

Global tropospheric effects of aromatic chemistry with the SAPRC-11 mechanism implemented in GEOS-Chem version 9-02

Yingying Yan^{1,2}, David Cabrera-Perez³, Jintai Lin², Andrea Pozzer³, Lu Hu⁴, Dylan B. Millet⁵, William C. Porter⁶, Jos Lelieveld³

¹ Department of Atmospheric Sciences, School of Environmental Studies, China University of Geosciences (Wuhan), 430074, Wuhan, China

² Laboratory for Climate and Ocean-Atmosphere Studies, Department of Atmospheric and Oceanic Sciences, School of Physics, Peking University, Beijing 100871, China

³ Max-Planck-Institute for Chemistry, Atmospheric Chemistry Department, Mainz, Germany

⁴ Department of Chemistry and Biochemistry, University of Montana, Missoula, MT, USA

⁵ Department of Soil, Water, and Climate, University of Minnesota, St. Paul, MN, USA

⁶ Department of Civil and Environmental Engineering, Massachusetts Institute of Technology, 77 Massachusetts Avenue, Cambridge, MA 02139-4307, USA

Correspondance : Jintai Lin, linjt@pku.edu.cn

Abstract

The GEOS-Chem model has been updated with the SAPRC-11 aromatics chemical mechanism, with the purpose of evaluating global and regional effects of the most abundant aromatics (benzene, toluene, xylenes) on the chemical species important for tropospheric oxidation capacity. The model evaluation based on surface and aircraft observations indicates good agreement for aromatics and ozone. A comparison between scenarios in GEOS-Chem with simplified aromatic chemistry (as in the standard setup, with no ozone formation from related peroxy radicals or recycling of NO_x) and with the SAPRC-11 scheme reveals relatively slight changes in ozone, hydroxyl radical, and nitrogen oxides on a global mean basis (1–4%), although remarkable regional differences (5–20%) exist near the source regions. NO_x decreases over the source regions and increases in the remote troposphere, due mainly to more efficient transport of peroxyacetyl nitrate (PAN), which is increased with the SAPRC aromatic chemistry. Model ozone mixing ratios with the updated aromatic chemistry increase by up to 5 ppb (more than 10%), especially in industrially polluted regions. The ozone change is partly due to the direct influence of aromatic oxidation products on ozone production rates, and in part to the altered spatial distribution of NO_x that enhances the tropospheric ozone production efficiency. Improved representation of aromatics is important to simulate the tropospheric oxidation.

1. Introduction

1 Non-methane volatile organic compounds (NMVOCs) play important roles in the tropospheric
2 chemistry, especially in ozone production (Atkinson, 2000; Seinfeld and Pandis, 2012). Aromatic
3 hydrocarbons such as benzene (C_6H_6), toluene (C_7H_8) and xylenes (C_8H_{10}) make up a large
4 fraction of NMVOCs (Ran et al., 2009; Guo et al., 2006; You et al., 2008) in the atmosphere of
5 urban and semi-urban areas. They are important precursors of secondary organic aerosol (SOA),
6 peroxyacetyl nitrate (PAN), and ozone (Kansal, 2009; Tan et al., 2012; Porter et al., 2017). In
7 addition, many aromatic compounds can cause detrimental effects on human health and plants
8 (Manuela et al., 2012; Sarigiannis and Gotti, 2008; Michalowicz and Duda, 2007).

9 Aromatics are released to the atmosphere by biomass burning as well as fossil fuel evaporation
10 and burning (Cabrera-Perez et al., 2016; Na et al., 2004). The dominant oxidation pathway for
11 aromatics is via reaction with hydroxyl radical (OH, the dominant atmospheric oxidant), followed
12 by reaction with nitrate radical (NO_3) (Cabrera-Perez et al., 2016; and references therein). The
13 corresponding aromatic oxidation products could be involved in many atmospheric chemical
14 processes, which can affect OH recycling and the atmospheric oxidation capacity (Atkinson and
15 Arey, 2003; Calvert et al., 2002; Bejan et al., 2006; Chen et al., 2011). A realistic model
16 description of aromatic compounds is necessary to improve our understanding of their effects on
17 the chemistry in the atmosphere. However, up to now few regional or global-scale chemical
18 transport models (CTMs) include detailed aromatic chemistry (Lewis et al., 2013; Cabrera-Perez
19 et al., 2016).

20 Despite the potentially important influence of aromatic compounds on global atmospheric
21 chemistry, their effect on global tropospheric ozone formation in polluted urban areas is less
22 analyzed with the model simulation. The main source and sink processes of tropospheric ozone
23 are photochemical production and loss, respectively (Seinfeld and Pandis 2006; Monks et al.,
24 2015; Yan et al., 2016). Observation-based approaches alone cannot provide a full picture of
25 ozone-source attribution for the different NMVOCs. Such ozone-source relationships are needed
26 to improve policymaking strategies to address hemispheric ozone pollution (Chandra et al.,
27 2006). Numerical chemistry-transport models allow us to explore the importance of impacts from
28 aromatics and to attribute observed changes in ozone concentrations to particular sources
29 (Stevenson et al., 2006; Stevenson et al., 2013; Zhang et al., 2014). Current global CTMs
30 reproduce much of the observed regional and seasonal variability in tropospheric ozone
31 concentrations. However, some systematic biases can occur, most commonly an overestimation
32 over the northern hemisphere (Fiore et al., 2009; Reidmiller et al., 2009; Yan et al., 2016, 2018a,
33 b; Ni et al., 2018) due to incomplete representation of physical and chemical processes, and
34 biases in emissions and transport, including the parameterization of small-scale processes and
35 their feedbacks to global-scale chemistry (Chen et al., 2009; Krol et al., 2005; Yan et al., 2014;
36 Yan et al., 2016).

37 Another motivation for the modeling comes from recent updates in halogen (bromine-chlorine)
38 chemistry, which when implemented in GEOS-Chem, a global chemical transport model being
39 used extensively for tropospheric chemistry and transport studies (Zhang and Wang, 2016; Yan et

al., 2014; Shen et al., 2015; Lin et al., 2016), decrease the global burden of ozone significantly (by 14%; 2–10 ppb in the troposphere) (Schmidt et al., 2017). This ozone burden decline is driven by decreased chemical ozone production due to halogen-driven nitrogen oxides ($\text{NO}_x = \text{NO} + \text{NO}_2$) loss; and the ozone decline lowers global mean tropospheric OH concentrations by 11%. Thus GEOS-Chem starts to exhibit low ozone biases compared to ozonesonde observations (Schmidt et al., 2017), particularly in the southern hemisphere, implying that some mechanisms (e.g., due to aromatics) are currently missing from the model.

A simplified aromatic oxidation mechanism has previously been employed in GEOS-Chem (e.g., Fischer et al., 2014; Hu et al., 2015), which is still used in the latest version v12.0.0. In that simplified treatment, oxidation of benzene (B), toluene (T), and xylene (X) by OH (Atkinson et al., 2000) is assumed to produce first-generation oxidation products (xRO_2 , $x = \text{B}, \text{T}, \text{or X}$). And these products further react with hydrogen peroxide (HO_2) or nitric oxide (NO) to produce LxRO_{2y} ($y = \text{H or N}$), passive tracers which are excluded from tropospheric chemistry. Thus in the presence of NO_x , the overall reaction is aromatic + OH + NO = inert tracer. While such a simplified treatment can suffice for budget analyses of the aromatic species themselves, it does not capture ozone production from aromatic oxidation products.

In this work, we update the aromatics chemistry in GEOS-Chem based on the SAPRC-11 mechanism, and use the updated model to analyze the global and regional scale chemical effects of the most abundant aromatics in the gas phase (benzene, toluene, xylenes) in the troposphere. Specifically, we focus on the impact on ozone formation (due to aromatics oxidation), as this is of great interest for urban areas and can be helpful for developing air pollution control strategies. Further targets are the changes to the NO_x spatial distribution and OH recycling. Model results for aromatics and ozone mixing ratios are evaluated by comparison with observations from surface and aircraft campaigns in order to constrain model accuracy. Finally, we discuss the global effects of aromatics on tropospheric chemistry including ozone, NO_x and HO_x ($\text{HO}_x = \text{OH} + \text{HO}_2$).

The rest of the paper is organized as follows. Section 2 describes the GEOS-Chem model setups, including the updates in aromatics chemical mechanism. A description of the observational datasets for aromatics and ozone is given in Sect. 3. Section 4 presents the model evaluation for aromatics based on the previously mentioned set of aircraft and surface observations, and evaluates modeled surface ozone with measurements from three networks. An analysis of the tropospheric impacts on ozone, NO_x , and OH, examining the difference between models results with simplified (as in the standard model setup) and with SAPRC-11 aromatic chemistry, is presented in Section 5. Section 6 concludes the present study.

2. Model description and setup

We use the GEOS-Chem CTM (version 9-02, available at <http://geos-chem.org/>) to interpret the importance of aromatics in tropospheric chemistry and ozone production. GEOS-Chem is a

global 3-D chemical transport model for a wide range of atmospheric composition problems. It is driven by meteorological data provided from the Goddard Earth Observing System (GEOS) of the NASA Global Modeling Assimilation Office (GMAO). A detailed description of the GEOS-Chem model is available at http://acmg.seas.harvard.edu/geos/geos_chem_narrative.html. Here, the model is run at a horizontal resolution of 2.5° long. \times 2° lat. with a vertical grid containing 47 layers (including 10 layers of ~ 130 m thickness each below 850 hPa), as driven by the GEOS-5 assimilated meteorological fields. The chemistry time step is 0.5 h, while the transport time step is 15 min in the model. A non-local scheme implemented by Lin and McElroy (2010) is used for vertical mixing in the planetary boundary layer. Model convection adopts the Relaxed Arakawa-Schubert scheme (Rienecker et al., 2008). Stratospheric ozone production employs the Linoz scheme (McLinden et al., 2000). Dry deposition for aromatic compounds is implemented following the scheme by Hu et al. (2015), which uses a standard resistance-in-series model (Wesely, 1989) and Henry's law constants for benzene (0.18 M atm^{-1}), toluene (0.16 M atm^{-1}), and xylenes (0.15 M atm^{-1}) (Sander, 1999).

2.1 Emissions

For anthropogenic NMVOCs emission including aromatic compounds (benzene, toluene, and xylenes), here we use emission inventory from the RETRO (REanalysis of the TROpospheric chemical composition) (Schultz et al., 2007). The global anthropogenic RETRO (version 2; available at <ftp://ftp.retro.enes.org/>) inventory includes monthly emissions for 24 distinct chemical species during 1960–2000 with a resolution of 0.5° long. \times 0.5° lat. (Schultz et al., 2007). It is implemented in GEOS-Chem by regridding to the model resolution (2.5° long. \times 2.0° lat.). Emission factors in RETRO are calculated on account of economic and technological considerations. In order to estimate the time dependence of anthropogenic emissions, RETRO also incorporate behavioral aspects (Schultz et al., 2007). The implementation of the monthly RETRO emission inventory in GEOS-Chem is described by Hu et al. (2015), which linked the RETRO species into the corresponding model tracers. Here the model speciation of xylenes includes m-xylene, p-xylene, o-xylene and ethylbenzene (Hu et al., 2015). The most recent RETRO data (for 2000) is used for the GEOS-Chem model simulation and the calculated annual global anthropogenic NMVOCs are $\sim 71 \text{ TgC}$. On a carbon basis, the global aromatics (benzene + toluene + xylenes) source accounts for $\sim 23\%$ (16 TgC) of the total anthropogenic NMVOCs. Figure 1 shows the spatial distribution of anthropogenic emissions for benzene, toluene, and xylenes, respectively. Anthropogenic benzene emissions in Asia (mainly over eastern China and India) are larger than those from other source regions (e.g., over the Europe and eastern US).

Global NO_x anthropogenic emissions are taken from the EDGAR (Emission Database for Global Atmospheric Research) v4.2 inventory. The global inventory has been replaced by regional inventories in China (MEIC, base year: 2008), Asia (excluding China; INTEx-B, 2006), the US (NEI05, 2005), Mexico (BRAVO, 1999), Canada (CAC, 2005), and Europe (EMEP, 2005). Details on these inventories and on the model NO_x anthropogenic emissions are shown in Yan et al. (2016).

Biomass burning emissions of aromatics and other chemical species (e.g., NO_x) in GEOS-Chem are calculated based on the monthly Global Fire Emission Database version 3 (GFED3) inventory (van der Werf et al., 2010). Natural emissions of NO_x (by lightning and soil) and of biogenic NMVOCs are calculated online by parameterizations driven by model meteorology. Lightning NO_x emissions are parameterized based on cloud top heights (Price and Rind, 1992), and are further constrained by the lightning flash counts detected from satellite instruments (Murray et al., 2012). Soil NO_x emissions are described in Hudman et al. (2012). Biogenic emissions of NMVOCs are calculated by MEGAN (Model of Emissions of Gases and Aerosols from Nature) v2.1 with the Hybrid algorithm (Guenther et al., 2012).

2.2 Updated aromatic chemistry

In the GEOS-Chem model setup, the current standard chemical mechanism with simplified aromatic oxidation chemistry is based on Mao et al. (2013), which is still the case for the latest version v12.0.0. As mentioned in the introduction, this simplified mechanism acts as strong sinks of both HO_x and NO_x , because no HO_x are regenerated in this reaction, and NO is consumed without regenerating NO_2 . However, it is reasonably well established that aromatics tend to be radical sources, forming highly reactive products that photolyze to form new radicals, and regenerating radicals in their initial reactions (Carter, 2010a, b; Carter and Heo, 2013). A revised mechanism that takes the general features of aromatics mechanisms into account would be much more reactive, given the reactivity of the aromatic products.

This work uses a more detailed and comprehensive aromatics oxidation mechanism: the State-wide Air Pollution Research Center version 11 (SAPRC-11) aromatics chemical mechanism. SAPRC-11 is an updated version of the SAPRC-07 mechanism (Carter and Heo, 2013) to give better simulations of recent environmental chamber experiments. The SAPRC-07 mechanism underpredicted NO oxidation and O_3 formation rates observed in recent aromatic- NO_x environmental chamber experiments. The new aromatics mechanism, designated SAPRC-11, is able to reproduce the ozone formation from aromatic oxidation that is observed in almost all environmental chamber experiments, except for higher (>100 ppb) NO_x (Carter and Heo, 2013). Table S1 lists new model species in addition to those in the standard GEOS-Chem model setup. Table S2 lists the new reactions and rate constants. In this mechanism, the tropospheric consumption process of aromatics is mainly reaction with OH .

As discussed by Carter (2010a, b), aromatic oxidation has two possible OH reaction pathways: OH radical addition and H-atom abstraction (Atkinson, 2000). In SAPRC-11, taking toluene as an example in Table S2, the reactions following abstraction lead to three different formation products: an aromatic aldehyde (represented as the *BALD* species in the model), a ketone (*PROD2*), and an aldehyde (*RCHO*). The largest yield of toluene oxidation is the reaction after OH addition of aromatic rings. The OH -aromatic adduct is reaction with O_2 to form an OH -aromatic- O_2 adduct or HO_2 and a phenolic compound (further consumed by reactions with OH and NO_3 radicals). The OH -aromatic- O_2 adduct further undergoes two competing unimolecular reactions to ultimately form OH , HO_2 , an α -dicarbonyl (such as glyoxal (*GLY*), methylglyoxal

(*MGLY*) or biacetyl (*BACL*)), a monounsaturated dicarbonyl co-product (*AFG1*, *AFG2*, the photoreactive products) and a di-unsaturated dicarbonyl product (*AFG3*, the non-photoreactive products) (Calvert et al., 2002).

Formed from the phenolic products, the SAPRC-11 mechanism includes species of cresols (*CRES*), phenol (*PHEN*), xylenols and alkyl phenols (*XYNL*), and catechols (*CATL*). Due to their different SOA and ozone formation potentials (Carter et al, 2012), these phenolic species are represented separately. Relatively high yields of catechol (*CATL*) have been observed in the reactions of OH radicals with phenolic compounds. Furthermore, their subsequent reactions are believed to be important for SOA and ozone formation (Carter et al, 2012).

