

Response to comments on “GOLUM-CNP v1.0: a data-driven modeling of carbon, nitrogen and phosphorus cycles in major terrestrial biomes” by Y. Wang et al.

We thank the referee for reviewing our manuscript. Please find attached a point-by point reply to each of the comments raised by the referee with legible text and figures organized along the text.

All the main points have been adequately addressed, though some smaller issues remain.

Response:

We would like to thank the referee for the valuable comments and suggestions for improving our manuscript. Following the reviewer’s comments, we carefully revised our manuscript. Please find below the point-to-point responses (in black) to all referee comments (in blue). For your convenience, changes in the revised manuscript are highlighted with dark red. All the pages and line numbers correspond to the original version of text.

The authors have significantly improved the explanation of what GOLUM is. However, it is still not sufficiently clear. On pages 2/3 where it is explained, no-where does it say that GOLUM is an offline component of CARDOMON - which is what it is if, as it seems from the description, it has to be run with CARDOMON output. Being an offline component is fine, you don't have to hide what the model is.

Response:

The GOLUM-CNP model is not an offline component of CARDAMOM. The GOLUM-CNP has a similar model structure as CARDAMOM C cycle for describing the different pools in the ecosystems (leaf, fine roots, wood, litter and SOM). In this study, GOLUM-CNP used CARDAMOM output because CARDAMOM is the only global data-driven C model that combined multiple observations of C pools and fluxes (satellite observations of leaf area and biomass, and global soil carbon data). GOLUM-CNP can also use the outputs from other observation-based C models if they have similar model structures. In this sense, GOLUM-CNP is not an offline component of any model, but an independent diagnostic model that combines multiple observation-based datasets.

The commenting of the code is much improved. The file names, "anal" etc. seem to be an abbreviation of "analysis". I suggest you choose another abbreviation, as the current one is a bit tasteless.

Response:

Thanks for the suggestion. We use `Globe_SteadyState.py` and `SteadyState_pixel.py` instead.

Page 3, line 4 The phrasing here suggests that GOLUM is providing "a new dataset", as if it were observations. It's not. It's a model. The second 'goal' is not achievable and it is misleading

to say this.

Response:

We change the word “dataset” to “observation-based estimates”.

We think the second goal “provide new observation-based estimates that can be used to evaluate the results of global terrestrial biosphere models with consistent state of C, N and P cycles” is feasible because other DGVMs can simulate the steady-state under present day conditions. For example, Wang et al. (2010) have shown their estimates of steady-state pool sizes and fluxes for 1990’s; CLM4CN has a specific equilibrium simulation with vegetation, CO₂, aerosol and nitrogen deposition data for the year 2000. Direct comparison between the results from these model and those from GOLUM-CNP is thus possible. In addition, these models were usually validated against some in-situ observations or subsets of some variables in the C, N and P cycles but were quite hard to evaluate *consistently* against multiple datasets with global coverage. In this sense, this study provides new estimates of all the pool sizes and fluxes that are consistent with available datasets. Considering these two aspects, we would like to highlight that this is the most important asset of this study.

The authors justify their approach by saying that Cleveland et al. (2013) used the same technique, but they use a different nomenclature (Ix rather than NPPnew). Introducing different nomenclature for the same concept is deeply unhelpful for the community and disrespectful of prior research. I understand that the authors will now be reluctant to change their nomenclature, though I think that is what they ought to do. At the very least, a brief note around page 7, line 13, saying that this is equivalent to Cleveland's NPPnew, is essential.

Response:

In Cleveland et al. (2013), they defined “the new terrestrial NPP” as the ratio between “the NPP fueled by new nutrient inputs” and “the total NPP nutrient demand”. Their definition is from the perspective of C cycle (NPP). In this study, we defined the “openness” from the perspective of nutrient cycles, as “the ratio of nutrients inputs (Ix, $X \in \{N,P\}$) over the amount of nutrients in production”.

In addition, although the general principle that the ratio between the “new” fluxes and “total” fluxes are the same in Cleveland et al. (2013) and in this study, the practical computations are not (see below the changes in the supplementary for detail). In Cleveland et al. (2013), they had C_j (NPP allocation fraction in their study) in the parentheses of their Eq. S7, which means the allocation fractions of new N in different vegetation pools are the same as the C allocation fractions. This assumption is neither right nor wrong, but lack evidences. In this study, we define the openness only based on the nutrient fluxes (equilibrium states) rather than converting them to C fluxes like what Cleveland et al. (2013) did. As long as indexes are precisely defined as they are in each study, one can be calculated from the other and this is not a source of confusion.

Our definition of the “openness” is thus not the same index than the “new NPP” of Cleveland et al. (2013). To make this point clearer, we revised our manuscript:

- Page 7 line 15: “... We calculated some ecologically relevant quantities from the GOLUM-CNP output. ~~Following Cleveland et al. (2013), w~~We defined the openness of N and P cycles as NO and PO that were calculated as the ratio of nutrients inputs (Ix, $X \in \{N,P\}$), ...”

- We add in page 7 line 20: “The openness quantifies how much nutrients is from external inputs, which is similar but not strictly equal to the “proportion of new NPP fueled by new nutrient inputs”. In general, the “openness” used this study and the “proportion of new NPP fueled by new nutrient inputs” in Cleveland et al. (2013) both quantify the ratios between fluxes that are related to external inputs and the “total” fluxes, but “openness” used in this study was defined from nutrient cycles, while the index used in Cleveland et al. (2013) was defined from NPP-carbon. In addition, the practical computation of the openness in this study are slightly different from that of Cleveland et al. (2013), which was quantitatively compared in the supplementary material.”
- In the supplementary, we add a paragraph that compares the computation of “openness” used in this study and that of the “proportion of new NPP fueled by new nutrient inputs” used in Cleveland et al. (2013):

“S1 Comparison of the “openness” in this study and in Cleveland et al. (2013)

In this section, we compare the computation of “openness” index used in this study versus “proportion of new NPP fueled by new nutrient inputs” used in Cleveland et al. (2013). We take the indexes for N as an example, but the computation of corresponding indexes for P are similar.

Assume NPP is allocated to leaf, wood and roots by 0.5, 0.3 and 0.2, those pools with C:N ratios of 25, 150 and 50. In Cleveland's index, N is allocated in fractions of 0.5, 0.3 and 0.2 in NPP-N (C_j can be moved before C_{toN_j} in their Eq. S7). The new NPP is $(0.5*25 + 0.3*150 + 0.2*50) * I_N = 67.5 I_N$, where I_N represent external nutrients inputs, i.e. the sum of deposition and biological fixation for N which is available to vegetation. They assume that the remaining NPP must be totally fueled by resorption and net mineralization in the soil. They modelled the amount of N resorption but computed the amount of mineralization as the difference between nutrient demand, new nutrient inputs, and nutrient resorption (their Eqs. S5). So the “proportion of new NPP fueled by new nutrient inputs” index in Cleveland et al. (2013) equals $67.5 I_N/NPP$.

In this study, the N allocation is $0.5/25 : 0.3/150 : 0.2/50$. Normalizing the values to ensure the sum of the fractions equals to 1 gives allocation fractions of N in NPP-N of $0.77 : 0.08 : 0.15$. Under steady state, we have the relationship that the total N demand = $NPP*0.5/25+NPP*0.3/150+NPP*0.2/50=F_N+RSB_N$, where F_N is the uptake from inorganic N soil pool and RSB_N is the flux of resorbed N. As a result, the openness index $NO = I_N / (F_N + RSB_N) = 38.5 I_N/NPP$.

Because there is no evidence that how much the new N inputs is allocated in the vegetation, we chose to define the openness index only based on N and P fluxes rather than to convert N and P fluxes into NPP like Cleveland et al. (2013) did.”

There is no figure "R1" included. I assume the authors mean the figure labelled R2. R2 is very substantially better than Figure 3 in the paper. It's a huge improvement and I suggest the authors consider their readers and use it in the main paper.

Response:

We think we depict the full C, N and P cycles including the pool sizes, residence times of different pools and fluxes with the original flow charts, which is not achievable by bar plots.

All the co-authors, who also have a wide range of expertise (measurements, models, etc.) in the community of global biochemical cycles, prefer the flow charts than the bar plots and we think that the flow charts contain critical information about how the different variables in the C, N and P cycles are linked.

To facilitate the readers, we put both figures in the main text (Figure R2 was put in the text as Figure 4). We re-ordered all the figures and change the texts that refer to different figures.

References:

- Cleveland, C. C., Houlton, B. Z., Smith, W. K., Marklein, A. R., Reed, S. C., Parton, W., Grosso, S. J. D. and Running, S. W.: Patterns of new versus recycled primary production in the terrestrial biosphere, *Proc. Natl. Acad. Sci.*, 110(31), 12733–12737, doi:10.1073/pnas.1302768110, 2013.
- Lawrence, D. M., Oleson, K. W., Flanner, M. G., Thornton, P. E., Swenson, S. C., Lawrence, P. J., Zeng, X., Yang, Z.-L., Levis, S., Sakaguchi, K., Bonan, G. B. and Slater, A. G.: Parameterization improvements and functional and structural advances in Version 4 of the Community Land Model, *Journal of Advances in Modeling Earth Systems*, 3(1), doi:10.1029/2011MS00045, 2011.
- Wang, Y. P., Law, R. M. and Pak, B.: A global model of carbon, nitrogen and phosphorus cycles for the terrestrial biosphere, *Biogeosciences*, 7(7), 2010.

GOLUM-CNP v1.0: a data-driven modeling of carbon, nitrogen and phosphorus cycles in major terrestrial biomes

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Abstract. Global terrestrial nitrogen (N) and phosphorus (P) cycles are coupled to the global carbon (C) cycle for net primary production (NPP), plant C allocation and decomposition of soil organic matter, but N and P have distinct pathways of inputs and losses. Current C-nutrient models exhibit large uncertainties in their estimates of pool sizes, fluxes and turnover rates of nutrients, due to a lack of consistent global data for evaluating the models. In this study, we present a new model-data fusion framework called Global Observation-based Land-ecosystems Utilization Model of Carbon, Nitrogen and Phosphorus (GOLUM-CNP) that combines the CARbon DATA MOdel fraMework (CARDAMOM) data-constrained C-cycle analysis with spatially explicit data-driven estimates of N and P inputs and losses and with observed stoichiometric ratios. We calculated the steady-state N- and P-pool sizes and fluxes globally for large biomes. Our study showed that new N inputs from biological fixation and deposition supplied >20% of total plant uptake in most forest ecosystems but accounted for smaller fractions in boreal forests and grasslands. New P inputs from atmospheric deposition and rock weathering supplied a much smaller fraction of total plant uptake than new N inputs, indicating the importance of internal P recycling within ecosystems to support plant growth. Nutrient-use efficiency, defined as the ratio of gross primary production (GPP) to plant nutrient uptake, were diagnosed from our model results and compared between biomes. Tropical forests had the lowest N-use efficiency and the highest P-use efficiency of the forest biomes. An analysis of sensitivity and uncertainty indicated that the NPP-allocation fractions to leaves, roots and wood contributed the most to the uncertainties in the estimates of nutrient-use efficiencies. Correcting for biases in NPP-allocation fractions produced more plausible gradients of N- and P-use efficiencies from tropical to boreal ecosystems and highlighted the critical role of accurate measurements of C allocation for understanding the N and P cycles.

1 **1 Introduction**

2 Nitrogen (N) and phosphorus (P) cycling are tightly coupled with the global carbon (C) cycle (Cleveland et al., 2013;
3 Elser et al., 2007; Gruber and Galloway, 2008; Ver et al., 1999; Turner et al., 2018) in terrestrial ecosystems. N and P
4 availability affects vegetation productivity, growth and other processes (Norby et al., 2010; Sutton et al., 2008; Vitousek and
5 Howarth, 1991). N and P also affect soil C by nutrient controls on the mineralization of litter and soil organic matter
6 (Gärdenäs et al., 2011; Melillo et al., 2011). Global vegetation models suggest that the coupling between the C, N and P
7 cycles is among the major factors determining projected changes in the terrestrial C balance under scenarios of climate
8 change and rising atmospheric CO₂, because additional productivity will only be realized if plants can increase their uptake
9 or recycling of nutrients (Hungate et al., 2003; Sun et al., 2017; Wang and Houlton, 2009; Zaehle et al., 2015). Estimates of
10 the magnitudes of these responses of ecosystems in the future, however, are highly uncertain (Peñuelas et al., 2013; Wieder
11 et al., 2015).

12 Nutrients are important for understanding the current perturbation and future projections of the global C cycle, so
13 several Dynamic Global Vegetation Models (DGVMs) have incorporated terrestrial N cycling (Goll et al., 2012; Medvigy et
14 al., 2009; Parton et al., 2010; Thornton et al., 2007; Wang et al., 2001, 2010; Weng and Luo, 2008; Xu-Ri and Prentice, 2008;
15 Yang et al., 2009; Zaehle et al., 2014; Zaehle and Friend, 2010). Fewer models have incorporated the cycling of P and its
16 interactions with C dynamics (Goll et al., 2012, 2017a; Wang et al., 2010). Many of the underlying processes are not fully
17 understood, and comprehensive data for evaluation are lacking to constrain the representation of some key processes (Zaehle
18 et al., 2014), so model structure, the processes included and the prescribed parameters differ widely among DGVMs (Zaehle
19 and Dalmonech, 2011). For example, some models assume constant stoichiometry (N:C and P:C ratios) in plant tissues
20 (Thornton et al., 2007; Weng and Luo, 2008), while others have a flexible stoichiometry (Wang et al., 2010; Xu-Ri and
21 Prentice, 2008; Yang et al., 2009; Zaehle and Friend, 2010). For the N cycle, for instance, some models do not include losses
22 of gaseous N from denitrification (Medvigy et al., 2009), some use the “hole-in-the-pipe” approach to simulate the
23 denitrification flux (Thornton et al., 2007; Wang et al., 2010), assuming it is proportional to net N mineralization, while
24 others calculate this flux as a function of soil N-pool size and soil conditions (temperature, moisture, pH, etc.) (Parton et al.,
25 2010; Xu-Ri and Prentice, 2008; Zaehle and Friend, 2010). For the P cycle, for instance, Jahnke (2000) estimated that the
26 global total amount of soil P was 200 Pg and that the P contained in plants was 3 Pg, based on empirical P content of soils
27 (0.1%) and soil thickness (60 cm). These estimates were questioned by recent studies from Wang et al. (2010) and Goll et al.
28 (2012), who estimated that P in plants ranged between 0.23 and 0.39 Pg and that P in soil was only 26.5 Pg based on P:C
29 ratios derived from more comprehensive stoichiometric data sets. Furthermore, terrestrial ecosystem models are usually only
30 evaluated for specific ecosystems or at a limited number of sites (Goll et al., 2017a; Yang et al., 2014a). The application of
31 these models for simulations with global coverage is thus highly uncertain (Goll et al., 2012; Wang et al., 2010; Zhang et al.,
32 2011).

33 A growing number of data sets in recent decades have addressed many aspects of the nutrient cycles and their
34 interactions with C dynamics. For example, Zechmeister-Boltenstern et al. (2015) synthesized the stoichiometry in different
35 ecosystem compartments and highlighted the latitudinal gradients of plant, litter and soil stoichiometry. Liu et al. (2017)
36 evaluated soil net N mineralization among different ecosystems at the global scale and found that net N mineralization
37 decreased with increasing latitude. They also found that the N mineralization at higher latitudes is more sensitive to
38 temperature changes than at lower latitudes, indicating potential alleviation of N limitation for plants' productivity at boreal
39 regions under global warming. Yang et al. (2013) provided spatially explicit estimates of different forms of soil P globally
40 and thus made it possible to assess the P content that is available for plant uptake. These data help to improve the
41 understanding of the global terrestrial biogeochemical cycles across large climatic and ecological gradients and can in
42 principle be combined to provide an integrated analysis of terrestrial C, N and P cycles. Estimates of C, N and P cycles
43 consistent with all these data sets, however, have not yet been successfully provided due to the difficulties in combining

1 these data sets with different uncertainties and inconsistent spatial/temporal representations.

2 We present a new global data-driven diagnostic of C, N and P pools and fluxes, called GOLUM-CNP (Global
3 Observation-based Land-ecosystems Utilization Model of Carbon, Nitrogen and Phosphorus) which is based on the
4 assumption that these cycles are equilibrated with present day conditions (see below for limitations of this approach). The
5 goals of this study are to: 1) establish a global data-driven diagnostics of C, N and P fluxes and pools in order to compare
6 nutrient use efficiencies, nutrient turnover rates and other relevant indicators across biomes; and 2) provide a new dataset
7 that can be used to evaluate the results of global terrestrial biosphere models with consistent state of C, N and P cycles. In
8 GOLUM-CNP, the C, N and P cycles were estimated for different biomes assuming steady state with present-day input of
9 carbon (NPP), nitrogen (N deposition and N fixation) and phosphorus (P deposition and release from rock weathering) (see
10 Sect. 3.2). The reason for this steady-state computation lies in the fact that only few global long-term observations associated
11 with N and P cycles are available and are insufficient to constrain a transient simulation under the model framework. For
12 example, field-scale manipulation experiments have shown that warming, elevated atmospheric CO₂, and N and P
13 fertilization can drive changes in stoichiometry and nutrient resorptions (Sistla and Schimel, 2012; Sardans et al., 2012;
14 Sardans and Peñuelas, 2012; Mayor et al., 2014; Yang et al., 2014b; Yuan and Chen, 2015; Sardans et al., 2016; Sardans et
15 al., 2017) in terrestrial ecosystems, but these data are insufficient to infer these changes in terrestrial ecosystems during the
16 past decades. As more data becomes available, the model framework can be adjusted to simulate a transient present day state.
17 Although, the steady-state assumption hampers the comparison of stocks with present day observations, a direct comparison
18 with simulated steady states of DGVM is possible as these model can simulate the steady-state for present day conditions.

19 Starting from a CARbon DAta MOdel fraMework (CARDAMOM) data-constrained analysis of the terrestrial C cycle
20 (Bloom et al., 2016), which is based on the Data Assimilation Linked Ecosystem Carbon Model version two (DALEC2,
21 Bloom and Williams, 2015; Williams et al., 2005) and on observations of biomass, soil C, leaf area index (LAI) and fire
22 emissions, we incorporated observed stoichiometric ratios (C:N:P) in each pool, N and P external input fluxes,
23 transformations and losses in ecosystems and the fraction of gaseous losses of N to total (gaseous and leaching) losses of N
24 from a global dataset of ¹⁵N measurements in soils. Although the diagnostics is presented for steady state, the methods used
25 to compute fluxes and pools are generic and could be extended to non-steady state (see Sect. 2 and equations in Appendix A-
26 C) when more data will become available in the future (see Sect. 5.3).

27 We first present the model structure (Sect. 2) and the data sets used to derive its outputs consisting of pools, fluxes and
28 turnovers of C, N and P (Sect. 3). The model results and their sensitivities to the input observation-based data sets are then
29 further analyzed in Sect. 4. In Sect. 5, we show examples of the application of this sensitivity analysis to identify the major
30 differences in the results from our model framework and a synthesis of *in situ* measurements, and a qualitative example of
31 how to compare the model and the independent data. These differences identify critical observations to reduce uncertainty in
32 global C and nutrient cycling and highlight the future demand for model development, calibration and evaluation.

33 2 Model structure

34 GOLUM-CNP describes the C, N and P cycles in natural (i.e. non-agricultural) terrestrial ecosystems (Fig. 1). We used
35 the same C pools and fluxes as in the CARDAMOM diagnostic (see Sect. 2.1 for details) to describe the C cycle and we
36 computed associated N and P pools and fluxes. Biomass is divided into three pools: foliage, fine roots and wood. The wood
37 pool includes woody stems, branches and coarse roots. The litter pool in Fig. 1 corresponds to fine litter from leaves and fine
38 roots. Soil organic matter (SOM) receives C from fine litter and woody biomass. Two additional pools not present in
39 CARDAMOM are added, representing soil inorganic N and labile soil P. These inorganic N and labile soil P pools are
40 assumed to represent nutrients accessible by plants (see Sect. 2.1 and 2.2). Of note is that these inorganic N and labile soil P
41 pools represent an integration of various forms of N and P. For example, P has various forms in the soil and can be
42 transformed between those forms (Wang et al., 2007; Yang and Post, 2011). Some forms of organic P (e.g. bicarbonate Po in

1 Hedley method, Yang and Post, 2011) can easily be mineralized and thus were implicitly included in our labile soil P pool.
2 Other forms of P that are not easily accessible to plants are referred to as “occluded P” and labile soil P can become occluded
3 P (Wang et al., 2010; Goll et al., 2017a). Fluxes connecting the pools are described by the differential equations given in
4 Appendices A-C. An overview of the C, N and P cycles and their interactions are presented in the following sections. A full
5 list of the symbols and their definitions is given in Table 1.

