover. Cloud cover hinders MODIS retrievals. The active fires in the product we 430 used are normalized to the fraction of cloud cover in a pixel. In highly cloudy pixels, the 431 product is set to zero. The local time of retrieval matters for fire detection, as fires are 432 driven by the daily cycle in solar heating. The largest number of active fires is detected 433 during daytime, with an order of magnitude difference between daytime detections and 434 nighttime detections [Ichoku et al., 2008]. Thus, differences are evident between the 435 Aqua and Terra retrievals. This motivated us to use data from the two satellites in our 436 analysis. We calculate and utilize climatological monthly means from the period 2003-437 2016.

## 438 **4.3 Burned area**

439 We used burned area from the Global Fire Emissions Database (GFED) version 440 4s that includes small fires [van der Werf et al., 2010, 2017; Randerson et al., 2012; 441 Giglio et al., 2013]. The GFED4s inventory is based on multi-sensor MODIS data, 442 involving both reflectance and thermal anomalies measurements from Aqua and Terra. 443 Retrievals must be free from cloud contamination and free from active fires within the 444 500 m MODIS grid cell. First, to generate the GFED4s data, MODIS burned area 445 collection 5.1 data (MCD64A1 product) are aggregated to a 0.25° grid. Then, burned area 446 from small fires is added. The burned area of small fires is statistically estimated using 447 active fires detected by MODIS (a composite of both Aqua and Terra). In this study we 448 use climatological monthly means of burned area from the period 2003-2016.

449

#### 4.4 Biomass burning emission inventory

GFED4s emissions are derived from the multiplication of burned area and fuel consumption [*van der Werf et al.*, 2010, 2017]. As such, they have the same spatial and temporal resolution as burned area, of 0.25° by 0.25° and a month. Fuel consumption is calculated using an estimation of fuel loss and combustion completeness, which are calculated using MODIS-based metrics such as differences in normalized burned area (dNBR), normalized vegetation index (NDVI), and land surface temperature (LST). The 456 satellite-based data are used as input to the Carnegie-Ames-Stanford Approach (CASA) 457 biogeochemical model [Randerson et al., 1996] to calculate the dry matter burned. Then, 458 emission factors [Andreae and Merlet, 2001; Akagi et al., 2011] are applied to convert 459 the dry matter burned to PFT-specific speciated gas and aerosol phase emissions. Kaiser 460 et al. (2012) and Pan et al. (2020) showed that there are regional biases in older and 461 current versions of GFED; being especially biased low in the Southern Hemisphere 462 compared to AERONET aerosol optical depth (AOD). In order to eliminate the strong 463 interannual BB variability, our analysis used GFED4s mean climatological data of 2000-464 2010.

### 465 **4.5 Fire regions**

The analysis we present below is based on the widely used fire regions (Fig. 1) as defined by GFED [*Giglio et al.*, 2006; *van der Werf et al.*, 2006]. The regions are defined based on climate and fire regimes, and are widely used as basis regions for global fire studies.

# 470 **4.6 Aerosol optical depth**

471 The impact of fire emissions on atmospheric composition is investigated by 472 comparing monthly Aqua and Terra MODIS retrievals of AOD at 550nm [Remer et al., 473 2005; Platnick et al., 2015]. AOD describes the entire atmospheric column-integrated 474 extinction of aerosols. MODIS AOD data are a useful tool in the study of simulated BB 475 plumes [Voulgarakis and Field, 2015; Johnson et al., 2016; Bauer et al., 2019]. The 476 AOD data we used has a 1° spatial resolution. The monthly mean data (MYD08 M3 and 477 MOD08 M3 products) have been averaged over the period 2003–2007 to create monthly 478 climatologies centered around the year 2005. The AOD product we use includes 479 improvements made via the Dark Target algorithm [Kaufman et al., 1997], which was 480 developed particularly for retrievals over dark vegetated surfaces [Wei et al., 2019]. 481 However, the algorithm fails at retrieving valid AOD data over bright surfaces like desert 482 areas [Levy et al., 2013], which we discard. Here we use collection 6.1 data.

# 483 **5 Results and discussion**

484 **5.1** Fire activity

485 **5.1.1 Regional suppression** 

486 First we want to demonstrate how the parameterization with regionally-dependent 487 fire suppression improves the simulation of fire activity compared to the original 488 simplified global fire suppression proposed by Pechony and Shindell (2009) (Fig. 4). Our 489 goal was to improve the fire parameterization in regions where the seasonality was 490 captured in timing but not in magnitude. We propose regional modifications to Africa 491 (NHAF, SHAF), a region that drives global fire activity, and had a distinct mismatch in 492 active fires compared to satellite retrievals. Originally, over NHAF the fire seasonality 493 was too flat, while over SHAF it matched MODIS-Terra, but was orders of magnitude 494 smaller than MODIS-Aqua. Since fire suppression for open BB is not commonly 495 practiced in rural Africa, eliminating it over NHAF and SHAF helped resolve the 496 seasonal cycle (Fig. 4 and Eq. 6). The two other regions we modified are TENA and 497 Middle East (MIDE). Over both of those regions the simulated fire seasonality was too 498 strong. Increasing fire suppression over MIDE and TENA greatly improved our 499 simulations compared to MODIS retrievals.

500 The pyrE module is skilled at capturing the fire seasonality in regions identified 501 by Forkel et al. (2017) as controlled by temperature and wetness (climate controls), like 502 Southern Hemisphere South America (SHSA) (Fig. A1). However, there are regions that 503 our parameterization does not simulate well, mainly due to the fact that the fire activity 504 there is driven by land use practices and intentional fire ignitions, which pyrE does not 505 resolve. For example, in TENA we are missing the spring peak of agricultural fires. 506 Similarly, over Europe and Boreal Asia (Fig. A1) we are missing the winter and spring 507 fires associated with intentional ignition [Dwyer et al., 2000; Ganteaume et al., 2013]. 508 Other regions where the seasonality is not well captured, likely due to the fact that it is 509 driven by intentional ignitions, include Central America, Northern Hemisphere South 510 America, Central Asia, Southeast Asia, and Equatorial Asia. Over Australia, the model 511 captures neither the magnitude nor the timing of the BB seasonality. This is in part due to 512 the model's poor performance of the simulated cloud-to-ground lightning ignitions in that 513 region (not shown).

514

In all simulations going forward we used the regional suppression scheme.

515 **5.1.2 Daily cycle** 

We looked at the active fires' daily cycle to see if it can explain the differences between Aqua, Terra, and the model. The monthly mean fire count detected by Aqua and Terra is expected to be different due to their different overpass times. In Fig. 5, pyrE simulates a distinct daily cycle in active fires in different locations. The simulated daily cycle is most strongly controlled by the simulated daily cycle in flammability (not presented here), matching the daily solar cycle. pyrE's ability to resolve a daily cycle of fire activity highlights the dynamic nature of a process-based fire model.

523 Using 30-minute simulation output, we sampled all surface grid cells at the 524 daytime overpass time of MODIS Terra, 10:30am local time, and MODIS Aqua, 1:30pm 525 local time. We focused on the daytime overpass time of Terra and Aqua since about 95% 526 of active fire detections occur then [Ichoku et al., 2008]. Our results in Fig. 6 and Fig. 7 527 indicate that, globally, simulated active fires sampled at daytime overpass are biased high 528 compared to MODIS retrievals from the respective satellite, for much of the year. On a 529 global annual mean, the active fires of the model sampled in daytime Terra overpass time 530 are higher than MODIS Terra by 45%, while the active fires of the model sampled in 531 daytime Aqua overpass time are higher than MODIS Aqua by 13%. However, this 532 behavior differs by region and maximizes in NH sub-Saharan Africa and SH central 533 Africa. The simulated fire activity is biased low compared to MODIS retrievals along the 534 coast of west Africa, in eastern southeast Asia and Australia. When simulated monthly 535 mean active fires values are in the range of Terra and Aqua (Fig. 4, A1), they are in fact 536 biased high, given the bias due to the overpass time of the satellite. Considering that the 537 actual number of active fires is likely higher than the number retrieved by MODIS, as 538 cloud contamination is decreasing its detection efficiency, it is conceivable that a model 539 weakly high-biased compare to the satellite retrievals is realistic. All results presented 540 later were not sampled according to a satellite overpass time, but instead were averaged 541 over the whole length of the day.

542 5.2 Burned area

543 The simulated burned area is biased low compared to the GFED4s inventory (Fig. 544 8, A2). The total annual simulated burned area (10-year climatological mean) is 380 Mha 545 while GFED4s burned area (mean of 2003-2016) is 460 Mha. However, this behavior is 546 region-specific. The simulated burned area is lower compared to GFED4s over northern 547 hemisphere Africa, particularly in November-December, over central and equatorial Asia, 548 and over Australia. The simulated burned area (Fig. 8, A2) reflects the spatial distribution 549 and seasonality of simulated active fires (Fig. 8, A1). GFED4s burned area and MODIS 550 fire count do not always have the same seasonality, for example during October-551 December. During this season the satellite-retrieved fires produce a higher burned area 552 relative to other seasons. The fire activity driving this behavior occurs in the NHAF 553 savanna, and northern hemisphere South America. In those regions and times of the year 554 the normalized mean bias of modeled burned area is at least twice the size of the 555 normalized mean bias of active fires, e.g. in NHAF a bias of 6.5 for burned area and 1-3 556 for active fires, depending on the MODIS satellite. This implies that for every fire 557 modeled in these regions and season a smaller area is simulated to burn compared to the 558 reference datasets.