2.3 Simulation setups

In order to investigate the global chemical effects of the most commonly emitted aromatics in the troposphere, two simulations were performed, one with the ozone related aromatic chemistry updates from SAPRC-11 (the SAPRC case), and the other with simplified aromatic chemistry as in the standard setup (the Base case). Both simulations (Base and SAPRC) at 2.5° long. × 2° lat. are conducted from July 2004 to December 2005, allowing for a 6-month spin-up for our focused analysis over the year of 2005 for comparison to the available observations (Sect. 3). Initial conditions of chemicals are regridded from a simulation at 5° long. × 4° lat. started from 2004 with another spin-up run from January to June 2004. For comparison with aromatics observations over the US in 2010–2011 (Sect. 3), we extend the simulations from July 2009 to December 2011 with July–December 2009 as the spin-up period.

3. Aromatics and ozone observations

We use a set of measurements from surface and aircraft campaigns to evaluate the model simulated aromatics and ozone.

3.1 Aromatic aircraft observations

For aromatics, we use airborne observations from CALNEX (California; May/June 2010) aircraft study. A proton transfer reaction quadrupole mass spectrometer (PTR-MS) was used to measure mixing ratios of aromatics (and an array of other primary and secondary pollutants) during CALNEX. Measurements are gathered mostly on a one-second time scale (approximately 100 m spatial resolution), which permits sampling of the source regions and tracking subsequent transport and transformation throughout California and surrounding regions. Further details of the CALNEX campaign, including the flight track, timeframe, location and instrument, are shown in Hu et al. (2015) and <https://www.esrl.noaa.gov/csd/projects/calnex>. For comparison to the model results, we averaged the high temporal-spatial resolution observations to the model resolution.

We also employ vertical profiles obtained in 2005 from the CARIBIC (Civil Aircraft for Regular Investigation of the atmosphere Based on an Instrument Container) project, which conducts

atmospheric measurements onboard a commercial aircraft (Lufthansa A340-600) (Brenninkmeijer et al., 2007; Baker et al., 2010). CARIBIC flights fly away from Frankfurt, Germany on the way to North America, South America, India and East Asia. Measurements are available in the upper troposphere (50% on average) and lower stratosphere (50%) (UTLS) at altitudes between 10–12 km. To evaluate our results, measurements are averaged to the model output resolution. Vertically, results from GEOS-Chem model simulations at the 250 hPa level are used to compare with observations between 200–300 hPa. Then the annual means of observations and model data sampled along the flight tracks are used in the comparison.

3.2 Aromatics surface measurements

To evaluate the ground-level mixing ratios of benzene, toluene, and xylenes as well as their seasonal cycles, surface observations of aromatics are collected from two networks (EMEP, data available at <http://www.nilu.no/projects/ccc/emepdata.html>, and the European Environmental Agency (EEA), data available at <http://www.eea.europa.eu/data-and-maps/data/airbase-the-european-air-quality-database-8>, both for the year 2005) over Europe and the KCMP tall tower dataset (data available at <https://atmoschem.umn.edu/data>, for the year 2011) over the US.

EMEP, which aims to investigate the long-range transport of air pollution and the flux through geographic boundaries (Torseth et al., 2012), locates measurement sites in locations where there are minimal local impacts, thus consequently the observations could represent the feature of large regions. EMEP has a daily resolution with a total of 14 stations located in Europe for benzene, 12 stations for toluene, and 8 stations for xylenes (Table 1). Here we use the monthly values calculated from the database to evaluate monthly model results. Note that measurement speciation of xylenes (o-xylene, m-xylene and p-xylene) in EMEP network does not exactly correspond with the model speciation of xylenes (m-xylene, p-xylene, o-xylene and ethylbenzene) (Hu et al., 2015). The speciation assumption probably can partly account for the xylene model-measurement discrepancy seen in Sect. 4.

EEA provides observations from a large number of sites over urban, suburban and background regions (EEA, 2014). However, here we use only rural background sites to do model comparison, as in Cabrera-Perez et al. (2016), because the model horizontal scale cannot simulate direct traffic or industrial influence. This leads to 22 stations available for benzene and 6 stations for toluene. Further details of the sites and location information of EEA (and EMEP) used here are described in Cabrera-Perez et al., 2016. For comparison, annual means for individual sites have been used.

The KCMP tall tower measurements (at 44.69°N, 93.07°W, Minnesota, US) have been widely used for studies of surface fluxes of tropospheric trace species and land-atmosphere interactions (Kim et al., 2013; Hu et al., 2015; Chen et al., 2018). A suite of NMVOCs including aromatics were observed at the KCMP tower during 2009–2012 with a high-sensitivity PTR-MS, sampling from a height of 185 m above ground level. We averaged the hourly observations of benzene, toluene and C₈ (xylenes + ethylbenzene; here consistent with the model speciation) aromatics to

monthly values and then used for our model evaluation. Monthly mean simulations at the 990 hPa level (~190 m) are used for comparison.

3.3 Ozone observations

Ozone observations are taken from the database of the World Data Centre for Greenhouse Gases (WDCGG, data available at <http://ds.data.jma.go.jp/gmd/wdogg/cgi-bin/wdogg/catalogue.cgi>), and the Chemical Coordination Centre of EMEP (EMEP CCC). These networks contain hourly ozone measurements over a total of 194 background sites in remote environments. We use monthly averaged observations of surface ozone in 2005 to examine the simulated surface ozone from the GEOS-Chem model. Simulated ozone from the lowest layer (centered at ~ 65 m) is sampled from the grid cells corresponding to the ground sites.

4. Evaluation of simulated aromatics and ozone

In this section, the SAPRC model simulation results of aromatics (benzene, toluene, xylenes and C₈ aromatics) and ozone from GEOS-Chem are evaluated with observations. Table 1 summarizes the statistical comparison between measured and simulated concentrations over the monitoring stations described in Sect. 3. For the statistical calculations, GEOS-Chem simulation results have been sampled along the geographical locations of the measurements. Table 1 includes the number of locations and time resolutions. The number of sites in EEA for xylenes is only 2, thus we do not include their comparison results in Table 1 due to the lack of representativeness.

4.1 Surface-level aromatics

For the aromatics near the surface mixing ratios over Europe, observed mean benzene (194.0 ppt for EEA and 166.4 ppt for EMEP) and toluene (240.3 ppt for EEA and 133.1 ppt for EMEP) mixing ratios are higher than observed mean xylene concentrations (42.3 ppt for EMEP). In general, the model underestimates EEA and EMEP observations of benzene (by 34% on average) and toluene (by 20% on average). For benzene, the model results systematically underestimate the annual means (36%) compared to the EMEP database, consistent with the model underestimate of the EEA dataset (32%). The model underestimate for toluene compared to the EMEP dataset (15%) is smaller than that relative to the EEA measurements (25%). The simulation overestimates the xylene measurements in EMEP by a factor of 1.9, in part because the model results include ethylbenzene but the observations do not (see Sect. 3.2). The fact that the anthropogenic RETRO emissions (for year 2000) do not correspond to the year of measurement (2005) may contribute to the above model-measurement discrepancies. Anthropogenic aromatics emissions are reported to have significant changes in emissions and their distributions over the decade by EDGARv4.3.2 (Crippa et al., 2018; [http://eccad.aeris-data.fr/#DatasetPlace:EDGARv4.3.2\\$DOI](http://eccad.aeris-data.fr/#DatasetPlace:EDGARv4.3.2$DOI)). It shows that the total aromatics emission from anthropogenic source are enhanced by 5% (2005) and 14% (2011) compared to the year 2000.

The model bias would be partly benefit from this emission increase with enhanced modeled mixing ratios of benzene and toluene.

The modeled spatial variability of aromatics (with standard deviations of 32.1–66.8 ppt) is 18–73% lower than that of the EMEP and EEA observations (41.9–118.4 ppt), probably due to the coarse model resolution. The spatial variability in benzene (46–73% lower) is the most strongly underestimated among the three aromatic species. Unlike benzene, simulated concentrations of toluene show a larger standard deviation (66.8 ppt) than the EEA measurements (59.4 ppt), indicating larger simulated spatial variability. Simulation results are thus poorly spatially correlated with observations ($R = 0.41\text{--}0.49$). However, the temporal variability of aromatics is well captured by GEOS-Chem with the correlations above 0.7 for most stations.

Figure 2 shows a comparison of model results with observations at six stations for benzene, toluene, and xylenes, respectively, following Cabrera-Perez et al. (2016). The sites are chosen as the first six stations with largest amount of data. Model results reproduce the annual cycle at the majority of sites. Aromatics are better simulated in summer than in winter. This feature has been previously found for the climate-chemistry model EMAC for aromatics (Cabrera-Perez et al., 2016) and simpler NMVOCs (Pozzer et al., 2007). In addition, the measurements show larger standard deviations than the GEOS-Chem simulations, with the ratios between the observed and the simulated standard deviations being 2–11.

Over the US, annual mean observed concentrations at the KCMP tall tower are 91.5 ppt for benzene, 56.7 ppt for toluene, and 90.3 ppt for C₈ aromatics (Table 1). The model biases for benzene (8.4 ppt; 9.2%) and C₈ aromatics (–1.4 ppt; –1.6%) are much lower than that for toluene (64.5 ppt; 114%). Figure 3 further shows the observed and simulated monthly averaged concentrations of benzene, toluene and C₈ aromatics. The SAPRC simulation reproduces their seasonal cycles, with higher concentrations in winter and lower mixing ratios in summer, consistent with Hu et al. (2015). The model-observation correlations are 0.89, 0.78 and 0.65 for monthly benzene, toluene, and C₈ aromatics, respectively. The large overestimation of modeled toluene is mainly due to simulated high mixing ratios during the cold season (Fig. 3, October to March).

4.2 Tropospheric aromatics

Table 1 shows that in the UTLS, both CARIBIC observed (16 ppt) and GEOS-Chem modeled (12.3 ppt) benzene mixing ratios are higher than toluene concentrations (3.6 ppt for CARIBIC and 1.5 ppt for GEOS-Chem). For benzene, the model underestimates appear to be smaller in the free troposphere (with an underestimate by 23%) than at the surface (36% for EMEP and 32% for EEA). In contrast to benzene, annual mean concentrations of toluene are underestimated by 58% in the UTLS. The geographical variability of benzene is larger than that for toluene (with standard deviation of 4.2 versus 0.7 ppt in model and 15.8 versus 7.5 ppt in observation), probably because of the shorter lifetime of benzene (between several hours and several days;

<http://www.nzdl.org/gsdmod?a=p&p=home&l=en&w=utf-8>) in combination with the lower concentrations in the UTLS for toluene. The model results show smaller spatial variability than the observations. This underestimation for spatial variability in the free troposphere (over 70%) is higher than that at the surface (not shown).

The black lines in Fig. 4 show the tropospheric aromatics profiles during the CALNEX campaign. The measured values peak at an altitude of 0.6–0.8 km, with concentrations decreasing at higher altitudes. Although the concentrations in the lower troposphere for benzene (40–100 ppt below 2 km) are lower than mixing ratios for toluene (70–160 ppt below 2 km) and C₈ aromatics (50–120 ppt below 2 km), the benzene mixing ratios (> 30 ppt) in the free troposphere are much higher than those of toluene and C₈ aromatics (< 10 ppt). The different profile shapes in the lower troposphere for benzene, toluene and C₈ aromatics are mainly due to their different emissions and lifetime. The SAPRC simulation (red lines in Fig. 4) captures the general vertical variations of CALNEX benzene and toluene, with statistically significant model-observation correlations of 0.74 and 0.65 for benzene and toluene, respectively. The model generally overestimates the measured C₈ aromatics below 0.5 km, albeit with an underestimate above 0.5 km, with lower model-observation correlation of 0.37. This overestimation below 0.5 km is also seen for benzene and toluene. The modeled overly rapid aromatics drop-off with altitude probably implies the modelled aromatics lifetime is short.

4.3 Surface ozone

Table 1 shows an average ozone mixing ratio of 34.1 ppb in 2005 over the regional background WDCGG sites. The annual mean ozone mixing ratios are lower over Europe (from the EMEP dataset), about 30.6 ppb. The SAPRC simulation tends to underestimate the mixing ratios over the sites of Europe and background regions with biases of –2.9 ppb and –5.5 ppb, respectively. Figure 5 shows the spatial distribution of the annual mean model biases with respect to the measurements. Unlike the modeled surface aromatics, the simulated ozone spatial variability can be either slightly lower or higher than the observed variability, depending on the compared database: the standard deviation is 12.8 ppb (simulated) versus 14.2 ppb (observed) for WDCGG sites, 13.2 versus 10.3 ppb for EMEP sites. The temporal variability (temporal correlations of 0.68–0.72) is better captured by the model than the spatial variability (spatial correlations of 0.52–0.54).

5. Global effects of aromatic chemistry

This section compares the Base and SAPRC simulations to assess to which extent the updated mechanism for aromatics affect the global simulation of ozone, HO_x and individual nitrogen species. Our focus here is on the large-scale impacts.

5.1 NO_y Species

Figure 6 and Table 2 show the changes from Base to SAPRC in annual average surface NO mixing ratios. A decrease in NO is apparent over NO_x source regions, e.g., by approximately 0.15 ppb (~20%) over much of the US, Europe and China (Fig. 6). In contrast, surface NO increases at locations downwind from NO_x source regions (up to ~0.1 ppb or 20%), including the oceanic area off the eastern US coast, the marine area adjacent to Japan, and the Mediterranean area. The change is negligible (by ~0.2%) for the annual global mean surface NO (Table 2). Seasonally, the decrease in spring, summer and fall is compensated partly by the increase in winter (Table 2). This winter increase versus decline in other seasons is probably attributed to the weakened photochemical reactions involving NO_x in winter.

The zonal average results in Fig. 7 show a clear decline in NO in the planetary boundary layer, in contrast to significant increases in the free troposphere, from Base to SAPRC. The free tropospheric NO increases are about the same from 30°S-90°N with an annual average enhancement up to 5% (Fig. 7), and are particularly large in winter (up to 10%, not shown). For the whole troposphere, the average NO increases by 0.6% from Base to SAPRC (Table 2).

Figure 6 shows that simulated surface NO₂ mixing ratios in the SAPRC scenario are enhanced over most locations across the globe, in comparison with the Base simulation. Over the source regions, the changes are mixed, with increases in some highly NO_x polluted regions (by up to 10%) and decreases in other polluted regions. On a global mean basis, NO₂ is increased (by 2.1% in the free troposphere and 1.0% at the surface, Table 2), due mainly to the recycling of NO_x from PAN associated with the aromatics, and the reactions of oxidation products from aromatics with NO or NO₃ (primarily) to form NO₂ and HO₂. Combining the changes in NO and NO₂ means that the total NO_x mixing ratios decrease in source regions but increase in the remote free troposphere (Fig. 8 and 9).

The NO₃ mixing ratios decrease at the global scale (−4.1% on average in the troposphere, Fig. 7 and Table 2) in the SAPRC simulation, except for an enhancement in surface NO₃ over the northern polar regions and most polluted areas like the eastern US, Europe and eastern China (Fig. 6). The NO₃ global decreases are mainly due to the consumption of NO₃ by reaction with the aromatic oxidation products. However, the NO₃ regional increases are probably caused by the enhanced regional atmospheric oxidation capacity.

Table 2 shows that nitric acid (HNO₃) increases in the SAPRC simulation, both near the surface (by approximately 1.1%) and in the troposphere (by 0.3%). The enhancement in HNO₃ appears uniformly over most continental regions in the northern hemisphere (not shown), due to the promotion of direct formation of HNO₃ from aromatics in the SAPRC simulation.