6 **2.1 C cycle**

7 The C cycle in the GOLUM-CNP model is based on the DALEC2 model (Bloom et al., 2016; Bloom and Williams,
8 2015). We used a similar structure to define the C pools of GOLUM-CNP but grouped the DALEC2 foliar and labile
9 vegetation C pools into a single foliar pool (Fig. 1). Net primary production (NPP) is allocated to the three biomass pools.
10 The outgoing fluxes from biomass pools include losses from fire, the transfer of foliage and root detritus to litter and the
11 transfer of wood debris directly to the SOM pool. The outgoing fluxes from litter include losses from fire and decomposition.
12 A fraction of decomposed litter is respired and returned to the atmosphere as CO₂, the remaining fraction being converted to
13 SOM. The SOM pool loses C by fire and decomposition. Differential equations governing the dynamics of C pools are given
14 in Appendix A.

15 **2.2 N cycle**

16 The N cycle in GOLUM-CNP is coupled to the C cycle: the pool sizes of N are determined by the C-pool sizes and
17 their respective N:C ratios; the N fluxes from different pools are determined by the N-pool sizes and corresponding turnover
18 rates. The N cycle includes a specific soil inorganic-N pool in addition to the five pools of the C cycle. The inputs of N to
19 ecosystems include atmospheric N deposition and N fixation ($N_d + N_{fix}$ in Fig. 1), both of which are assumed to enter the
20 inorganic-N pool. The total N-fixation flux in this study includes both symbiotic and asymbiotic fixation (see Sect. 3.1),
21 separately estimated from a previous study (see Sect. 3.1). We do not separate the two fixation processes and assume that
22 they together contribute to the inorganic-N pool, although these two pathways of N fixation are differed in terms of the
23 relationships between N₂-fixing microorganisms and plants. We did not consider the flux of N mobilized from near-surface
24 rocks, although a recent paper by Houlton et al. (2018) pointed out this flux may be an important N sources in montane and
25 high-latitude ecosystems. N uptake (F_N) by plants is assumed to be solely from the inorganic-N pool. Organic N is an
26 important N supply for plants (Näsholm et al., 2009) in boreal-forest and tundra ecosystems (Schimel and Bennett, 2004;
27 Schimel and Chapin, 1996; Zhu and Zhuang, 2013), but the quantitative importance of this process is still unknown for other
28 ecosystems globally. We thus ignored the uptake of organic soil N. N uptake by plants from the inorganic-N pool is modeled
29 from the N:C ratio of NPP allocated to biomass pools minus the resorbed N. In the real world, N is only resorbed at the end
30 of the growing season or leaf lifespan and then stored in plant organs and remobilized during the next growing season. Here,
31 because our model does not have a sub-annual time step, rates of resorption described by a resorption coefficient (Appendix
32 B) are assumed to be constant over time. We also assumed that N is not resorbed from fine roots or wood, because evidence
33 for this process is inconclusive (Gordon and Jackson, 2000; Zechmeister-Boltenstern et al., 2015). N mineralization is
34 modeled along with litter and SOM decomposition. N immobilization due to the uptake of inorganic soil N by soil organisms
35 is modeled to match the higher N:C ratio of the SOM pool than its donor (wood and litter) pools. Loss of N from ecosystems
36 occurs through fire, denitrification and leaching. The N lost due to fire is assumed to be emitted only in gaseous form,
37 because the proportion of N retained in the residual ash is very small (Niemeyer et al., 2005; Qian et al., 2009). We consider
38 the gaseous loss of inorganic N from denitrification but ignore the volatilization of N in the form of NH₄⁺. This flux usually
39 occurs at a soil pH >8 (Freney et al., 1983) or after application of N fertilizers (Yang et al., 2009), and NH₃ emissions from
40 soils under non-agricultural vegetation are relatively small globally (5 Tg N y⁻¹, Bouwman et al., 1997; Houlton et al., 2015),
41 representing <5% of total gaseous loss, so the omission of NH₄⁺ N volatilization will not introduce large biases in our model

1 for most regions. The dynamics of N in the pools are summarized by the differential equations in Appendix B.

2 **2.3 P cycle**

3 The P cycle, like the N cycle, is also coupled to the C cycle; the dynamics of P in the pools are described by the
4 differential equations in Appendix C. The external inputs of P to ecosystems include atmospheric P deposition and P
5 released from P-bearing minerals by chemical weathering (P_d+P_w in Fig. 1). P from deposition and rock weathering enters
6 the soil inorganic-P pool. The structure of the P cycle is the same as for the N cycle described above for foliar-P resorption,
7 P released from the decomposition of litter and SOM and the immobilization of inorganic soil P by soil organisms. Inorganic
8 P, unlike inorganic N, can be sorbed onto/into soil particles and subsequently become occluded. This form is assumed to be
9 unavailable to plants. We modeled the flux from the labile soil P to occluded P with a constant rate. Loss pathways of P
10 include fire, leaching and conversion to occluded P. Notably, not all P mobilized by fire is emitted in gaseous form, but is
11 partly retained in the residual ash (Niemeyer et al., 2005; Qian et al., 2009). We used a constant fraction of 75% (Niemeyer
12 et al., 2005; Qian et al., 2009) to model the P retained in the residual ash during a fire, and this fraction of P enters the labile
13 soil P pool.

14 **3 Methods**

15 **3.1 Input data sets**

16 All parameters used as inputs for the calibration of GOLUM-CNP are listed in Table 1. A steady state was assumed to
17 infer remaining variables (also listed in Table 1). The estimates of fluxes and C-pool sizes were based on mass balances, and
18 the estimates of N- and P-pool sizes were derived from the C-pool size and stoichiometric data (see below and Appendix E).
19 We used the C fluxes and turnover times of C pools derived from CARDAMOM for the C cycle (Bloom et al., 2016), which
20 offered a data-consistent analysis of terrestrial C cycling on a global $1^\circ \times 1^\circ$ grid for 2001-2010 by optimizing the DALEC2
21 model parameters to match the state and process variables with the global observations of MODIS LAI (Myneni et al., 2015),
22 soil C (Hiederer and Köchy, 2011), burned area (Giglio et al., 2013) and tropical biomass (Saatchi et al., 2011). Although the
23 CARDAMOM data-driven analysis only reported the C pools and fluxes, the impacts of N and P on the C cycle have been
24 implicitly reflected in CARDAMOM through the constraints by some of the observations. For example, the availability of N
25 and/or P limits the growth of vegetation and thus the LAI observed (Klodd et al., 2016; Reich et al., 2010); the N and P
26 contents in soil control the decomposition of soil C and thus the soil C pool observed (Manzoni et al., 2010). In this sense, it
27 is appropriate to use C cycle from CARDAMOM as inputs to estimate the pool and fluxes of N and P.

28 Different indices have been used to describe nutrient cycling from different perspectives (soil, individual plant,
29 vegetation, ecosystem, etc) (Augusto et al., 2017; Cleveland et al., 2013; Gill and Finzi, 2016). In this study, we focused on
30 the openness, nutrient use efficiencies and the residence time (Sect. 3.2) which are defined at ecosystem scale and thus
31 correspond to the scale at which DGVMs are typically defined. For the presentation of results, we distinguish seven biomes:
32 tropical rainforests (TRF), temperate deciduous forests (TEDF), temperate coniferous forests (TECF), boreal coniferous
33 forests (BOCF), tundra (TUN), tropical/C4 grasslands (TRG) and temperate/C3 grasslands (TEG). Note that similar
34 empirical land-cover maps have been also used in previous studies to simulate C, N and P cycles (Cleveland et al., 2013;
35 Wang et al., 2010). We applied observed biome-specific N:C ratios for each pool from the synthesis by Zechmeister-
36 Boltenstern et al. (2015).

37 We used the spatially explicit estimates of N deposition (Wang et al., 2017) for 2001-2010 (Fig. S2a), which were
38 evaluated with globally distributed *in-situ* measurements. The spatially explicit estimate for N fixation (Fig. S2b) was taken
39 from the CABLE model simulation for 2001-2010 (Peng et al., submitted) with a N fixation model developed by Wang et al.
40 (2007). The simulation result matches the relative abundance of N_2 -fixing legumes in different ecosystems. Globally, the N

1 fixation is 116 Tg N yr⁻¹ and is within the range of empirical estimates (100-290 Tg N yr⁻¹; Cleveland et al., 1999; Galloway
2 et al., 2004), but larger than the estimate of 44 Tg N yr⁻¹ by Vitousek et al. (2013) for pre-industrial conditions. The large
3 range (44-290 TgN yr⁻¹) in the estimates of nitrogen fixation reflects both a paucity of measurements of N fixation, as well as
4 incomplete understanding of the biophysical and biochemical controls on N fixation. And to our knowledge, CABLE
5 simulation is the only product that has spatially explicit and processed-based estimates of N fixation, and is therefore used in
6 this study. The resorption coefficients of leaves for the seven biomes were derived from the N:C ratios of leaves and leaf
7 litter reported by Zechmeister-Boltenstern et al. (2015). The rate of loss of inorganic N by leaching was determined from
8 data for total soil moisture and runoff (Eq. B7). The spatially explicit estimate of total soil moisture was derived from the
9 European Centre for Medium-Range Weather Forecasts (ECMWF) Interim Reanalysis (ERA-Interim/Land; Albergel et al.,
10 2013; Balsamo et al., 2015). Gridded soil water content was provided in ERA-Interim/Land in four discretized layers until
11 2.89 m below ground, which is the soil depth we considered for the total soil moisture. Although some uncertainties exist at
12 grid scale, the large-scale patterns in soil moisture from ERA-Interim/Land are consistent with other products (Rötzer et al.,
13 2015), enabling us to use it to represent the large-scale spatial gradients in soil water moisture. The global gridded estimate
14 of runoff data was obtained from the Global Runoff Data Centre (GRDC, <http://www.grdc.sr.unh.edu/>), which is constrained
15 by observed river discharges from 663 stations globally. We used observation-based estimates of the fraction of N lost by
16 denitrification to the total inorganic-N loss (denitrification + leaching) pathways (Goll et al., 2017b) to calibrate the
17 denitrification-loss flux. This fraction of denitrification loss (f_{denit}) was derived using a process-based statistical model fitted
18 to global soil $\delta^{15}\text{N}$ data sets, based on the distinct ¹⁵N fractionation effect of denitrification versus loss from leaching (Bai et
19 al., 2012; Houlton and Bai, 2009).

20 We constrained the P cycle using spatially explicit estimates by Wang et al. (2017) for P deposition for 2001-2010 (Fig.
21 S3a). Spatially explicit estimates of P input from rock weathering (Fig. S3b) were derived from data for river discharge and
22 the chemical composition of minerals by Hartmann et al. (2014). The P:C ratios and resorption coefficients for the seven
23 biomes were obtained from Zechmeister-Boltenstern et al. (2015). Only a fraction of total inorganic P can be lost by leaching,
24 and this fraction of dissolved inorganic P in total labile P was derived based on the observations of Hedley soil P fractions as
25 the resin-extractable P divided by total labile P reported by Yang and Post (2011) for the twelve USDA soil orders. The
26 constant rate at which inorganic P becomes strongly sorbed (f_{sorb} , Eq. C6) was fixed at 0.04 y⁻¹ (Goll et al., 2017a).

27 Zechmeister-Boltenstern et al. (2015) only reported the stoichiometric ratios and the N and P resorption coefficients for
28 seven large non-agricultural biomes, but other input variables were grid-based products. A land-cover map was used to
29 aggregate the grid-based C-cycle variables from Bloom et al. (2016) into the biomes used by Zechmeister-Boltenstern et al.
30 (2015). The land-cover map was derived from the dominant land-cover type for each grid cell for the globe, excluding
31 croplands, from the land-cover map of the Climate Change Initiative (LC_CCI) established by the European Space Agency
32 (ESA) (Bontemps et al., 2013) at 0.25 °×0.25 ° resolution. Specifically, we used the 2010 map to classify all grid cells into
33 one of the seven non-agricultural biomes of Zechmeister-Boltenstern et al. (2015) (Fig. 2), following the methodology
34 presented by Poulter et al. (2015).

35 3.2 Model integration and output diagnostics

36 We applied the model framework described in Sect. 2 to derive a data-driven estimate of steady-state C, N and P
37 cycling. A steady state indicates that annual mean input fluxes for all pools are assumed to be balanced by annual mean
38 outgoing fluxes, with the annual mean outgoing fluxes from organic pools calculated as the quotient of the pool sizes to the
39 corresponding turnover times. Assuming that all pool sizes were in a steady state, the left side of the equations in Appendices
40 A-C (Eqs. A1-A5, B1-B6 and C1-C6) are all equal to zero. Adding the constraints in Appendix D (Eqs. D1-D11), we
41 derived a system with 28 equations and 28 unknown variables (Table 1), thereby defining all unknowns in GOLUM-CNP.
42 The unknown variables were solved by applying the 33 global spatially explicit observation-based estimates listed in Table 1

1 in these equations (Appendix E). The set of equations of the GOLUM-CNP model was solved for each 0.25 °×0.25 ° grid cell
 2 using biome-mean N:C and P:C stoichiometric ratios, grid-cell specific values of C variables from Bloom et al. (2016) and
 3 the gridded external N- and P-input and -output fields described above. In this computation, some processes were only
 4 solved by mass balance and the steady-state assumption instead of explicitly being calibrated. For example, we did not
 5 explicitly simulate various pathways of N and P mineralization and immobilization. The N and P mineralization fluxes are
 6 computed as the product of the decomposition of C in litter and SOM and their respective stoichiometries, and N and P
 7 immobilization fluxes are computed by mass balance to match the higher N:C and P:C ratios in the SOM pool compared to
 8 the ratios in inputs to SOM from wood and litter decomposition. For instance, N and P mineralization were computed as the
 9 difference between nutrient demand of vegetation and the sum of external inputs and resorption in Cleveland et al. (2013,
 10 Eqs. S5 and S6), assuming that the nutrients available to plants in soil do not change significantly at current stage. Some
 11 variables were computed by mass balance and do not rely on steady state assumption, e.g. the uptake of N (F_N) and P (F_P) by
 12 plants (Table 1). Such computations based on mass balance and steady-state assumptions allow us to have a diagnostic
 13 modelling framework, but at the same time capture observations of carbon fluxes, pools and pool stoichiometries.

14 The inputs of the C cycle from the original CARDAMOM dataset were provided as probability distributions, while
 15 other datasets were provided only as mean values. In this study, we compute the GOLUM-CNP using the mean values of all
 16 the input datasets to represent the mean behaviour of the C, N and P cycling.

17 We present the C, N and P pools and the fluxes between them for each biome. We also aggregated the results at the
 18 global scale and compared them with previous studies. We calculated some ecologically relevant quantities from the
 19 GOLUM-CNP output. ~~Following Cleveland et al. (2013), we~~ We defined the *openness* of N and P cycles as NO and PO that
 20 were calculated as the ratio of nutrients inputs (I_x , $X \in \{N,P\}$), taken as the sum of deposition (N_d) and biological fixation
 21 (N_{fix}) for N and as the sum of deposition (P_d) and release from rock weathering (P_w) for P, over the amount of nutrients in
 22 production, taken as the sum of uptake from inorganic N or labile soil P pools (F_x , $X \in \{N,P\}$) and resorbed nutrients (RSB_x ,
 23 $X \in \{N,P\}$), leading to:

$$24 \quad XO = \frac{I_x}{F_x + RSB_x} \quad (1)$$

25 The openness quantifies how much nutrients is from external inputs, which is similar but not strictly equal to the
 26 “proportion of new NPP fueled by new nutrient inputs”. In general, the “openness” used this study and the “proportion of
 27 new NPP fueled by new nutrient inputs” in Cleveland et al. (2013) both quantify the ratios between fluxes that are related to
 28 external inputs and the “total” fluxes, but “openness” used in this study was defined from nutrient cycles, while the index
 29 used in Cleveland et al. (2013) was defined from NPP-carbon. In addition, the practical computation of the openness in this
 30 study are slightly different from that of Cleveland et al. (2013), which was quantitatively compared in the supplementary
 31 material S1.

32 The mean residence time of N and P for the entire ecosystem ($\tau_{X,eco}$, $X \in \{N,P\}$) was defined as the ratio of total
 33 modeled pool mass (including plant, litter, SOM and inorganic pools) to all outgoing fluxes. The sum of all steady-state
 34 outgoing fluxes was set equal to the sum of external input fluxes, so we calculated the mean residence time of N and P by:

$$35 \quad \tau_{N,eco} = \frac{\sum_{i=1}^5 N_i + N_{inorg}}{N_d + N_{fix}} \quad (2)$$

$$\tau_{P,eco} = \frac{\sum_{i=1}^5 P_i + P_{inorg}}{P_d + P_w}$$

The *nutrient-use efficiencies* (NUE and PUE) were defined by:

$$XUE = \frac{GPP}{F_X} = \frac{NPP / f_{NPP}}{F_X} \quad (3)$$

where F_X ($X \in \{N, P\}$) is the annual uptake of inorganic soil N or P by plants, and f_{NPP} is the ratio of NPP to gross primary production (GPP) from CARDAMOM. Our model used NPP as the input C flux for ecosystems, but we used GPP instead of NPP in Eq. (3) to calculate XUE for comparing with the estimates based on *in situ* measurements by Gill and Finzi (2016), which were also based on GPP. We thus used f_{NPP} only as an external variable in our modeling framework, and f_{NPP} was not targeted when evaluating the sensitivities and uncertainties of the results (see below).

We tested the steady-state sensitivity (SS) of the model results to the observational data sets (inputs of the model listed in Table 1) by linearizing the GOLUM-CNP model and its solver for calculating the first-order partial derivative of all outputs relative to each input parameter:

$$SS = \partial \mathbf{O} / \partial \mathbf{I} \quad (4)$$

where \mathbf{I} is the vector of the input variables, and \mathbf{O} is the vector of the output variables. This approach directly provided a sensitivity matrix, which allowed us to test the effect of the accuracy of the measurement of each input variable on the model results for the N and P cycles. This method was similar to the “one-at-a-time” (OAT) approach used for sensitivity analysis in previous C-N coupled modeling studies (Orwin et al., 2011; Shi et al., 2016; Zaehle and Friend, 2010) but did not require running simulations by changing the inputs one at a time. This approach did not fully explore the possible range of values for a given parameter, but provided comparable SS values for different parameters, which is useful when the full uncertainty ranges of some parameters are unknown, e.g. uncertainty due to the inconsistent definitions between the measured pools in the real world and the conceptual pools in the model, or the large uncertainty due to sparse observations for some biomes. The input parameters had distinct magnitudes (and units), so we used the relative sensitivities, e.g. $SS = \partial \mathbf{O} / \mathbf{O} / (\partial \mathbf{I} / \mathbf{I})$, to compare the sensitivities to different model inputs. For the sensitivity analysis, an SS of 1 indicates that a 1% increase (or decrease) in a model input produces a 1% increase (or decrease) in the model output, and an SS of -0.5 indicates that a 1% increase (or decrease) in the model input produces a -0.5% decrease (or increase) in the model output. The results of this sensitivity analysis could be further used to investigate the sources of uncertainty in the outputs and to evaluate variances of the model outputs using error propagation:

$$\begin{aligned} \boldsymbol{\varepsilon}_{\mathbf{O},i} &= \frac{\partial \mathbf{O}}{\partial \mathbf{I}_i} \boldsymbol{\varepsilon}_{\mathbf{I},i} \\ \boldsymbol{\varepsilon}_{\mathbf{O}} &= \sum_{i=1}^n \frac{\partial \mathbf{O}}{\partial \mathbf{I}_i} \boldsymbol{\varepsilon}_{\mathbf{I},i} \\ \Sigma_{\mathbf{O}} &= E(\boldsymbol{\varepsilon}_{\mathbf{O}}^T \boldsymbol{\varepsilon}_{\mathbf{O}}) \end{aligned} \quad (5)$$

where $\boldsymbol{\varepsilon}_{\mathbf{I},i}$ is the error in the i th input data, $\boldsymbol{\varepsilon}_{\mathbf{O},i}$ is the error propagated from the error in input i , $\boldsymbol{\varepsilon}_{\mathbf{O}}$ is the error that accounts for errors in all input data, E represents the expectation of a variable and $\boldsymbol{\varepsilon}_{\mathbf{O}}$ is the covariance matrix whose diagonal entries are the variances of the outputs.

3.3 Adjustments of CARDAMOM C cycle

In CARDAMOM, there was no explicit separation between forests and grasslands and CARDAMOM provided low woody biomass in grassland dominated regions (Saatchi et al., 2011; Williams et al., 2013), while grasslands are considered as biomes and have no woody biomass in GOLUM-CNP. In order to represent the grassland biomes in GOLUM-CNP and to conserve the global NPP from CARDAMOM, we approximated the C-cycle state of the non-forest biomes (TRG, TEG and TUN) by partitioning half of CARDAMOM woody NPP to foliar NPP and half to fine roots, in order to better represent grassland C, N and P cycling across these biomes.

