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PCR-GLOBWB 2: a 5 arc-minute global hydrological

8 and water resources model

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Abstract.

We present PCR-GLOBWB 2, a global hydrology and water resources model. Compared to previous versions of PCR-GLOBWB, this version fully integrates water use. Sector-specific water demand, groundwater and surface water withdrawal, water consumption and return flows are dynamically calculated at every time step and interact directly with the simulated hydrology. PCR-GLOBWB 2 has been fully rewritten in Python and PCRaster-Python and has a modular structure, allowing easier replacement, maintenance, and development of model components. PCR-GLOBWB 2 has been implemented at 5 arc-minute resolution, but a version parameterized at 30 arc-minute resolution is also available. Both versions are available as open source codes on https://github.com/UU-Hydro/PCR-GLOBWB model. PCR-GLOBWB 2 has its own routines for groundwater dynamics and surface water routing. These relatively simple routines can alternatively be replaced by dynamically coupling PCR-GLOBWB 2 to a global two-layer groundwater model and 1D-2D-hydrodynamic models, respectively. Here, we describe the main components of the model, compare results of the 30 arc-minute and the 5 arc-minute versions and evaluate their model performance using GRDC discharge data. Results show that model performance of the 5 arc-minute version is notably better than that of the 30 arc-minute version. Furthermore, we compare simulated time series of total water storage (TWS) of the 5 arc-minute model with those observed with GRACE, showing similar negative trends in areas of prevalent

groundwater depletion. Also, we find that simulated water withdrawal, by source and sector, matches

reasonably well with reported water withdrawal from AQUASTAT.

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1 Introduction

50 51 The last decades saw the development of an increasing number of global hydrological models (GHMs), e.g. 52 VIC (Liang et al., 1994; Nijssen et al., 2001), WMB (Fekete et al., 2002), WaterGAP (Döll et al., 2003), H08 53 (Hanasaki et al., 2008a), MAC-PDM (Gosling and Arnell, 2011) (see Bierkens et al., 2014, Bierkens, 2015 and 54 Kauffeldt et al. 2016 for a more extensive list, also including land surface models). GHMs have become 55 essential tools to quantify and understand the global terrestrial water cycle, as they simulate the distributed 56 hydrological response to weather and climate variations at higher resolution (typically $0.5^{\circ} \times 0.5^{\circ}$) than used 57 previously in general circulation models (GCMs), with more sophisticated runoff generation processes and 58 river routing. As such, global hydrological models have been used for medium-range to seasonal flood 59 forecasting (Bierkens and van Beek, 2009; Alfieri et al., 2013; Candogan Yossef et al., 2013) as well as for a 60 myriad of water-related global change assessments. Examples are: the projection or estimation of future flood 61 and drought events (Sperna-Weiland et al., 2012; Dankers et al., 2013; Prudhomme et al., 2013, Wanders et al. 62 2015, Wanders and Wada, 2016), current and future flood hazard and risk (Pappenberger et al., 2012; 63 Hirabayashi et al., 2013; Ward et al., 2013; Winsemius et al., 2013; 2016), global groundwater depletion (Wada et al., 2010; Gleeson et al., 2012), the contribution of terrestrial water stores to global sea level change 64 65 (Konikow, 2011; Wada et al., 2012; Pohkrel et al., 2013), current and future water scarcity under climate 66 change and increasing population growth (Hanasaki et al., 2008b; Wada et al., 2011a, 2011b; Schewe et al., 67 2013; Haddeland et al., 2013; Wada and Bierkens, 2014), tele-connections between climate oscillations and 68 water availability (Wanders and Wada, 2015), the impact of land use change om global water resources (Rost 69

et al., 2008; Sterling et al., 2015; Bosmans et al., 2016) and trends in surface water temperature and cooling 70 water potential (van Beek et al., 2012; van Vliet et al., 2012). More recently, the output from global 71 hydrological models has been extended to study socioeconomic impacts, such as virtual water trade (Konar et

72 al., 2013; Dalin et al., 2017) and future agricultural production (Elliott et al., 2013). These applications show 73

that GHMs have become invaluable tools in support of global water management and policy assessments.

PCR-GLOBWB (PCRaster GLOBal Water Balance) (van Beek and Bierkens, 2009; van Beek et al. 2011) is one of the recently developed GHMs. PCR-GLOBWB is a grid-based global hydrological model developed at the Department of Physical Geography, Faculty of Geosciences, Utrecht University, the Netherlands. The model, describing the terrestrial part of the hydrological cycle, was first introduced in a technical report by van Beek and Bierkens (2009) and then formally published in a paper of Van Beek et al. (2011), focusing on global water availability issues. PCR-GLOBWB was originally developed to solve the global daily surface water balance with a spatial resolution of 30 arc-minutes (about 50 km by 50 km at the equator) and compare the resulting fresh water availability with monthly sectoral water demand in order to assess global-scale water scarcity (van Beek et al., 2011; Wada et al., 2011a,b). In this first version of PCR-GLOBWB (called PCR-GLOBWB 1 hereafter), similar to other global-scale hydrological models, water demand and water availability

are treated independently, i.e. without direct feedback between human water use and other terrestrial water