559 Why is the burned area per fire relationship in simulations much weaker than it is 560 in the reference datasets? Two contributing factors are: prescribed PFT and simulated 561 wind. The prescribed PFT distribution present in the model is rudimentary; it is 562 comprised of 11 flammable vegetation types (Table 1). As for surface winds, the 563 simulated wind patterns driving burned area are averaged over a coarse grid cell 564 (2°x2.5°). Simulated wind does not represent sub-grid scale processes and is not fueled 565 by the fire's energy, which is likely contributing to an underestimation of the spread of 566 burned area. However, though wind directly impacts burned area, it does not play a major 567 role in the distribution of simulated fires, since burned area itself has a minor impact on 568 fires through flammability due to its small percentage in a grid cell. At most burned area 569 reaches less than 18% of the naturally vegetated fraction of a grid cell, and is on average 570 less than 1%.

#### **571 5.3 Emissions**

572 Due to limitations in the current capabilities of the simulated terrestrial biosphere 573 in ModelE, emissions are generated from active fires, similar to the approach of *Pechony* 574 *and Shindell* (2009, 2010) and *Pechony et al.* (2013). The main source regions for fire 575 emissions are NHAF, EQAS, SHSA, and SHAF. Emissions are well simulated over 576 SHSA and SHAF (Fig. A3-A5), both in terms of timing of the seasonality and in 577 magnitude. The main regions where simulated emissions are lower than GFED4s are 578 NHAF and EQAS, mainly Indonesia (Fig. 8, A3-A5). However, more generally, 579 simulated gaseous and particulate emissions are globally biased low compared to 580 GFED4s emissions (Table 2). To a lesser degree, simulated fire emissions are also 581 weaker compared to GFED4s in the boreal regions (Fig. A3-A5). The contribution from 582 these regions to the global total is an order of magnitude smaller compared to the main 583 source regions.

584 The weaker emissions compared to GFED4s are responding to the following inputs: 585 offline emissions factors, lack of crop and peat fires, LAI, and prescribed PFTs. The 586 emission factors that generate fire emissions are derived using multivariate statistical 587 analysis. Though we used seven full years (2003-2009) of data to derive the factors, it 588 might have generated biases in emissions. Areas that burn annually are properly sampled, 589 but areas that have a fire cycle that is longer than a seven year might be biased high or 590 low, depending on whether they were included in the training dataset or not. Also, crop 591 and peat fires are not explicitly included in the simulated emissions, as intentional 592 ignition is not parameterized in pyrE. Specifically, fires are not applied to the crop faction 593 of a grid cell, and peat surfaces are not included in the PFTs. However, our method of 594 deriving the offline emission factors uses MODIS fire count and GFED4s emissions, and 595 does not distinguish between intentional and accidental fires. Hence, intentional fires are 596 indirectly accounted for in the global sum. However, this indirect inclusion of intentional 597 fires does not necessarily add missing fire emissions in the correct locations. The LAI in 598 Ent, ModelE's DGVM, is based on 2005 MODIS retrievals. Though we cannot estimate 599 the role that the lack of interactive LAI plays, it is certainly not optimal, neither for fire 600 activity simulation, nor for fire emissions that are derived from active fires. Unlike 601 simulated active fires, simulated fire emissions are strongly tied to the map of PFTs. The 602 offline emission factors are based on prescribed PFTs, and the interactive emissions 603 themselves are applied according to the sub-grid PFT distribution. The prescribed PFT 604 distribution present in the model might be different than reality, and those differences 605 affect emissions. In the model, the PFTs in areas where emissions are biased high 606 compared to GFED4s there is a high percentage (>50%) of the following PFTs: 607 evergreen broadleaf trees (Amazon, central Africa), cold broadleaf trees (northeast 608 America, Europe), and drought broadleaf trees (central Africa and northern India). In 609 EQAS, a region with biased low simulated emissions, close to 100% of the prescribed

610 PFTs is evergreen broadleaf trees, which in reality is replaced by crops. The biased low

611 emissions in EQAS are very likely tied to the lack of prescribed peat PFT. In areas with

612 biased low emissions modeled PFTs are mainly (>50%) c4 grass (NHAF, Australia),

613 deciduous needle leaf trees (boreal regions), and arid shrubs (S Africa, Australia).

614 **5.4 Composition** 

# 615 **5.4.1 Column load**

616 In order to quantify how the model skill changes with the inclusion of pyrE instead of prescribed emission inventory data in ModelE2.1, we compare a simulation 617 618 with interactive fires to a simulation with prescribed BB sources. Though emissions are 619 mostly biased-low compared to GFED4s, this behavior is less evident in the column 620 density (Fig. 9). For most BB emitted species, the simulation with interactive fires has 621 lower column densities than the simulation with prescribed emissions (Table 2) with a 622 bias ranging from -6.3-0.5% for gaseous species, -4.8% for black carbon and -16% for 623 organic aerosol. However, the column densities are only partly driven by fire emissions, 624 as those make up less than 35% of total global emissions of either CO, organic aerosol, 625 and black carbon emissions. Non-emissions production-and-loss mechanisms also impact 626 column densities. Having a weak global impact on composition does not imply that 627 regionally fires are not important.

The difference in column densities between the two simulations is greatest over north sub-Saharan Africa, Indonesia, and the boreal regions. The behavior is regionspecific, and some regions like central Africa and northern hemisphere South America have higher column densities compared to the simulation with prescribed emissions. The differences between the two simulations are more prominent for organic aerosol than any of the other species (Fig. 9, Table 2), while the differences in the spatial distribution of CO are marginal.

# 635 5.4.2 Aerosol optical depth (AOD)

In Fig. 10 we compare climatologically-simulated clear-sky AOD with MODIS AOD (Aqua) for January, April, July, and October. The conclusions from Terra products are similar to Aqua's, and will not be presented here, for brevity. In a regional perspective, simulated AOD is able to reproduce the seasonality and spatial distribution of MODIS-retrieved pollution over west and central Africa, east and southeast Asia, and
the Arabian sea. The simulations of ModelE2.1 has higher AOD compared to MODIS
over the tropical eastern Pacific, an artifact due to the model's skill in simulating
stratocumulus cloud decks, which have been improved in a newer version of the ESM
(ModelE3).

645 Model performance as a function of interactive versus offline fire emissions is 646 similar in terms of AOD (Fig. 11). Both simulations have persistently lower (0-30%) 647 AODs over central Africa and central South America compared to MODIS. The locations 648 with an outstanding difference in performance between the simulations are in central sub-649 Saharan Africa in January and July, and over a small area in Indonesia (Kalimantan) 650 during October. In January over central sub-Saharan Africa the simulation with pyrE has 651 AOD values (NHAF regional mean AOD of 0.26) closer to MODIS (NHAF regional 652 mean AOD of 0.2) than a simulation with prescribed fire emissions (NHAF regional 653 mean AOD of 0.33), while in July it is the simulation with pyrE (NHAF regional mean 654 AOD of 0.53) that is more biased high than the prescribed one (NHAF regional mean 655 AOD of 0.46). Over EQAS in October the simulation with prescribed fires has an AOD of  $\sim 0.28$  while the simulation with pyrE has an AOD of  $\sim 0.18$ . AOD in this region is 656 657 sensitive to peat fires, which are not included in ModelE, strongly impacting pyrE's 658 results. Globally, mean AOD simulated with interactive fire emissions is 0.142 while 659 mean AOD simulated with prescribed fire emissions is 0.146. The fact that pyrE has a 660 marginal performance in climatological runs when compared against a simulation with 661 the more accurate offline emissions is a strong indication that it is a robust module that 662 can be used with confidence at time periods where offline emissions are not available.

663 Finally, we demonstrate the contribution of BB emissions to total clear-sky AOD 664 by comparing the simulations with both prescribed and interactive fire emissions to a 665 simulation that has no fire emissions at all (Fig. 12). In the simulation with prescribed fire 666 emissions, clear sky AOD is on average 10% higher than it is in a simulation with no fire 667 emissions. In a simulation with pyrE clear sky AOD is about 7.5% higher than it is in a 668 simulation with no fire emissions. The impact of BB emissions on AOD is most pronounced in the source regions of Africa and the Amazon. In those regions the 669 670 difference in AOD varies between 0.15-0.3. It is important to note that the differences in AOD are not only due to impact of BB emissions, but also reflect climate variability,which impacts aerosol lifetime and interactive dust emissions.

## 673 6 Conclusions

674 The development of pyrE allowed us for the first time to interactively simulate 675 climate and fire activity with GISS-ModelE2.1. The pyrE module, which is based on a 676 the fire parameterizations of Pechony and Shindell (2009), was expanded to include fire 677 spread and burned area, following the approach of Li et al. (2012). This study set out to 678 simulate the climatology of fires, and not individual fire events. Like only a few other fire 679 models [Zou et al., 2019], pyrE was developed with consideration of regional behavior. 680 The new fire suppression scheme depends on population density, but also on geographic 681 regions. The new scheme reflects more intense fire suppression in the USA and Middle 682 East, and revokes fire suppression in Africa, which improved the fire activity seasonality 683 simulated by pyrE compared to satellite retrievals. Active fires' seasonality is well 684 simulated in the fire source regions: the Amazon, SH Africa, and NH Africa, with the 685 exception of being biased low compared to MODIS during November-December. This is 686 due to the lack in parameterization of intentional ignitions and agricultural fires.