1 The CARDAMOM terrestrial C analysis did not assume steady states. Our goal, however, was to describe the steady
2 states of C, N and P cycling, because few global long-term observations associated with N and P were available to constrain
3 the model. We recalculated the C cycle based on a subset of the CARDAMOM results. Specifically, we used NPP and
4 turnover times of the C pools for 2001-2010 (Table 1) and recalculated the steady-state sizes of these pools and the transfers
5 of C between the pools represented in Fig. 1, solving Eqs. A1-A5 with their left sides as zeros. This steady-state
6 transformation of the CARDAMOM C cycle is assessed in Sect. 4.1.

7 **4 Results**

8 **4.1 Steady state C cycle**

9 Table 2 shows the global C-pool sizes and main fluxes of the steady-state C cycle transformed from CARDAMOM
10 under the climate conditions of 2001-2010.. Although the steady-state C stocks do not exactly represent the C stocks at
11 present day, the differences between the steady-state transformed pool sizes and the original non-steady-state CARDAMOM
12 results were within 10% for most C pools and fluxes. The largest differences were for biomass (foliar, fine-root and wood)
13 pools. The larger foliar and fine-root pools in the steady-state GOLUM-CNP model were due to adjustments done for grass
14 dominated grid cells for which CARDAMOM provided some wood growth inconsistent with the biome distribution used in
15 GOLUM-CNP. In these cases, we allocated wood growth from CARDAMOM into growth of fine root and foliage. These
16 pools in GOLUM-CNP, however, remained within the [5, 95th] percentile range of the original CARDAMOM values. The
17 pool size for global woody biomass was 37% smaller in the steady-state model (469 Pg) than the original CARDAMOM
18 results but remained within its inter-quartile range (364-984 Pg). The differences between the gridded maps from original
19 CARDAMOM and GOLUM-CNP are shown in Fig S1. The steady-state transformed C stocks in biomass, litter and SOM
20 were within the 25th and 75th percentiles of the original CARDAMOM results at more than 90% of forest grid cells,
21 indicating that our steady-state transformed C stocks are close to the actual C stocks at present day, given the large
22 uncertainties in the state-of-the-art estimates. Due to the adjustment made for the grass dominated grid cells (see above), the
23 C pools for grassland differ more strongly than forest-dominated area from the original CARDAMOM.

24 **4.2 Steady-state nutrient stocks and fluxes**

25 Figure 3 and Fig. S4 summarizes the stocks and fluxes of N and P for the seven biomes and for the globe. The uptake
26 fluxes of N and P were largest for tropical forests, mainly driven by the large NPP of this biome. Rates of N and P uptake
27 were lower for temperate and boreal forests than tropical forests and for non-forest biomes than forests. The pool sizes of N
28 and P in plants tended to decrease from tropical to boreal regions, following the C-pool sizes and their observed
29 stoichiometries. Conversely, N and P contents in litter were larger for boreal forests, temperate grasslands and tundra
30 ecosystems than the other biomes, mainly due to a longer turnover of the litter pool in these biomes. The N-pool size of
31 SOM was also larger in boreal forests, temperate grasslands and tundra than the other biomes. The P-pool size in SOM,
32 however, was smaller for boreal forests and tundra than the other biomes, consistent with the differences between the N:C
33 and P:C ratios of boreal biomes compared to other biomes (Table S1). Inorganic N and labile soil P pools and leaching rates
34 of N and P were higher in tropical forests, where runoff was higher than in the other biomes. Semi-arid tropical grassland
35 (TRG) had high losses of N and P by fire and a low loss from leaching. The internal N and P fluxes within ecosystems were
36 usually much larger than the external input fluxes and the output fluxes for all biomes, highlighting the dominant role of
37 internal cycling of N and P, which differed from C cycles where NPP and losses by respiration were larger than any internal
38 C flux.

39 Here we compared the estimates of N and P stocks for global terrestrial biosphere with other studies. Our estimate of N
40 in plants (3.9 Pg N) was close to the estimate modeled by Zaehle et al. (2013) (3.5 Pg N), and was within the range of other

1 studies, from 1.8 Pg N by Yang et al. (2009) and 6.57 Pg N by Wang et al. (2010) (6.57 Pg N). Our estimate of N in litter
2 and SOM was lower than the estimate of 65 Pg N by Xu-ri et al. (2008) and Yang et al. (2009), but smaller than the estimate
3 of 126 Pg N by Wang et al. (2010). Our estimate of the P mass in plants (0.17 Pg P) was smaller than the estimates modeled
4 by Wang et al. (2010) (0.39 Pg P) and Goll et al. (2012) (0.23 Pg P). Our estimate of the litter P mass (0.03 Pg P) was similar
5 to the estimate of 0.04 Pg P by the CABLE model (Wang et al. 2010) but was two-fold lower than the estimate (0.08 Pg P)
6 modeled by Goll et al. (2012).

7 The rate of total N input (deposition and fixation) aggregated to global scale was 0.19 Pg N y⁻¹ and equated (by
8 construction) to the steady-state rate of total N loss. Total N uptake by plants was 0.68 Pg N y⁻¹. Our estimate of N
9 denitrification was 0.10 Pg N y⁻¹, consistent with the independent estimate of global soil denitrification of 0.12 Pg N y⁻¹ by
10 Seitzinger et al. (2006) and within the range reported by other studies, from 0.04 Pg N y⁻¹ (Houlton and Bai, 2009) to 0.29 Pg
11 N y⁻¹ (Galloway et al., 2013). The global loss of N was 0.05 Pg from fire and 0.04 Pg N y⁻¹ from leaching, the latter being
12 similar to the independent estimates by Galloway et al. (2004, 2013) of 0.013-0.18 Pg N y⁻¹ and by Houlton and Bai (2009)
13 of 0.09 Pg N y⁻¹. Globally, the loss of N by fire accounted for 26% of the total N loss. The total input of P to the terrestrial
14 ecosystem was 0.007 Pg P y⁻¹, 86% from deposition (range from 71% for BOCF to 92% for TRG); only a small fraction was
15 from rock weathering (ranging from 8 to 29% across biomes). The loss of P is mainly from leaching and the loss by fire
16 accounted for only 18% of the total P loss, much smaller than the fraction for N.

17 4.3 Implications for ecological research

18 Figure 4-5 shows the latitudinal distribution of foliar N:P ratios in our model. This result reflects the distribution of the
19 seven biomes and respective C:N and C:P ratios – both of which are prescribed here. Foliar N:P ratios decreased on average
20 from low to high latitudes. Estimates from previous studies also followed this trend (Kerkhoff et al., 2005; McGroddy et al.,
21 2004; Reich and Oleksyn, 2004) based on foliar measurements. The mean N:P ratios in our study were in the middle of the
22 range of observations for all latitudes. The results of GOLUM-CNP better indicated the high N:P ratios between 20 ° to 40 °,
23 where grassland is the dominant biome, than the monotonic regressions (colored lines in Fig. 4) derived by Reich and
24 Oleksyn (2004) and Kerkhoff et al. (2005) for foliar data, implying that the use of stoichiometries at the scale of large
25 biomes can identify the general features of the spatial gradients of N and P cycling.

26 Figure 5a-6a and 5b-6b show the distribution of the openness (defined as the ratio of new nutrient inputs to the total
27 plant uptake of nutrients, Sect. 3.2) for N and P in different ecosystems and Fig. S54a and S45b show the gridded maps of
28 these indices. New N in forest ecosystems (sum of deposition and biological fixation) accounted for 10% (BOCF) to 51%
29 (TECF) of the total plant uptake of N, and new P (due to deposition and rock weathering) accounted for only 3.5% (BOCF)
30 to 15% (TRF) of the total plant uptake of P. The openness of both N and P in grassland ecosystems decreased from the
31 tropics to high latitudes. The residence times of N and P in ecosystems were much longer than those of C (Table S2) and
32 decreased from the tropics to boreal areas (Fig. 5e5c, 5d5d, 56a-55a and 56b55b).

33 The openness and residence times of N and P together allow to assess the relative importance of external inputs and
34 internal cycling to support plant growth. For example, TECF are characterized by a more open N cycle and a longer N
35 residence times compared to TRF. The difference in the openness of N cycle indicates that the TECF intends to invest more
36 resources to obtain N from external inputs than TRF. The differences in residence times indicate that N is more efficiently
37 conserved within the ecosystem in TECF compared to TRF, and such a conservation within ecosystems is primarily driven
38 by differences in the turnover of dead organic matter (Fig. 3). The P cycle is less open than N cycle in all ecosystems,
39 highlighting the importance of ecosystem P recycling within ecosystems to support plant growth.

40 Figure 6-7 shows the diagnosed nutrient-use efficiencies from GOLUM-CNP outputs for the seven biomes and Fig. S7a
41 S6a and S7b-S6b show the gridded maps of nutrient-use efficiencies. Among forest biomes, tropical forest had the lowest
42 NUE and the highest PUE compared to other forest biomes (Fig. 6a7a), consistent with the higher P and lower N stresses in

1 tropical ecosystems (Gill and Finzi, 2016; Reich and Oleksyn, 2004). The values of NUE and PUE were similar to each
2 other for TEDF, TECF and BOCF. Nutrient-use efficiencies were about 3-fold lower for non-forest biomes (Fig. 6b7b) than
3 forest biomes, and both NUE and PUE decreased from tropical/C4 grassland to tundra.

4 4.4 Sensitivity analysis

5 Figure 7-8 shows the mean sensitivity of the nutrient-uptake fluxes (F_N and F_P), nutrient-use efficiencies (NUE and
6 PUE), pool sizes of inorganic N and P (N_{inorg} and P_{inorg}), N and P openness, residence times of N and P in the ecosystem
7 ($\tau_{N,eco}$ and $\tau_{P,eco}$) and residence times of N and P in plants ($\tau_{N,plant}$ and $\tau_{P,plant}$) to the input variables for the tropical-rainforest
8 biome (TRF). The sensitivities were similar for the other biomes (Figs. 88S7-S13S12). The uptake of nutrients in GOLUM-
9 CNP was determined by NPP, NPP-allocation fractions, observation-based nutrient:C ratios and resorption coefficients (Eqs
10 E7 and E18), so N uptake for tropical forest (Fig. 7a8a) was highly sensitive to NPP (1.0), NPP-allocation fractions (0.3) and
11 the N:C ratio (0.4) of the woody pool, and P uptake was sensitive to NPP (1.0) and foliar variables (0.5 for $\gamma_{C,1}$ and 0.5 for
12 $\rho_{P,1}$; see Table 1 for the definition of these variables). The nutrient-use efficiencies, defined in Eq. (1) as the ratio between
13 GPP and the nutrient-uptake fluxes (Eq. 3), were negatively sensitive to the input variables mentioned above. Estimates of
14 the openness of N and P were sensitive to input fluxes, NPP, NPP-allocation fractions and stoichiometric inputs. The
15 residence times of nutrients in the ecosystem were influenced by variables affecting vegetation growth (e.g. NPP and
16 allocation fractions of NPP) and those affecting inputs (e.g. deposition, N fixation and P release from rock weathering) to
17 about equal extent. They were also very sensitive to variables related to soil, e.g. the N:C and P:C ratios in soil and residence
18 times of soil. This reflect the large stocks of C and nutrients in soils than in the vegetation. The residence times of nutrients
19 in whole plants ($\tau_{N,plant}$ and $\tau_{P,plant}$) were more sensitive to the variables affecting woody biomass than those affecting foliage
20 and fine roots. The sensitivity of residence times in the ecosystem and whole plants suggested that the nutrient cycling in the
21 terrestrial biosphere were primarily determined by the largest pools.

22 5 Discussion

23 We developed a new observation-based modeling framework of global terrestrial N and P cycling built on a data-driven
24 C-cycle model and observed N:C and P:C stoichiometric ratios in different pools spatially averaged at the scale of large
25 biomes and observation-based estimates of the external input and output fluxes of N and P. This model was then used to
26 estimate the pool sizes and fluxes in N and P cycles and indicators of the coupling between nutrient and C cycling, including
27 nutrient openness, residence times in ecosystems and nutrient-use efficiencies. The data-driven estimates of steady-state
28 global C, N and P cycles are the first that are fully consistent with a large set of global observation-based data sets, under the
29 condition of current climate, deposition and CO₂ concentration. The indicators for the coupling between nutrient and C
30 cycling, which are emerging properties of GOLUM-CNP, are used to evaluate the capabilities of GOLUM-CNP to capture
31 observed patterns among biomes. We found that there are some differences between our data-driven estimates and previous
32 studies about the nutrient efficiencies at biome scales (Gill and Finzi, 2016) and the openness (Cleveland et al., 2013). In this
33 section, we discussed the major uncertainties in our model and show how these uncertainties affect the computation of
34 nutrient efficiencies and the openness (Sect. 5.1). Of note is that most of our discussions are for the C cycle (based on the
35 sensitivity analysis, see below), and since CARDAMOM is the only data-driven C cycle to our knowledge, the modifications
36 of the CARDAMOM dataset we made in this section are more qualitative and diagnostic rather than deterministic. Such an
37 example highlights some important variables that should be investigated or considered in future data-driven products.

38 5.1 Sensitivity to C variables

39 Our estimates of nutrient-use efficiencies differed significantly from those estimated from *in situ* measurements (red
40 squares and diamonds in Fig. 47) by Gill and Finzi (2016), particularly the values of PUE for all biomes and NUE for

1 temperate and boreal forests. NUE and PUE were determined by NPP-allocation fractions, stoichiometric ratios, resorption
2 coefficients and fractions of fire in the total outgoing C flux (Eqs. 3 and E7, where NPP is canceled by the division). The
3 CARDAMOM observation-based analysis of the C cycle is the basis of the GOLUM-CNP modeling framework, so that
4 errors and uncertainties in CARDAMOM for the C cycle translate into errors and uncertainties of GOLUM-CNP.
5 Quantitatively, the sensitivity analysis (Figs. 78, S8S7-S13S12) indicated that F_N and F_P , and thus NUE and PUE, were most
6 sensitive to the NPP-allocation fractions (especially to woody biomass) and foliar stoichiometry. We applied the sensitivity
7 matrix (Eq. 5) to further calculate the contribution of variances from each of these input variables, in which the uncertainties
8 in the NPP-allocation and fire fractions were obtained from CARDAMOM and the uncertainties (1-sigma) in the N:C and
9 P:C stoichiometric ratios and resorption coefficients were assumed to be 40%. This 40% uncertainty was larger than the
10 uncertainty (20%) of the N:C ratios used by Wang et al. (2010), so our estimate of the contribution of uncertainties from the
11 stoichiometric ratios was relatively large. The contribution of these different sources of uncertainty to the variances of NUE
12 and PUE is shown in Fig. 8-9 for temperate coniferous forests whose NUE and PUE deviated the most from the estimate by
13 Gill and Finzi (2016). Fig. 8-9 shows that the NPP-allocation fractions were the largest contributors to the total variances in
14 NUE and PUE, which totaled >80%.

15 The NPP-allocation coefficients in CARDAMOM were only constrained indirectly by the satellite observations of LAI
16 and tropical aboveground biomass. The uncertainty of the CARDAMOM allocation fractions was thus substantial, especially
17 for non-tropical biomes where no biomass data were used (allocation-fraction 25th – 75th percentile ranges are typically >50%
18 of the mean). For example, the mean fraction of NPP allocated to woody biomass in CARDAMOM was >60% in most grids
19 (Fig. S14aS13a), which is rare for field measurements (Chen et al., 2013; Doughty et al., 2015). The mean allocation of NPP
20 to fine roots may have been underestimated, characterized by too long a turnover time in CARDAMOM (range from <1 to
21 10 y) compared to field measurements (<3 y for all ecosystems, Gill and Jackson, 2000; Green et al., 2005). The
22 CARDAMOM results indicated a turnover time of leaves in temperate and boreal biomes of <1 y, while Reich et al. (2014)
23 indicated that the typical life span of conifer needles in evergreen coniferous forests depended on temperature and ranged
24 from 2.5 to >10 y, this inconsistency being attributed by Bloom et al. (2016) to the potential roles of seasonal MODIS LAI
25 biases and to the presence of understory vegetation across high-latitude ecosystems (Heiskanen et al., 2012).

26 Considering these inconsistencies between mean CARDAMOM values and *in situ* measurements, we conducted an
27 additional experiment in which the CARDAMOM fields were further adapted: 1) the mean NPP-allocation fractions to
28 woody biomass and the turnover time of woody biomass was divided by 1.5 to make sure that the NPP-allocation fractions
29 to woody biomass fall in the range of field measurements, 2) the foliage turnover time of TECF and BOCF forests and
30 associated NPP-allocation fractions were adjusted (keeping foliar biomass not changed) to match *in situ* observations (Reich
31 et al., 2014) based on the fitted relationship between the needle longevity and mean annual temperature from Reich et al.
32 (2014), assuming that understory vegetation plays a minimal role in C, N and P cycling and 3) Since in CARDAMOM, the
33 NPP was constrained by GPP (GPP being constrained by the observation of LAI and the relationship between LAI and GPP)
34 and the observation of biomass, additional adjustments were made to conserve the total NPP and pool sizes estimated from
35 CARDAMOM by allocating the residual NPP after the modifications from step 1 and 2 to fine roots, and adjusting the
36 turnover time of fine roots to conserve exactly the pool size of CARDAMOM (see Figs. S14S13-S17-S16 for the adjusted
37 variables and original CARDAMOM values).. Fig. 9-10 shows the NUE and PUE from this new experiment based on this
38 modified version of the C cycle from CARDAMOM. The NUE and PUE were lower than those in Fig. 6-7 for the forest
39 biomes, especially for TECF and BOCF, which tended to decrease PUE from tropical to boreal forest. This distribution of
40 PUE among the biomes in Fig. 9-10 better matched the differences between biomes presented by Gill and Finzi (2016).
41 Remaining inconsistencies could be attributed to the different methods used in this study and by Gill and Finzi (2016) are
42 different. For example, Gill and Finzi (2016) notably used the net mineralization rates of N and P to approximate plant
43 uptake, because their differences were an order of magnitude smaller than net nutrient mineralization. These authors used *in*

1 *situ* measurements of net N mineralization but used a statistical model to estimate P mineralization based on a soil-order-
2 specific soil-P pool due to the lack of data (Yang and Post, 2011) and a regression between soil-P turnover times and mean
3 annual temperature. Their estimate of plant uptake was thus independent of vegetation stoichiometry, which differed from
4 our study. Gill and Finzi (2016) also used bootstrapping to sample the NPP and net N (or P) mineralization from independent
5 studies. Their estimates of NUE and PUE were thus not based on paired data, so their estimates may contain some sampling
6 errors.

7 **5.2 Uncertainty of nutrient-cycle openness**

8 The distribution of nutrient-cycle openness in the seven biomes was presented in Sect. 4.3 and Fig. 56. Our estimate of
9 a small openness of N and P in BOCF, and that the openness was smaller for the P than the N cycle, were consistent with the
10 estimates by Cleveland et al. (2013). Our estimates of the N openness, however, were about twice as large as the estimates of
11 Cleveland et al. (2013). This difference was due to the larger deposition fluxes in our study (globally 72 Tg N yr⁻¹) than those
12 used by Cleveland et al. (2013) (33 Tg N yr⁻¹; from Dentener, 2006), because Wang et al. (2017) used an atmospheric model
13 with higher horizontal resolution and an updated inventory of reactive-N (e.g. NO_x and NH₃) emission (Wang et al., 2017)
14 and also because Cleveland et al. (2013) assumed that only 15% of deposited N was available to plants. Cleveland et al.
15 (2013) demonstrated that changing the fraction of biologically available deposited N to 100% did not significantly change
16 the openness, because N-deposition fluxes were generally smaller than N fixation and accounted for a small fraction of
17 external N inputs in their study. Our estimates of P openness were also larger than those of Cleveland et al. (2013), which we
18 attributed to the large differences in the estimates of P deposition between the two studies. Cleveland et al. (2013) used P
19 deposition (0.26 Tg yr⁻¹) from Mahowald et al. (2008), which were an order of magnitude lower than recent estimates from
20 Wang et al. (2017) used in this study (5.8 Tg yr⁻¹), because Wang et al. (2017) revised the contribution of anthropogenic P
21 emissions and P in particles with diameters >10 μm (Wang et al., 2015, 2017). We also found that the P-cycle openness
22 decreased from the tropics to the boreal region, in contrast to the results by Cleveland et al. (2013). This also derives from
23 the differences in the spatial gradients of P deposition in the two studies. Mahowald et al. (2008) found that P deposition was
24 largest in northern Africa and that P deposition was within the same order of magnitude for tropical and temperate forests.
25 Wang et al. (2017), however, found that P deposition was much larger over tropical forests than other regions. The
26 contrasting spatial gradients in P deposition was likely due to the different models of atmospheric transport used by Wang et
27 al. (2017) and Mahowald et al. (2008). More importantly, most stations measuring total P deposition are in temperate regions,
28 and measurements of P deposition over tropical forests are very limited (Mahowald et al., 2008; Wang et al., 2017), so the
29 estimates of P deposition in the tropics were not well constrained by *in situ* observations and thus had large uncertainties.
30 Differences in the spatial gradients in nutrient-cycle openness between our study and the study by Cleveland et al. (2013)
31 demonstrated the impact of uncertain input data sets on the estimate of ecologically relevant quantities. The quantitative
32 assessment of the uncertainties in our estimates of openness, however, was difficult, because the potential uncertainties in
33 these data sets were not systematically evaluated within and between different estimates, and should therefore be addressed
34 in future studies.