5.2 OH and HO₂

Compared to the Base simulation, OH increases slightly by 1.1% at the surface in the SAPRC simulation, with that declines over the tropics (30°S–30°N) are compensated by enhancements over other regions (Fig. 10 and Table 2). The largest increases in OH concentrations are found

over source regions dominated by anthropogenic emissions (i.e., the US, Europe, and Asia) and in subtropical continental regions with large biogenic aromatic emissions. In these locations, the peroxy radicals formed by aromatic oxidation react with NO and HO₂, which can have a significant effect on the ambient ozone and NO_x mixing ratios. This in turn influences OH, as the largest photochemical sources of OH in the model are the photolysis of O₃ as well as the reaction of NO with HO₂. Seasonally, a few surface locations see OH concentration increases of more than 10% during April–August (not shown), including parts of the eastern US, central Europe, eastern Asia and Japan.

The OH enhancement (0.2%) is also seen in the free troposphere in the SAPRC simulation (Fig. 11 and Table 2). OH is increased in the troposphere of the northern hemisphere, in contrast to the decline in the troposphere of tropics and southern hemisphere (Fig. 11). These OH changes correspond to the hemispherically distinct changes in aromatics (benzene, toluene, and xylenes), which show a decrease in the northern hemisphere, an increase in the southern hemisphere (Fig. 12 and 13), and an increase in global mean (by 1%) (Table 2). Despite the overall increase in tropospheric OH, CO is increased by ~1% (Table 2) due to additional formation from aromatics oxidation.

Table 2 shows that from Base to SAPRC, HO₂ shows a significant increase at the global scale: 3.0% at the surface and 1.3% in the troposphere, due to regeneration of HO_x from aromatics oxidation products. Correspondingly, the OH/HO₂ ratio decreases slightly. These changes mean that, compared to the simplified aromatic chemistry in the standard model setup, the SAPRC mechanism are associated with higher OH (i.e., more chemically reactive troposphere) and even higher HO₂.

5.3 Ozone

From Base to SAPRC, the global average surface ozone mixing ratio increases by less than 1% (Table 2). This small difference is comparable to the result calculated by Cabrera-Perez et al. (2017) with the EMAC model, which is based on a reduced version of the aromatic chemistry from the Master Chemical Mechanism (MCMv3.2). Figure 10 shows that the 1% increase in surface ozone occurs generally over the northern hemisphere. Similar to the changes in OH, the most notable ozone increase occurs in industrially-polluted regions. These regions show significant local ozone photochemical formation in both the Base case and the SAPRC simulation. The updated aromatic chemistry increases ozone by up to 5 ppb in these regions. Increases of ozone are much smaller (less than 0.2 ppb) over the tropical oceans than in the continental areas. In contrast, ozone declines in regions of South America, Central Africa, Australia and Indonesia over the tropics (30°S–30°N). Changes elsewhere in the troposphere are similar in magnitude, as shown in Figure 11.

Two general factors likely contribute to the ozone change from Base to SAPRC. In the SAPRC simulation, the addition of aromatic oxidation products (i.e., peroxy radicals) can contribute

1 directly to ozone formation in NO_x-rich source regions and also in the NO_x-sensitive remote
2 troposphere (i.e., from PAN to NO_x and to ozone). The second factor is a change in the NO_x
3 spatial distribution, with an overall enhancement in average NO₂ concentrations. The
4 redistribution is mainly caused by enhanced transport of NO_x to the remote troposphere (see Sect.
5 5.1). The enhanced NO_x in the remote troposphere enhances the overall ozone formation because
6 this process is more efficient in the remote regions (e.g., Liu et al., 1987). The increased ozone,
7 NO₂ and NO_x transport all lead to the aforementioned changes. This is described in detail in
8 section 5.4.

9 There are notable decreases (more than 5%, Fig. 11) in simulated ozone and OH in the free
10 troposphere (above 4 km) over the tropics (30°S–30°N). A similar decrease is found in modeled
11 NO_x (above 6 km, Fig. 9). These decreases are probably related to the upward transport of
12 aromatics by tropical convection processes. The aromatics transported to the upper troposphere
13 may cause net consumption of tropospheric OH and NO_x, which can further reduce ozone
14 production.

15 From Base to SAPRC, the modeled ozone concentrations are close to the WDCGG and EMEP
16 network measurements (Table 3). For the WDCGG background sites, the annual and seasonal
17 model biases are ~10% smaller in the SAPRC simulation compared to the Base case. For the
18 EMEP stations, although the model results are not improved in summer and fall, the annual
19 model bias is 25% smaller (–2.8 ppb versus –3.5 ppb) in the SAPRC simulation.

20 **5.4 Discussion of SAPRC aromatic-ozone chemistry**

21 As discussed in Sect. 5.3, the increased O₃ mixing ratios from Base to SAPRC are due to the
22 direct impact of aromatic oxidation products (i.e., peroxy radicals) and to the effect of increased
23 NO₂ concentrations. The simulated odd oxygen family ($O_x = O_3 + O(^1D) + O(^3P) + NO_2 +$
24 $2 \times NO_3 + 3 \times N_2O_5 + HNO_3 + HNO_4 + PAN$, Wu et al., 2007; Yan et al., 2016) formation
25 increases by 1–10%, both over the source regions and in the remote troposphere (Fig. 10 and 11).
26 Although the percentage changes are similar, the driving factors over the source regions are
27 different from the drivers in the remote troposphere.

28 Regions with large aromatics emissions show a significant increase of oxidation products from
29 Base to SAPRC. The modeled ozone in these regions increases with increasing NO₂ and its
30 oxidation products. NO and NO₃ are often lower in these regions in the SAPRC scenario because
31 of their reactions with the aromatic-OH oxidation products to form NO₂ and HO₂. In remote
32 regions and in the free troposphere, ozone production is also enhanced by both NO₂ and HO₂
33 increases in the SAPRC simulation, but the increase in ozone formation is mainly attributed to
34 the increase in NO_x mixing ratios.

35 NO_x concentrations decrease in source regions and increase in the remote regions because of
36 more efficient transport of PAN and its analogues (represented by *PBZN* here in SAPRC-11).
37 From Base to SAPRC, modeled PAN has been enhanced in a global scale (Fig. 8 and 9) via

1 reactions of aromatic-OH oxidation products with NO₂ (equation of BR13 in Table S2). In the
2 SAPRC-11 aromatics chemical scheme the immediate precursor of PAN (peroxyacetyl radical)
3 has five dominant photochemical precursors. They are acetone (CH₃COCH₃, model species:
4 *ACET*), methacrolein (*MACR*), biacetyl (*BACL*), methyl glyoxal (*MGLY*) and other ketones (e.g.,
5 *PROD2*, *AFG1*). These compounds explain the increased rate of PAN formation. For example,
6 the SAPRC simulation has increased the concentration of *MGLY* by a factor of 2. In addition,
7 production of organic nitrates (*PBZN* (reactions of BR30 and BR31 in Table S2) and *RNO3*
8 (*PO36*)) in the model with SAPRC aromatics chemistry may also explain the increase in ambient
9 NO_x in the remote regions, due to the re-release of NO_x from organic nitrates (as opposed to
10 removal by deposition). Due to such re-release of NO_x from PAN-like compounds and also
11 transport of NO_x, NO_x increases by up to 5% at the surface in most remote regions and by ~1% in
12 the troposphere as a whole. This then leads to increased ozone due to the effectiveness of ozone
13 formation in the free troposphere.

14 SAPRC is a highly efficient and compact chemical mechanism with the use of maximum ozone
15 formation as a primary metric in the chamber experiment benchmark. The mechanism has been
16 primarily used and evaluated in regional CTMs such as CMAQ and CAMx, at much finer
17 resolution (i.e., a few kilometers). Our study has significant application to use it in a global
18 model. Implementing SAPRC-11 aromatic chemistry would add ~3% more computational effort
19 in terms of model simulation times.

20 SAPRC is based on lumped chemistry, which is partly optimized on empirical fitting to smog
21 chamber experiments that are representative to one-day photochemical smog episodes typical of,
22 for example, Los Angeles and other US urban centers. However, SAPRC-11 gives better
23 simulations of ozone formation in almost all conditions, except for higher (>100 ppb) NO_x
24 experiments where O₃ formation rates are consistently over predicted (Carter and Heo, 2013).
25 This over prediction can be corrected if the aromatics mechanism is parameterized to include a
26 new NO_x dependence on photoreactive product yields, but that parameterization is not
27 incorporated in SAPRC-11 because it is inconsistent with available laboratory data.

28 Other option, such as the condensed MCM mechanism, which are based upon more fundamental
29 laboratory and theoretical data and used for policy and scientific modelling multi-day
30 photochemical ozone formation, is experienced over Europe by Cabrera-Perez. (2016). Our
31 results are consistent with the simulation of EMAC model implemented with a reduced version of
32 the MCM aromatic chemistry. Moreover, aromatic chemistry is still far from being completely
33 understood. For example, Bloss et al., (2005) show that for alkyl substituted mono-aromatics,
34 when comparisons to chamber experiment over a range of VOC/NO_x conditions, the chemistry
35 under predicts the reactivity of the system but over predicts the amount of O₃ formation (model
36 shows more NO to NO₂ conversion than on the experiments).

37 6. Conclusions

1 A representation of tropospheric reactions for aromatic hydrocarbons in the SAPRC-11
2 mechanism has been added to GEOS-Chem, to provide a more realistic representation of their
3 atmospheric chemistry. The GEOS-Chem simulation with the SAPRC-11 aromatics mechanism
4 has been evaluated against measurements from aircraft and surface campaigns. The comparison
5 with observations shows reasonably good agreement for aromatics (benzene, toluene, and
6 xylenes) and ozone. Model results for aromatics can reproduce the seasonal cycle, with a general
7 underestimate over Europe for benzene and toluene, and an overestimate of xylenes; while over
8 the US a positive model bias for benzene and toluene and a negative bias for C₈ aromatics are
9 found. From the Base to the SAPRC simulation, the model ozone bias is reduced by 10% relative
10 to WDCGG observations and by 25% relative to EMEP observations.

11 The simplified aromatics chemistry in the Base simulation under-predicts NO and NO₃ oxidation,
12 and it does not represent ozone formed from aromatic-OH-NO_x oxidation. Although the global
13 average changes in simulated chemical species are relatively small (1%–4% from Base to
14 SAPRC), on a regional scale the differences can be much larger, especially over aromatics and
15 NO_x source regions. From Base to SAPRC, NO₂ is enhanced by up to 10% over some highly
16 polluted areas, while reductions are notable in other polluted areas. Although the simulated
17 surface NO decreases by approximately 0.15 ppb (~20%) or more in the northern hemispheric
18 source regions, including most of the US, Europe and China, increases are found (~0.1 ppb, up to
19 20%) at locations downwind from these source regions. The total NO_x mixing ratios decrease in
20 source regions but increase in the remote free troposphere. This is mainly due to the addition of
21 aromatics oxidation products in the model that lead to PAN, which facilitates the transport of
22 nitrogen oxides to downwind locations remote from the sources. Finally, the updated aromatic
23 chemistry in GEOS-Chem increases ozone concentrations, especially over industrialized regions
24 (up to 5 ppb, or more than 10%). Ozone changes in the model are partly explained by the direct
25 impact of increased aromatic oxidation products (i.e., peroxy radical), and partly by the effect of
26 the altered spatial distribution of NO_x. Overall, our results suggest that a better representation of
27 aromatics chemistry is important to model the tropospheric oxidation capacity.

28 **Data Availability**

29 The aircraft and surface data used in this paper is already publically available. Airborne
30 observations of aromatics from CALNEX (<https://www.esrl.noaa.gov/csd/projects/calnex>) and
31 CARIBIC project. Surface observations of aromatics are collected from EMEP
32 (<http://www.nilu.no/projects/ccc/emepdata.html>) and EEA ([http://www.eea.europa.eu/data-and-](http://www.eea.europa.eu/data-and-maps/data/airbase-the-european-air-quality-database-8)
33 [maps/data/airbase-the-european-air-quality-database-8](http://www.eea.europa.eu/data-and-maps/data/airbase-the-european-air-quality-database-8)) over Europe and the KCMP tall tower
34 dataset (<https://atmoschem.umn.edu/data>) over the US. Ozone observations are taken from
35 WDCGG (<http://ds.data.jma.go.jp/gmd/wdcgg/cgi-bin/wdcgg/catalogue.cgi>).

36 **Code Availability**

The GEOS-Chem code of version 9-02 used to generate this paper and the model results are available upon request. We are submitting the code for inclusion into the standard model. The revised aromatics chemistry will be incorporated in the current version 12.0.0 and the later versions.

Acknowledgements

This research is supported by the National Natural Science Foundation of China (41775115), the 973 program (2014CB441303) and the Key Program of Ministry of Science and Technology of the People's Republic of China (2016YFA0602002; 2017YFC0212602). The research was also funded by the Start-up Foundation for Advanced Talents (162301182756). We acknowledge the free use of ozone data from networks of WDCGG (<http://ds.data.jma.go.jp/gmd/wdcgg/cgi-bin/wdcgg/catalogue.cgi>), EMEP (<http://www.nilu.no/projects/ccc/emepdata.html>) and aromatic compounds observations from EEA (<http://www.eea.europa.eu/data-and-maps/data/airbase-the-european-air-quality-database-8>) and EMEP. We also want to thank Angela Baker for providing the CARIBIC data. DBM acknowledges support from NASA (Grant #NNX14AP89G).

Reference

- Atkinson, R.: Atmospheric chemistry of VOCs and NO_x, *Atmospheric Environment*, 34, 2063-2101, 10.1016/s1352-2310(99)00460-4, 2000.
- Atkinson, R. and Arey, J.: Atmospheric degradation of volatile organic compounds, *Chem. Rev.*, 103, 4605–4638, <https://doi.org/10.1021/cr0206420>, 2003.
- Bloss, C., Wagner, V., Bonzanini, A., Jenkin, M. E., Wirtz, K., Martin-Reviejo, M., and Pilling, M. J. Evaluation of detailed aromatic mechanisms (MCMv3 and MCMv3.1) against environmental chamber data, *Atmos. Chem. Phys.*, 5, 623-639, 2005.
- Baker, A. K., Slemr, F., and Brenninkmeijer, C. A. M.: Analysis of non-methane hydrocarbons in air samples collected aboard the CARIBIC passenger aircraft, *Atmospheric Measurement Techniques*, 3, 311-321, 2010.
- Bejan, I., Abd El Aal, Y., Barnes, I., Benter, T., Bohn, B., Wiesen, P., and Kleffmann, J.: The photolysis of ortho-nitrophenols: a new gas phase source of HONO, *Physical Chemistry Chemical Physics*, 8, 2028-2035, 10.1039/b516590c, 2006.
- Brenninkmeijer, C. A. M., Crutzen, P., Boumard, F., Dauer, T., Dix, B., Ebinghaus, R., Filippi, D., Fischer, H., Franke, H., Friess, U., Heintzenberg, J., Helleis, F., Hermann, M., Kock, H. H., Koepfel, C., Lelieveld, J., Leuenberger, M., Martinsson, B. G., Miemczyk, S., Moret, H. P., Nguyen, H. N., Nyfeler, P., Oram, D., O'Sullivan, D., Penkett, S., Platt, U., Pupek, M., Ramonet, M., Randa, B., Reichelt, M., Rhee, T. S., Rohwer, J., Rosenfeld, K., Scharffe, D., Schlager, H., Schumann, U., Slemr, F., Sprung, D., Stock, P., Thaler, R., Valentino, F., van Velthoven, P., Waibel, A., Wandel, A., Waschitschek, K., Wiedensohler, A., Xueref-Remy, I., Zahn, A., Zech,

1 U., and Ziereis, H.: Civil Aircraft for the regular investigation of the atmosphere based on an
 2 instrumented container: The new CARIBIC system, *Atmospheric Chemistry and Physics*, 7,
 3 4953-4976, 2007.