687 The regional model skill of fire activity was also demonstrated in the simulated 688 burned area. Burned area in southern hemisphere Africa was well simulated by the model, 689 while less active fire regions like temperate and boreal North America, Boreal Asia, 690 Europe, and Middle East were biased high compared to GFED4s. Other regions like 691 Australia, northern sub-Saharan Africa in November-December, Central Asia and 692 Southeast Asia in January-March were biased low. Though the seasonality of simulated 693 burned area reflects that of simulated active fires, the bias of burned area compared to 694 GFED4s data is at least double that of active fires. Burned area is a quantity that most fire 695 models struggle with. Wind speed, a driver of burned area, is averaged over a coarse grid 696 cell, with no feedback from fire heat and energy, which can be a contributing factor to the 697 lower simulated burned area values. The prescribed rudimentary PFTs of the model are a 698 simplified version of the real world and thus can be a source of additional uncertainty. 699 Finally, the rate of spread of burned area, a function of the burning vegetation type, that 700 pyrE and other fire models use is on the lower end of field observations. A higher rate of

spread could help to both override the scaling factor used for burned area, and to reducethe negative bias compared to GFED4s.

703 Unlike other fire models, fire emissions in pyrE are driven directly by fires 704 instead of burned area. Emissions are based on online active fires calculations and offline 705 emission factors derived as described in Sect. 2.6. In contrast to the fact that simulated 706 active fires are biased high compared to MODIS, globally, fire emissions are biased low 707 compared to GFED4s. Fire emissions are well-simulated over the southern hemisphere 708 with the exception of Australia. Emissions are biased low over the northern hemisphere 709 including northern sub-Sahara, with the exception of NH South America, which is biased 710 high. The bias of active fires compared to MODIS in Australia and in northern sub-711 Saharan Africa during November-December propagates to emissions. The emission 712 factors, which were calculated offline using MODIS fire count and GFED4s fire 713 emissions and were applied based on the prescribed PFTs of the model, have their own 714 limitations. They are based on a training dataset of seven years, which would introduce 715 biases in regions where fire cycle is longer than seven years. Also, they rely on the 716 modeled PFTs, enhancing the emissions dependency on the prescribed PFT and the lack 717 of peat. Emission factors do not distinguish between intentional and accidental fires, thus 718 they indirectly account for all fire emissions, which reduce existing biases, although the 719 regional distribution of them will not match the locations of intentional fires, unless 720 natural vegetation burning occurs in the vicinity.

Less emissions compared to GFED4s means lower column densities and lower AOD when comparing a simulation with interactive fires to one with prescribed fires. However, as these quantities depend on climate feedbacks including processes other than fire, e.g. additional emission sources, precipitation, deposition, transport, and chemistry, the differences between the two simulations dilute. Nonetheless, a comparison with MODIS AOD demonstrates that AOD from a simulation with interactive fire emissions is comparable to AOD from a simulation with prescribed fire emissions.

The work presented here highlights that timing matters just as much as magnitude. This is true for fire distribution, emissions, and atmospheric composition. Timing is also the reason why intentional ignition was excluded from pyrE. Intentional ignition, namely land clearing and agricultural fires, depends on region and crop specific planting and 732 harvesting times. To include it would require crop functionality in ModelE, which was 733 not present during the time of our development. Further future development should focus 734 on the inclusion of intentional ignition and agricultural fires which are seasonal in nature, 735 derived from crop planting and land clearing times. This addition could perhaps improve 736 model performance over regions like equatorial Asia, Southeast Asia, and Central 737 America as well as override the global scaling factors applied to active fires and burned 738 area. The use of scaling factors is a common practice among fire models, and should be 739 carefully and transparently documented. Also, enhancing the prescribed PFTs, especially 740 via the addition of peat is imperative when studying fires. Peat exists as well outside of 741 tropical Asia. There are immense reservoirs of peat in Africa [Dargie et al., 2017], as 742 well as the boreal regions [Yu, 2012], where it used to be trapped under permafrost. Peat 743 will likely become an even bigger source of fire emissions in the future. Improvement of 744 the cloud to ground lightning parameterization may also prove useful, as changes to 745 natural ignition will likely have significant impacts on Australian and boreal fire 746 emissions. Finally, given that the heat component of fires interact with the climate system, 747 and can also be used to derive more accurate emissions, as demonstrated by Ichoku and 748 Ellison (2014) and three of the eleven FireMIP models (Rabin et al., 2017), it is 749 worthwhile taking it into consideration when developing new fire modeling capabilities.

#### 750 **7 Code availability**

Information on ModelE, including access to online data and descriptions are available at
http://www.giss.nasa.gov/tools/modelE. The pyrE module is included in ModelE version
2.1. The source code, along with documentation, can be downloaded from the NASA
Goddard Institute of Space Studies website: https://simplex.giss.nasa.gov/snapshots/.

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### 762 **References**

- Akagi, S. K., R. J. Yokelson, C. Wiedinmyer, M. J. Alvarado, J. S. Reid, T. Karl, J. D.
  Crounse, and P. O. Wennberg: Emission factors for open and domestic biomass
  burning for use in atmospheric models, *Atmos. Chem. Phys.*, *11*(9), 4039–4072,
  doi:10.5194/acp-11-4039-2011, 2011.
- Andela, N., and G. R. Van Der Werf : Recent trends in African fires driven by cropland
  expansion and El Niño to La Niña transition, 4(September), 791–795,
- 769 doi:10.1038/NCLIMATE2313, 2014.
- Andela, N., D. C. Morton, L. Giglio, Y. Chen, G. R. van der Werf, P. S. Kasibhatla, R. S.
- DeFries, G. J. Collatz, S. Hantson, S. Kloster, D. Bachelet, M. Forrest, G. Lasslop, F.
  Li, S. Mangeon, J. R. Melton, C. Yue, J. T. Randerson: A human-driven decline in
- 773 global burned area, , *1362*(June), 1356–1362, 2017.
- Andreae, M. O.: Biomass burning: Its history, use, and distribution and its impact on
  environmental quality and global climate, in Global Biomass Burining: Atmospheirc,
  Climate and Biospheric implications, edited by J, S. Levine, *MIT Press. Cambridge, Mass.*, 3–21, 1991.
- Andreae, M. O., and P. Merlet: Emission of trace gases and aerosols from biomass
  burning, *Global Biogeochem. Cycles*, *15*(4), 955–966, doi:10.1029/2000GB001382,
  2001.
- Andreae, M. O., D. Rosenfeld, P. Artaxo, A. A. Costa, G. P. Frank, K. M. Longo, and M.
  A. F. Silva-Dias: Smoking rain clouds over the Amazon., *Science*, *303*, 1337–1342,
  doi:10.1126/science.1092779, 2004.
- Andreae, M. O. (2019), Emission of trace gases and aerosols from biomass burning an
  updated assessment, *Atmos. Chem. Phys.*, *19*, *8523-8546*,
  https://doi.org/10.5194/acp-19-8523-2019.
- Archibald, S., A. C. Staver, and S. A. Levin: Evolution of human-driven fire regimes in
  Africa, *Proc. Natl. Acad. Sci.*, *109*(3), 847–852, doi:10.1073/pnas.1118648109,
  2012.
- Archibald S. 2016: Managing the human component of fire regimes: lessons from Africa.
  Phil. Trans. R. Soc. B 371: 20150346. http://dx.doi.org/10.1098/rstb.2015.0346
- 792 Arora, V. K., and G. J. Boer: Fire as an interactive component of dynamic vegetation

- Balch, J. K., B. A. Bradley, J. T. Abatzoglou, R. C. Nagy, and E. J. Fusco: Human-started
  wildfires expand the fire niche across the United States, , *114*(11),
- 796 doi:10.1073/pnas.1617394114, 2017.
- Bachelet, D., Ferschweiler, K., Sheehan, T. J., Sleeter, B. M., and Zhu, Z.: Projected
  carbon stocks in the conterminous USA with land use and variable fire regimes,
  Glob. Change Biol., 21, 4548–4560, doi:10.1111/gcb.13048, 2015.

800 Bauer, S.E., D. Wright, D. Koch, E.R. Lewis, R. McGraw, L.-S. Chang, S.E. Schwartz,

and R. Ruedy: MATRIX (Multiconfiguration Aerosol TRacker of mIXing state): An
aerosol microphysical module for global atmospheric models. Atmos. Chem. Phys.,

803 8, 6603-6035, doi:10.5194/acp-8-6003-2008, 2008.

- Bauer, S. E., and S. Menon: Aerosol direct, indirect, semidirect, and surface albedo
  effects from sector contributions based on the IPCC AR5 emissions for preindustrial
  and present-day conditions, 117, 1–15, doi:10.1029/2011JD016816, 2012.
- Bauer, S. E., U. Im, K. Mezuman, and C. Y. Gao: Desert dust, industrialization and
  agricultural fires: Health impacts of outdoor air pollution in Africa, *J. Geophys. Res. Atmos.*, 1–17, doi:10.1029/2018JD029336, 2019.
- Bellouin, N., A. Jones, J. Haywood, and S. A. Christopher: Updated estimate of aerosol
  direct Radiative forcing from satellite observations and comparison against the
  centre climate model, *J. Geophys. Res. Atmos.*, *113*(10), 1–15,
- 813 doi:10.1029/2007JD009385, 2008.
- Bond, T. C., and R. W. Bergstrom: Light Absorption by Carbonaceous Particles: An
  Investigative Review, *Aerosol Sci. Technol.*, 40(1), 27–67,
- 816 doi:10.1080/02786820500421521, 2006.
- 817 Bowman, D. M. J. S., J. K. Balch, P. Artaxo, W. J. Bond, J. M. Carlson, M. A. Cochrane,
- 818 C. M. D'Antonio, R. S. DeFries, J. C. Doyle, S. P. Harrison, F. H. Johnston, J. E.
- 819 Keeley, M. A. Krawchuk, C. A. Kull, J. B. Marston, M. A. Moritz, I. C. Prentice, C.
- 820 I. Roos, A. C. Scott, T. W. Swetnam, G. R. van der Werf, and S. J. Pyne: Fire in the
- 821 Earth System, *Science*, *324*(5926), 481–484, doi:10.1126/science.1163886, 2009.
- 822 Bowman, D. M. J. S., J. Balch, P. Artaxo, W. J. Bond, M. A. Cochrane, C. M. D'Antonio,
- 823 R. DeFries, F. H. Johnston, J. E. Keeley, M. A. Krawchuk, C. A. Kull, M. Mack, M.