35 **5.3 Future research and data needs**

36 Our estimates of global N and P cycles were at the scale of large biomes. Recent studies of N and P cycles have relied
37 on biome-specific stoichiometry (Cleveland et al., 2013; Wang et al., 2010). Stoichiometry, however, is also highly variable
38 within biomes (Reich and Oleksyn, 2004). For example, Kattge et al. (2011) found that 40% of the variability in foliar N
39 content was within species (finer scale than that of large biomes) and suggested that these stoichiometric ratios may be better
40 represented by future trait-based estimates rather than fixed species-specific values. Some improvements have been made on
41 the variation of stoichiometric ratios across climatic and ecological gradients within and across biomes, and on the

1 contribution of plant traits and environmental conditions to these variations (Dong et al., 2017; Han et al., 2005; Meyerholt
2 and Zaehle, 2015). However, it is still not sufficient to derive a globally gridded overview of the N and P cycles on current
3 knowledge. A better understanding of the stoichiometric variability and its drivers is still needed in terms of not only
4 representing the large-scale gradients but also reducing the uncertainties at local scale. New and spatially interpolated
5 stoichiometric data sets should partly overcome this problem, although uncertainties in the interpolation will need to be
6 carefully propagated on GOLUM-CNP outputs.

7 We assumed that all terrestrial ecosystems were at a steady state for 2001-2010 due to a lack of global constraints on
8 the dynamics of N and P cycling over a long period. Terrestrial ecosystems, however, are not currently at steady states (Luo,
9 2017; Luo and Weng, 2011), due to climate change, increasing atmospheric CO₂, anthropogenic disturbance etc,
10 (Friedlingstein et al., 2006; Sitch et al., 2015). Zaehle (2013) reported that the terrestrial biosphere has accumulated 1.2 Pg N
11 and 134.0 Pg C since the pre-industrial period. Wang et al. (2017) also found that N and P deposition have changed
12 dramatically over time. The simulations by different models varies considerably, e.g. the responses of the biosphere to the
13 increasing atmospheric CO₂ (Zaehle et al., 2014) and thus in future projections, because the current data sets have had little
14 success in constraining all key processes in most DGVMs. Our results contribute to evaluating models simulating global
15 biogeochemical cycles. Although our steady-state C pool sizes (given the NPP and residence time at the condition of current
16 climate) were within the [25, 75th] percentile range of the original non-steady-state CARDAMOM results (Fig. S1) at most
17 grid cells, the biomass C stocks at 5%-10% of forest grid cells exceed the uncertainty range of CARDAMOM. In addition,
18 independent remote-sensing estimates for 30°N to 80°N were 4.76 ± 1.78 kg C m⁻² for mean forest C density and 79.8 ± 29.9
19 Pg C for total forest C (Thurner et al., 2014), which were lower than the GOLUM-CNP estimates (6.51 kg C m⁻² for mean
20 forest C density across pixels defined as forest in Fig. 2, and 181 Pg C for total forest C) for this region. This inconsistency
21 was largely due to the fact that northern temperate and boreal forests may deviate substantially from their equilibrium for the
22 current NPP (Pan et al., 2011), because of climate change and elevated CO₂. Residual overestimation could be also due to
23 the fact that biomass removal by harvesting and from disturbance other than fires was not explicitly constrained in
24 CARDAMOM and thus not represented in GOLUM-CNP. A transient simulation of N and P cycling will be needed in future
25 studies as more constraints on N and P cycles emerge to study the effects of climate change, increasing CO₂ levels and
26 disturbance on N and P cycles and their feedbacks. In such a transient simulation, a key process would be to simulate both
27 the short-term and long-term responses of plants to the changing environment, e.g. how the plants would react when the
28 inorganic N or labile soil P was not sufficient. Different models assumed different hypotheses under these conditions. For
29 instance, N:C and P:C ratios are fixed and the photosynthesis rate is reduced to meet the low uptake of nutrients in Thornton
30 et al., (2007). In Wang et al. (2010), the N:C and P:C ratios in biomass can vary within certain ranges, insufficient nutrient
31 uptake would first result in a low concentration of N and/or P in plant tissues and the low concentration of nutrient would
32 then limit the photosynthesis according to an empirical relationship between nutrient concentration and NPP. Similarly,
33 when the N:C and P:C ratios in litter change, the decomposition rate of litter would change as a result of altered activity of
34 microbes (Manzoni et al., 2017). In the future, more data are required to test these hypotheses and the transient simulation of
35 next version of GOLUM-CNP should incorporate these interactions between the plants and environments.

36 In addition, some processes, such as the N inputs from rock weathering (Houlton et al., 2018) were not considered in
37 this study, because 1) as stated in Houlton et al. (2018), it is still unknown how much of rock-released N can be used by
38 plants when rock weathering happens deep beneath the soils; 2) in GOLUM-CNP, adding rock N inputs has the same effect
39 than N fixation and N deposition (Eqs. B6 and E17); and 3) the estimate of total input of N to ecosystems (188 Tg N yr⁻¹) in
40 this study are already at the higher end of the estimate (mean 147 Tg yr⁻¹, and range between 99.1 and 185.1 Tg yr⁻¹) of
41 Houlton et al. (2018), even if rock N inputs are not accounted for, due to our larger estimates of N fixation and N deposition
42 than Houlton et al. (2018). In the future, the rock N inputs and the fraction of these N inputs are accessible to plants should
43 be further quantified and the quantity of total N inputs to the ecosystems should be reconciled between different studies.

1 With these improvements, the future development of data-driven GOLUM-CNP should take into all these processes and
2 fluxes.

3 The model structure of GOLUM-CNP is mainly described by the inputs (NPP for C cycle, N deposition and fixation for
4 N cycle, P deposition and release from rock weathering for P cycle) and residence times. Most DGVMs (e.g. Goll et al.,
5 2012, 2017a; Medvigy et al., 2009; Parton et al., 2010; Thornton et al., 2007; Wang et al., 2010; Weng and Luo, 2008; Xu-Ri
6 and Prentice, 2008; Yang et al., 2009; Zaehle et al., 2014; Zaehle and Friend, 2010) can be summarized by these two
7 components, although these models have more processes and use complex equations to describe the dynamics controlling
8 carbon and nutrient distribution among pools and the turnover of each pool. In this context, the output of the GOLUM-CNP
9 provides a traceable tool that can be used in the future to compare the results between GOLUM-CNP and different DGVMs.
10 As DGVMs are capable of computing the steady state of the biogeochemical cycles for present conditions, a direct
11 comparison between GOLUM-CNP estimate and DGVMs' estimates is possible.

12 At last, the sensitivity matrix presented in Sect. 4.4 provides a useful tool for assessing the uncertainties in model
13 outputs by propagating the uncertainties in the model inputs. We applied this method to quantitatively assess the sources of
14 uncertainties in the estimated nutrient-use efficiencies (Sect. 5.1 and Fig. 89), but we also found that the uncertainties for
15 some other quantities were currently difficult to obtain, because the estimates of uncertainties were not available for all
16 spatially explicit input data. This sensitivity analysis can be used in future studies to quantify the contribution of each input
17 data set to the uncertainty in other model outputs, to characterize the dominant sources of uncertainties in the estimated C, N
18 and P processes, to identify the major differences between different models (e.g. GOLUM-CNP versus DGVMs) and thus to
19 identify priorities for future data syntheses to fill the largest gaps in uncertainty. Future studies that provide global data sets
20 will need to include systematic evaluations and spatially explicit estimates of uncertainties in their data sets.

21 **6 Concluding remarks**

22 This study is a first attempt to combine observation-based estimates of C, N, P fluxes and pools in terrestrial ecosystems
23 into a consistent (steady-state) diagnostic model. Although there are considerable uncertainties in our results due to uncertain
24 and incomplete carbon cycle and nutrient observations, the main findings are: 1) external inputs of P from outside the
25 ecosystem contributes to a smaller plant P uptake than that of N, indicating a more important role of internal P recycling than
26 that of internal N recycling in supporting plant growth, 2) tropical forests have the lowest N use efficiency and the largest P
27 use efficiency, suggesting the adaptive response of this biome to the low P availability in the tropics. The structure of
28 GOLUM-CNP is analogous to most other process-based DGVMs describing carbon and nutrient interactions. The output of
29 the GOLUM-CNP provides a traceable tool and can be used in the future to test the performance of complex DGVMs in the
30 simulation of interactions between C, N and P cycling.

31 **Code and data availability**

32 The source code and the map of the classification of seven large biomes are included in the Supplement. For the other
33 datasets that are listed in Table 1, it is encouraged to contact the first authors of the original references.

34 **Appendix A Equations for carbon cycle**

35 The carbon cycle framework is based on DALEC2 model (Bloom and Williams, 2015), except that we combined the
36 labile and foliage pools together since the labile pool in DALEC2 only transfer to foliage. There are five pools in the C cycle
37 (1: foliage; 2: fine roots; 3: wood; 4: litter; 5: SOM). The equations governing the change of C pools are given by:

$$38 \frac{dC_1}{dt} = -\tau_1^{-1}C_1 + \gamma_1Fc \quad (\text{A1})$$

$$1 \quad \frac{dC_2}{dt} = -\tau_2^{-1}C_2 + \gamma_2Fc \quad (A2)$$

$$2 \quad \frac{dC_3}{dt} = -\tau_3^{-1}C_3 + \gamma_3Fc \quad (A3)$$

$$3 \quad \frac{dC_4}{dt} = \tau_1^{-1}C_1(1 - f_{fireC,1}) + \tau_2^{-1}C_2(1 - f_{fireC,2}) - \tau_4^{-1}C_4 \quad (A4)$$

$$4 \quad \frac{dC_5}{dt} = \tau_3^{-1}C_3(1 - f_{fireC,3}) + \eta\tau_4^{-1}C_4 - \tau_5^{-1}C_5 \quad (A5)$$

5 The definitions of the symbols are listed in Table 1.

6 Appendix B Equations for nitrogen cycle

7 There are five organic N pools and one inorganic soil N pool. The N cycle are described by the following equations:

$$8 \quad \frac{dN_1}{dt} = -\tau_1^{-1}N_1(1 - \varepsilon_1) + \beta_1Fn \quad (B1)$$

$$9 \quad \frac{dN_2}{dt} = -\tau_2^{-1}N_2 + \beta_2Fn \quad (B2)$$

$$10 \quad \frac{dN_3}{dt} = -\tau_3^{-1}N_3 + \beta_3Fn \quad (B3)$$

$$11 \quad \frac{dN_4}{dt} = \tau_1^{-1}N_1(1 - \varepsilon_1)(1 - f_{fireC,1}) + \tau_2^{-1}N_2(1 - f_{fireC,2}) - \tau_4^{-1}N_4 \quad (B4)$$

$$12 \quad \frac{dN_5}{dt} = \tau_3^{-1}N_3(1 - f_{fireC,3}) + \eta\tau_4^{-1}N_4 + N_{imob} - \tau_5^{-1}N_5 \quad (B5)$$

$$13 \quad \frac{dN_{inorg}}{dt} = \tau_5^{-1}N_5(1 - f_{fireC,5}) + \tau_4^{-1}N_4(1 - \eta - f_{fireC,4}) + N_d + N_{fix} - N_{imob} - f_{leach}N_{inorg} - f_{denit}N_{inorg} - Fn \quad (B6)$$

14

15 The definitions of the symbols are listed in Table 1.

16 In Eq. B6, the fraction of inorganic N (f_{leach}) that is lost due to leaching is computed by soil water (Θ) and the sum of
 17 drainage and surface runoff (q). We use the spatially explicit estimate of daily soil moisture derived from the European
 18 Centre for Medium-Range Weather Forecasts (ECMWF) Interim Reanalysis (ERA-Interim/Land; Albergel et al., 2013;
 19 Balsamo et al., 2015) (see Table 1), and the global gridded estimate of monthly mean runoff data from the Global Runoff
 20 Data Centre (GRDC, <http://www.grdc.sr.unh.edu/>). Since the runoff data only have a monthly time step, we use the same
 21 value of runoff for each day within one month. The leaching fraction at annual scale is thus computed by:

$$22 \quad f_{leach} = \sum_{d=1}^{365} \frac{q_i}{\Theta_i + q_i} \quad (B7)$$

23 Of note is that in this computation, f_{leach} can exceed one, meaning that the turnover time of inorganic N pool is smaller
 24 than one year (Wang et al., 2010).

25 Appendix C Equations for phosphorus cycle

26 There are five organic P pools and one inorganic soil P pool. The P cycle are described by the following equations:

$$27 \quad \frac{dP_1}{dt} = -\tau_1^{-1}P_1(1 - \theta_1) + \varphi_1Fp \quad (C1)$$

$$28 \quad \frac{dP_2}{dt} = -\tau_2^{-1}P_2 + \varphi_2Fp \quad (C2)$$

$$29 \quad \frac{dP_3}{dt} = -\tau_3^{-1}P_3 + \varphi_3Fp \quad (C3)$$

$$30 \quad \frac{dP_4}{dt} = \tau_1^{-1}P_1(1 - \theta_1)(1 - f_{fireC,1}) + \tau_2^{-1}P_2(1 - f_{fireC,2}) - \tau_4^{-1}P_4 \quad (C4)$$

$$31 \quad \frac{dP_5}{dt} = \tau_3^{-1}P_3(1 - f_{fireC,3}) + \eta\tau_4^{-1}P_4 + P_{imob} - \tau_5^{-1}P_5 \quad (C5)$$

$$32 \quad \frac{dP_{inorg}}{dt} = \tau_5^{-1}P_5(1 - f_{fireC,5}) + \tau_4^{-1}P_4(1 - \eta - f_{fireC,4}) + P_d + P_w + 0.75Fire_p - P_{imob} - f_{leach}f_{dissolve}P_{inorg} -$$

$$33 \quad f_{sorb}P_{inorg} + Fp \quad (C6)$$

1 Where $Fire_P$ represent the P in the ecosystem that suffers from fire events:

$$2 \quad Fire_P = \tau_1^{-1}P_1(1 - \theta_1)f_{fireC,1} + \tau_2^{-1}P_2f_{fireC,2} + \tau_3^{-1}P_3f_{fireC,3} + \tau_4^{-1}P_4f_{fireC,4} + \tau_5^{-1}P_5f_{fireC,5} \quad (C7)$$

3 Appendix D Additional constraints

4 1) Under steady-state, the N:C and P:C ratios for the plants and soil are assumed to be constant, so that N_i and P_i can be
5 calculated by the production of the C pool size from CARDAMOM and the stoichiometry ratios for each pool from
6 Zechmeister-Boltenstern et al. (2015), except litter which has different definitions in CARDAMOM and Zechmeister-
7 Boltenstern et al. (2015):

$$8 \quad N_i = \rho_{N,i}C_i \quad (i = 1,2,3,5) \quad (D1-D4)$$

$$9 \quad P_i = \rho_{P,i}C_i \quad (i = 1,2,3,5) \quad (D5-D8)$$

10 2) The fraction of NPP, F_N and F_P allocations sum up to 1:

$$11 \quad \beta_1 + \beta_2 + \beta_3 = 1 \quad (D9)$$

$$12 \quad \varphi_1 + \varphi_2 + \varphi_3 = 1 \quad (D10)$$

13 3) The fraction of gaseous loss of N due to denitrification to the total inorganic N loss should satisfy the estimates by
14 using global $\delta^{15}N$ observations (f_{gasN} , Goll et al., 2017b):

$$15 \quad \frac{f_{denit}N_{inorg}}{f_{leach}N_{inorg} + f_{denit}N_{inorg}} = f_{gasN} \quad (D11)$$

16 7 Appendix E Solutions under steady-state assumption

$$17 \quad C_i = F_C \gamma_{C,i} \tau_i \quad (i = 1,2,3) \quad (E1-E4)$$

$$18 \quad C_4 = \left[\frac{C_1}{\tau_1} (1 - f_{fireC,1}) + \frac{C_2}{\tau_2} (1 - f_{fireC,2}) \right] \tau_4 \quad (E5)$$

$$19 \quad C_5 = \left[\frac{C_3}{\tau_3} (1 - f_{fireC,3}) + \frac{C_4}{\tau_4} (1 - f_{fireC,4}) \right] \tau_5 \quad (E6)$$

$$20 \quad F_N = F_C \left[\rho_{N,1} \gamma_{C,1} (1 - f_{fireC,1}) (1 - \varepsilon_{N,1}) + \rho_{N,1} \gamma_{C,1} f_{fireC,1} + \rho_{N,2} \gamma_{C,2} + \rho_{N,2} \gamma_{C,3} \right] \quad (E7)$$

$$21 \quad \gamma_{N,2} = \frac{\rho_{N,2} C_2}{\tau_2 F_N} \quad (E8)$$

$$22 \quad \gamma_{N,3} = \frac{\rho_{N,3} C_3}{\tau_3 F_N} \quad (E9)$$

$$23 \quad \gamma_{N,1} = 1 - \gamma_{N,2} - \gamma_{N,3} \quad (E10)$$

$$24 \quad N_i = \rho_{N,i} C_i \quad (i = 1,2,3,5) \quad (E11-E14)$$

$$25 \quad N_4 = \frac{\frac{\rho_{N,1} C_1}{\tau_1} (1 - f_{fireC,1}) + \frac{\rho_{N,2} C_2}{\tau_2} (1 - f_{fireC,2})}{\frac{C_1}{\tau_1} (1 - f_{fireC,1}) + \frac{C_2}{\tau_2} (1 - f_{fireC,2})} C_4 \quad (E15)$$

$$26 \quad N_{imorb} = \eta \left(\rho_{N,5} - \frac{N_4}{C_4} \right) \frac{C_4}{\tau_4} + \left(\rho_{N,5} - \frac{N_3}{C_3} \right) \frac{C_3}{\tau_3} (1 - f_{fireC,3}) \quad (E16)$$

$$27 \quad N_{inorg} = \frac{N_d + N_{fix} - \sum_{i=1}^5 \left(\frac{N_i}{\tau_i} f_{fireC,i} \right)}{f_{leach}} \quad (E17)$$

$$28 \quad F_P = F_C \left[\rho_{P,1} \gamma_{C,1} (1 - f_{fireC,1}) (1 - \varepsilon_{P,1}) + \rho_{P,1} \gamma_{C,1} f_{fireC,1} + \rho_{P,2} \gamma_{C,2} + \rho_{P,2} \gamma_{C,3} \right] \quad (E18)$$

$$29 \quad \gamma_{P,2} = \frac{\rho_{P,2} C_2}{\tau_2 F_P} \quad (E19)$$

$$30 \quad \gamma_{P,3} = \frac{\rho_{P,3} C_3}{\tau_3 F_P} \quad (E20)$$

$$31 \quad \gamma_{P,1} = 1 - \gamma_{P,2} - \gamma_{P,3} \quad (E21)$$

$$32 \quad P_i = \rho_{P,i} C_i \quad (i = 1,2,3,5) \quad (E22-E25)$$

$$P_4 = \frac{\frac{\rho_{P,1}C_1(1-f_{fireC,1})}{\tau_1} + \frac{\rho_{P,2}C_2(1-f_{fireC,2})}{\tau_2}}{\frac{C_1(1-f_{fireC,1})}{\tau_1} + \frac{C_2(1-f_{fireC,2})}{\tau_2}} C_4 \quad (E26)$$

$$P_{imorb} = \eta \left(\rho_{P,5} - \frac{P_4}{C_4} \right) \frac{C_4}{\tau_4} + \left(\rho_{P,5} - \frac{P_3}{C_3} \right) \frac{C_3}{\tau_3} (1 - f_{fireC,3}) \quad (E27)$$

$$P_{inorg} = \frac{P_d + P_w - \sum_{i=1}^5 \left(\frac{P_i}{\tau_i} f_{fireC,i} \right)}{f_{leach} f_{dissolve} + f_{sorb}} \quad (E28)$$

4 Acknowledgement

5 The idea of GOLUM was initially discussed at a workshop held at the Northwest Agricultural and Forestry, China. We
6 are grateful for the financial support of the workshop by the State Key Laboratory of Soil Erosion and Dryland Farming on
7 the Loess Plateau of Northwest A & F University and the National Basic Research Programme of China grant
8 2013CB956602. Funding was provided by the Laboratory for Sciences of Climate and Environment (LSCE), CEA, CNRS
9 and UVSQ. PC, DG, SP, JP and JS acknowledge support from the European Research Council Synergy grant ERC-2013-
10 SyG-610028 IMBALANCE-P. Contribution by AAB was carried out at the Jet Propulsion Laboratory, California Institute of
11 Technology, under a contract with the National Aeronautics and Space Administration. BDS was funded by ERC H2020-
12 MSCA-IF-2015, grant number 701329.