4 Crippa, M., Guizzardi, D., Muntean, M., Schaaf, E., Dentener, F., van Aardenne, J. A., Monni, S.,
 5 Doering, U., Olivier, J. G. J., Pagliari, V., and Janssens-Maenhout, G.: Gridded Emissions of Air
 6 Pollutants for the period 1970–2012 within EDGAR v4.3.2, *Earth Syst. Sci. Data Discuss.*, null,
 7 in review, 2018

8 Cabrera-Perez, D., Taraborrelli, D., Sander, R., and Pozzer, A.: Global atmospheric budget of
 9 simple monocyclic aromatic compounds, *Atmospheric Chemistry and Physics*, 16, 6931-6947,
 10 10.5194/acp-16-6931-2016, 2016.

11 Chandra, A., Pradhan, P., Tewari, R., Sahu, S., and Shenoy, P.: An observation-based approach
 12 towards self-managing web servers, *Computer Communications*, 29, 1174-1188,
 13 10.1016/j.comcom.2005.07.003, 2006.

14 Calvert, J. G., Atkinson, R., Becker, K. H., Kamens, R. M., Seinfeld, J. H., Wallington, T. J., and
 15 Yarwood, G.: *The mechanisms of atmospheric oxidation of aromatic hydrocarbons*, Oxford
 16 University Press, New York, 2002.

17 Chen, D., Wang, Y., McElroy, M. B., He, K., Yantosca, R. M., and Le Sager, P.: Regional CO
 18 pollution and export in China simulated by the high-resolution nested-grid GEOS-Chem model,
 19 *Atmos. Chem. Phys.*, 9, 3825–3839, doi:10.5194/acp-9-3825-2009, 2009.

20 Chen, J., Wenger, J. C., and Venables, D. S.: Near-Ultraviolet Absorption Cross Sections of
 21 Nitrophenols and Their Potential Influence on Tropospheric Oxidation Capacity, *Journal of*
 22 *Physical Chemistry A*, 115, 12235-12242, 10.1021/jp206929r, 2011.

23 Chen, Z., T.J. Griffis, J.M. Baker, D.B. Millet, J.D. Wood. E.J. Dlugokencky, A.E. Andrews, C.
 24 Sweeney, C. Hu, and R.K. Kolka (2018), Source partitioning of methane emissions and its
 25 seasonality in the U.S. Midwest, *J. Geophys. Res.*, 123, doi:10.1002/2017JG004356.

26 Carter, W. P. L., Goo young Heo, David R. Cocker III, and Shunsuke Nakao (2012): “SOA
 27 Formation: Chamber Study and Model Development,” Final report to CARB contract 08-326,
 28 May 21. Available at <http://www.cert.ucr.edu/~carter/absts.htm>

29 Calvert, J. G., R. Atkinson, K. H. Becker, R. M. Kamens, J. H. Seinfeld, T. J. Wallington and G.
 30 Yarwood (2002): “The Mechanisms of Atmospheric Oxidation of Aromatic Hydrocarbons,”
 31 Oxford University Press, New York, 566p.

32 Carter, W. P. L. (2010a): “Development of the SAPRC-07 Chemical Mechanism and Updated
 33 Ozone Reactivity Scales,” Final report to the California Air Resources Board Contract No. 03-
 34 318, 06-408, and 07-730. January 27. Available at www.cert.ucr.edu/~carter/SAPRC.

1 Carter, W. P. L. (2010b): "Development of the SAPRC-07 chemical mechanism," *Atmospheric*
2 *Environment* 44, 5324-5335.

3 Carter, W. P. L., and Heo, G.: Development of revised SAPRC aromatics mechanisms,
4 *Atmospheric Environment*, 77, 404-414, 10.1016/j.atmosenv.2013.05.021, 2013.

5 EEA: Air quality in Europe - 2014 report, Report No. 5/2014, European Environment Agency,
6 Copenhagen, DK, available at: [http://www.eea.europa.eu/publications/air-quality-in-europe-](http://www.eea.europa.eu/publications/air-quality-in-europe-2014)
7 2014, 2014.

8 Fiore, A. M., Dentener, F. J., Wild, O., Cuvelier, C., Schultz, M. G., Hess, P., Textor, C., Schulz,
9 M., Doherty, R. M., Horowitz, L. W., MacKenzie, I. A., Sanderson, M. G., Shindell, D. T.,
10 Stevenson, D. S., Szopa, S., Van Dingenen, R., Zeng, G., Atherton, C., Bergmann, D., Bey, I.,
11 Carmichael, G., Collins, W. J., Duncan, B. N., Faluvegi, G., Folberth, G., Gauss, M., Gong, S.,
12 Hauglustaine, D., Holloway, T., Isaksen, I. S. A., Jacob, D. J., Jonson, J. E., Kaminski, J. W.,
13 Keating, T. J., Lupu, A., Marmer, E., Montanaro, V., Park, R. J., Pitari, G., Pringle, K. J., Pyle, J.
14 A., Schroeder, S., Vivanco, M. G., Wind, P., Wojcik, G., Wu, S., and Zuber, A.: Multimodel
15 estimates of intercontinental source-receptor relationships for ozone pollution, *Journal of*
16 *Geophysical Research*, 114, 10.1029/2008jd010816, 2009.

17 Fischer, E. V., Jacob, D. J., Yantosca, R. M., Sulprizio, M. P., Millet, D. B., Mao, J., Paulot, F.,
18 Singh, H. B., Roiger, A., Ries, L., Talbot, R. W., Dzepina, K., and Pandey Deolal, S.:
19 Atmospheric peroxyacetyl nitrate (PAN): a global budget and source attribution, *Atmos. Chem.*
20 *Phys.*, 14, 2679-2698, <https://doi.org/10.5194/acp-14-2679-2014>, 2014.

21 Guenther, A.: Estimates of global terrestrial isoprene emissions using MEGAN (Model of
22 Emissions of Gases and Aerosols from Nature) (vol 6, pg 3181, 2006), *Atmospheric Chemistry*
23 *and Physics*, 7, 4327-4327, 2007.

24 Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and
25 Wang, X.: The Model of Emissions of Gases and Aerosols from Nature version 2.1
26 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions,
27 *Geoscientific Model Development*, 5, 1471-1492, 10.5194/gmd-5-1471-2012, 2012.

28 Guo, H., Wang, T., Blake, D. R., Simpson, I. J., Kwok, Y. H., and Li, Y. S.: Regional and local
29 contributions to ambient non-methane volatile organic compounds at a polluted rural/coastal site
30 in Pearl River Delta, China, *Atmospheric Environment*, 40, 2345-2359,
31 10.1016/j.atmosenv.2005.12.011, 2006.

32 Hu, L., Millet, D. B., Baasandorj, M., Griffis, T. J., Travis, K. R., Tessum, C. W., Marshall, J. D.,
33 Reinhart, W. F., Mikoviny, T., Muller, M., Wisthaler, A., Graus, M., Warneke, C., and de Gouw,
34 J.: Emissions of C-6-C-8 aromatic compounds in the United States: Constraints from tall tower
35 and aircraft measurements, *Journal of Geophysical Research-Atmospheres*, 120, 826-842,
36 10.1002/2014jd022627, 2015.

1 Hudman, R. C., Moore, N. E., Mebust, A. K., Martin, R. V., Russell, A. R., Valin, L. C., and
2 Cohen, R. C.: Steps towards a mechanistic model of global soil nitric oxide emissions:
3 implementation and space based-constraints, *Atmospheric Chemistry and Physics*, 12, 7779-
4 7795, 10.5194/acp-12-7779-2012, 2012.

5 Kansal, A.: Sources and reactivity of NMHCs and VOCs in the atmosphere: A review, *Journal of*
6 *Hazardous Materials*, 166, 17-26, 10.1016/j.jhazmat.2008.11.048, 2009.

7 Kim, S.Y., D.B. Millet, L. Hu, M.J. Mohr, T.J. Griffis, D. Wen, J.C. Lin, S.M. Miller, and M.
8 Longo (2013), Constraints on carbon monoxide emissions based on tall tower measurements in
9 the US Upper Midwest, *Environ. Sci. Technol.*, 47, 8316-8324.

10 Krol, M., Houweling, S., Bregman, B., van den Broek, M., Segers, A., van Velthoven, P., Peters,
11 W., Dentener, F., and Bergamaschi, P.: The two-way nested global chemistry-transport zoom
12 model TM5: algorithm and applications, *Atmos. Chem. Phys.*, 5, 417–432, doi:10.5194/acp-5-
13 417-2005, 2005.

14 Lewis, A. C., Evans, M. J., Hopkins, J. R., Punjabi, S., Read, K. A., Purvis, R. M., Andrews, S.
15 J., Moller, S. J., Carpenter, L. J., Lee, J. D., Rickard, A. R., Palmer, P. I., and Parrington, M.: The
16 influence of biomass burning on the global distribution of selected non-methane organic
17 compounds, *Atmos. Chem. Phys.*, 13, 851–867, doi:10.5194/acp-13-851-2013, 2013.

18 Lin, J. T., and McElroy, M. B.: Impacts of boundary layer mixing on pollutant vertical profiles in
19 the lower troposphere: Implications to satellite remote sensing, *Atmospheric Environment*, 44,
20 1726-1739, 10.1016/j.atmosenv.2010.02.009, 2010.

21 Lin, J.-T., Tong, D., Davis, S., Ni, R.-J., Tan, X., Pan, D., Zhao, H., Lu, Z., Streets, D., Feng, T.,
22 Zhang, Q., Yan, Y.-Y., Hu, Y., Li, J., Liu, Z., Jiang, X., Geng, G., He, K., Huang, Y., and Guan,
23 D.: Global climate forcing of aerosols embodied in international trade, *Nature Geoscience*, 9,
24 790-794, doi:10.1038/NGEO2798, 2016.

25 Monks, P.S., A.T. Archibald, A. Colette, O. Cooper, M. Coyle, R. Derwent, D. Fowler, C.
26 Granier, K.S. Law, G.E. Mills, D.S. Stevenson, O. Tarasova, V. Thouret, E. von Schneidmesser,
27 R. Sommariva, O. Wild, and M.L. Williams: Tropospheric ozone and its precursors from the
28 urban to the global scale from air quality to short-lived climate forcer. *Atmos. Chem. Phys.*
29 15:8889–973. doi:10.5194/acp-15-8889-2015, 2015. Manuela, C., Gianfranco, T., Maria, F.,
30 Tiziana, C., Carlotta, C., Giorgia, A., Assunta, C., Pia, S. M., Jean-Claude, A., Francesco, T., and
31 Angela, S.: Assessment of occupational exposure to benzene, toluene and xylenes in urban and
32 rural female workers, *Chemosphere*, 87, 813-819, 10.1016/j.chemosphere.2012.01.008, 2012.

33 McLinden, C. A., Olsen, S. C., Hannegan, B., Wild, O., Prather, M. J., and Sundet, J.:
34 Stratospheric ozone in 3-D models: A simple chemistry and the cross-tropopause flux, *Journal of*
35 *Geophysical Research-Atmospheres*, 105, 14653-14665, 10.1029/2000jd900124, 2000.

1 Michalowicz, J., and Duda, W.: Phenols - Sources and toxicity, Polish Journal of Environmental
2 Studies, 16, 347-362, 2007.

3 Murray, L. T., Jacob, D. J., Logan, J. A., Hudman, R. C., and Koshak, W. J.: Optimized regional
4 and interannual variability of lightning in a global chemical transport model constrained by
5 LIS/OTD satellite data, Journal of Geophysical Research-Atmospheres, 117,
6 10.1029/2012jd017934, 2012.

7 Murray, L. T., Logan, J. A., and Jacob, D. J.: Interannual variability in tropical tropospheric
8 ozone and OH: The role of lightning, Journal of Geophysical Research-Atmospheres, 118,
9 11468-11480, 10.1002/jgrd.50857, 2013.

10 M.M. Rienecker, M.J. Suarez, R. Todling, J. Bacmeister, L. Takacs, H.-C. Liu, W. Gu, M.
11 Sienkiewicz, R.D. Koster, R. Gelaro, I. Stajner, E. Nielsen The GEOS-5 Data Assimilation
12 System - Documentation of Versions 5.0.1, 9.1.0, and 5.2.0, NASA, 2008

13 Na, K., Kim, Y. P., Moon, I., and Moon, K. C.: Chemical composition of major VOC emission
14 sources in the Seoul atmosphere, Chemosphere, 55, 585-594,
15 10.1016/j.chemosphere.2004.01.010, 2004.

16 Ni, R., Lin, J., Yan, Y., and Lin, W.: Foreign and domestic contributions to springtime ozone
17 over China, Atmos. Chem. Phys. Discuss., <https://doi.org/10.5194/acp-2017-1226>, in review,
18 2018.

19 Porter, W. C., Safieddine, S. A., and Heald, C. L.: Impact of aromatics and monoterpenes on
20 simulated tropospheric ozone and total OH reactivity, Atmospheric Environment, 169, 250-257,
21 10.1016/j.atmosenv.2017.08.048, 2017

22 Price, C., and Rind, D.: A simple lightning parameterization for calculating global lightning
23 distributions, J. Geophys. Res.-Atmos., 97, 9919-9933, <https://doi.org/10.1029/92JD00719>, 1992.

24 Ran, L., Zhao, C. S., Geng, F. H., Tie, X. X., Tang, X., Peng, L., Zhou, G. Q., Yu, Q., Xu, J. M.,
25 and Guenther, A.: Ozone photochemical production in urban Shanghai, China: Analysis based on
26 ground level observations, Journal of Geophysical Research-Atmospheres, 114,
27 10.1029/2008jd010752, 2009.

28 Reidmiller, D. R., Fiore, A. M., Jaffe, D. A., Bergmann, D., Cuvelier, C., Dentener, F. J.,
29 Duncan, B. N., Folberth, G., Gauss, M., Gong, S., Hess, P., Jonson, J. E., Keating, T., Lupu, A.,
30 Marmer, E., Park, R., Schultz, M. G., Shindell, D. T., Szopa, S., Vivanco, M. G., Wild, O., and
31 Zuber, A.: The influence of foreign vs. North American emissions on surface ozone in the US,
32 Atmospheric Chemistry and Physics, 9, 5027-5042, 2009.

33 Sarigiannis, D. A., and Gotti, A.: BIOLOGY-BASED DOSE-RESPONSE MODELS FOR
34 HEALTH RISK ASSESSMENT OF CHEMICAL MIXTURES, Fresenius Environmental
35 Bulletin, 17, 1439-1451, 2008.

1 Schmidt, J. A. ; Jacob, D. J. ; Horowitz, H. M. ; Hu, L. ; Sherwen, T. ; Evans, M. J. ; Liang, Q. ;
2 Suleiman, R. M. ; Oram, D. E. ; LeBreton, M. ; et al. Modeling the tropospheric BrO background:
3 importance of multiphase chemistry and implications for ozone, OH, and mercury. *J. Geophys.*
4 *Res.* Submitted.

5 Shen, L., Mickley, L. J., and Tai, A. P. K.: Influence of synoptic patterns on surface ozone
6 variability over the eastern United States from 1980 to 2012, *Atmospheric Chemistry and*
7 *Physics*, 15, 10925-10938, 10.5194/acp-15-10925-2015, 2015.

8 Stevenson, D. S., Dentener, F. J., Schultz, M. G., Ellingsen, K., van Noije, T. P. C., Wild, O.,
9 Zeng, G., Amann, M., Atherton, C. S., Bell, N., Bergmann, D. J., Bey, I., Butler, T., Cofala, J.,
10 Collins, W. J., Derwent, R. G., Doherty, R. M., Drevet, J., Eskes, H. J., Fiore, A. M., Gauss, M.,
11 Hauglustaine, D. A., Horowitz, L. W., Isaksen, I. S. A., Krol, M. C., Lamarque, J. F., Lawrence,
12 M. G., Montanaro, V., Müller, J. F., Pitari, G., Prather, M. J., Pyle, J. A., Rast, S., Rodriguez, J.
13 M., Sanderson, M. G., Savage, N. H., Shindell, D. T., Strahan, S. E., Sudo, K., and Szopa, S.:
14 Multimodel ensemble simulations of present-day and near-future tropospheric ozone, *Journal of*
15 *Geophysical Research*, 111, 10.1029/2005jd006338, 2006.