<sup>793</sup> models, J. Geophys. Res., 110, doi:10.1029/2005JG000042, 2005.

824	A. Moritz, S. Pyne, C. I. Roos, A. C. Scott, N. S. Sodhi, and T. W. Swetnam: The
825	human dimension of fire regimes on Earth, J. Biogeogr., 38(12), 2223-2236,
826	doi:10.1111/j.1365-2699.2011.02595.x, 2011.
827	Buchholz, R. R., D. Hammerling, H. M. Worden, M. N. Deeter, L. K. Emmons, D. P.
828	Edwards, and S. A. Monks: Links Between Carbon Monoxide and Climate Indices
829	for the Southern Hemisphere and Tropical Fire Regions, J. Geophys. Res. Atmos.,
830	123(17), 9786–9800, doi:10.1029/2018JD028438, 2018.
831	Butsic, V., M. Kelly, and M. Moritz: Land Use and Wildfire: A Review of Local
832	Interactions and Teleconnections, Land, 4(1), 140-156, doi:10.3390/land4010140,
833	2015.
834	Carslaw, K. S., L. A. Lee, C. L. Reddington, K. J. Pringle, A. Rap, P. M. Forster, G. W.
835	Mann, D. V. Spracklen, M. T. Woodhouse1, L. A. Regayre, and J. R. Pierce: Large
836	contribution of natural aerosols to uncertainty in indirect forcing., Nature, 503(7474),
837	67-71, doi:10.1038/nature12674, 2013.
838	Chuvieco, E., C. Yue, A. Heil, F. Mouillot, I. Alonso-canas, M. Padilla, J. M. Pereira, D.
839	Oom, and K. Tansey: METHODS A new global burned area product for climate
840	assessment of fire impacts, , 45, 619-629, doi:10.1111/geb.12440, 2016.
841	Crutzen, P. J., L. E. Heidt, J. P. Krasnec, W. H. Pollock, and W. Seiler: Biomass burning
842	as a source of atmospheric gases CO, H2, N2O, NO, CH3Cl and COS, Nature, 282,
843	253-256, doi:10.1038/282253a0.
844	Crutzen, P. J., and M. O. Andreae (1990), Biomass burning in the tropics: impact on
845	atmospheric chemistry and biogeochemical cycles., Science, 250, 1669–1678,
846	doi:10.1126/science.250.4988.1669, 1979.
847	Díaz-Avalos, C., D. L. Peterson, E. Alvarado, S. a Ferguson, and J. E. Besag: Space-time
848	modelling of lightning-caused ignitions in the Blue Mountains, Oregon, Can. J. For.
849	Res., 31, 1579–1593, doi:10.1139/cjfr-31-9-1579, 2001.
850	Dwyer, E., S. Pinnock, J. M. Gregoire, and J. M. C. Pereira: Global spatial and temporal
851	distribution of vegetation fire as determined from satellite observations, Int. J.
852	Remote Sens., 21(6-7), 1289-1302, doi:10.1080/014311600210182, 2000.
853	Feingold, G., L. A. Remer, J. Ramaprasad, and Y. J. Kaufman: Analysis of smoke impact
854	on clouds in Brazilian biomass burning regions: An extension of Twomey's

855 approach, J. Geophys. Res., 106(D19), 22907, doi:10.1029/2001JD000732, 2001. 856 Field, R. D., G. R. van der Werf, T. Faninc, E. J. Fetzerd, R. Fullerd, H. Jethvae, R. 857 Levye, N. J. Liveseyd, M. Luod, O. Torrese, and H. M. Worden: Indonesian fire 858 activity and smoke pollution in 2015 show persistent nonlinear sensitivity to El 859 Niño-induced drought, Proc. Natl. Acad. Sci., 113(33), 9204–9209, 860 doi:10.1073/pnas.1524888113, 2016. 861 Fischer, A. P., T. A. Spies, T. A Steelman, C. Moseley, B. R. Johnson, J. D. Bailey, 862 A. A. Ager, P. Bourgeron, S. Charnley, B. M. Collins, J. D. Kline, J. E. Leahy, 863 J. S. Littell, J. D. A. Millington, M. Nielsen-Pincus, C. S. Olsen, T. B. Paveglio, C. I. 864 Roos, M. M. Steen-Adams, F. R. Stevens, J. Vukomanovic, E. M. White, and D. M. 865 J. S. Bowman: Wildfire risk as a socioecological pathology, Front. Ecol. Environ., 866 14(5), 276–284, doi:10.1002/fee.1283, 2016. 867 Forkel, M., W. Dorigo, G. Lasslop, I. Teubner, E. Chuvieco, and K. Thonicke: A data-868 driven approach to identify controls on global fire activity from satellite and climate 869 observations (SOFIA V1), Geosci. Model Dev., 10(12), 4443-4476, 870 doi:10.5194/gmd-10-4443-2017, 2017. 871 Forkel, M., N. Andela, S. P. Harrison, G. Lasslop, M. van Marle, E. Chuvieco, W. Dorigo, 872 M. Forrest, S. Hantson, A. Heil, F. Li, J. Melton, S. Sitch, C. Yue, and A. Arneth: 873 Emergent relationships with respect to burned area in global satellite observations 874 and fire-enabled vegetation models, *Biogeosciences*, 16(1), 57-76, doi:10.5194/bg-875 16-57-2019, 2019. 876 Friedl, M. A., D. Sulla-Menashe, B. Tan, A. Schneider, N. Ramankutty, A. Sibley, and X. 877 Huang: MODIS Collection 5 global land cover: Algorithm refinements and 878 characterization of new datasets, Remote Sens. Environ., 114, 168-182, 879 doi:10.1016/j.rse.2009.08.016, 2010. 880 Ganteaume, A., A. Camia, M. Jappiot, J. San-Miguel-Ayanz, M. Long-Fournel, and C. 881 Lampin: A review of the main driving factors of forest fire ignition over Europe, 882 Environ. Manage., 51(3), 651–662, doi:10.1007/s00267-012-9961-z, 2013. 883 Giglio, L.: MODIS Collection 5 Active Fire Product User's Guide Version 2.5, Sci. Syst. 884 Appl. Inc, (March), 61, 2013. 885 Giglio, L., J. D. Kendall, and R. Mack: A multi-year active fire dataset for the tropics

- derived from the TRMM VIRS, Int. J. Remote Sens., 24(22), 4505–4525,
- doi:10.1080/0143116031000070283, 2003a.
- Giglio, L., J. Descloitres, C. O. Justice, and Y. J. Kaufman: An enhanced contextual fire
  detection algorithm for MODIS, *Remote Sens. Environ.*, 87(2–3), 273–282,
- 890 doi:10.1016/S0034-4257(03)00184-6, 2003b.
- Giglio, L., I. Csiszar, and C. O. Justice: Global distribution and seasonality of active fires
  as observed with the Terra and Aqua Moderate Resolution Imaging
- 893 Spectroradiometer (MODIS) sensors, *J. Geophys. Res. Biogeosciences*, *111*(2), 1–12,
   894 doi:10.1029/2005JG000142, 2006.
- Giglio, L., J. T. Randerson, and G. R. Van Der Werf: Analysis of daily, monthly, and
  annual burned area using the fourth-generation global fire emissions database
- 897 (GFED4), J. Geophys. Res. Biogeosciences, 118(1), 317–328,
- 898 doi:10.1002/jgrg.20042, 2013.
- Goff, J. A.: Saturation pressure of water on the new Kelvin temperature scale, in *Transactions of the American Society of Heating and Ventilating Engineers, 63rd Semi-Annual Meeting*, pp. 347–354, Am. Soc. of Heating and Ventilating Eng.,
  Murray Bay, Quebec, Canada, 1957.
- Goff, J. A., and S. Gratch: Low-pressure properties of water from 160 to 212F, in
   *Transactions of the American Society of Heating and Ventilating Engineers*, 52nd
- 905 *Annual Meeting*, pp. 95–122, Am. Soc. of Heating and Ventilating Eng., New York,
  906 1946.
- Hamilton, D. S., S. Hantson, C. E. Scott, J. O. Kaplan, K. J. Pringle, L. P. Nieradzik, A.
  Rap, G. A. Folberth, D. V. Spracklen, and K. S. Carslaw: Reassessment of preindustrial fire emissions strongly affects anthropogenic aerosol forcing, *Nat.*
- 910 *Commun.*, *9*(1), doi:10.1038/s41467-018-05592-9, 2018.
- Hantson, S., G. Lasslop, S. Kloster, and E. Chuvieco: Anthropogenic effects on global
  mean fire size, *Int. J. Wildl. Fire*, 24(5), 589–596, doi:10.1071/WF14208, 2015.
- 913 Hantson, S., A, Arneth, S. P. Harrison, D. I. Kelley, I. C. Prentice, S. S. Rabin, S.
- 914 Archibald, F. Mouillot, S. R. Arnold, P. Artaxo, D. Bachelet, P. Ciais, M. Forrest, P.
- 915 Friedlingstein, T. Hickler, J. O. Kaplan, S. Kloster, W. Knorr, G. Lasslop, F. Li, S.
- 916 Mangeon, J. R. Melton, A. Meyn, S. Sitch, A. Spessa, G. R. van der Werf, A.