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1 **Table 1** Global spatially explicit observation-based estimates of model variables used as input data sets and the unknowns estimated in this study (including the symbols for each
2 variable/parameter).

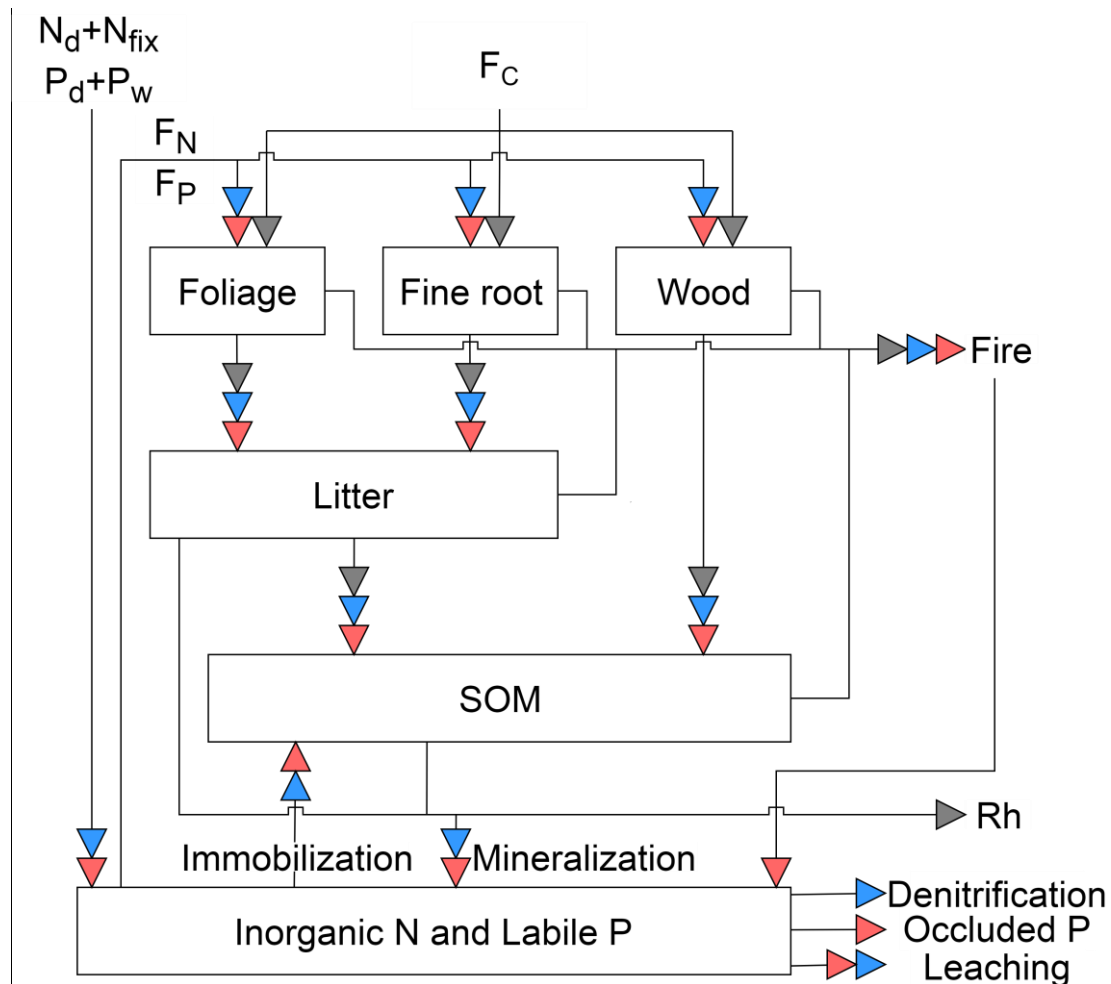
Variable	Definition	Description	Computation method	References
Inputs: carbon cycle				
F _c	NPP	Spatially resolved model-data fusion estimates	Input	CARDAMOM; Bloom et al., 2016
$\tau_{i=1,2,3,4,5}$	Residence time of foliage, fine roots, wood, litter and SOM	Spatially resolved model-data fusion estimates	Input	CARDAMOM; Bloom et al., 2016
$\gamma_{C, i=1,2,3}$	Fraction of NPP allocated to foliage, fine roots and wood	Spatially resolved model-data fusion estimates	Input	CARDAMOM; Bloom et al., 2016
$f_{\text{fireC}, i=1,2,3,4,5}$	Fraction of fire to total outgoing flux from foliage, fine roots, wood, litter and SOM	Spatially resolved model-data fusion estimates	Input	CARDAMOM; Bloom et al., 2016
η	Fraction of litter outflux that enters SOM	Spatially resolved model-data fusion estimates	Input	CARDAMOM; Bloom et al., 2016
Inputs: nitrogen cycle				
$\rho_{N, i=1,2,3,5}$	N:C ratio in foliage, fine roots, wood and SOM	Biome-scale synthesis based on <i>in situ</i> measurements	Input	Zechmeister-Boltenstern et al., 2015
f_{leach}	Fraction of inorganic N (or P) lost due to leaching (Eq. B7)	Spatially resolved reanalysis by model; Model result, scaled to match measurements	Input	Balsamo et al., 2015 Fekete et al., 2002
$\epsilon_{N, 1}$	Resorption coefficient of N in foliage	Biome-scale synthesis based on <i>in situ</i> measurements	Input	Zechmeister-Boltenstern et al., 2015
N_d	N deposition	Spatially resolved model result, scaled to match <i>in situ</i> measurements	Input	Wang et al., 2017
N_{fix}	N fixation	Spatially resolved model result, scaled to match the estimates of NPP and N:C ratios	Input	Peng et al., submitted
f_{gas}	Fraction of denitrification to the total loss of inorganic N	Spatially resolved process-based statistical model result	Input	Goll et al., 2017b
Inputs: phosphorus cycle				
$\rho_P, i=1,2,3,5$	P:C ratio in foliage, fine roots, wood and SOM	Biome-scale synthesis based on <i>in situ</i> measurements	Input	Zechmeister-Boltenstern et al., 2015
$f_{\text{dissolved}}$	Fraction of labile soil P that is dissolved in the soil water	<i>In situ</i> measurements, averaged based on soil order	Input	Yang and Post, 2011
f_{sorb}	Fraction of inorganic P that is transformed to strongly sorbed P	Assumed constant	Input	Goll et al., 2017a
$\epsilon_P, 1$	Resorption coefficient of P in foliage	Biome-scale synthesis based on <i>in situ</i> measurements	Input	Zechmeister-Boltenstern et al., 2015
P_d	P deposition	Spatially resolved model result, scaled to match <i>in situ</i> measurements	Input	Wang et al., 2017
P_d	P weathering	Spatially resolved model result, scaled to match observed data	Input	Hartmann et al., 2014
Unknowns estimated from mass balance assuming steady state				
$C_{i=1,2,3,4,5}$	C pool of foliage, fine roots, wood, litter and SOM	Pools	Based on steady-state assumption	

F_N	N uptake from inorganic-N pool by vegetation	Flux	Mass balance approach based on NPP (input) and stoichiometry ratios (input)
$\gamma_{N, i=1,2,3}$	Fraction of F_N allocated to foliage, fine roots and wood	Allocation fractions	Mass balance approach based on NPP (input) and stoichiometry ratios (input)
$N_{i=1,2,3,4,5}$	N in foliage, fine roots, wood, litter and SOM	Pools	Mass balance approach based on stoichiometry ratios (input) and steady-state C pools ($C_{i=1,2,3,4,5}$), assuming N:C ratios do not change over time
N_{imob}	N immobilization flux	Pools	Based on steady-state assumption that stoichiometry ratios (input), litter C and soil C do not change at annual scale
f_{denit}	Annual denitrification rate	Rate	Mass balance approach, assuming annual mean inorganic N pool size does not change at annual scale
N_{inorg}	Inorganic-N pool	Pool	Based on steady-state assumption that inorganic N do not change at annual scale
F_P	P uptake from inorganic-P pool by vegetation	Flux	Mass balance approach based on NPP (input) and stoichiometry ratios (input)
$\gamma_{P, i=1,2,3}$	Fraction of F_P allocated to foliage, fine roots and wood	Allocation fractions	Mass balance approach based on NPP (input) and stoichiometry ratios (input)
$P_{i=1,2,3,4,5}$	P in foliage, fine roots, wood, litter and SOM	Pools	Mass balance approach based on stoichiometry ratios (input) and steady-state C pools ($C_{i=1,2,3,4,5}$)
P_{imob}	P immobilization flux	Flux	Based on steady-state assumption that stoichiometry ratios (input), litter C and soil C do not change at annual scale
P_{inorg}	Inorganic-P pool	Pool	Based on steady-state assumption that labile P do not change at annual scale

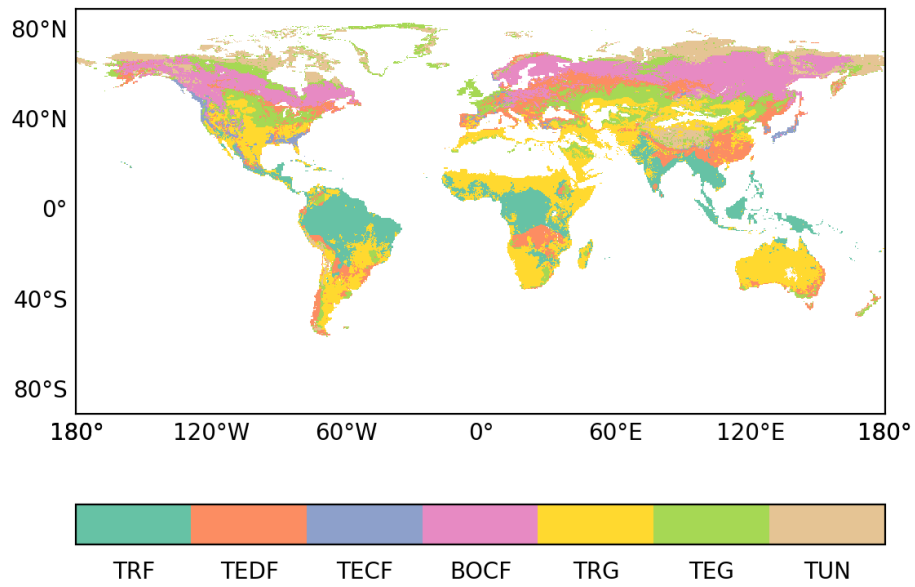
1 **Table 2** Global annual mean C-pool sizes, NPP and heterotrophic-respiration fluxes in the C-cycle model assuming steady
 2 states under the climate conditions of 2001-2010, compared to the means and percentile ranges from the original
 3 CARDAMOM results during 2001-2010.

	This study	Original CARDAMOM				
		5 th percentile	25 th percentile	Mean	75 th percentile	95 th percentile
Foliage-pool size (Pg C)	23	3.2	7	15	21	34
Fine-root-pool size (Pg C)	27	1.9	5	18	25	56
Wood-pool size (Pg C)	493	193	364	755	984	1850
Litter-pool size (Pg C)	20	1.3	4	22	26	88
SOM-pool size (Pg C)	1421	749	1100	1557	1882	2771
NPP (Pg C y ⁻¹)	52.5	Not given	39	52	63	Not given
Fire (Pg C y ⁻¹)	1.5	Not given	1.3	1.7	2.0	Not given
Heterotrophic respiration (Pg C y ⁻¹)	51	Not given	37	54	67	Not given

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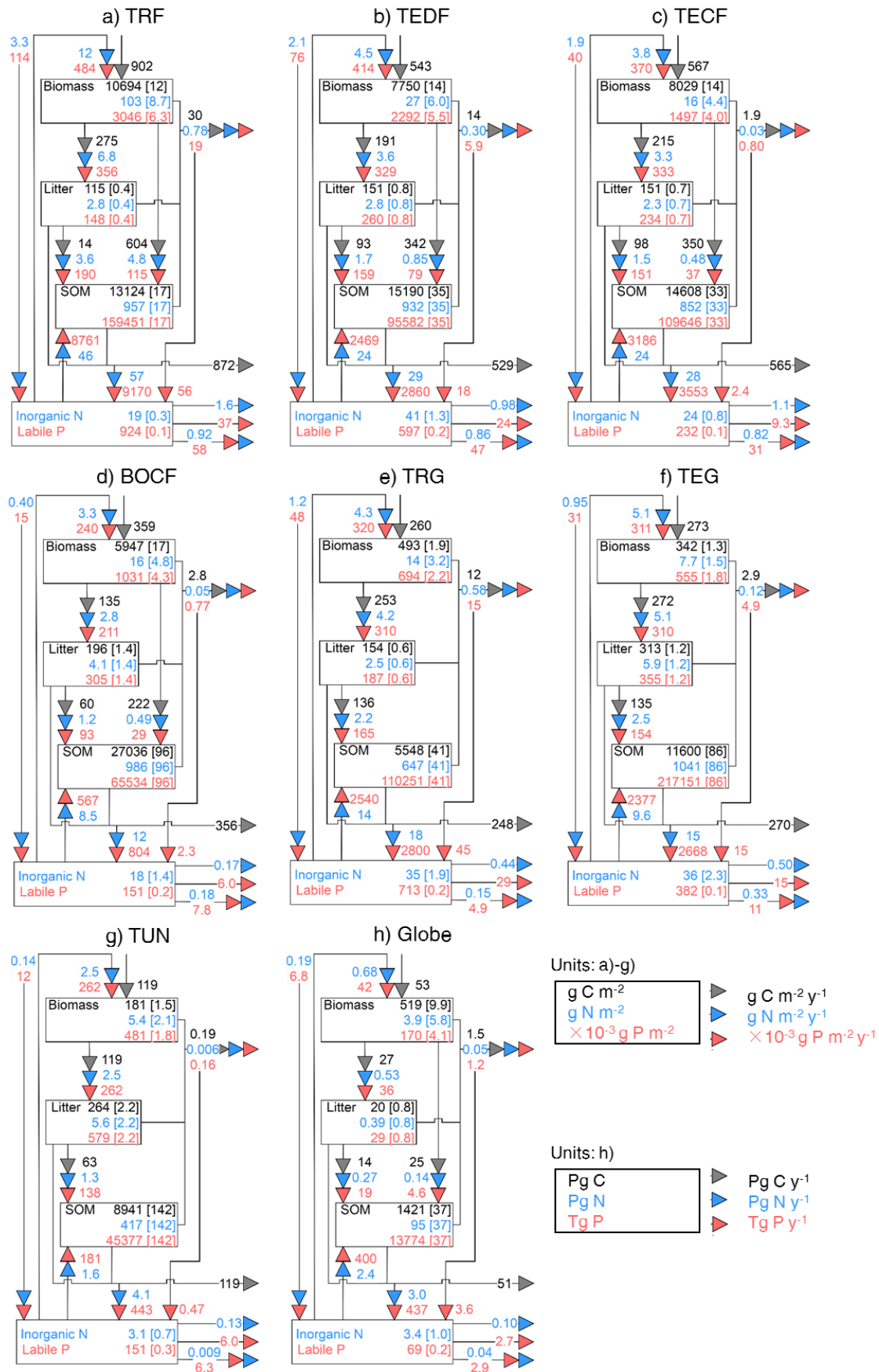


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2 **Figure 1** Schematic representation of the pools and fluxes in the C, N and P cycles within GOLUM-CNP. The gray, blue and
3 red arrows represent C, N and P fluxes, respectively. Plants are divided into foliar, fine root and wood pools, where the wood
4 pool includes woody stems and coarse roots. Litter and soil are two separate pools. The inorganic pool represents the nutrient
5 sources in the soil that are available for plant uptake. Arrows between the pools represent the directions of C, N and P flow
6 between pools. External inputs of N are atmospheric deposition (N_d) and biological N fixation (N_{fix}). External inputs of P are
7 atmospheric deposition (P_d) and P released by rock weathering (P_w). F_C is net primary production (NPP). F_N and F_P are plant
8 uptake of N and P from the inorganic N and labile P pools, respectively. R_h is release of C due to heterotrophic respiration.
9 Mineralization of N and P is modeled along with litter and SOM decomposition, and N and P immobilization is modeled by
10 a flux from the inorganic pool to SOM. External losses of N occur by fire, leaching and denitrification. External losses of P
11 occur by fire, leaching and transfer to occluded P in the soil.
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Figure 2 ESA CCI land-cover map classified into the seven large biomes for which average N:C and P:C ratios for each carbon pool are available, at 0.25°×0.25° resolution: tropical rainforests (TRF), temperate deciduous forests (TEDF), temperate coniferous forests (TECF), boreal coniferous forests (BOCF), tropical/C4 grasslands (TRG), temperate/C3 grasslands (TEG) and tundra (TUN).



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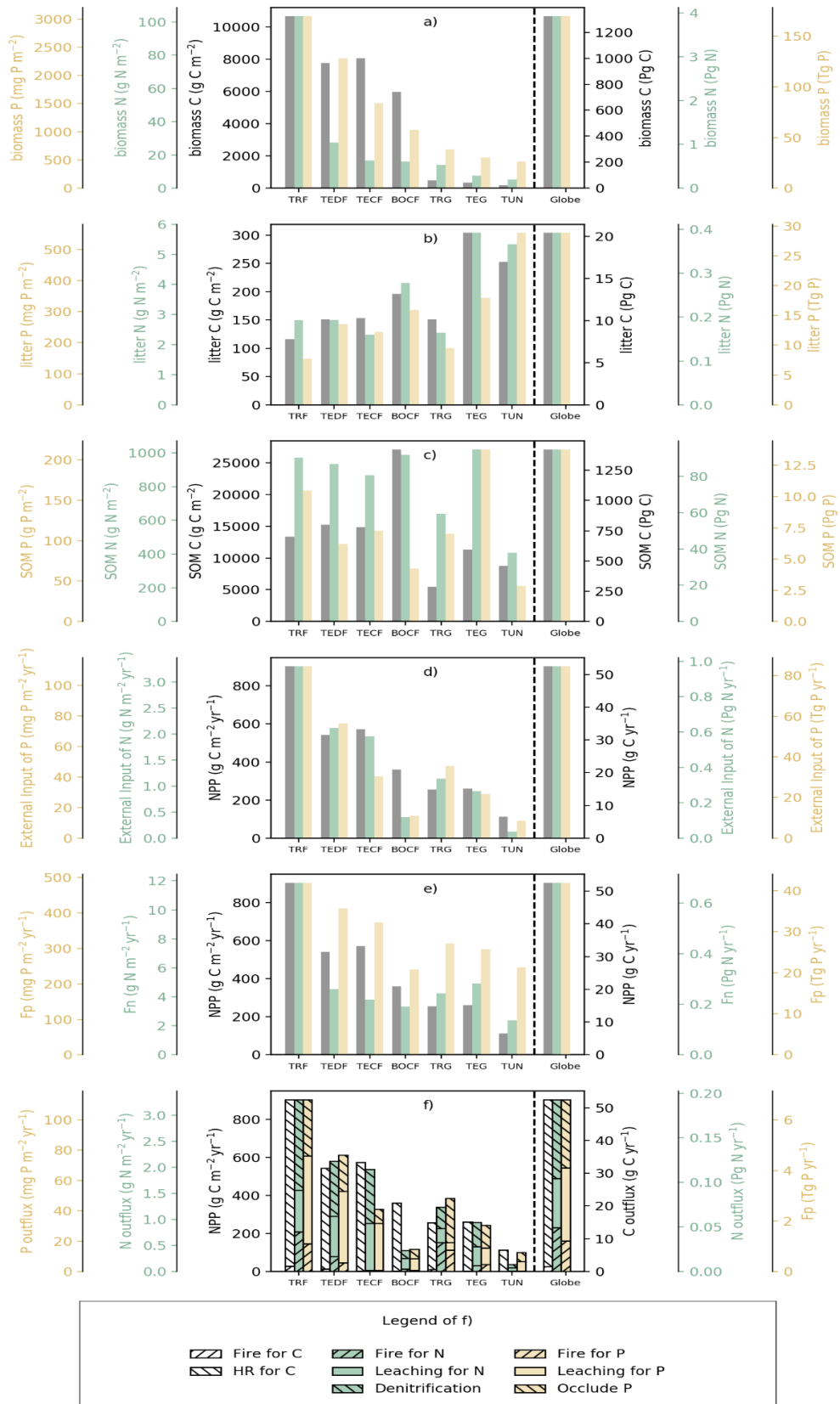
3 **Figure 3** Fluxes (numbers along arrows), mean residence times (in parentheses) and pool sizes of the N (blue) and P (red)

4 cycles in the terrestrial biosphere at steady state for the large biomes (a-g) and globe (h). The targeted biomes are tropical

5 rainforests (TRF, a), temperate deciduous forests (TEDF, b), temperate coniferous forests (TECF, c), boreal coniferous

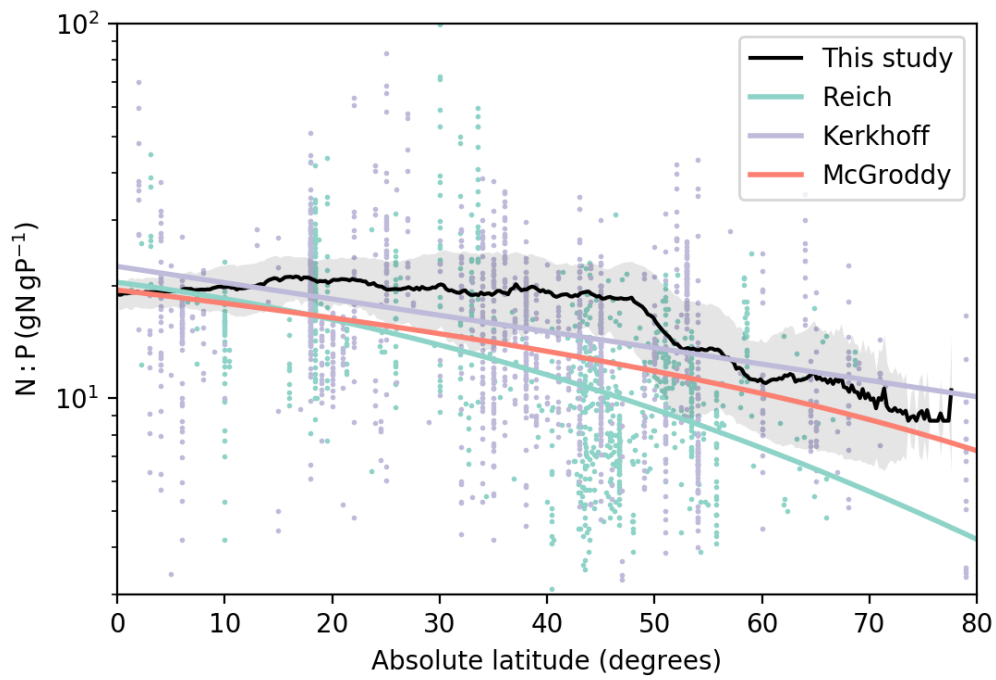
6 forests (BOCF, d), tropical/C4 grasslands (TRG, e), temperate/C3 grasslands (TEG, f) and tundra (TUN, g).