16 Stevenson, D. S., Young, P. J., Naik, V., Lamarque, J. F., Shindell, D. T., Voulgarakis, A., Skeie,
17 R. B., Dalsoren, S. B., Myhre, G., Berntsen, T. K., Folberth, G. A., Rumbold, S. T., Collins, W.
18 J., MacKenzie, I. A., Doherty, R. M., Zeng, G., van Noije, T. P. C., Strunk, A., Bergmann, D.,
19 Cameron-Smith, P., Plummer, D. A., Strode, S. A., Horowitz, L., Lee, Y. H., Szopa, S., Sudo, K.,
20 Nagashima, T., Josse, B., Cionni, I., Righi, M., Eyring, V., Conley, A., Bowman, K. W., Wild,
21 O., and Archibald, A.: Tropospheric ozone changes, radiative forcing and attribution to emissions
22 in the Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP),
23 *Atmospheric Chemistry and Physics*, 13, 3063-3085, 10.5194/acp-13-3063-2013, 2013.

24 Sander, R. (1999), Compilation of Henry's law constants for inorganic and organic species of
25 potential importance in environmental chemistry (Version 3), available at [http://www.henrys-](http://www.henrys-law.org)
26 [law.org](http://www.henrys-law.org).

27 Schultz, M., L. Backman, Y. Balkanski, S. Bjoerndalsaeter, R. Brand, J. Burrows, S. Dalsoeren,
28 M. de Vasconcelos, B. Grodtmann, and D. Hauglustaine (2007), REanalysis of the TROpospheric
29 chemical composition over the past 40 years (RETRO)—A long-term global modeling study of
30 tropospheric chemistry: Final report, Jülich/Hamburg, Germany.

31 Seinfeld, J. H. and Pandis, S. N.: *Atmospheric chemistry and physics: from air pollution to*
32 *climate change*, John Wiley & Sons, 2012.

33 Seinfeld, J.H. and Pandis, S.N.: *Atmospheric Chemistry and Physics: from air pollution to*
34 *climate change*, 2nd Edition, John Wiley & Sons, New Jersey, 2006.

1 Tan, J. H., Guo, S. J., Ma, Y. L., Yang, F. M., He, K. B., Yu, Y. C., Wang, J. W., Shi, Z. B., and
2 Chen, G. C.: Non-methane Hydrocarbons and Their Ozone Formation Potentials in Foshan,
3 China, *Aerosol and Air Quality Research*, 12, 387-398, 10.4209/aaqr.2011.08.0127, 2012.

4 Torseth, K., Aas, W., Breivik, K., Fjaeraa, A. M., Fiebig, M., Hjellbrekke, A. G., Myhre, C. L.,
5 Solberg, S., and Yttri, K. E.: Introduction to the European Monitoring and Evaluation Programme
6 (EMEP) and observed atmospheric composition change during 1972-2009, *Atmospheric*
7 *Chemistry and Physics*, 12, 5447-5481, 10.5194/acp-12-5447-2012, 2012.

8 van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M., Kasibhatla, P. S.,
9 Morton, D. C., DeFries, R. S., Jin, Y., and van Leeuwen, T. T.: Global fire emissions and the
10 contribution of deforestation, savanna, forest, agricultural, and peat fires (1997-2009),
11 *Atmospheric Chemistry and Physics*, 10, 11707-11735, 10.5194/acp-10-11707-2010, 2010.

12 Wesely, M. L.: PARAMETERIZATION OF SURFACE RESISTANCES TO GASEOUS DRY
13 DEPOSITION IN REGIONAL-SCALE NUMERICAL-MODELS, *Atmospheric Environment*,
14 23, 1293-1304, 10.1016/0004-6981(89)90153-4, 1989.

15 Yan, Y., Lin, J., Chen, J., and Hu, L.: Improved simulation of tropospheric ozone by a global-
16 multi-regional two-way coupling model system, *Atmospheric Chemistry and Physics*, 16, 2381-
17 2400, 10.5194/acp-16-2381-2016, 2016.

18 Yan, Y. Y., Lin, J. T., Kuang, Y., Yang, D., and Zhang, L.: Tropospheric carbon monoxide over
19 the Pacific during HIPPO: two-way coupled simulation of GEOS-Chem and its multiple nested
20 models, *Atmospheric Chemistry and Physics*, 14, 12649-12663, 10.5194/acp-14-12649-2014,
21 2014.

22 Yan, Y., Lin, J., and He, C.: Ozone trends over the United States at different times of day, *Atmos.*
23 *Chem. Phys.*, 18, 1185-1202, <https://doi.org/10.5194/acp-18-1185-2018>, 2018a.

24 Yan, Y., Pozzer, A., Ojha, N., Lin, J., and Lelieveld, J.: Analysis of European ozone trends in the
25 period 1995–2014, *Atmos. Chem. Phys.*, 18, 5589-5605, [https://doi.org/10.5194/acp-18-5589-](https://doi.org/10.5194/acp-18-5589-2018)
26 2018, 2018b.

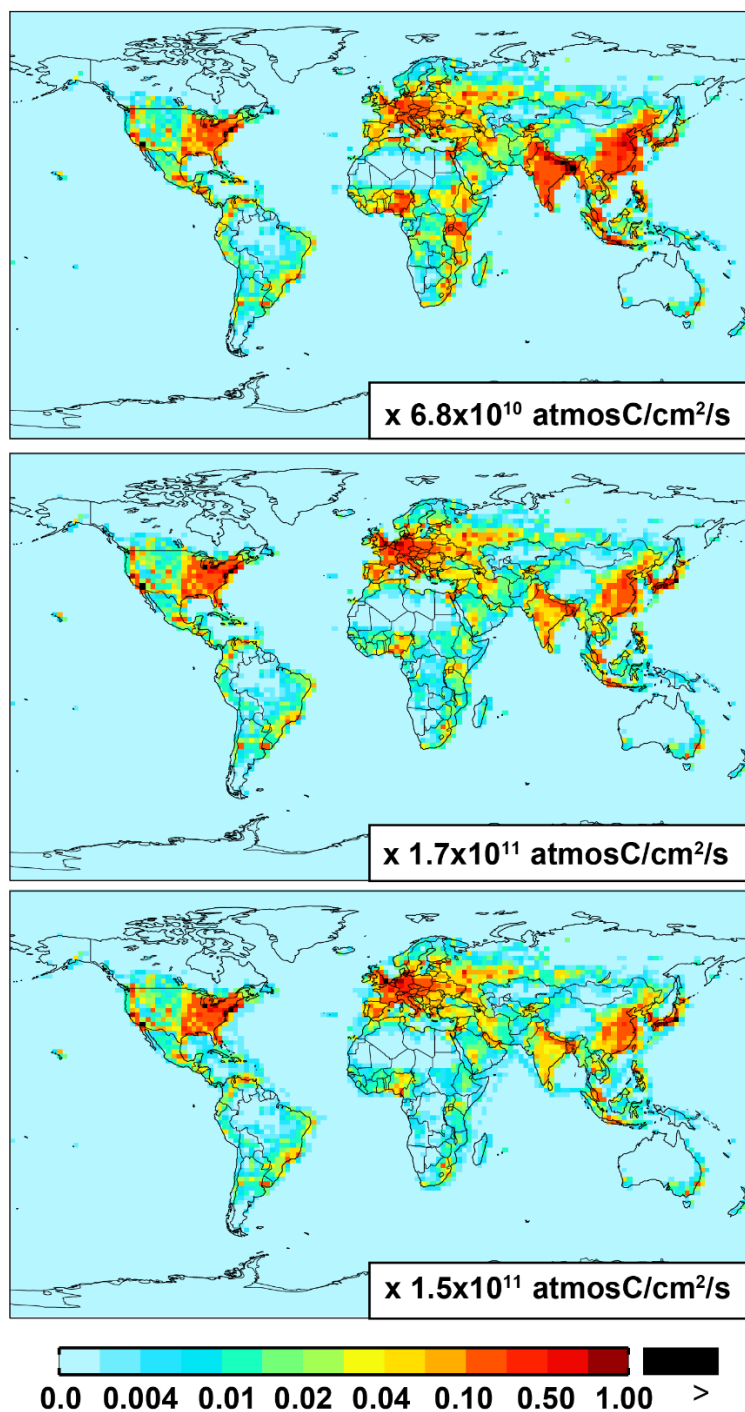
27 You, X. Q., Senthilselvan, A., Cherry, N. M., Kim, H. M., and Burstyn, I.: Determinants of
28 airborne concentrations of volatile organic compounds in rural areas of Western Canada, *Journal*
29 *of Exposure Science and Environmental Epidemiology*, 18, 117-128, 10.1038/sj.jes.7500556,
30 2008.

31 Zhang, L., Jacob, D. J., Yue, X., Downey, N. V., Wood, D. A., and Blewitt, D.: Sources
32 contributing to background surface ozone in the US Intermountain West, *Atmospheric Chemistry*
33 *and Physics*, 14, 5295-5309, 10.5194/acp-14-5295-2014, 2014.

1 Zhang, Y. Z., and Wang, Y. H.: Climate-driven ground-level ozone extreme in the fall over the
2 Southeast United States, Proceedings of the National Academy of Sciences of the United States
3 of America, 113, 10025-10030, 10.1073/pnas.1602563113, 2016.

4

1



2

3 Figure 1. Spatial distribution of anthropogenic emissions from RETRO for benzene (top), toluene
4 (middle), and xylenes (bottom), respectively.

5

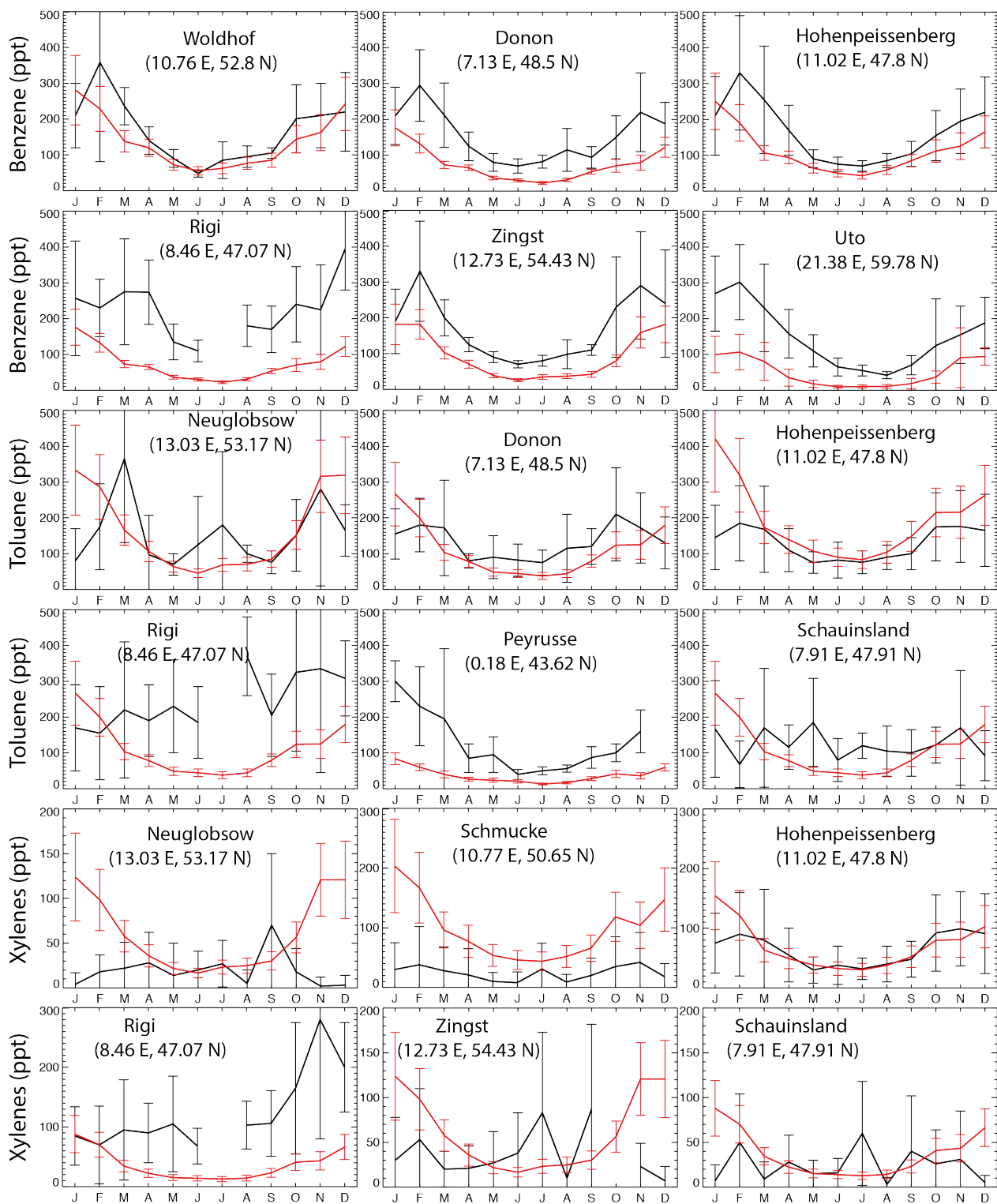


Figure 2. Monthly average EMEP observations (in black) of benzene (first two rows), toluene (middle two rows) and xylenes (last two rows) at six different locations for the year 2005, as well as the model results in the SAPRC simulation (in red), both in ppt. Error bars show the standard deviations.

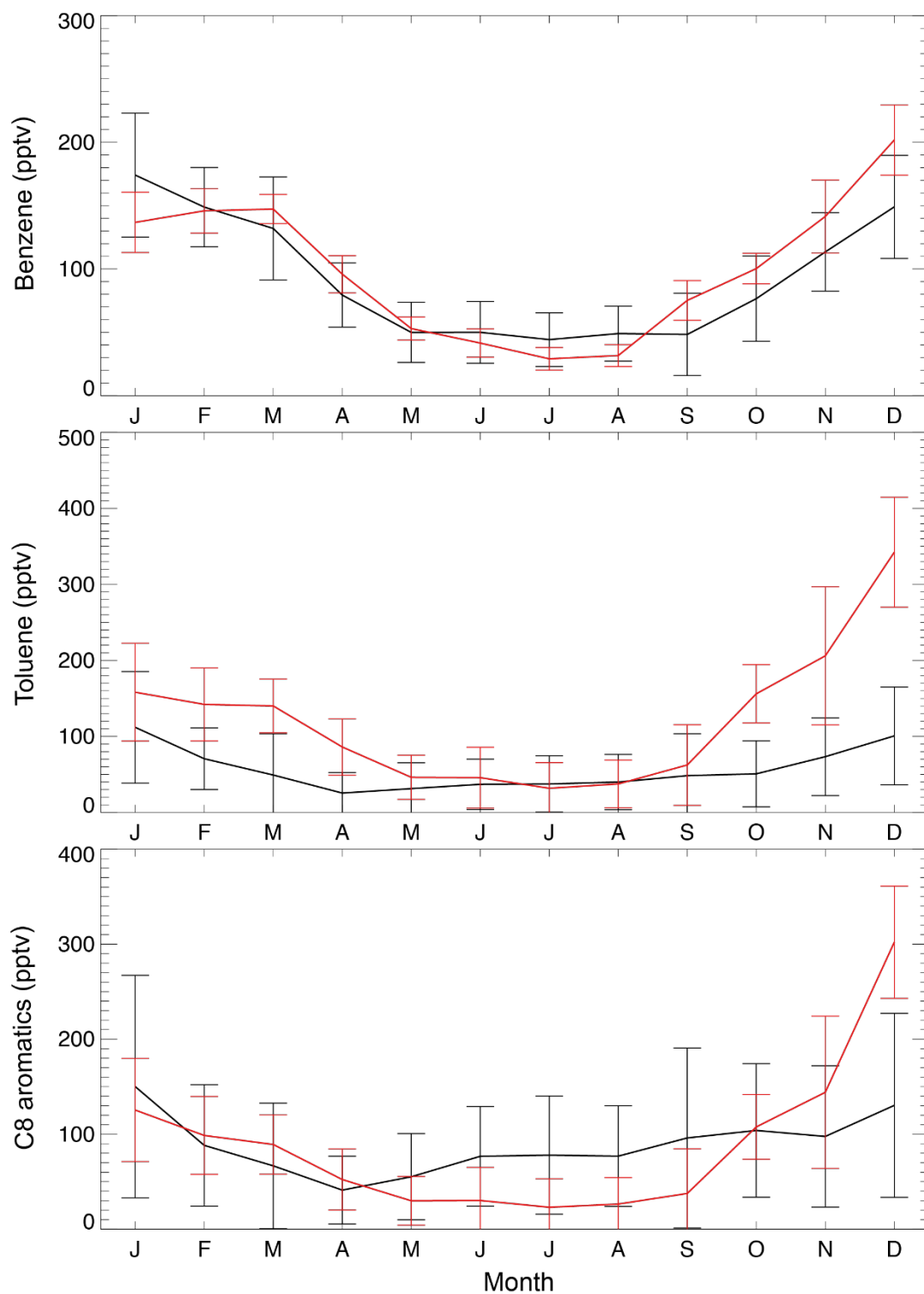


Figure 3. Monthly average KCMP tall tower observations (in black) of benzene, toluene and C₈ (xylenes + ethylbenzene) aromatics in the year 2011 and the model results in the SAPRC simulation (in red). Error bars show the standard deviations.