917	Voulgarakis, and C. Yue: The status and challenge of global fire modelling,
918	Biogeosciences, 13(11), 3359–3375, doi:10.5194/bg-13-3359-2016, 2016.
919	Hantson, S., M. Scheffer, S. Pueyo, C. Xu, G. Lasslop, E. H. Van Nes, M. Holmgren, and
920	J. Mendelsohn: Rare, Intense, Big fires dominate the global tropics under drier
921	conditions, Sci. Rep., 7(1), 7–11, doi:10.1038/s41598-017-14654-9, 2017.
922	Hoesly, R. M., S. J. Smith, L. Feng, Z. Klimont, G. Janssens-Maenhout, T. Pitkanen, J. J.
923	Seibert, L. Vu, R. J. Andres, R. M. Bolt, T. C. Bond, L. Dawidowski, N. Kholod, J.
924	Kurokawa, M. Li, L. Liu, Z. Lu, M. C. P. Moura, P. R. O'Rourke, and Q. Zhang:
925	Historical (1750-2014) anthropogenic emissions of reactive gases and aerosols from
926	the Community Emissions Data System (CEDS), Geosci. Model Dev., 11(1), 369-
927	408, doi:10.5194/gmd-11-369-2018, 2018.
928	Ichoku, C., L. Giglio, M. J. Wooster, and L. A. Remer: Global characterization of
929	biomass-burning patterns using satellite measurements of fire radiative energy,
930	Remote Sens. Environ., 112(6), 2950-2962, doi:10.1016/j.rse.2008.02.009, 2008.
931	Ichoku, C., and L. Ellison: Global top-down smoke-aerosol emissions estimation using
932	satellite fire radiative power measurements, Atmos. Chem. Phys., 14, 6643-6667,
933	doi:10.5194/acp-14-6643-2014, 2014.
934	Ichoku, C., R. Kahn, and M. Chin: Satellite contributions to the quantitative
935	characterization of biomass burning for climate modeling, Atmos. Res., 111, 1-28,
936	doi:10.1016/j.atmosres.2012.03.007, 2012.
937	Ito, A., and J. E. Penner: Historical emissions of carbonaceous aerosols from biomass and
938	fossil fuel burning for the period 1870-2000, Global Biogeochem. Cycles, 19(2), 1-
939	14, doi:10.1029/2004GB002374, 2005.
940	Jiang, Y., Z. Lu, X. Liu, Y. Qian, K. Zhang, Y. Wang, and XQ. Yang: Impacts of
941	Global Wildfire Aerosols on Direct Radiative, Cloud and Surface-Albedo Forcings
942	Simulated with CAM5, Atmos. Chem. Phys., 16, 14805–14824, doi:10.5194/acp-16-
943	14805-2016, 2016.
944	Johnson, B. T., J. M. Haywood, J. M. Langridge, E. Darbyshire, W. T. Morgan, K. Szpek,
945	J. K. Brooke, F. Marenco, H. Coe, P. Artaxo, K. M. Longo, J. P. Mulcahy, G. W.
946	Mann, M. Dalvi, and N. Bellouin: Evaluation of biomass burning aerosols in the
947	HadGEM3 climate model with observations from the SAMBBA field campaign, ,

949 Johnston, F. H., S. B. Henderson, Y. Chen, J. T. Randerson, M. Marlier, R. S. Defries, P. 950 Kinney, D. M. J. S. Bowman, and M. Brauer: Estimated Global Mortality 951 Attributable to Smoke from Landscape Fires, , 120(5), 695–701, 2012. 952 Johnston, F. H., S. Purdie, B. Jalaludin, K. L. Martin, S. B. Henderson, and G. G. Morgan: 953 Air pollution events from forest fires and emergency department attendances in 954 Sydney, Australia 1996-2007: A case-crossover analysis, Environ. Heal. A Glob. 955 Access Sci. Source, 13(1), 1-9, doi:10.1186/1476-069X-13-105, 2014. 956 Johnston, F. H., S. Melody, and D. M. J. S. Bowman: The pyrohealth transition: How 957 combustion emissions have shaped health through human history, Philos. Trans. R. 958 Soc. B Biol. Sci., 371(1696), doi:10.1098/rstb.2015.0173, 2016. 959 Justice, C. ., L. Giglio, S. Korontzi, J. Owens, J. . Morisette, D. Roy, J. Descloitres, S. 960 Alleaume, F. Petitcolin, and Y. Kaufman: The MODIS fire products, Remote Sens. 961 *Environ.*, 83(1–2), 244–262, doi:10.1016/S0034-4257(02)00076-7, 2002. 962 Kaiser, J. W., A. Heil, M. O. Andreae, A. Benedetti, N. Chubarova, L. Jones, J.-J. 963 Morcrette, M. Razinger, M. G. Schultz, M. Suttie, and G. R. van der Werf: Biomass 964 burning emissions estimated with a global fire assimilation system based on 965 observed fire radiative power, Biogeosciences, 9(1), 527–554, doi:10.5194/bg-9-966 527-2012, 2012. 967 Kaufman, Y. J., A. E. Wald, L. A. Remer, B. C. Gao, R. R. Li, and L. Flynn: MODIS 968 2.1-µm channel - correlation with visible reflectance for use in remote sensing of 969 aerosol, IEEE Trans. Geosci. Remote Sens., 35(5), 1286-1298, 970 doi:10.1109/36.628795, 1997. 971 Keetch, J. J. J. J., and G. M. G. M. Byram: A drought index for forest fire control, Notes, 972 *E-38. Ashe*, 35, doi:10.1016/j.accpm.2015.04.007, 1968. 973 Kim, Y., P. R. Moorcroft, I. Aleinov, M. J. Puma, and N. Y. Kiang: Variability of 974 phenology and fluxes of water and carbon with observed and simulated soil moisture 975 in the Ent Terrestrial Biosphere Model (Ent TBM version 1.0.1.0.0), Geosci. Model 976 Dev., 8(12), 3837–3865, doi:10.5194/gmd-8-3837-2015, 2015. 977 Klein Goldewijk, K., a. Beusen, and P. Janssen: Long-term dynamic modeling of global 978 population and built-up area in a spatially explicit way: HYDE 3.1, *The Holocene*,

14657-14685, doi:10.5194/acp-16-14657-2016, 2016.

948

- Lack, D. A., J. M. Langridge, R. Bahreini, C. D. Cappa, and A. M. Middlebrook: Brown
  carbon and internal mixing in biomass burning particles, *109*(37),
  doi:10.1073/pnas.1206575109/-
- 983 /DCSupplemental.www.pnas.org/cgi/doi/10.1073/pnas.1206575109, 2012.
- Lack, D. a., and J. M. Langridge: On the attribution of black and brown carbon light
  absorption using the Ångström exponent, *Atmos. Chem. Phys.*, *13*(20), 10535–10543,
  doi:10.5194/acp-13-10535-2013, 2013.
- 987 Lamarque, J. F., T. C. Bond, V. Eyring, C. Granier, A. Heil, Z. Klimont, D. Lee, C.
- 988 Liousse, A. Mieville, B. Owen, M. G. Schultz, D. Shindell, S. J. Smith, E. Stehfest, J.
- 989 Van Aardenne, O. R. Cooper, M. Kainuma, N. Mahowald, J. R. McConnell, V. Naik,
- 990 K. Riahi, and D. P. van Vuuren: Historical (1850–2000) gridded anthropogenic and
- biomass burning emissions of reactive gases and aerosols: methodology and
- 992 application, Atmos. Chem. Phys., 10(15), 7017–7039, doi:10.5194/acp-10-7017-
- 993 2010, 2010.
- Landry, J.-S., and H. D. Matthews: Non-deforestation fire vs. fossil fuel combustion: the
  source of CO<sub>2</sub> emissions affects the global carbon cycle and climate responses, *Biogeosciences*, *13*(7), 2137–2149, doi:10.5194/bg-13-2137-2016, 2016.
- Laskin, A., J. Laskin, and S. A. Nizkorodov: Chemistry of Atmospheric Brown Carbon,
   *Chem. Rev.*, *115*(10), 4335–4382, doi:10.1021/cr5006167, 2015.
- Lasslop, G., K. Thonicke, and S. Kloster: SPITFIRE within the MPI Earth system model:
  Model development and evaluation, *J. Adv. Model. Earth Syst.*, *6*, 740–755,
- 1001 doi:10.1002/2013MS000284.Received, 2014.
- Lasslop, G., A. I. Coppola, A. Voulgarakis, C. Yue, and S. Veraverbeke: Influence of
  Fire on the Carbon Cycle and Climate, *Curr. Clim. Chang. Reports*,
- 1004 doi:10.1007/s40641-019-00128-9, 2019.
- Lelieveld, J., J. S. Evans, M. Fnais, D. Giannadaki, and A. Pozzer: The contribution of
  outdoor air pollution sources to premature mortality on a global scale., *Nature*,
- 1007 525(7569), 367–71, doi:10.1038/nature15371, 2015.
- Levy, R. C., S. Mattoo, L. A. Munchak, L. A. Remer, A. M. Sayer, F. Patadia, and N. C.
  Hsu: The Collection 6 MODIS aerosol products over land and ocean, *Atmos. Meas.*

<sup>979 20(4), 565–573,</sup> doi:10.1177/0959683609356587, 2010.