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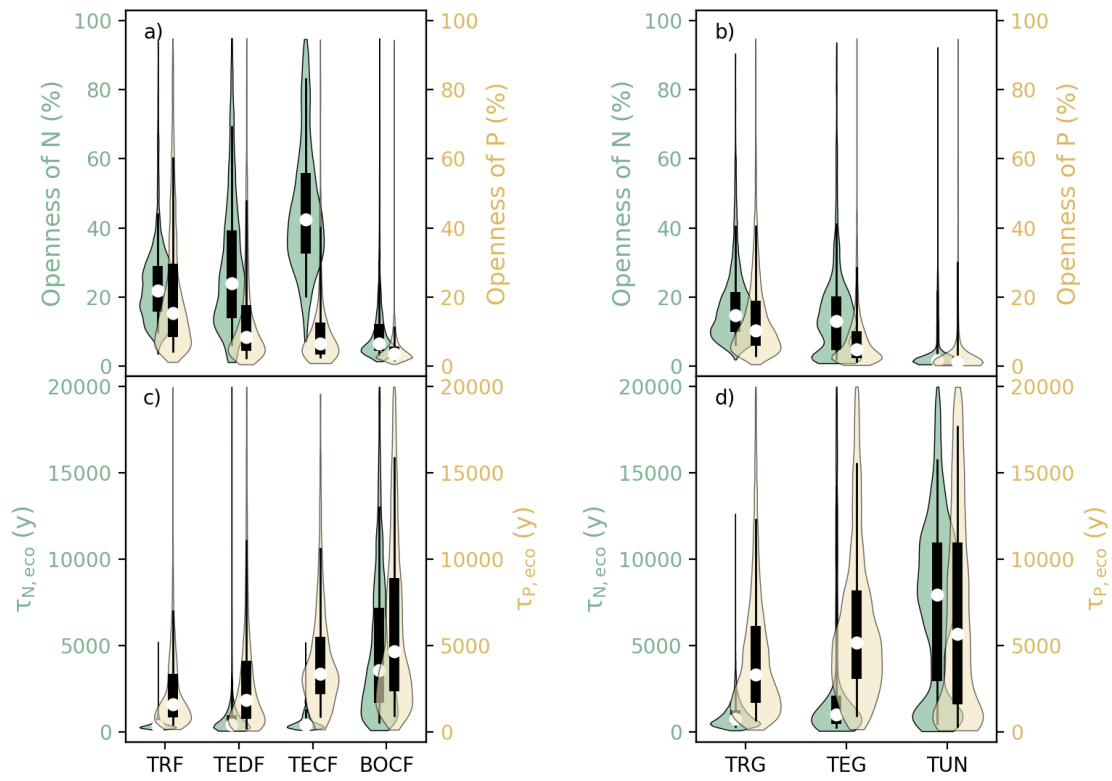
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Figure 4 Pool sizes and fluxes of C (black), N (green) and P (yellow) computed from GOLUM-CNP.



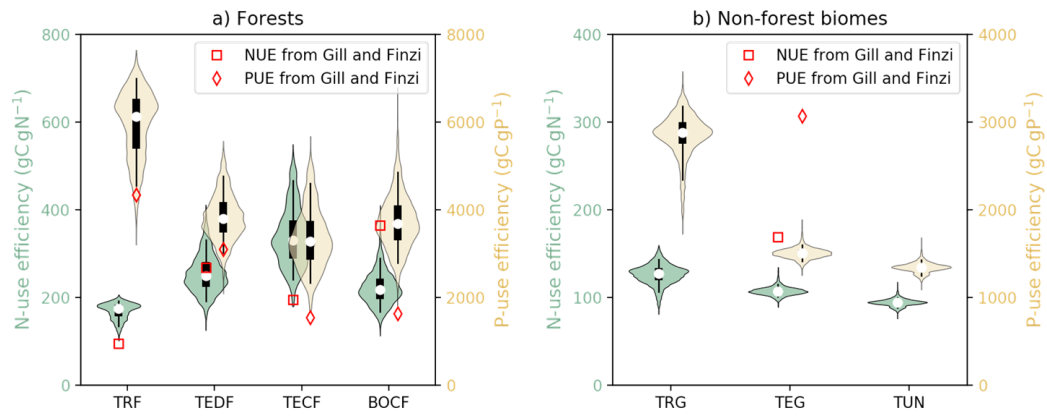
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Figure 4-5 Relationship between foliar N:P ratios (gN gP^{-1}) and absolute latitude. The black line is the mean N:P ratios from this study, and the shaded area is the one-sigma standard deviation of the N:P ratios for specific latitude. Colored lines are the regression trends of foliar N:P ratios as a function of absolute latitude from Reich and Oleksyn (2004; green), Kerkhoff et al. (2005; blue) and McGroddy et al. (2004; red). Dots are the raw data that Reich and Oleksyn (2004; green) and Kerkhoff et al. (2005; blue) used to derive their regression trends.



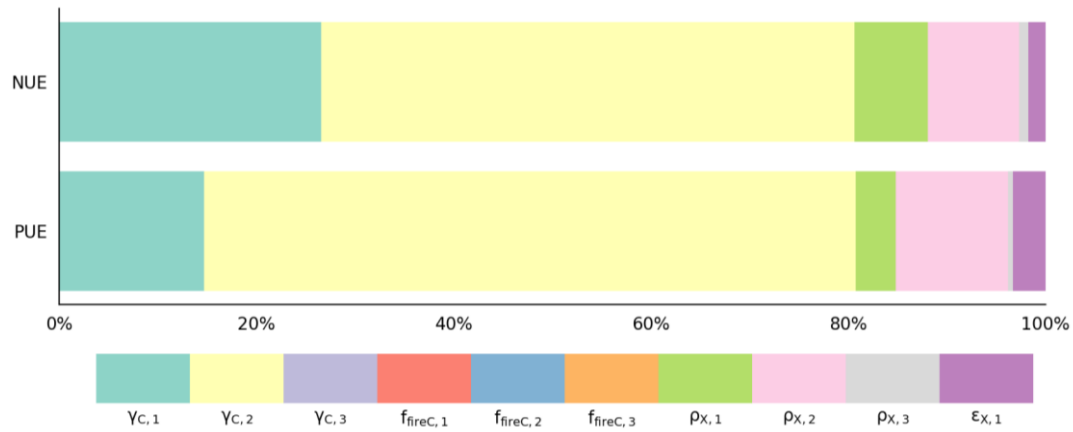
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2 **Figure 5-6** Violin plots of the openness of N and P cycling (the percentage of total plant uptake of N and P attributed to new
 3 nutrient inputs) for a) forest and b) grassland biomes. Residence times of N ($\tau_{N,eco}$) and P ($\tau_{P,eco}$) in c) forest ecosystems and d)
 4 grassland biomes. Open circles are medians of all grid cells within each biome, with balloons representing the probability
 5 density distribution of each value. Black whiskers indicate interquartile (thick) and 95% confidence intervals (thin). The
 6 biomes are tropical rainforests (TRF), temperate deciduous forests (TEDF), temperate coniferous forests (TECF), boreal
 7 coniferous forests (BOCF), tropical/C4 grasslands (TRG), temperate/C3 grasslands (TEG) and tundra (TUN).
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2 **Figure 6-7** Violin plots of N- and P-use efficiencies (NUE and PUE, the nutrient uptake by plants divided by GPP) of seven
3 biomes. Open circles are medians of all grid cells within each biome, with balloons representing the probability density
4 distribution of each value. Black whiskers indicate interquartile (thick) and 95% confidence intervals (thin). a) Forest biomes,
5 including tropical rainforests (TRF), temperate deciduous (TEDF), temperate coniferous (TECF) and boreal coniferous
6 forests (BOCF). b) Grassland biomes, including tropical/C4 (TRG), temperate/C3 grasslands (TEG) and tundra (TUN). Red
7 squares (NUE) and diamonds (PUE) are the independent estimates from site observations and other generic data sets
8 compiled and harmonized by Gill and Finzi (2016) based on site measurements of GPP and net N/P mineralization.
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$\gamma_{C,1}$: NPP-allocation fraction to foliage
 $\gamma_{C,2}$: NPP-allocation fraction to fine roots
 $\gamma_{C,3}$: NPP-allocation fraction to wood
 $f_{\text{fireC},1}$: fraction of fire to total outgoing flux from foliage
 $f_{\text{fireC},2}$: fraction of fire to total outgoing flux from fine roots
 $f_{\text{fireC},3}$: fraction of fire to total outgoing flux from wood
 $\rho_{X,1}(X \in \{N, P\})$: X:C ratio of foliage
 $\rho_{X,2}(X \in \{N, P\})$: X:C ratio of fine roots
 $\rho_{X,3}(X \in \{N, P\})$: X:C ratio of wood
 $\varepsilon_{X,1}(X \in \{N, P\})$: resorption coefficient of foliar nutrients

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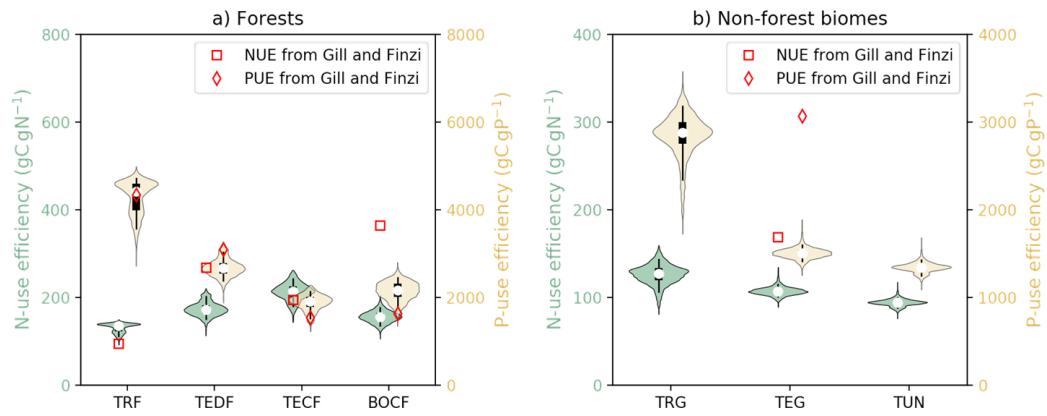
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Figure 8-9 Contribution of input data to the variance of the estimates of nutrient-use efficiencies ($X \in \{N, P\}$) for temperate coniferous forests. $\gamma_{C, i=1,2,3}$ are NPP-allocation fractions to foliage, fine roots and wood, respectively. $f_{\text{fireC}, i=1,2,3}$ are fractions of fire to total outgoing flux from foliage, fine roots and wood, respectively. $\rho_{X, i=1,2,3}$ ($X \in \{N, P\}$) are X:C ratios of foliage, fine roots and wood, respectively. $\varepsilon_{X,1}$ ($X \in \{N, P\}$) is the resorption coefficient of foliar nutrients.



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Figure 9-10 Violin plots of the nutrient-use efficiencies of the seven biomes from the experiment in which the allocation fraction of NPP to woody biomass and to leaves in coniferous forests is reduced. Open circles are the medians of all grid cells within each biome, with balloons representing the probability density distribution of each value. Black whiskers indicate interquartile (thick) and 95% confidence intervals (thin). The biomes are tropical rainforests (TRF), temperate deciduous forests (TEDF), temperate coniferous forests (TECF), boreal coniferous forests (BOCF), tropical/C4 grasslands (TRG), temperate/C3 grasslands (TEG) and tundra (TUN). The red squares (NUE) and diamonds (PUE) are the independent estimates from site observations and other generic data sets compiled and harmonized by Gill and Finzi (2016) based on site measurements of GPP and net N/P mineralization.

Supporting information of “GOLUM-CNP v1.0: a data-driven modeling of carbon, nitrogen and phosphorus cycles in major terrestrial biomes” by Yilong Wang, Philippe Ciais, Daniel Goll, Yuanyuan Huang, Yiqi Luo, Ying-Ping Wang, A. Anthony Bloom, Grégoire Broquet, Jens Hartmann, Shushi Peng, Josep Penuelas, Shilong Piao, Jordi Sardans, Benjamin D. Stocker, Rong Wang, Sönke Zaehle, Sophie Zechmeister-Boltenstern

S1 Comparison of the “openness” in this study and in Cleveland et al. (2013)

In this section, we compare the computation of “openness” index used in this study versus “proportion of new NPP fueled by new nutrient inputs” used in Cleveland et al. (2013). We take the indexes for N as an example, but the computation of corresponding indexes for P are similar.

Assume NPP is allocated to leaf, wood and roots by 0.5, 0.3 and 0.2, those pools with C:N ratios of 25, 150 and 50. In Cleveland's index, N is allocated in fractions of 0.5, 0.3 and 0.2 in NPP-N (C_j can be moved before C_{toN_j} in their Eq. S7). The new NPP is $(0.5*25 + 0.3*150 + 0.2*50) * I_N = 67.5 I_N$, where I_N represent external nutrients inputs, i.e. the sum of deposition and biological fixation for N which is available to vegetation. They assume that the remaining NPP must be totally fueled by resorption and net mineralization in the soil. They modelled the amount of N resorption but computed the amount of mineralization as the difference between nutrient demand, new nutrient inputs, and nutrient resorption (their Eqs. S5). So the “proportion of new NPP fueled by new nutrient inputs” index in Cleveland et al. (2013) equals $67.5 I_N/NPP$.

In this study, the N allocation is 0.5/25 : 0.3/150 : 0.2/50. Normalizing the values to ensure the sum of the fractions equals to 1 gives allocation fractions of N in NPP-N of 0.77 : 0.08 : 0.15. Under steady state, we have the relationship that the total N demand = $NPP*0.5/25+NPP*0.3/150+NPP*0.2/50=F_N+RSB_N$, where F_N is the uptake from inorganic N soil pool and RSB_N is the flux of resorbed N. As a result, the openness index $NO = I_N / (F_N + RSB_N) = 38.5 I_N/NPP$.

Because there is no evidence that how much the new N inputs is allocated in the vegetation, we chose to define the openness index only based on N and P fluxes rather than to convert N and P fluxes into NPP like Cleveland et al. (2013) did.

Table S1 C:N, C:P and N:P molar (atomic) ratios across major biomes from Zechmeister-Boltenstern et al. (2015). Targeted biomes are: tropical rain forests (TRF), temperate deciduous forests (TEDF), temperate coniferous forests (TECF), boreal coniferous forests (BOCF), tundra (TUN), tropical/C4 grasslands (TRG), and temperate/C3 grasslands (TEG). Note that N:P ratios in Zechmeister-Boltenstern et al. (2015) are not exactly equal to the ratio of C:P to C:N, but the differences are small.

		TRF	TEDF	TECF	BOCF	TRG	TEG	TUN
C:N	Foliage	25	25	59	49	39	25	49
	Root	47	59	67	57	39	88	54
	Wood	148	471	844	525	--	--	--
	Soil	16	19	20	32	25	10	13
C:P	Foliage	1027	867	1232	1049	753	1278	2167
	Root	3125	1962	1186	1574	1300	2829	1300
	Wood	13574	11179	24297	19734	--	--	--
	Soil	159	366	302	960	509	130	138
N:P	Foliage	43	36	23	23	20	53	45
	Root	52	22	18	30	32	27	20
	Wood	93	24	29	38	--	--	--
	Soil	13	20	15	31	31	--	11

Table S2 Mean residence time of C, N and P in ecosystems (unit: years). Targeted biomes are:

tropical rain forests (TRF), temperate deciduous forests (TEDF), temperate coniferous forests (TECF), boreal coniferous forests (BOCF), tundra (TUN), tropical/C4 grasslands (TRG), and temperate/C3 grasslands (TEG).

	TRF	TEDF	TECF	BOCF	TRG	TEG	TUN	Globe
C	29	49	48	106	40	67	101	38
N	382	1016	637	4834	987	3075	7896	1586
P	2520	3263	4413	6167	4483	6291	7077	6092

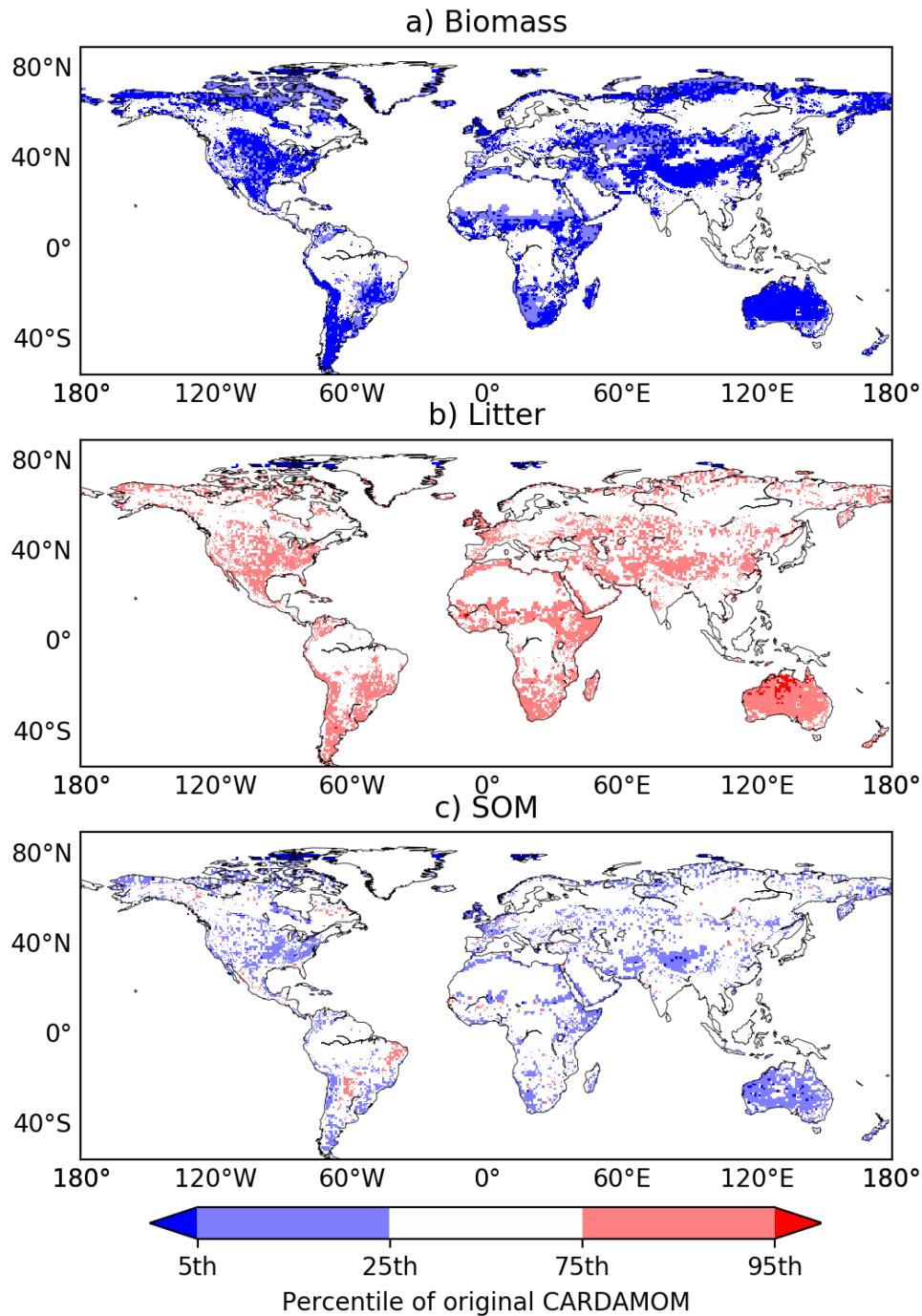


Fig. S1 Comparison between C pool sizes of transformed steady-state C cycle and original CARDAMOM results.