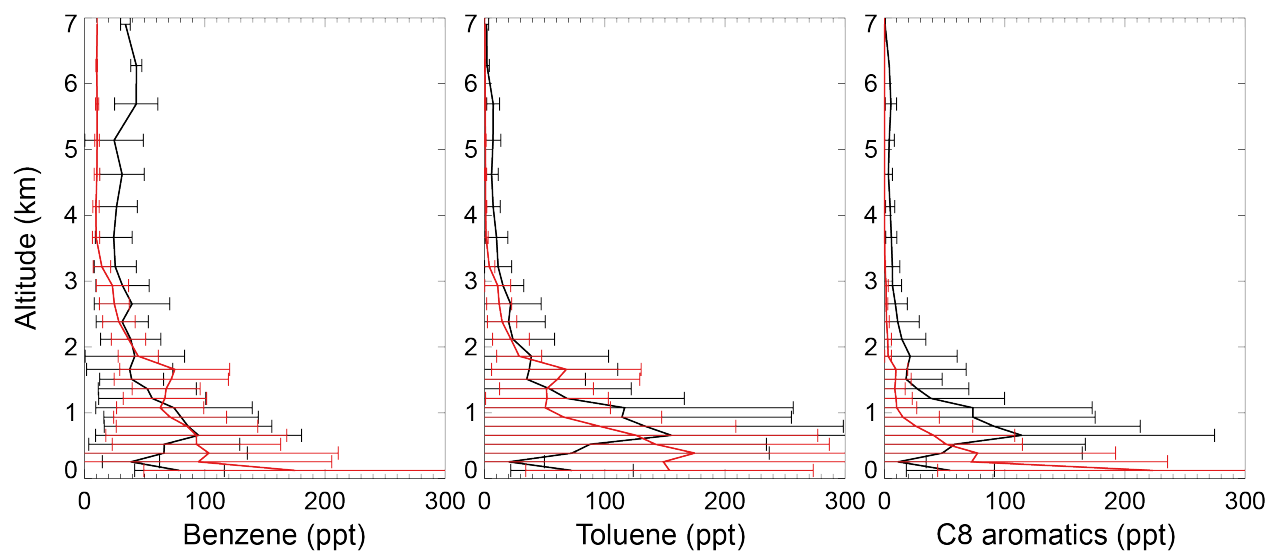


Figure 4. Measured (black) and simulated (red for the SAPRC case) vertical profiles of aromatics in May/June 2010 for the CALNEX campaigns. Model results are sampled at times and locations coincident to the measurements. Horizontal lines indicate the standard deviations.

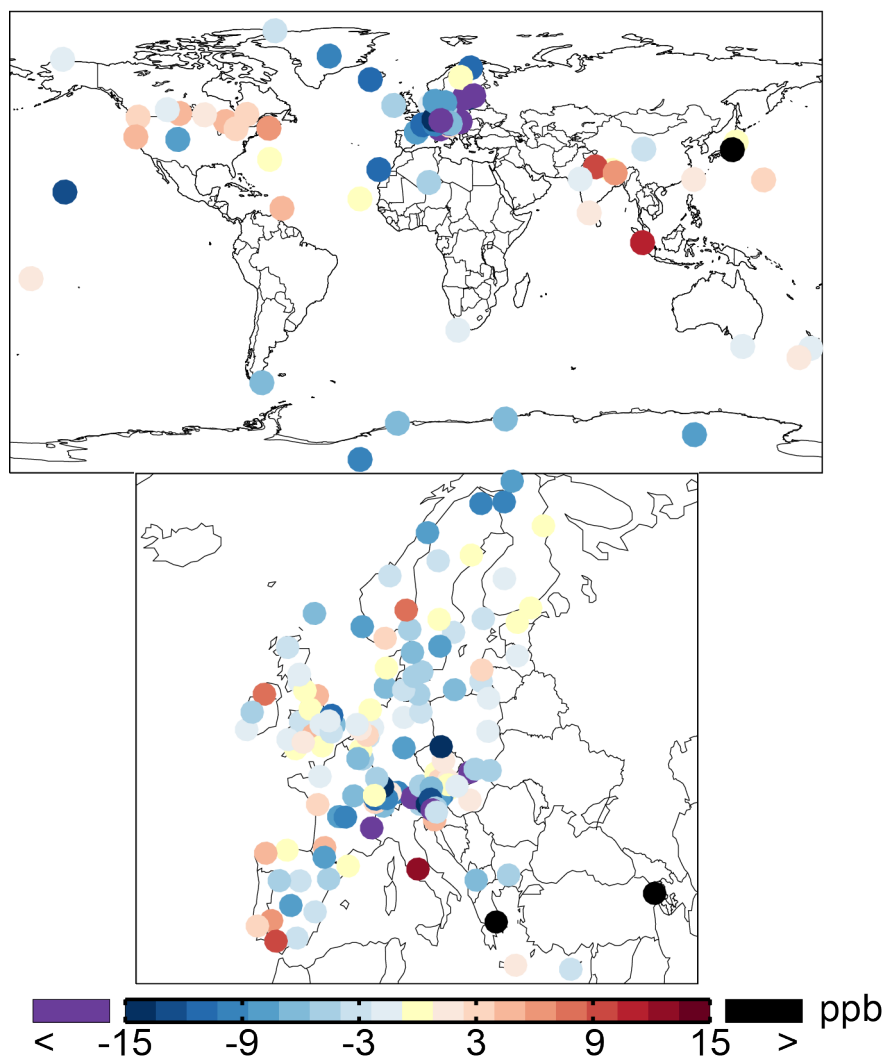


Figure 5. Annual mean model biases for surface ozone in the SAPRC simulation, with respect to measurements from WDCGG (top panel), and EMEP (bottom panel) networks.

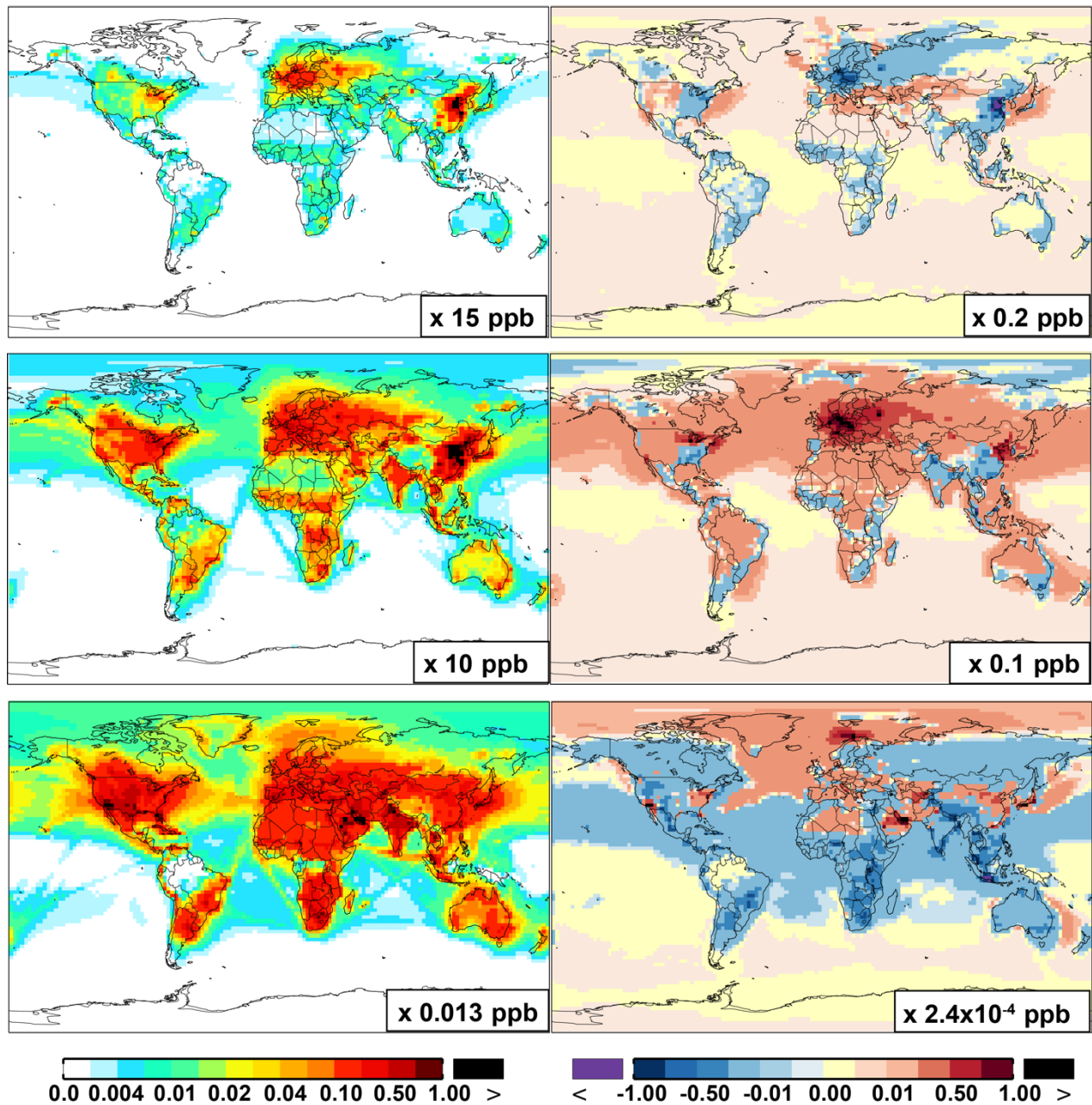
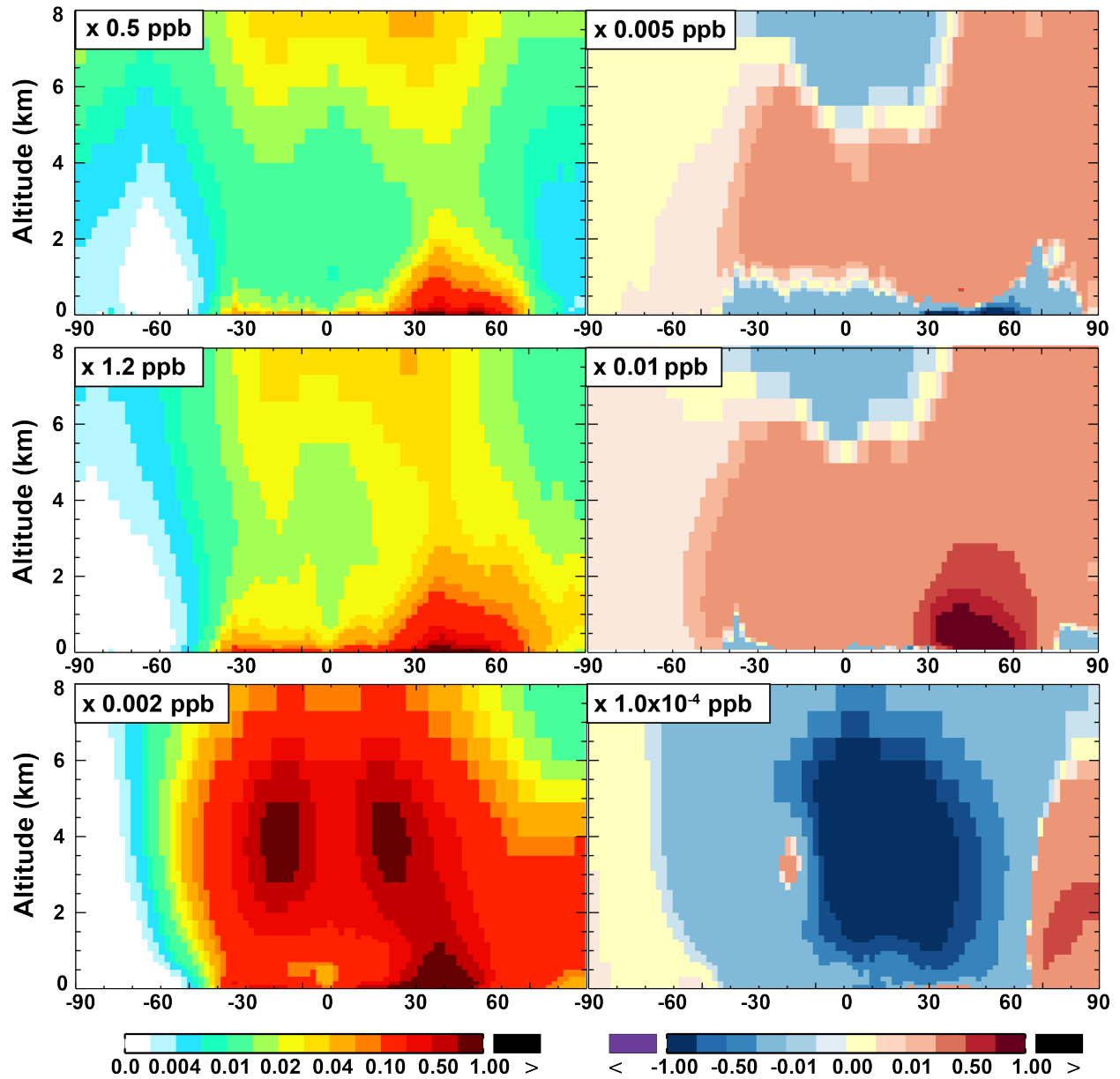


Figure 6. (Left column) Modeled spatial distributions of annual mean surface NO (top), NO_2 (middle), and NO_3 (bottom) simulated in the Base case for the year 2005. (Right column) The respective changes from Base to SAPRC.