- Li, F., X. D. Zeng, and S. Levis: A process-based fire parameterization of intermediate
  complexity in a dynamic global vegetation model, *Biogeosciences*, 9(7), 2761–2780,
  doi:10.5194/bg-9-2761-2012, 2012.
- Lindeskog, M., Arneth, A., Bondeau, A., Waha, K., Seaquist, J., Olin, S., and Smith, B.:
  Implications of accounting for land use in simulations of ecosystem carbon cycling
- 1016 in Africa, Earth Syst. Dynam., 4, 385–407, doi:10.5194/esd-4-385-2013, 2013.
- 1017 Mangeon, S., A. Voulgarakis, R. Gilham, A. Harper, S. Sitch, and G. Folberth:
- 1018 INFERNO : a fire and emissions scheme for the UK Met Office's Unified Model,
  1019 2685–2700, doi:10.5194/gmd-9-2685-2016, 2016.
- Mao, J., L. W. Horowitz, V. Naik, S. Fan, J. Liu, and A. M. Fiore: Sensitivity of
  tropospheric oxidants to biomass burning emissions: Implications for radiative
  forcing, Geophys. Res. Lett., 40(2), 1241–1246, doi:10.1002/grl.50210, 2013.
- Marlon , J. R., P. J. Bartlein, D G. Gavin, C. J. Long, R. S. Anderson, C. E. Briles, K. J.
  Brown, D. Colombaroli, D. J. Hallett, M. J. Power, E. A. Scharf, M. K. Walsh:
  Long-term perspective on wildfires in the western USA. Proceedings of the National
- 1026 Academy of Sciences, 109 (9) E535-E543; DOI: 10.1073/pnas.1112839109, 2012.
- 1027 Van Marle, M.J.E., S. Kloster, B.I. Magi, J.R. Marlon, A.-L. Daniau, R.D. Field, A.
- 1028 Arneth, M. Forrest, S. Hantson, N.M. Kehrwald, W. Knorr, G. Lasslop, F. Li, S.
- 1029 Mangeon, C. Yue, J.W. Kaiser, and G.R. van der Werf: Historic global biomass
- 1030 burning emissions for CMIP6 (BB4CMIP) based on merging satellite observations
- 1031 with proxies and fire models (1750-2015). Geosci. Model Dev., 10, 3329-3357,
- 1032 doi:10.5194/gmd-10-3329-2017, 2017.
- Moritz, M. A., E. Batllori, R. A. Bradstock, A. M. Gill, J. Handmer, P. F. Hessburg, J.
  Leonard, S. McCaffrey, D. C. Odion, T. Schoennagel, and A. D. Syphard: Learning
  to coexist with wildfire, *Nature*, *515*(7525), 58–66, doi:10.1038/nature13946, 2014.
- 1036 Murray, L. T.: Lightning NOx and Impacts on Air Quality, *Curr. Pollut. Reports*, (x),
- 1037 doi:10.1007/s40726-016-0031-7, 2016.
- Pan, X., Ichoku, C., Chin, M., Bian, H., Darmenov, A., Colarco, P., Ellison, L., Kucsera,
  T., da Silva, A., Wang, J., Oda, T., and Cui, G. (2020): Six Global Biomass Burning
  Emission Datasets: Inter-comparison and Application in one Global Aerosol Model,

<sup>1010</sup> *Tech.*, *6*(11), 2989–3034, doi:10.5194/amt-6-2989-2013, 2013.

- 1041 *Atmos. Chem. Phys.*, 20, 969-994, https://doi.org/10.5194/acp-20-969-2020.
- M.-A. Parisien, M.A. Moritz: Environmental controls on the distribution of wildfire at
   multiple spatial scales
- 1044 Ecol. Monogr., 79 (2009), pp. 127-154, 10.1890/07-1289.1
- Pechony, O., and D. T. Shindell: Fire parameterization on a global scale, *J. Geophys. Res. Atmos.*, *114*, doi:10.1029/2009JD011927, 2009.
- 1047 Pechony, O., and D. T. Shindell: Driving forces of global wildfires over the past
- 1048 millennium and the forthcoming century, *Proc. Natl. Acad. Sci.*, 107(45), 19167–
- 1049 19170, doi:10.1073/pnas.1003669107, 2010.
- Pechony, O., D. T. Shindell, and G. Faluvegi: Direct top-down estimates of biomass
  burning CO emissions using TES and MOPITT versus bottom-up GFED inventory,
- 1052 J. Geophys. Res. Atmos., 118, 8054–8066, doi:10.1002/jgrd.50624, 2013.
- 1053 Pfeifer, E. M., a. Spessa, and J. O. Kaplan: A model for global biomass burning in
- preindustrial time: LPJ-LMfire (v1.0), *Geosci. Model Dev.*, *6*, 643–685,
  doi:10.5194/gmd-6-643-2013, 2013.
- 1056 Platnick, S., M. D. King, K. G. Meyer, G. Wind, N. Amarasinghe, B. Marchant, G. T.
- 1057 Aronold, Z. ZHANG, P. A. Hubanks, B. Ridgway, J. Riedi: MODIS Cloud Optical
- 1058 Properties: User Guide for the Collection 6/6.1 Level-2 MOD06/MYD06 Product
- and Associated Level-3 Datasets,
- 1060 doi:https://doi.org/10.5067/MODIS/MOD08\_M3.006, 2015.
- 1061 Pongratz, J., C. Reick, T. Raddatz, and M. Claussen: A reconstruction of global
- agricultural areas and land cover for the last millennium, *Global Biogeochem*. *Cycles*, *22*, doi:10.1029/2007GB003153, 2008.
- Price, C., and D. Rind: A Simple Lightning Parameterization for Calculating Global
  Lightning Distributions, J. Geophys. Res., 97(D9), 9919–9933, 1992.
- Price, C., and D. Rind: What Determines The Cloud-to-Ground Lightning Fraction, *Geophys. Res. Lett.*, 20(6), 463–466, 1993.
- 1068 Rabin, S. S., J. R. Melton, G. Lasslop, D. Bachelet, M. Forrest, and S. Hantson: The Fire
- 1069 Modeling Intercomparison Project (FireMIP), phase 1 : experimental and analytical
- 1070 protocols with detailed model descriptions, , 1175–1197, doi:10.5194/gmd-10-1175-
- 1071 2017, 2017.