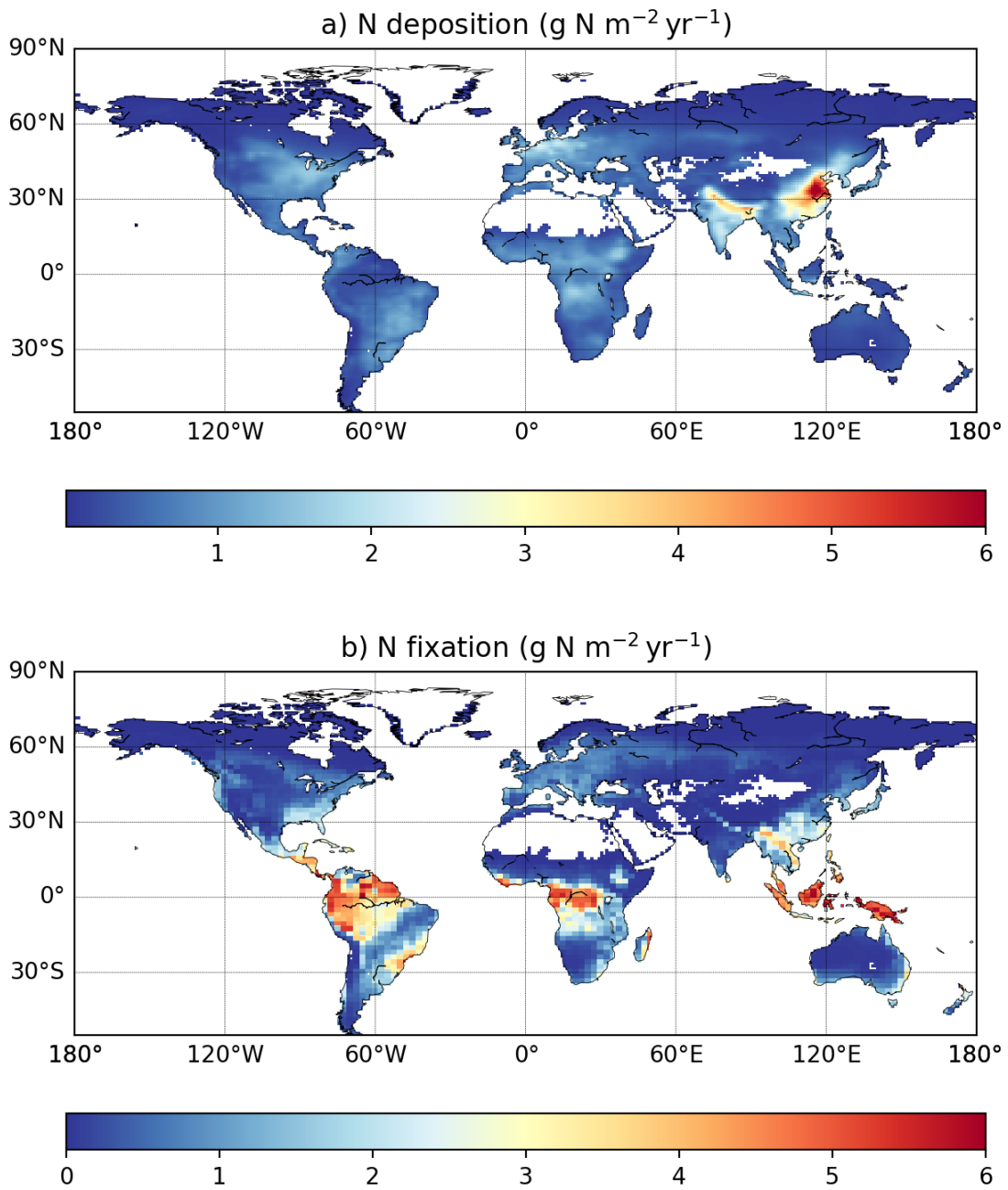


Fig. S2 Global external N inputs. a) N deposition from Wang et al. (2017). b) N fixation from Peng et al. (submitted)

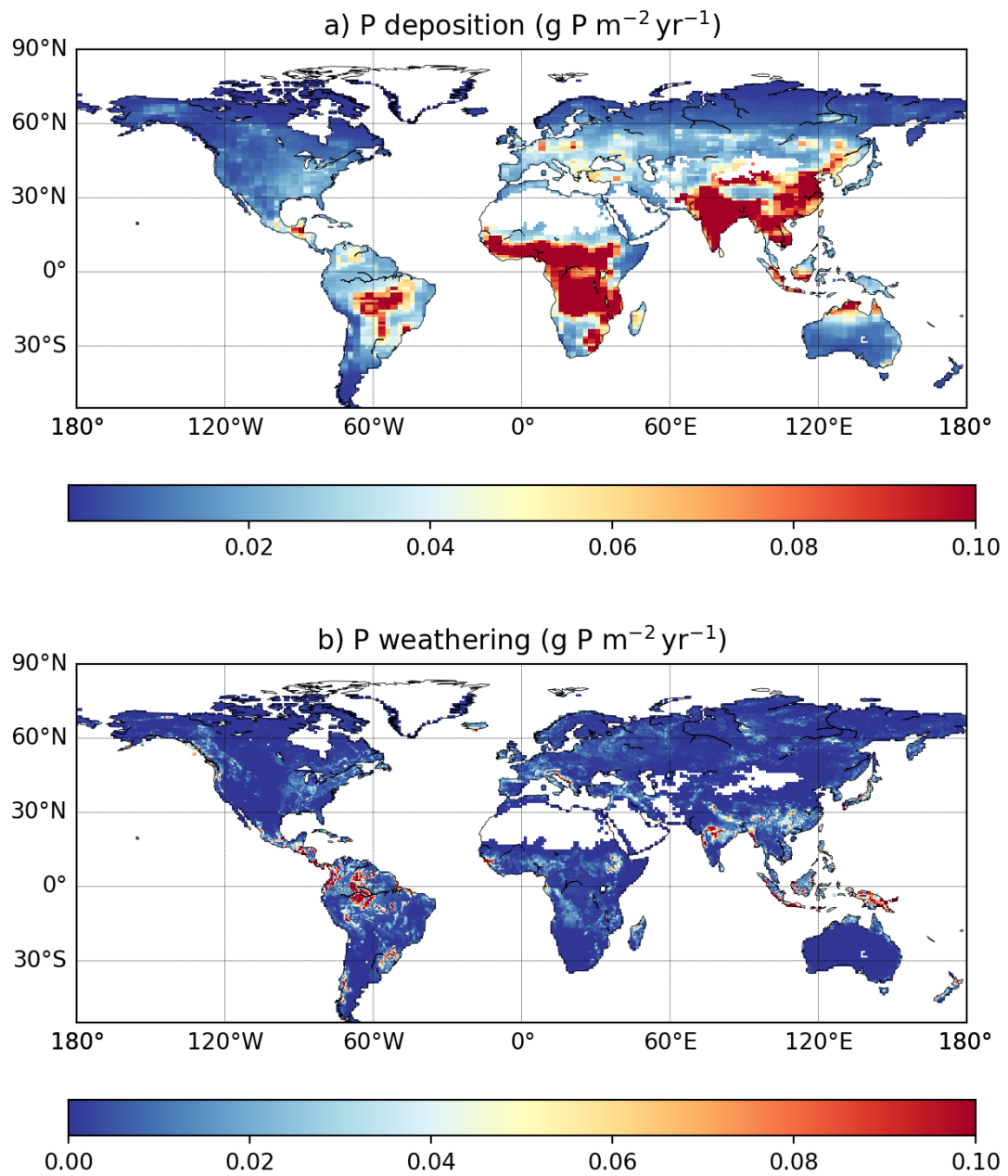


Fig. S3 Global external P inputs. a) P deposition from Wang et al. (2017). b) Release of P from rock weathering from Hartmann et al. (2014)

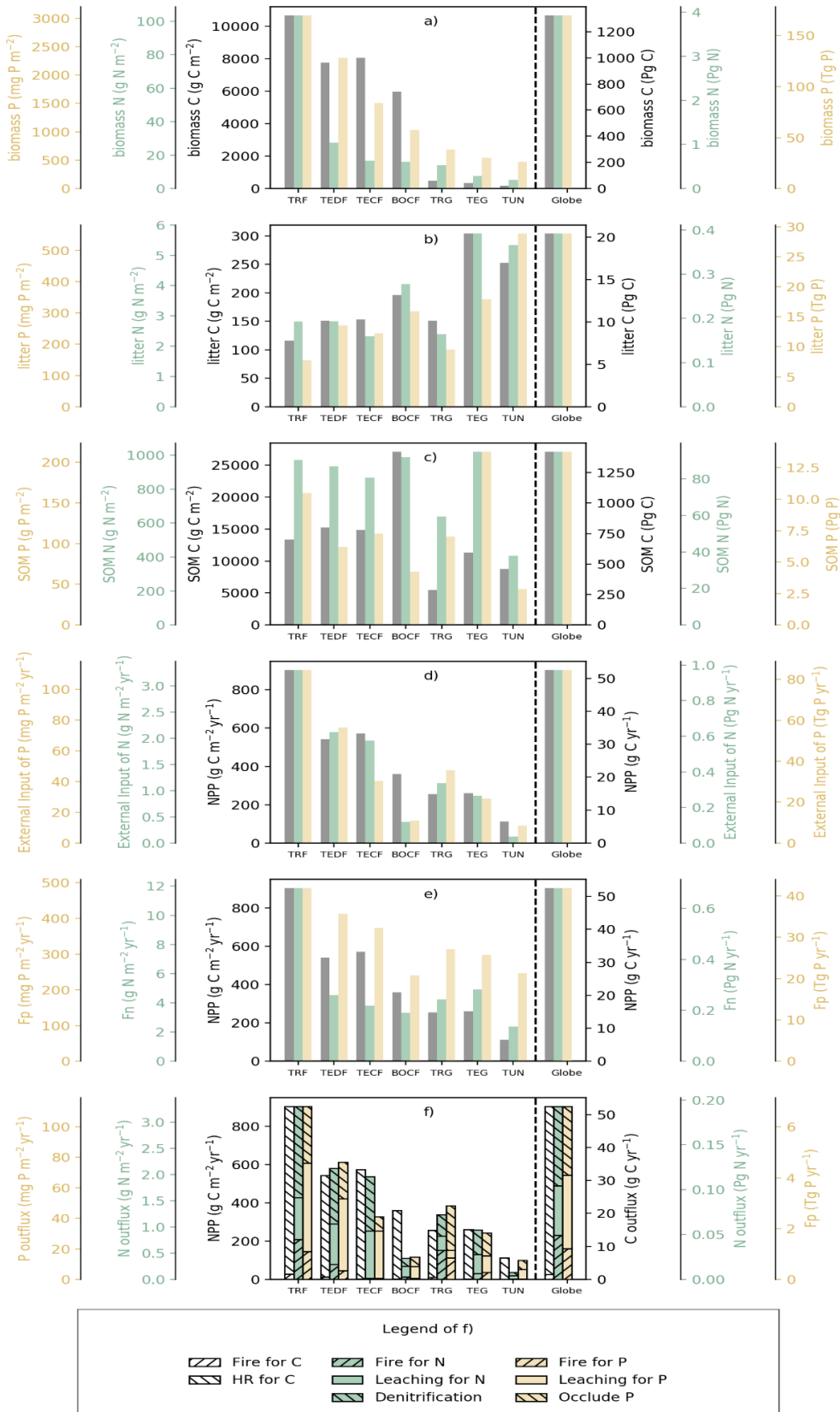


Figure S4 Pool sizes and fluxes of C (black), N (green) and P (yellow) computed from GOLUM CNP. The targeted biomes are tropical rainforests (TRF, a), temperate deciduous forests (TEDF, b), temperate coniferous forests (TECF, c), boreal coniferous forests (BOCF, d), tropical/C4 grasslands (TRG, e), temperate/C3 grasslands (TEG, f) and tundra (TUN, g).

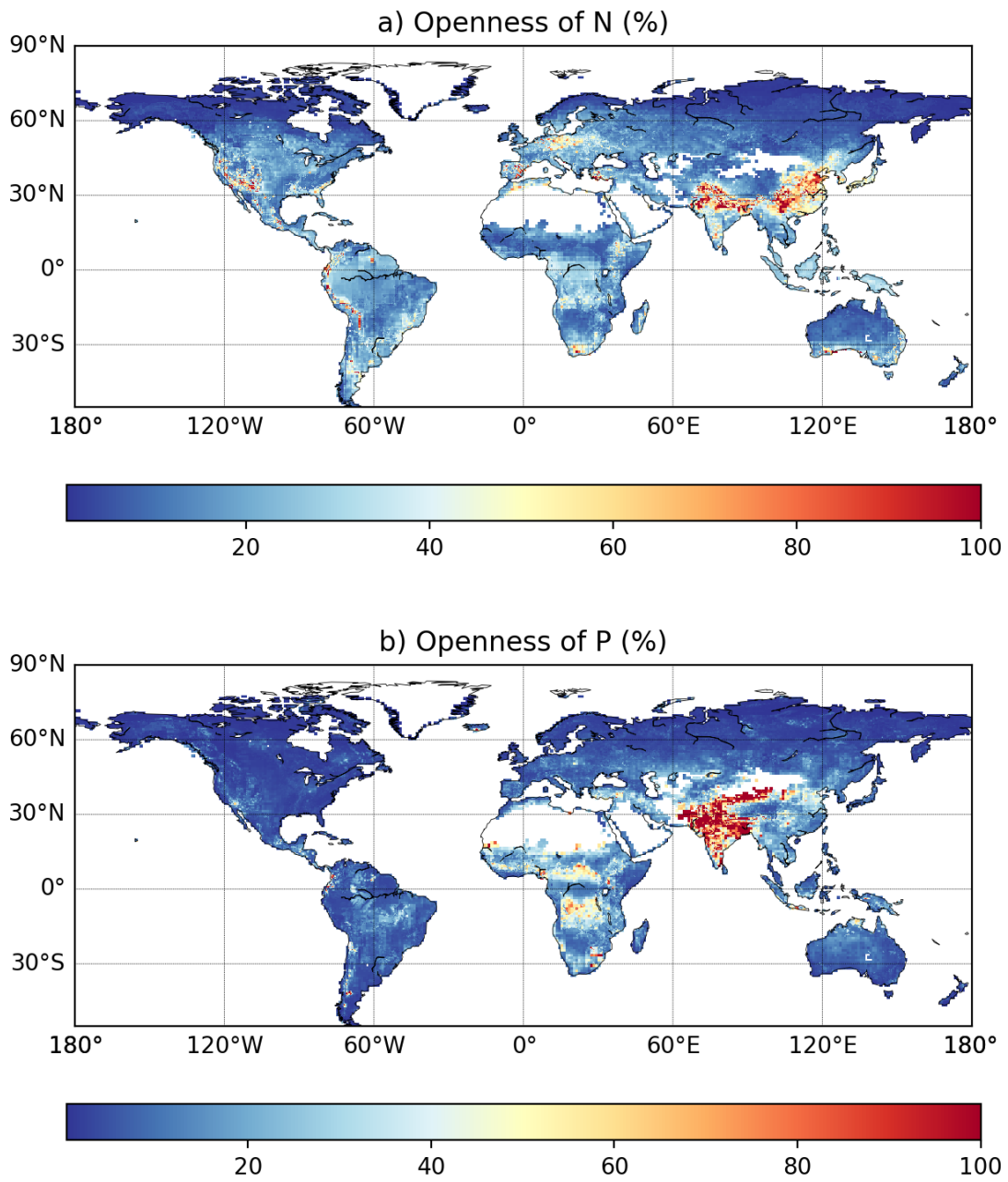


Fig. S5-S4 Global nutrient openness computed from GOLUM-CNP. a) Openness of N. b) Openness of P.

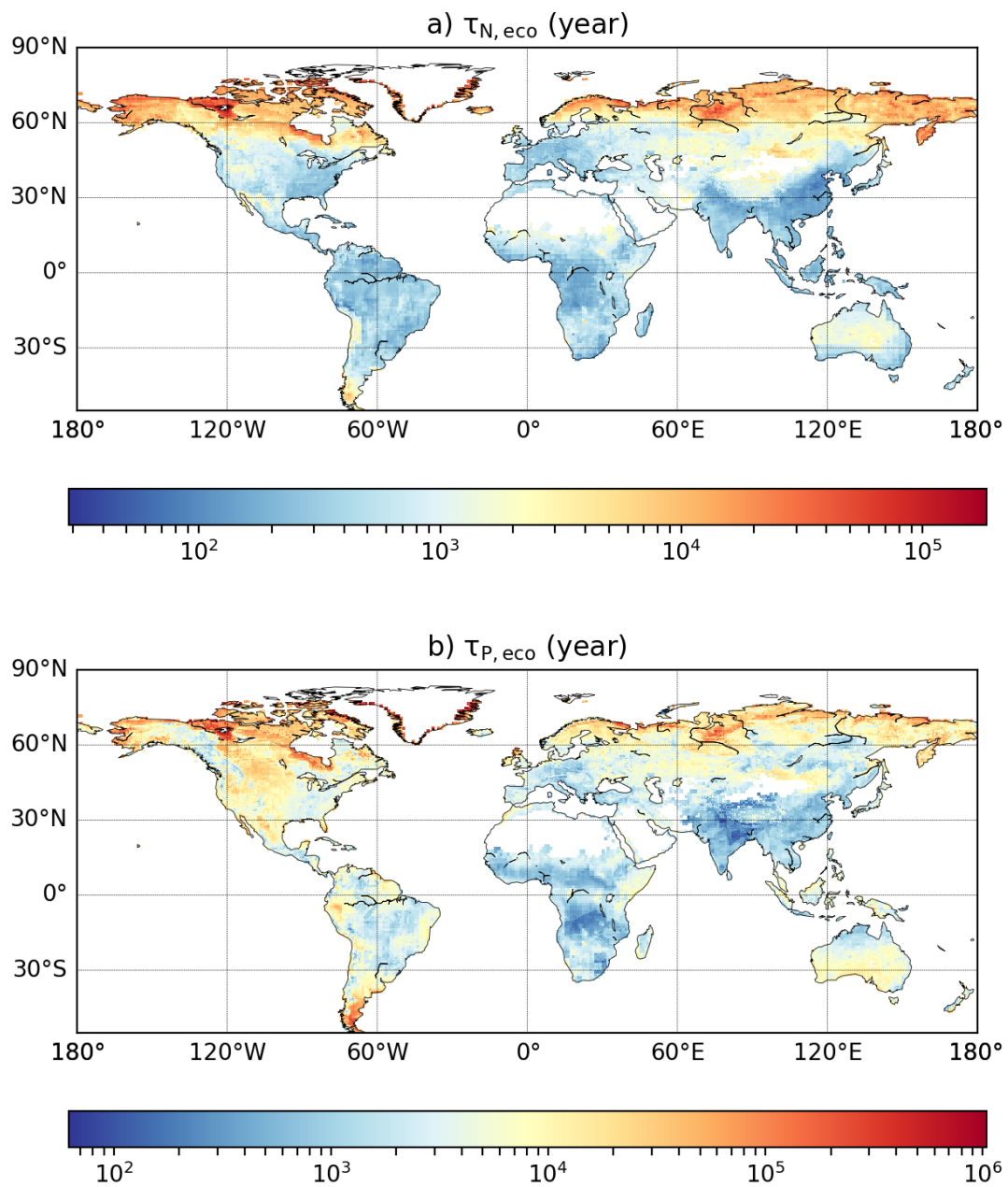


Fig. S6-S5 Global residence times of nutrients in the ecosystems. a) Residence times of N in the ecosystems. b) Residence times of P in the ecosystems