1



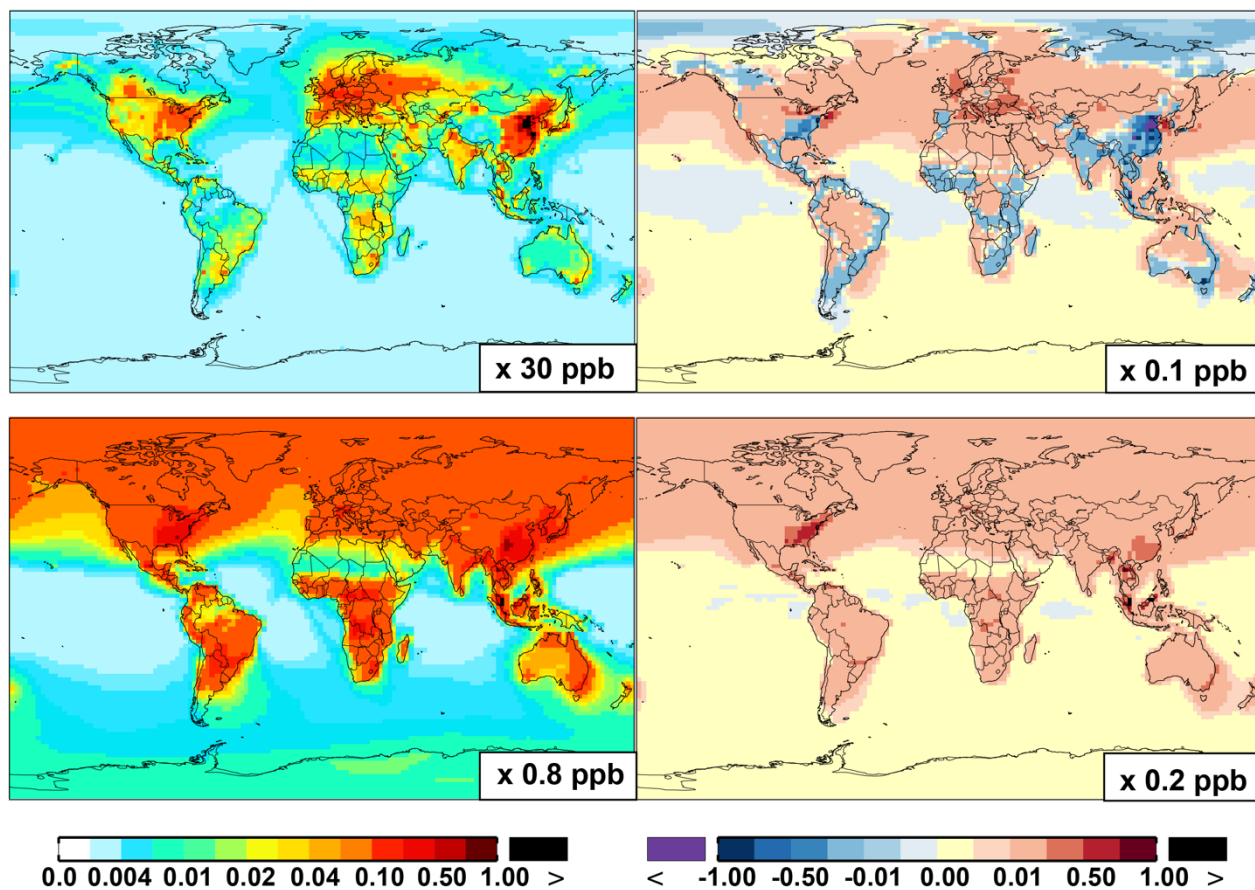
2

3 Figure 7. (Left column) Modeled zonal average latitude-altitude distributions of annual mean NO (top)
 4 and NO₂ (middle), and NO₃ (bottom) simulated in the Base scenario for the year 2005. (Right column)
 5 The respective changes from Base to SAPRC.