- 1072 Radeloff, V. C., David P. H., H. A. Kramera, M. H. Mockrinb, P. M. Alexandrea, A. Bar-
- 1073 Massadac, V. Butsicd, T. J. Hawbakere, S. Martinuzzia, A. D. Syphardf, and S. I.
- Stewart: Rapid growth of the US wildland-urban interface raises wildfire risk, *Proc. Natl. Acad. Sci.*, 201718850, doi:10.1073/pnas.1718850115, 2018.
- 1076 Randerson, J. T., M. V. Thompson, C. M. Malmstrom, C. B. Field, and I. Y. Fung:
- 1077 Substrate limitations for heterotrophs: Implications for models that estimate the 1078 seasonal cycle of atmospheric CO 2, *Global Biogeochem. Cycles*, *10*(4), 585–602,
- 1079 doi:10.1029/96GB01981, 1996.
- 1080 Randerson, J. T., Y. Chen, G. R. Van Der Werf, B. M. Rogers, and D. C. Morton: Global
- burned area and biomass burning emissions from small fires, *J. Geophys. Res. Biogeosciences*, *117*(4), doi:10.1029/2012JG002128, 2012.
- 1083 Rayner, N. A., D. E. Parker, E. B. Horton, C. K. Folland, L. V. Alexander, D. P. Rowell,
- 1084 E. C. Kent, and A. Kaplan: Global analyses of sea surface temperature, sea ice, and
- 1085 night marine air temperature since the late nineteenth century, J. Geophys. Res.,
- 1086 *108*(D14), 4407, doi:10.1029/2002JD002670, 2003.
- 1087 Remer, L. A., Y. J. Kaufman, D. Tanré, S. Mattoo, D. A. Chu, J. V. Martins,
- 1088 R.-R. Li, C. Ichoku, R. C. Levy, R. G. Kleidman, T. F. Eck, E. Vermote, and B. N.
- Holben: The MODIS Aerosol Algorithm, Products, and Validation, *J. Atmos. Sci.*,
  62(4), 947–973, doi:10.1175/JAS3385.1, 2005.
- Ryan, K. C., E. E. Knapp, and J. M. Varner: Prescribed fire in North American forests
  and woodlands: History, current practice, and challenges, *Front. Ecol. Environ.*, *11*(SUPPL. 1), doi:10.1890/120329, 2013.
- 1094 Schmidt, G.A., M. Kelley, L. Nazarenko, R. Ruedy, G.L. Russell, I. Aleinov, M. Bauer,
- 1095 S.E. Bauer, M.K. Bhat, R. Bleck, V. Canuto, Y.-H. Chen, Y. Cheng, T.L. Clune, A.
- 1096 Del Genio, R. de Fainchtein, G. Faluvegi, J.E. Hansen, R.J. Healy, N.Y. Kiang, D.
- 1097 Koch, A.A. Lacis, A.N. LeGrande, J. Lerner, K.K. Lo, E.E. Matthews, S. Menon,
- 1098 R.L. Miller, V. Oinas, A.O. Oloso, J.P. Perlwitz, M.J. Puma, W.M. Putman, D. Rind,
- 1099 A. Romanou, M. Sato, D.T. Shindell, S. Sun, R.A. Syed, N. Tausnev, K. Tsigaridis,
- 1100 N. Unger, A. Voulgarakis, M.-S. Yao, and J. Zhang: Configuration and assessment
- 1101 of the GISS ModelE2 contributions to the CMIP5 archive. J. Adv. Model. Earth
- 1102 Syst., 6, no. 1, 141-184, doi:10.1002/2013MS000265, 2014.
| 1103 | Schoennagel, T., T. T. Veblen, and W. H. Romme: The Interaction of Fire, Fuels, and           |
|------|-----------------------------------------------------------------------------------------------|
| 1104 | Climate across Rocky Mountain Forests, Bioscience, 54(JULY), 393-402,                         |
| 1105 | doi:10.1641/0006-3568(2004)054, 2004.                                                         |
| 1106 | Schultz, M. G., A. Heil, J. J. Hoelzemann, A. Spessa, K. Thonicke, J. G. Goldammer, A.        |
| 1107 | C. Held, J. M. C. Pereira, and M. van het Bolscher: Global wildland fire emissions            |
| 1108 | from 1960 to 2000, Global Biogeochem. Cycles, 22(2), doi:10.1029/2007GB003031,                |
| 1109 | 2008.                                                                                         |
| 1110 | Scott, A. C., and I. J. Glasspool: The diversification of Paleozoic fire systems and          |
| 1111 | fluctuations in atmospheric oxygen concentration, Proc. Natl. Acad. Sci., 103(29),            |
| 1112 | 10861–10865, doi:10.1073/pnas.0604090103, 2006.                                               |
| 1113 | Seager, R., A. Hooks, A. P. Williams, B. Cook, J. Nakamura, and N. Henderson:                 |
| 1114 | Climatology, variability, and trends in the U.S. Vapor pressure deficit, an important         |
| 1115 | fire-related meteorological quantity, J. Appl. Meteorol. Climatol., 54(6), 1121-1141,         |
| 1116 | doi:10.1175/JAMC-D-14-0321.1, 2015.                                                           |
| 1117 | Seiler, W., and P. J. Crutzen: Estimates of gross and net fluxes of carbon between the        |
| 1118 | biosphere and the atmosphere from biomass burning, Clim. Change, 2, 207–247,                  |
| 1119 | doi:10.1007/BF00137988, 1980.                                                                 |
| 1120 | Sheehan, T., Bachelet, D., and Ferschweiler, K.: Projected major fire and vegetation          |
| 1121 | changes in the Pacific Northwest of the conterminous United States under selected             |
| 1122 | CMIP5 climate futures, Ecol. Model., 317, 16–29,                                              |
| 1123 | doi:10.1016/j.ecolmodel.2015.08.023, 2015.                                                    |
| 1124 | Simard, M., N. Pinto, J. B. Fisher, and A. Baccini: Mapping forest canopy height globally     |
| 1125 | with spaceborne lidar, J. Geophys. Res. Biogeosciences, 116, 1-12,                            |
| 1126 | doi:10.1029/2011JG001708, 2011.                                                               |
| 1127 | Smith, B., Prentice, I. C., and Sykes, M. T.: Representation of vegetation dynamics in the    |
| 1128 | modelling of terrestrial ecosystems: comparing two contrasting approaches within              |
| 1129 | European climate space, Global Ecol. Biogeogr., 10, 621-637, doi:10.1046/j.1466-              |
| 1130 | 822X.2001.t01-1-00256.x, 2001.                                                                |
| 1131 | Smith, B., Wårlind, D., Arneth, A., Hickler, T., Leadley, P., Silt- berg, J., and Zaehle, S.: |
| 1132 | Implications of incorporating N cycling and N limitations on primary production in            |
| 1133 | an individual based dynamic vegetation model, Biogeosciences, 11, 2027–2054,                  |
|      |                                                                                               |

1134

doi:10.5194/bg-11-2027-2014, 2014.

- Thonicke, K., S. Venevsky, S. Sitch, and W. Cramer: The role of fire disturbance for
  global vegetation dynamics: coupling fire into a Dynamic Global Vegetation Model, *Glob. Ecol. Biogeogr.*, 10, 661–677, doi:10.1046/j.1466-822X.2001.00175.x, 2001.
- 1138 Tian, Y., C. E. Woodcock, Y. Wang, J. L. Privette, N. V. Shabanov, L. Zhou, Y. Zhang,
- 1139 W. Buermann, J. Dong, B. Veikkanen, Tuomas Häme, K. Andersson, M. Ozdogan,
- 1140 Y. Knyazikhin, R. B. Myneni: Multiscale analysis and validation of the MODIS LAI
- 1141 product I. Uncertainty assessment, *Remote Sens. Environ.*, *83*, 414–430,

1142 doi:10.1016/S0034-4257(02)00047-0, 2002a.

- Tian, Y., C. E. Woodcock, Y. Wang, J. L. Privette, N. V. Shabanov, L. Zhou, Y. Zhang,
  W. Buermann, J. Dong, B. Veikkanen, Tuomas Häme, K. Andersson, M. Ozdogan,
- 1145 Y. Knyazikhin, R. B. Myneni: Multiscale analysis and validation of the MODIS LAI
- 1146 product II. Sampling strategy, *Remote Sens. Environ.*, 83, 431–441,
- 1147 doi:10.1016/S0034-4257(02)00058-5, 2002b.
- 1148Tosca, M. G., D. J. Diner, M. J. Garay, and O. V. Kalashnikova: Human-caused fires1149limit convection in tropical Africa: First temporal observations and attribution,

1150 Geophys. Res. Lett., 42(15), 6492–6501, doi:10.1002/2015GL065063, 2015.

- Venevsky, S., K. Thonicke, S. Sitch, and W. Cramer: Simulating fire regimes in humandominated ecosystems: Iberian Peninsula case study, *Glob. Chang. Biol.*, *8*, 984–998,
  doi:10.1046/j.1365-2486.2002.00528.x, 2002.
- 1154 Voulgarakis, A., and R. D. Field: Fire Influences on Atmospheric Composition, Air
  1155 Quality and Climate, *Curr. Pollut. Reports*, 1(2), 70–81, doi:10.1007/s40726-0151156 0007-z, 2015.
- 1157 van Wagner, C. E.: A simple fire-growth model, *For. Chron.*, 45(2), 103–104,
  1158 doi:10.5558/tfc45104-2, 1969.
- Ward, D. S., S. Kloster, N. M. Mahowald, B. M. Rogers, J. T. Randerson, and P. G. Hess:
  The changing radiative forcing of fires: global model estimates for past, present and
  future, *Atmos. Chem. Phys.*, *12*(22), 10857–10886, doi:10.5194/acp-12-10857-2012,
  2012.
- Wei, J., Z. Li, Y. Peng, and L. Sun: MODIS Collection 6.1 aerosol optical depth products
  over land and ocean: validation and comparison, *Atmos. Environ.*, 201(October

1165	2018), 428–440, doi:10.1016/j.atmosenv.2018.12.004, 2019.

- van der Werf, G. R.: Continental-Scale Partitioning of Fire Emissions During the 1997 to
  2001 El Nino/La Nina Period, *Science*, *303*(5654), 73–76,
- 1168 doi:10.1126/science.1090753, 2004.
- 1169 van der Werf, G. R., J. T. Randerson, L. Giglio, G. J. Collatz, P. S. Kasibhatla, and a. F.
- Arellano: Interannual variability of global biomass burning emissions from 1997 to
  2004, *Atmos. Chem. Phys.* 6, 3423-3441, https://doi.org/10.5194/acp-6-3423-2006,
  2006.
- 1173 van der Werf, G. R., J. T. Randerson, L. Giglio, G. J. Collatz, M. Mu, P. S. Kasibhatla, D.
- 1174C. Morton, R. S. DeFries, Y. Jin, and T. T. van Leeuwen: Global fire emissions and1175the contribution of deforestation, savanna, forest, agricultural, and peat fires (1997–
- 1176 2009), Atmos. Chem. Phys., 10(23), 11707–11735, doi:10.5194/acp-10-11707-2010,
  1177 2010.
- 1178 van der Werf, G. R., J. T. Randerson, L. Giglio, T. T. van Leeuwen, Y. Chen, B. M.
- Rogers, M. Mu, M. J. E. van Marle, D. C. Morton, G. J. Collatz, R. J. Yokelson, and
  P. S. Kasibhatla: Global fire emissions estimates during 1997–2016, *Earth Syst. Sci. Data*, 9(2), 697–720, doi:10.5194/essd-9-697-2017, 2017.
- Whitburn, S., M. Van Damme, L. Clarisse, S. Turquety, C. Clerbaux, and P. Coheur:
  Doubling of annual ammonia emissions from the peat fires in Indonesia during the
- 1184 2015 El Niño, , doi:10.1002/2016GL070620.Received, 2016.
- 1185 Williams, A. P., R. Seager, A. K. Macalady, M. Berkelhammer, M. A. Crimmins, T. W.
- 1186 Swetnam, A. T. Trugman, N. Buenning, D. Noone, N. G. McDowell, N. Hryniw, C.
- 1187 I. Mora, and T. Rahn: Correlations between components of the water balance and
- burned area reveal new insights for predicting forest fire area in the southwest
- 1189 United States, Int. J. Wildl. Fire, 24(1), 14, doi:10.1071/WF14023, 2015.
- Wooster, M. J., and Y. H. Zhang: Boreal forest fires burn less intensely in Russia than in
  North America, *Geophys. Res. Lett.*, *31*(20), 2–4, doi:10.1029/2004GL020805, 2004.
- 1192 Zou, Y., Y. Wang, Z. Ke, H. Tian, J. Yang, and Y. Liu: Development of a REgion-
- 1193 Specific Ecosystem Feedback Fire (RESFire) Model in the Community Earth
- 1194 System Model, J. Adv. Model. Earth Syst., 11(2), 417–445,
- doi:10.1029/2018MS001368, 2019.