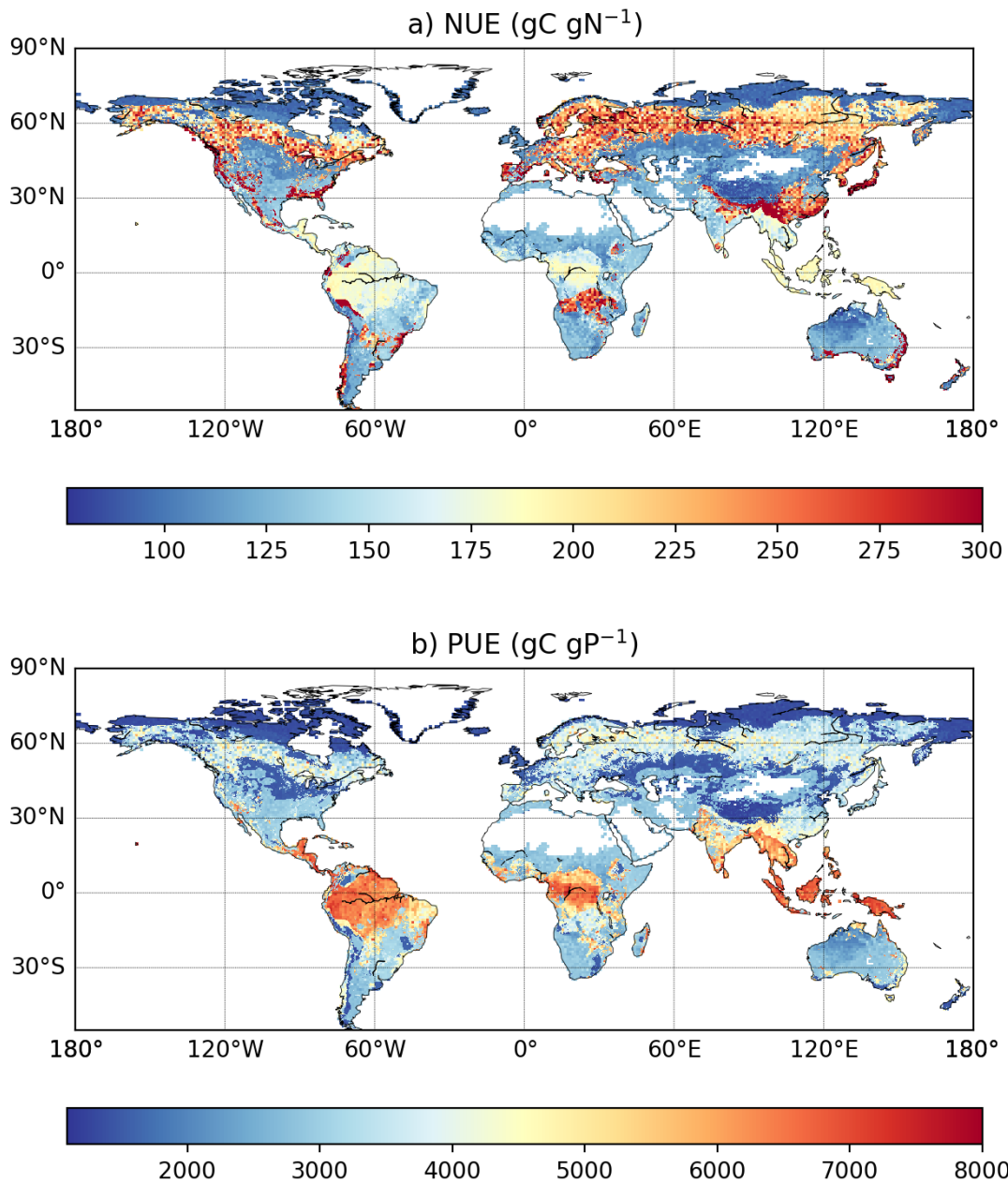


Fig. S7-S6 Global nutrient use efficiencies (the nutrient uptake by plants divided by GPP). a) N use efficiency. b) P use efficiency.

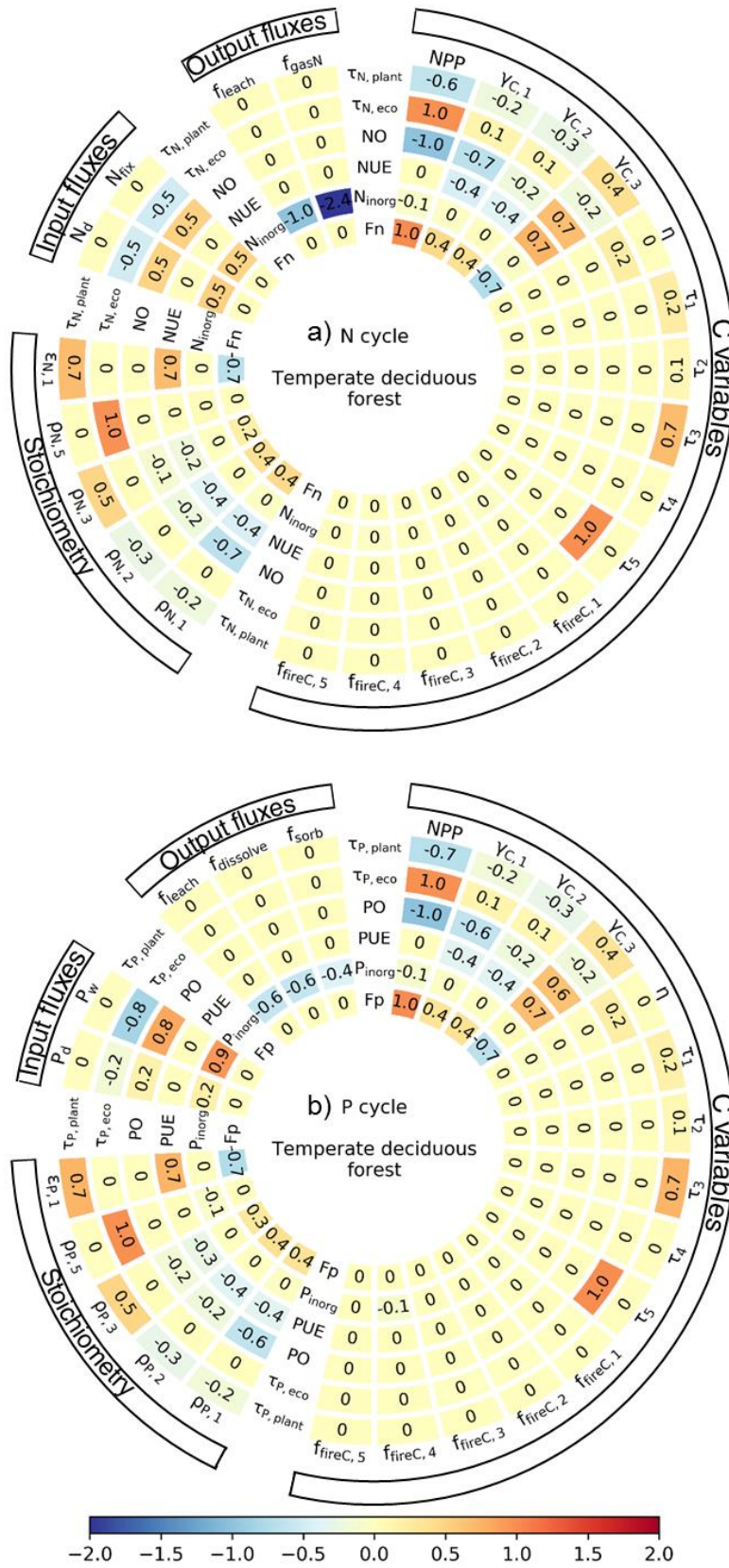


Figure S8-S7 Mean sensitivity of the estimates of rates of nutrient uptake, inorganic nutrients, nutrient-use efficiencies, openness, turnover time of nutrients in the ecosystem and turnover time of nutrients in plants to the input variables for temperate deciduous forests.

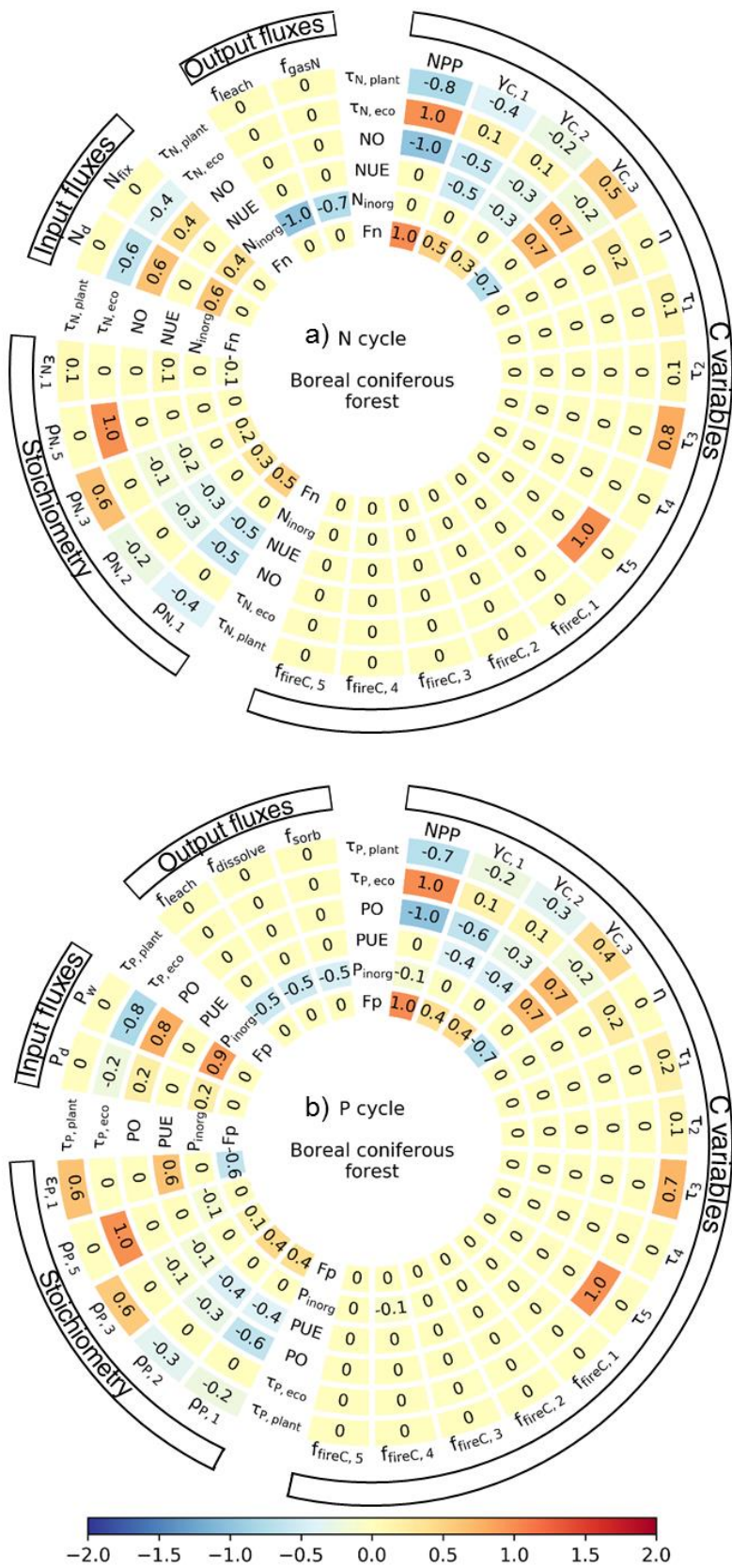


Figure S10-S9 Mean sensitivity of the estimates of rates of nutrient uptake, inorganic nutrients, nutrient-use efficiencies, openness, turnover time of nutrients in the ecosystem and turnover time of nutrients in plants to the input variables for boreal coniferous forests.

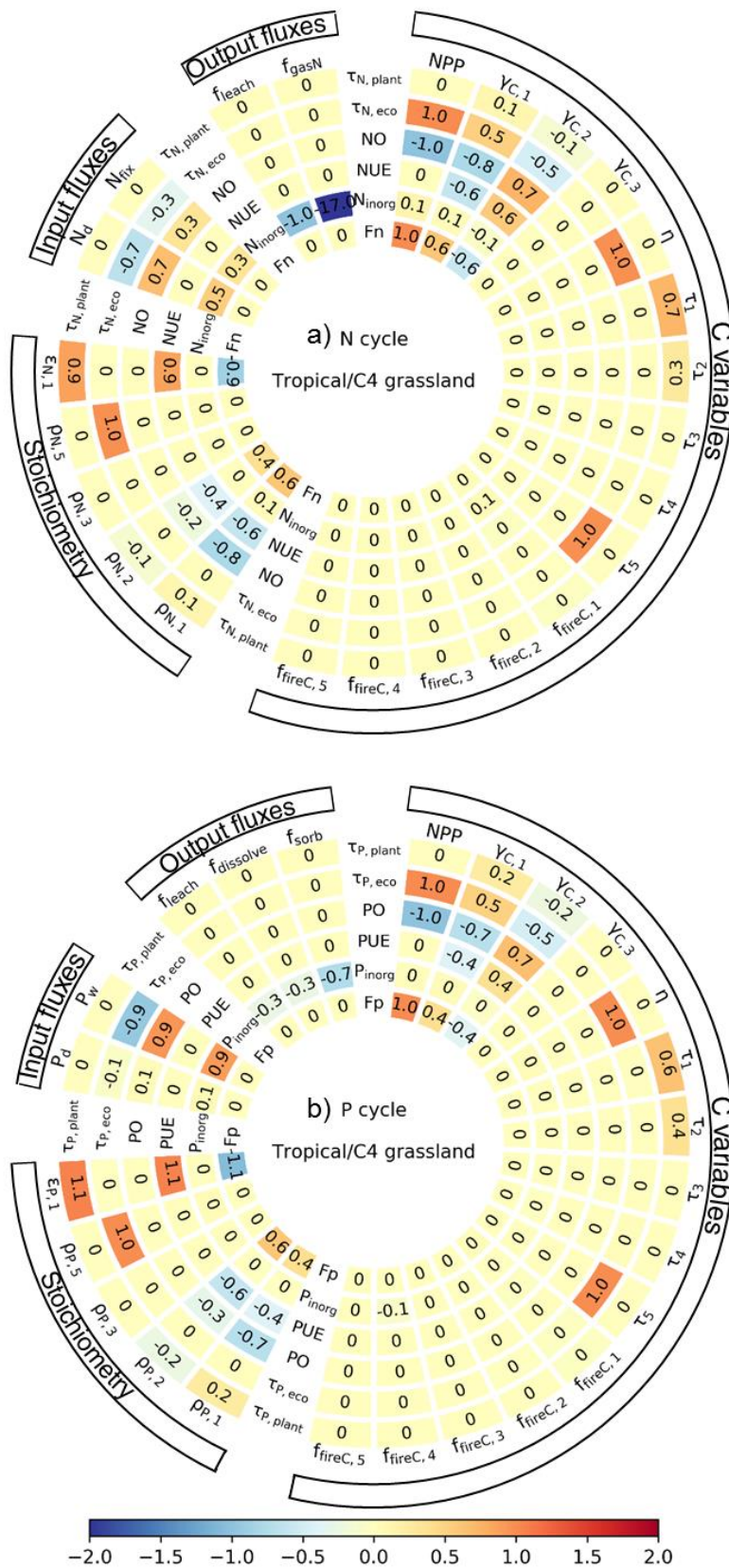


Figure S11-S10. Mean sensitivity of the estimates of rates of nutrient uptake, inorganic nutrients, nutrient-use efficiencies, openness, turnover time of nutrients in the ecosystem and turnover time of nutrients in plants to the input variables for tropical/C4 grasslands.

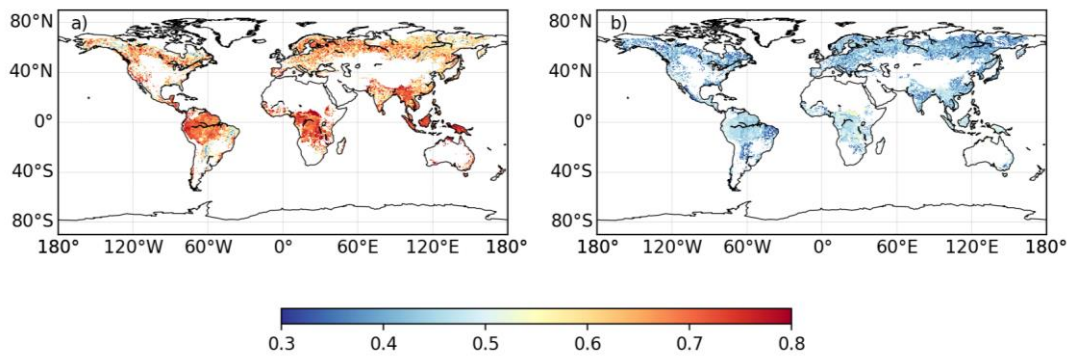


Figure S14-S13 NPP allocation fractions to woody biomass in original CARDAMOM (a) and adjusted (see Sect. 5.1) carbon cycle model (b).

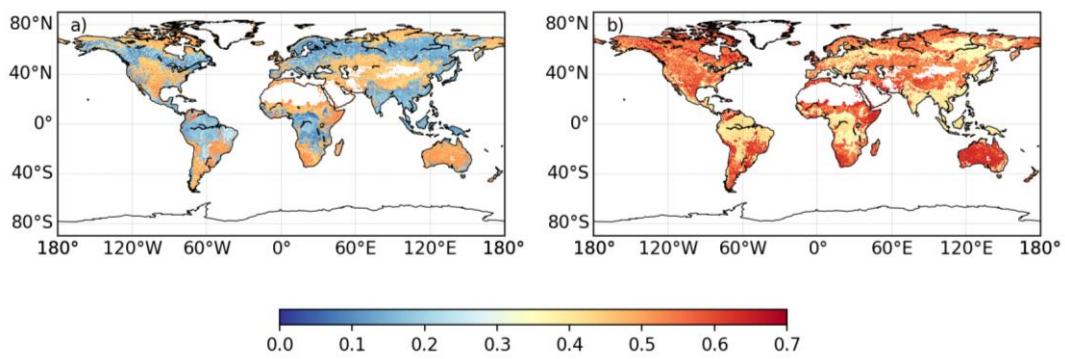


Figure S15-S14 NPP allocation fractions to fine roots in original CARDAMOM (a) and adjusted (see Sect. 5.1) carbon cycle model (b).

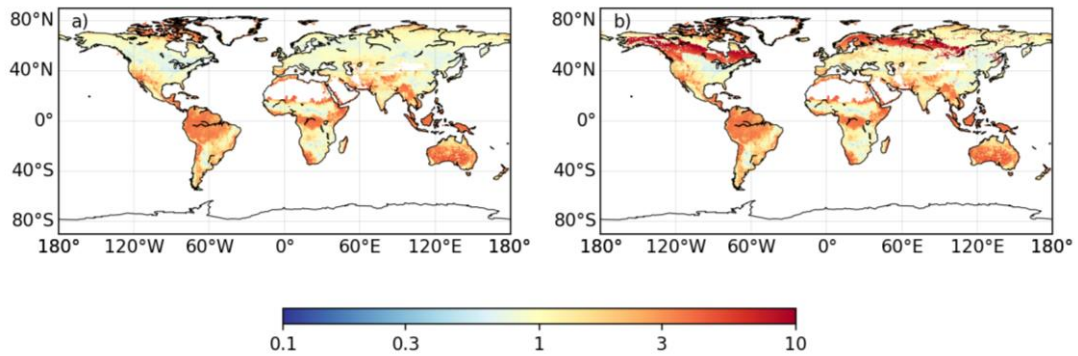


Figure S16-S15 Leaf longevity in original CARDAMOM (a) and adjusted (see Sect. 5.1) carbon cycle model (b).

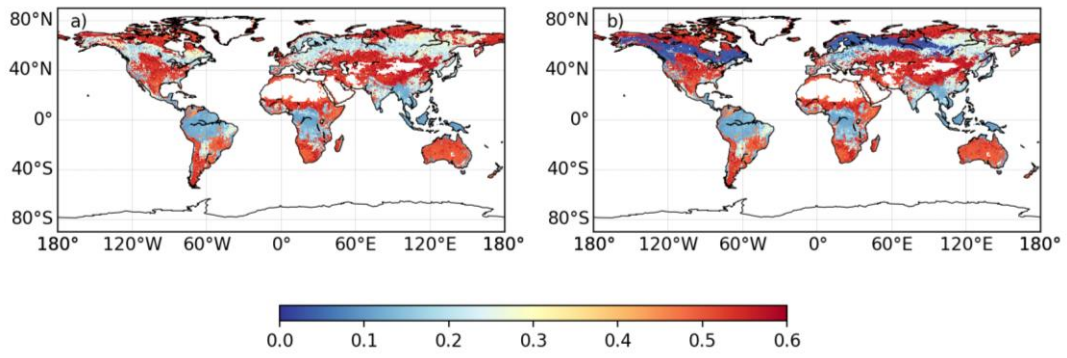


Figure S17-S16 NPP allocation fractions to foliage in original CARDAMOM (a) and adjusted (see Sect. 5.1) carbon cycle model (b).