6

7

1



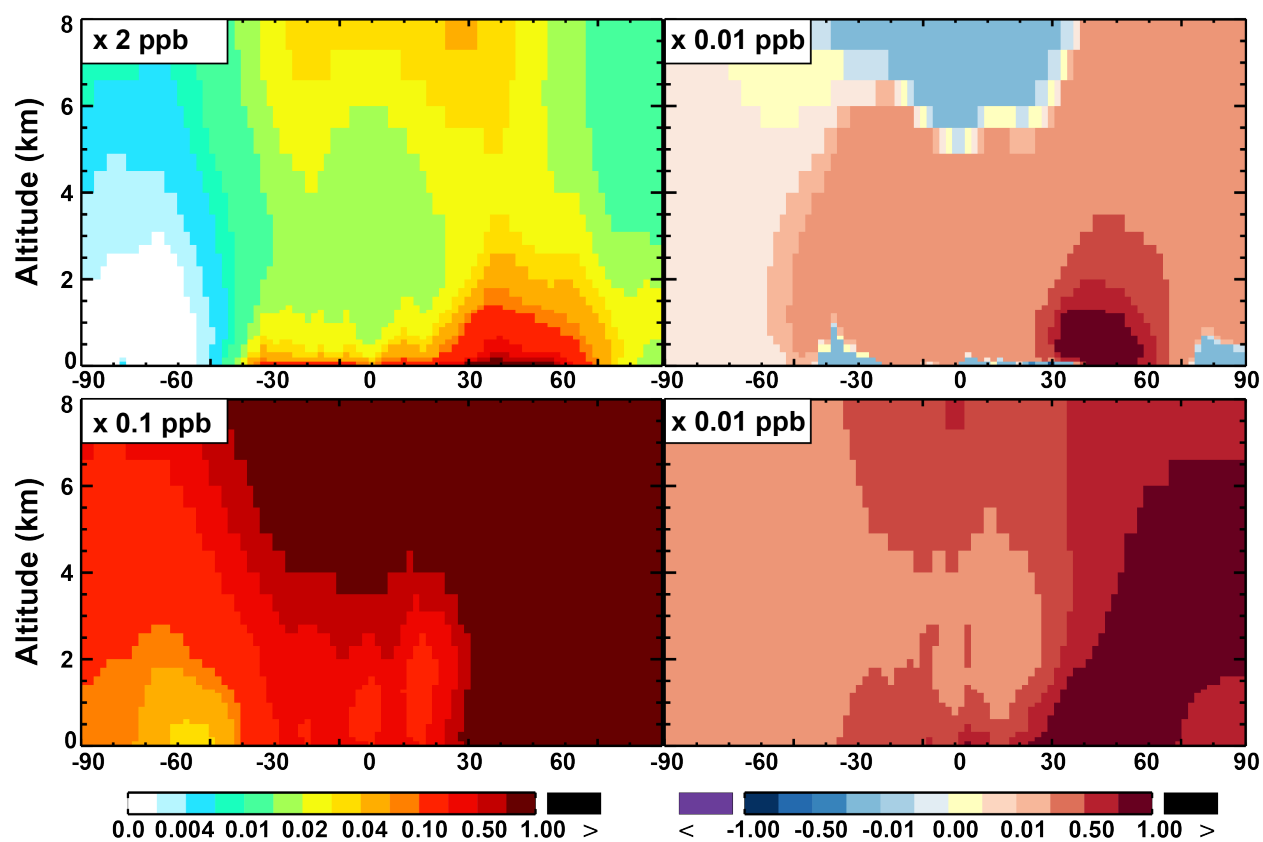
2

3 Figure 8. Same as Fig. 6 but for NO_x (top panels) and PAN (bottom panels).

4

5

1

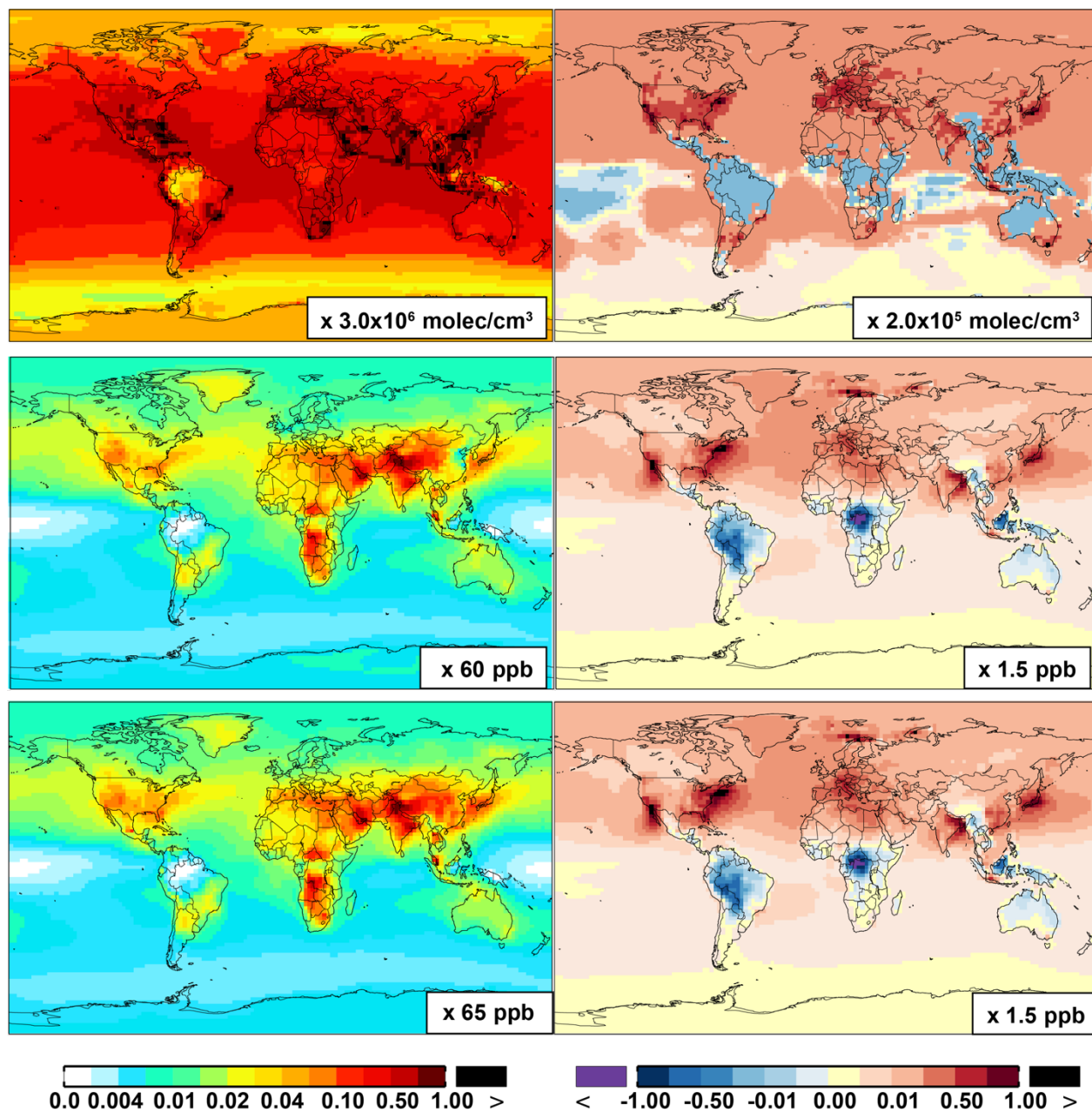


2

3 Figure 9. Same as Fig. 7 but for NO_x (top panels) and PAN (bottom panels).

4

1

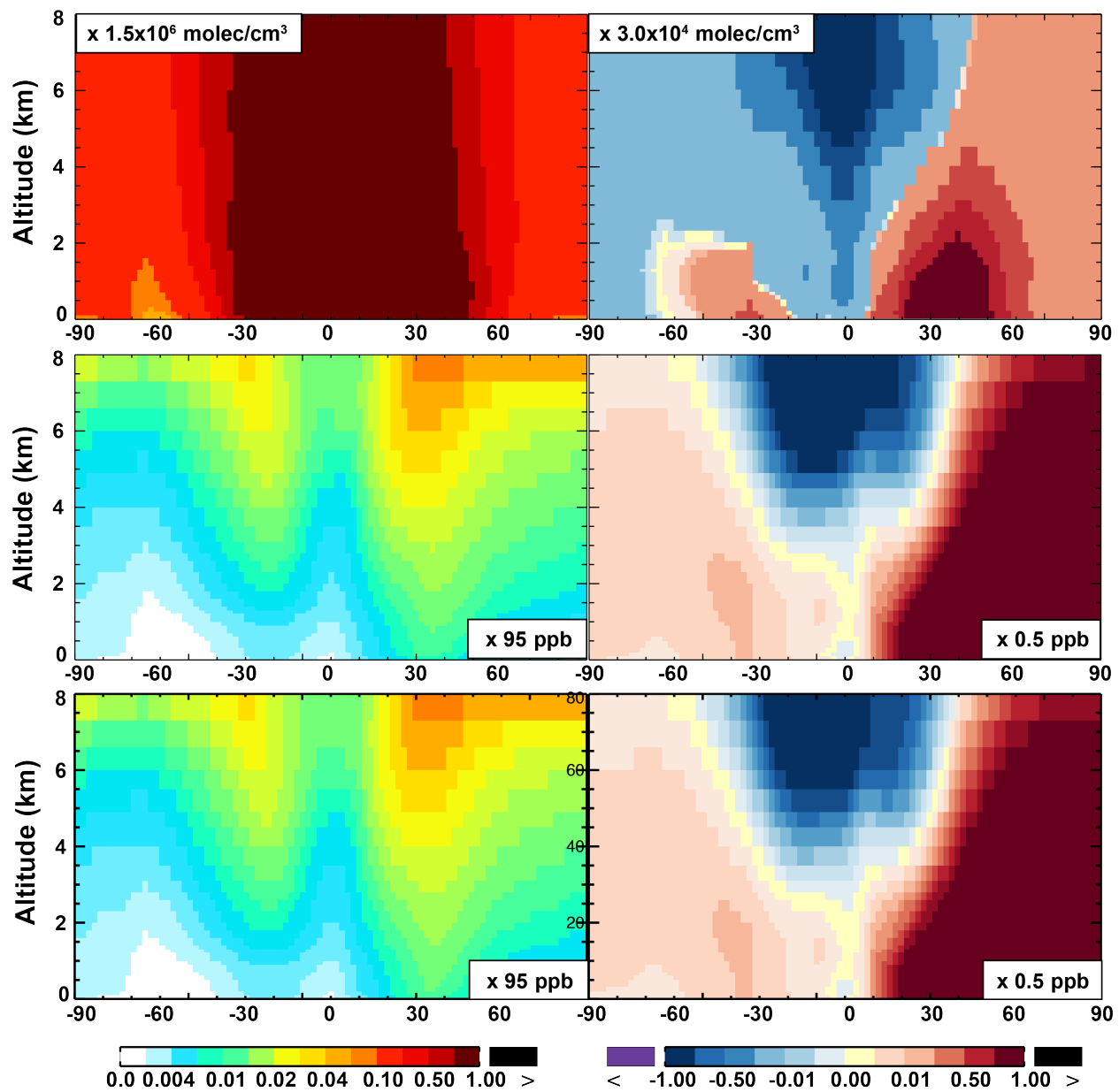


2

3 Figure 10. Same as Fig. 6 but for OH (top panels), O_3 (middle panels) and O_x (bottom panels).

4

1



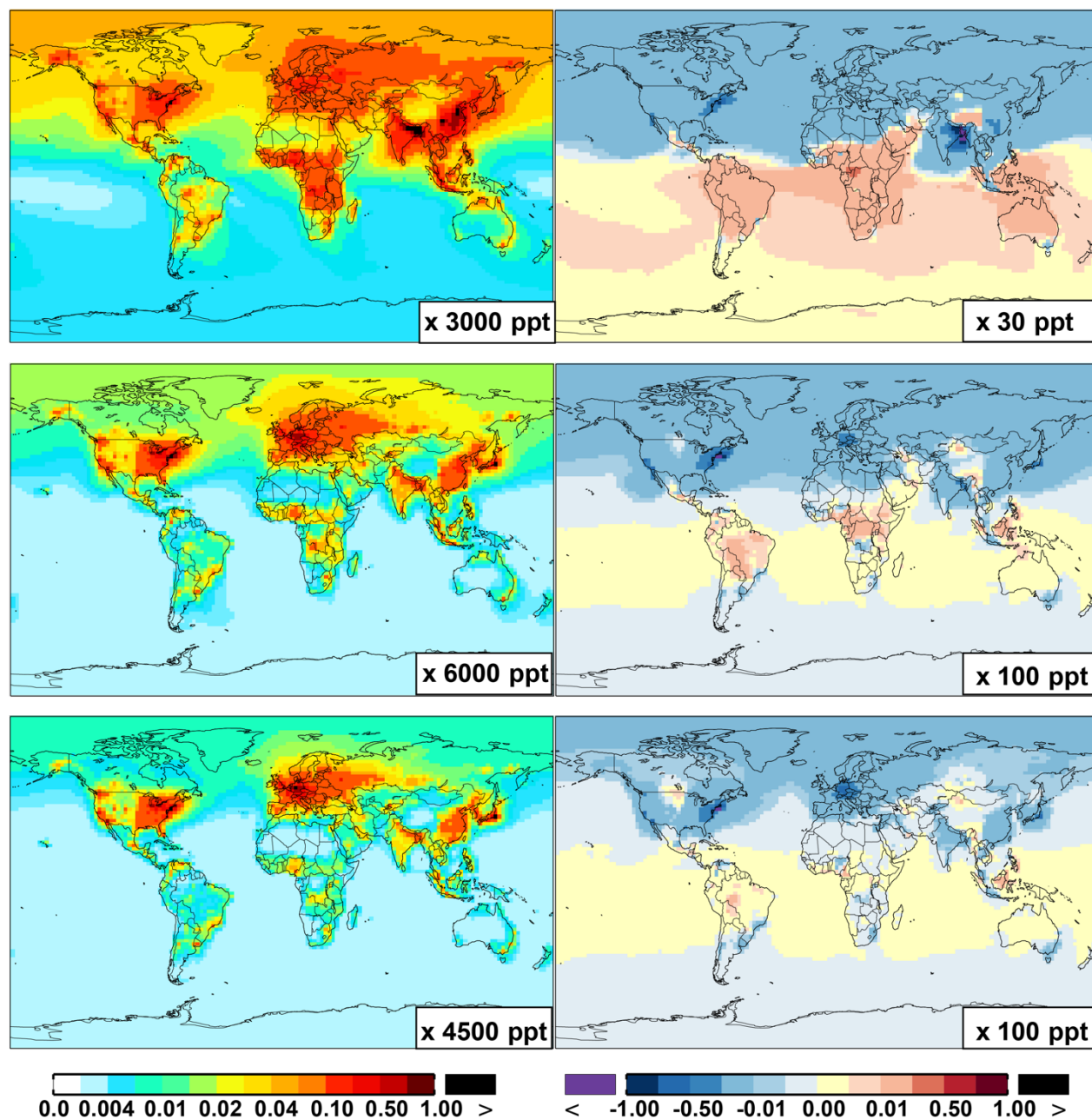
2

3 Figure 11. Same as Fig. 7 but for OH (top panels), O₃ (middle panels) and O_x (bottom panels).

4

5

1

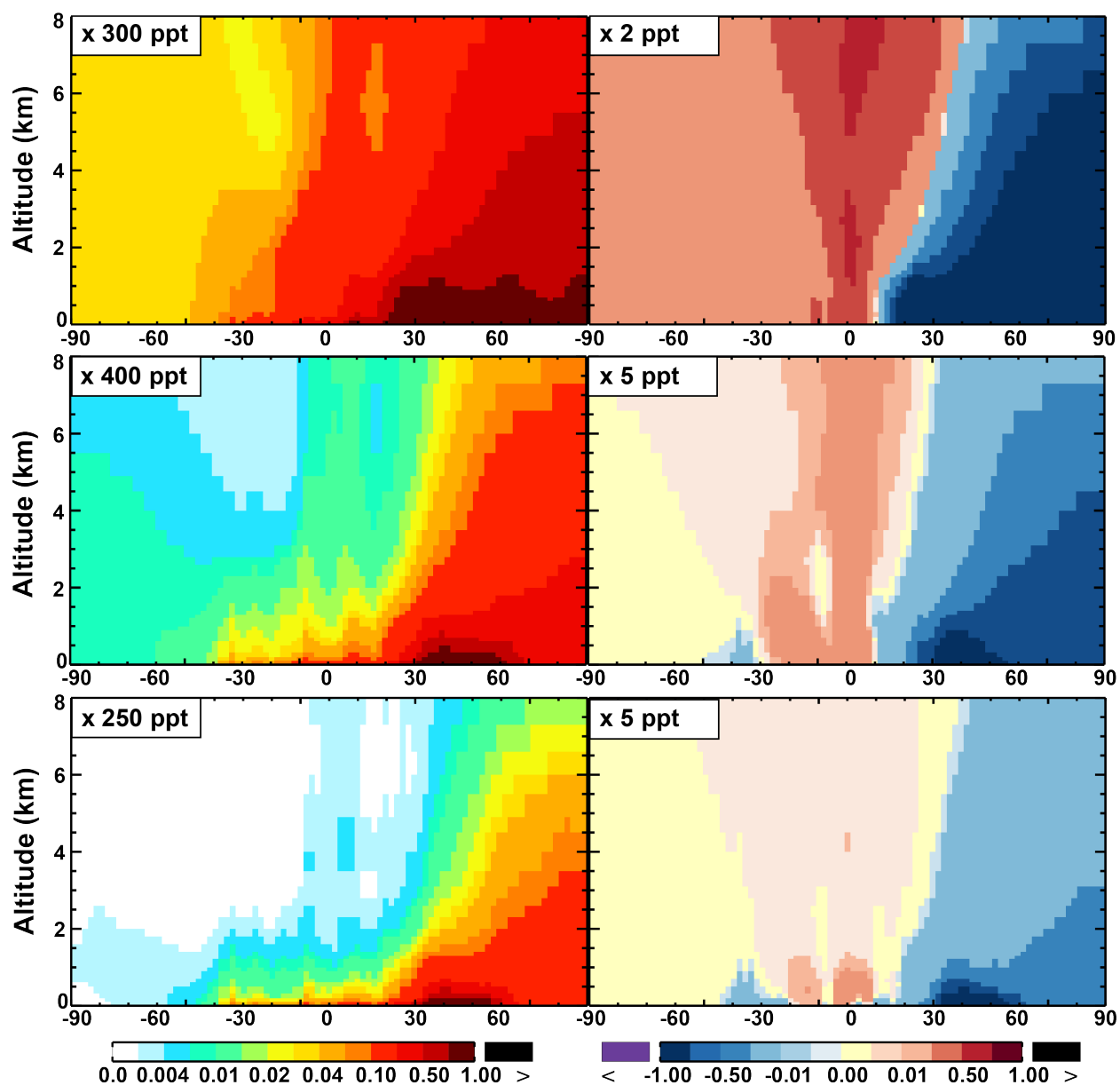


2

3 Figure 12. Same as Fig. 6 but for benzene (top panels), toluene (middle panels) and xylene (bottom
4 panels).

5

1



2

3 Figure 13. Same as Fig. 7 but for benzene (top panels), toluene (middle panels) and xylene (bottom
4 panels).

5

6

1
2 Table 1. Summary of the statistical comparison between observed and simulated concentrations (ppt for
3 aromatics, ppb for ozone). MMOD and MOBS represent the mean values for the SAPRC simulation and
4 the observation, respectively. MRB is the relative bias of model results defined as: $(\text{MMOD} - \text{MOBS})/\text{MOBS}$. SMOD and SOBS are their standard deviations. TCOR and SCOR are the temporal and
5
6 spatial correlations between model results and measurements.

Species	Network	Num of sites	Time resolution	MMOD (MRB)	MOBS	SMOD	SOBS	TCOR	SCOR
Benzene	CARIBIC	1241	Instantaneous	12.3 (-23%)	16.0	4.2	15.8	-	0.31
	EEA	22	Annual mean	131.6 (-32%)	194.0	32.1	118.4	-	0.49
	EMEP	14	Monthly	106.5 (-36%)	166.4	38.7	71.7	0.77	0.44
	CALNEX	7708	Instantaneous	66.1 (15%)	57.7	78.3	57.7	-	0.51
	KCMP	1	Hourly	99.9 (9%)	91.5	92.6	56.7	0.65	-
Toluene	CARIBIC	789	Instantaneous	1.5 (-58%)	3.6	0.7	7.5	-	0.36
	EEA	6	Annual mean	180.9 (-25%)	240.3	66.8	59.4	-	0.41
	EMEP	12	Monthly	113.2 (-15%)	133.1	47.3	66.2	0.81	0.47
	CALNEX	7708	Instantaneous	80.6 (10%)	73.2	179.7	131.9	-	0.46
	KCMP	1	Hourly	121.2 (114%)	56.7	191.4	54.7	0.51	-
Xylenes	EMEP	8	Monthly	78.4 (85%)	42.3	34.5	41.9	0.78	0.48
C ₈ aromatics	CALNEX	7708	Instantaneous	28.8 (-41%)	48.6	112.2	97.2	-	0.39
	KCMP	1	Hourly	88.9 (-2%)	90.3	119.2	79.5	0.46	-
Ozone	WDCGG	64	Monthly	28.6 (-16%)	34.1	12.8	14.2	0.68	0.54
	EMEP	130	Monthly	27.7 (-9%)	30.6	13.2	10.3	0.76	0.52

1 Table 2. Annual and seasonal mean changes (%) in modeled surface as well as tropospheric
2 concentrations from the Base to the SAPRC simulation. Also shown are the numbers for northern
3 hemisphere (NH) and southern hemisphere (SH).

Species	Annual		MAM		JJA		SON		DJF	
	Surface (NH, SH)	Trop (NH, SH)	Surface (NH, SH)	Trop (NH, SH)	Surface (NH, SH)	Trop (NH, SH)	Surface (NH, SH)	Trop (NH, SH)	Surface (NH, SH)	Trop (NH, SH)
NO	-0.2% (-0.2%, -1.4%)	0.6% (0.8%, - 0.2%)	-0.4% (-0.3%, -1.7%)	0.7% (0.9%, - 0.3%)	-1.3% (-1.3%, -1.2%)	-0.1% (-0.1%, -0.1%)	-1.5% (-1.5%, -1.3%)	-0.5% (-0.5%, -0.3%)	0.8% (0.9%, - 1.6%)	1.6% (2.0%, - 0.3%)
O ₃	0.9% (1.2%, 0.3%)	0.4% (0.6%, - 0.1%)	1.1% (1.6%, 0.3%)	0.5% (0.8%, - 0.1%)	0.6% (0.9%, 0.2%)	0.3% (0.5%, - 0.1%)	0.8% (1.1%, 0.4%)	0.4% (0.6%, - 0.1%)	1.0% (1.3%, 0.3%)	0.4% (0.6%, - 0.1%)
CO	0.8% (0.5%, 1.3%)	1.0% (0.7%, 1.4%)	0.5% (0.2%, 1.1%)	0.7% (0.4%, 1.3%)	1.1% (0.8%, 1.4%)	1.2% (1.0%, 1.5%)	1.1% (0.9%, 1.5%)	1.3% (1.1%, 1.6%)	0.5% (0.3%, 1.0%)	0.7% (0.5%, 1.2%)
HNO ₃	1.1% (1.3%, -0.6%)	0.3% (0.7%, - 0.9%)	1.2% (1.3%, -0.4%)	0.4% (0.7%, - 0.9%)	0.7% (0.9%, - 0.6%)	-0.1% (0.2%, - 1.0%)	1.0% (1.4%, - 0.7%)	0.2% (0.7%, - 1.0%)	1.4% (1.6%, - 0.7%)	0.6% (1.1%, - 0.8%)
NO ₂	1.0% (1.0%, 0.2%)	2.1% (2.4%, 0.7%)	0.8% (0.8%, 0.3%)	1.8% (2.0%, 0.8%)	-0.2% (-0.3%, 0.1%)	0.6% (0.6%, 0.8%)	0.5% (0.6%, 0.2%)	1.3% (1.5%, 0.5%)	2.0% (2.1%, 0.2%)	3.6% (4.0%, 0.5%)
NO ₃	-0.9% (-0.6%, -2.7%)	-4.1% (- 4.5%, - 3.5%)	-1.5% (-1.3%, -2.7%)	-5.6% (-7.0%, -3.0%)	-0.9% (-0.5%, -2.5%)	-3.7% (-4.3%, -3.0%)	-0.5% (-0.1%, -2.6%)	-3.4% (-3.4%, -3.6%)	-0.8% (-0.5%, -3.6%)	-4.1% (-4.2%, -4.5%)
BENZ	-0.5% (-0.6%, 0.6%)	-0.4% (- 0.6%, 1.4%)	-0.9% (-1.0%, 0.7%)	-1.0% (-1.1%, 1.7%)	0.1% (-0.1%, 0.5%)	0.7% (0.5%, 1.0%)	-0.1% (-0.2%, 0.8%)	0.2% (-0.1%, 1.6%)	-0.6% (-0.6%, 0.9%)	-0.6% (-0.7%, 2.0%)
TOLU	-1.2%	-1.9%	-1.5%	-2.8%	-0.8%	-0.9%	-1.0%	-1.5%	-1.3%	-1.9%

	(-1.3%, 0.1%)	(- 2.0%, 0.4%)	(-1.6%, 0.3%)	(-3.0%, 0.8%)	(-1.0%, -0.2%)	(-1.2%, -0.1%)	(-1.1%, 0.2%)	(-1.6%, 0.6%)	(-1.3%, 0.4%)	(-2.0%, 1.3%)
XYLE	-1.4% (-1.5%, -0.3%)	-2.3% (- 2.3%, - 0.2%)	-1.2% (-1.2%, -0.2%)	-2.1% (-2.2%, 0.3%)	-1.2% (-1.3%, -0.6%)	-1.5% (-1.6%, -0.9%)	-1.6% (-1.7%, -0.1%)	-2.3% (-2.4%, 0.2%)	-1.5% (-1.5%, -0.1%)	-2.4% (-2.4%, 0.5%)
OH	1.1% (1.6%, 0.3%)	0.2% (0.6%, - 0.3%)	1.4% (1.9%, 0.3%)	0.4% (0.8%, - 0.4%)	1.2% (1.3%, 0.5%)	0.3% (0.5%, - 0.2%)	0.9% (1.5%, 0.3%)	0.1% (0.4%, - 0.4%)	1.0% (2.1%, 0.2%)	0.1% (0.9%, - 0.3%)
HO ₂	3.0% (3.2%, 2.8%)	1.3% (1.4%, 1.2%)	2.9% (2.8%, 3.1%)	1.4% (1.5%, 1.2%)	3.3% (3.2%, 3.6%)	1.3% (1.2%, 1.6%)	3.1% (3.4%, 2.8%)	1.3% (1.5%, 1.2%)	2.8% (3.7%, 2.2%)	1.2% (1.9%, 0.9%)
OH/ HO ₂	-1.4% (-1.0%, -1.7%)	-0.9% (- 0.7%, - 1.3%)	-1.2% (-1.1%, -1.9%)	-0.8% (-0.5%, -1.4%)	-1.6% (-1.1%, -2.0%)	-1.0% (-0.7%, -1.6%)	-1.4% (-0.9%, -1.9%)	-1.0% (-0.8%, -1.4%)	-1.1% (-0.5%, -2.1%)	-0.8% (-0.6%, -1.3%)

1

2

1 Table 3. Annual and seasonal mean model ozone biases for the Base and the SAPRC case, compared to
 2 measurements from WDCGG and EMEP.

Species (ppb)	Annual		MAM		JJA		SON		DJF	
	Base	SAPRC	Base	SAPRC	Base	SAPRC	Base	SAPRC	Base	SAPRC
WDCGG	-6.0	-5.4	-9.0	-8.4	-0.4	0.1	-2.5	-2.1	-11.9	-11.5
EMEP	-3.5	-2.8	-5.5	-4.7	4.5	5.2	0.3	0.8	-13.1	-12.8

3

4

5

6