## 1196 Tables

1197 Table 1 Fire emission factors for the different plant functional types (PFTs) in ModelE2.1.

1198 Factors are in units of kg per fire per PFT in the grid cell. For organic and black carbon1199 units kg is substituted with kg of carbon.

PFT	СО	NO <sub>x</sub>	$SO_2$	NH <sub>3</sub>	Alkenes	Paraffin	OC	BC
Cold Broadleaf	113392	1529	555	2101	106	69.8	3437	767
Deciduous	481485	1559	4168	10722	422	373	26752	1844
Needle leaf	401403	1339	4108	10722	422	575	36753	1044
Drought	230829	4835	1687	2340	214	108	10667	1382
Broadleaf	230829			2340				
Evergreen	249906	4905	1438	2847	220	102	10941	1434
Broadleaf	247700							
Evergreen	146622	1197	972	2277	137	89.1	6537	821
Needle leaf	140022			2211				
Cold Shrub	105936	241	878	2006	104	72.1	6562	357
Arid Shrub	39268	1009	262	378	36.6	18.5	1479	238
C3 Annual	26761	690	147	313	25.1	13.9	728	173
Grass	20701							
C3 Arctic	251702	1094	2315	5065	489	226	15551	1159
Grass	231702			5005	407			
C3 Perennial	41043	908	270	438	38.8	20.7	1504	257
Grass	71073	900	270	430	30.0	20.7	1304	231
C4 Grass	117577	3152	795	1196	110	57	4339	726

1208 Table 2: Total fire emissions and global mean column loads of fire emitted species.

1209	Modeled annual	emissions and	l column i	load means	are based	on an ensemble of	10

Species	Variable	pyrE	GFED4s	Bias [%]
СО	Emissions [Tg a <sup>-1</sup> ]	2.14E+02	3.51E+02	-39
CO	Column Load [kg m <sup>-2</sup> ]	7.22E-04	7.71E-04	-6.3
OA	Emissions [TgC a <sup>-1</sup> ]	1.31E+01	2.29E+01	-42
	Column Load [kg m <sup>-2</sup> ]	8.52E-07	1.02E-06	-16
BC	Emissions [TgC a <sup>-1</sup> ]	1.25E+00	1.84E+00	-32
	Column Load [kg m <sup>-2</sup> ]	7.25E-09	7.62E-09	-4.8
NO <sub>x</sub>	Emissions [Tg a <sup>-1</sup> ]	4.27E+00	6.76E+00	-36
	Column Load [kg m <sup>-2</sup> ]	5.94E-07	5.91E-07	0.5
NH <sub>3</sub>	Emissions [Tg a <sup>-1</sup> ]	2.43E+00	4.15E+00	-41
	Column Load [kg m <sup>-2</sup> ]	2.15E-07	2.23E-07	-3.5
$SO_2$	Emissions [Tg a <sup>-1</sup> ]	1.34E+00	2.25E+00	-40
	Column Load [kg m <sup>-2</sup> ]	2.67E-06	2.69E-06	-0.7
Alkenes	Emissions [Tg a <sup>-1</sup> ]	1.94E-01	3.18E-01	-39
	Column Load [kg m <sup>-2</sup> ]	5.73E-08	5.70E-08	0.5
D ()	Emissions [Tg a <sup>-1</sup> ]	9.79E-02	1.65E-01	-40
Paraffin	Column Load [kg m <sup>-2</sup> ]	2.36E-07	2.42E-07	-2.4

1210 simulations. GFED4s emissions are based on a 2000-2010 climatological mean.

## 1221 FIGURES



NHAF Northern Hemisphere Africa
SHAF Southern Hemisphere Africa
BOAS Boreal Asia
CEAS Central Asia
SEAS Southeast Asia
EQAS Equatorial Asia
AUST Australia and New Zealand

1224 Figure 1. GFED basis regions regrided to the resolution of ModelE2.1 of 2° in latitude by

## $2.5^{\circ}$ in longitude.



1237 Figure 2. Structure of the fire parameterization of pyrE. Processes related to atmospheric

1238 properties in blue, surface properties in green, ignition and suppression in yellow and

- 1239 gray, and fire properties in red.



1260 Figure 3. Approximation of a single fire spread. Based on van Wagner (1969) and Arora

- *and Boer* (2005).



1263

Figure 4: Seasonality of total active fires for NHAF (a), SHAF (b), TENA (c) and MIDE (d) observed by MODIS Aqua (red) and Terra (orange) and simulated with explicit regional suppression (blue) and generic global suppression parameterization (green); Eq. 6. Error bars represent the range over 10-year climatological simulations. Note that TERRA and AQUA have different overpass times, and the model data presented here are monthly means. Also, note the different scale in each panel.



Figure 5: Daily mean cycle in active fires (FC, blue line) and daily mean (black line) at 4
locations (Russia (a), India (b), Brazil (c), Nigeria (d)) during the month of January. The
daytime overpass times of Terra (10:30am) and Aqua (13:30pm) are marked with a red
star. Error bars represent the range during the month. Note the different scale in each
panel.





Figure 6: Global seasonality of total active fires (FC) by MODIS Aqua (red) and Terra (orange) and simulated by the model: monthly mean (blue), monthly mean sampled at the daytime Terra overpass time (green), and sampled at the daytime Aqua overpass time (purple). Error bars represent the 10-year range in the simulation.

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Figure 7: Annual mean model (left) and MODIS (right) active fires. Modeled annual mean is based on an ensemble of 10 simulations. Simulated fires sampled at the daytime Terra overpass time, 10:30am local time (a) and daytime Aqua overpass time, 1:30pm local time (c). MODIS active firs are based on MODIS Terra (b) and MODIS Aqua (d)

1294 from 2003-2016.



1295

Figure 8: Annual mean model (left) and satellite based (right) active fires (a, b), burned area (c, d), and CO emissions (e, f). Modeled annual mean is based on an ensemble of 10 simulations. Satellite detected active fires are based on MODIS Aqua retrievals of 2003-2016, burned area is based on GFED4s inventory of 2003-2016, and CO emissions are based on climatological GFED4s emissions of 2000-2010.



Figure 9: Modeled annual mean column density using pyrE fire emissions (left), and the
difference in column densities with a simulation using offline GFED4s emissions (pyrE –
GFED4s; right). CO (a, b), OA (c, d), and BC (e, f). Data based on an ensemble of 10
simulations.





Figure 10: Monthly modeled clear-sky aerosol optical depth (AOD) simulated using pyrE
fire emissions (left), and detected by Aqua-MODIS (right). January (a, b), April (c, d),
July (e, f), and October (g, h). Monthly mean simulated AOD is based on an ensemble of
10 simulations, and climatologically monthly MODIS AOD is based on 2003-2007 data.
Missing MODIS data is shaded in light gray.





Figure 11: The difference in monthly modeled clear-sky aerosol optical depth (AOD) and MODIS Aqua (model – satellite). Model simulations using pyrE fire emissions (left) and model simulations using offline GFED4s emissions (right). January (a, b), April (c, d), July (e, f), and October (g, h). The difference is based on an ensemble of 10 simulations and 2003-2007 MODIS climatological monthly data. Missing MODIS data is shaded in light gray.



Figure 12: The difference in annual modeled clear-sky aerosol optical depth (AOD) between a simulation with no fire emissions to a simulation using pyrE fire emissions (a), and a simulation with offline GFED4s emissions (b). The difference (model with no fire emissions – model with fire emissions) is based on an ensemble of 10 simulations.

1341 APPENDIX



Figure A1: Seasonality of total active fires (FC) detected by MODIS Aqua (red) and
Terra (orange) and simulated (blue) in all GFED regions (Fig. 1). Error bars represent the
10-year range in the simulations. Note the different scale in each panel.





Figure A2: Seasonality of total burned area; simulated (blue) and reported by GFED4s
(red) in GFED regions. Error bars represent the 10-year range in the simulations. Note the
different scale in each panel.





Figure A3: Seasonality of total fire CO emissions; simulated (blue) and reported by
GFED4s (red) in GFED regions. Error bars represent the 10-year range in the simulations.
Note the different scale in each panel.



1354

Figure A4: Seasonality of total fire organic aerosol (OA) emissions; simulated (blue) and reported by GFED4s (red) in all GFED regions. Error bars represent the 10-year range in the simulations. Note the different scale in each panel.



1358

Figure A5: Seasonality of total fire BC emissions; simulated (blue) and reported by GFED4s (red) in all GFED regions. Error bars represent the 10-year range in the simulations. Note the different scale in each panel.