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86 fluxes (e.g. Döll and Siebert, 2002; Wisser et al., 2010). Since it was first introduced, PCR-GLOBWB has 87 been applied extensively in global water resources assessment studies. For instance, a recent search on Scopus (accessed on 30 October 2017) on the key-word "PCR-GLOBWB" yielded 97 publications with collectively 88 89 over 2100 references. Since the first version, several new model features have been introduced such as a 90 comprehensive water demand and irrigation module (Wada et al., 2011b, 2014), a scheme for dynamic 91 allocation of sectoral water demand to available surface water and groundwater resources and the associated 92 calculation of return flow (de Graaf et al., 2014). These features essentially introduced a two-way interaction 93 between water demand, water withdrawal, water consumption and availability, particularly over irrigated areas 94 where water demand is large and return flow is significant. Nevertheless, all of these preceding studies using 95 PCR-GLOBWB were performed at a relatively coarse resolution of 30 arc-minutes, limiting their sub-regional 96 or local applications. Additionally, some added functionalities, such as the possibility to couple the land 97 surface component of PCR-GLOBWB to a global MODFLOW-based groundwater model (Sutanudjaja et al., 98 2011; 2014; de Graaf et al., 2015; 2017) and an extension to simulate surface water temperature (Van Beek et 99 al., 2012), were incorporated in different versions based on the original PCR-GLOWB 1, leading to divergent 100 model code development. 101 102 The objective of this paper is to summarize and present the new version of the model, PCR-GLOBWB 2, 103 which consolidates all components that have been developed since the original version of the model was first 104 introduced (van Beek et al., 2011). The new version of the model, PCR-GLOBWB 2, simulates the water 105 balance at a finer spatial resolution of 5 arc-minutes and supersedes the original PCR-GLOBWB 1 that has a resolution of 30 arc-minutes1. The finer resolution of PCR-GLOBWB 2 allows a much better representation of 106 107 the effects of spatial heterogeneity in topography, soils, and vegetation on terrestrial hydrological dynamics 108 (Wood et al., 2011; Bierkens et al., 2014). Likewise, it provides a better resolution for visualization that allows 109 stakeholders and decision makers to assess model simulation output more easily and directly for the places 110 they are specifically interested in (Sheffield et al., 2010; Beven and Cloke, 2012). To evaluate the possible 111 improvements, this paper also presents the first validation results from the simulation of PCR-GLOBWB 2 at 5 112 arc-minute resolution and compares this with a 30 arc-minutes version. As discharge data are commonly used 113 in hydrological model performance evaluation, the simulated river discharge of PCR-GLOBWB 2 is compared 114 to in situ discharge observations from the Global Runoff Data Centre (GRDC, 2014). 115 116 The paper is organized as follows. Section 2 provides a global description of PCR-GLOBWB 2, including its 117 model structure and the new components and functionalities that have been added since PCR-GLOBWB 1. In 118 section 3 the global application of PCR-GLOBWB 2 is demonstrated and the results from a 58-year simulation 119 (1958-2015) are validated against observations of discharge, total water storage and reported withdrawal data. 120 Section 4 summarizes and concludes this paper and discusses possible future developments. Section 5 provides

Note that Wada et al. (2016) made a preliminary version of the model that operates at 6 arc-minutes.

information about availability of the model code and the underlying data.

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2. PCR-GLOBWB 2 – Model description

and to simulate the surface water temperature.

2.1 General overview

129 PCR-GLOBWB 2 is a state-of-the-art grid-based global hydrology and water resources model. It is a 130 component-based model implementation in Python using open source PCRaster Python routines (Karssenberg 131 et al., 2010; http://pcraster.geo.uu.nl/). The code is distributed through Github. The computational grid covers 132 all continents except Greenland and Antarctica. Currently two versions are available: one with a spatial 133 resolution of 5 arc-minutes in latitude and longitude and one with a coarser resolution of 30 arc-minutes. 134 Typical time steps for hydrology and water use are one-day while sub-daily time stepping is used for hydrodynamic river routing. For all dynamic processes involved, PCR-GLOWB 2 uses a time-explicit scheme. 135 136 For each grid cell and each time step, PCR-GLOBWB 2 simulates moisture storage in two vertically stacked 137 upper soil layers (S_1+S_2 in Figure 1), as well as the water exchange between the soil, the atmosphere and the 138 underlying groundwater reservoir (S_3 in Figure 1). The exchange with the atmosphere comprises of 139 precipitation, evaporation from soils, open water, snow and soils and plant transpiration, while the model also 140 simulates snow accumulation and snowmelt. Sub-grid variability of land use, soils and topography is included 141 and influences the schemes for runoff-infiltration partitioning, interflow, groundwater recharge (from S₂ to S₃) 142 and capillary rise (from S_3 to S_2). Runoff, generated by snowmelt, surface runoff, interflow and baseflow, is 143 routed across the river network to the ocean or endorheic lakes and wetlands. Routing can either be simple 144 accumulation, simplified dynamic routing using a method of characteristics, or kinematic wave routing. In

PCR-GLOBWB 2 includes a simple reservoir operation scheme that is applied to over roughly 6000 manmade reservoirs from the GranD database (Lehner et al., 2011), which are progressively introduced according to their construction year. Human water use is fully integrated within the hydrological model, meaning that at each time step: 1) water demands are estimated for irrigation, livestock, industry and households; 2) these demands are translated into actual withdrawals from groundwater, surface water (rivers, lakes and reservoirs) and desalinization, subject to availability of these resources and maximum groundwater pumping capacity in place; 3) consumptive water use and return flows are calculated per sector.

case the kinematic wave routing is used, it is also possible to use a (simplified) floodplain inundation scheme

As an option PCR-GLOBWB 2 can be partially or fully coupled to a two-layer global groundwater model based on MODFLOW (de Graaf et al, 2017). Recent work (Hoch et al., 2017a,b) also includes coupling PCR-GLOBWB 2 to either Delft3D Flexible Mesh (Kernkamp et al., 2011) or LISFLOOD-FP (Bates et al., 2010)

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which are model codes that can be used to solve the 1D-2D shallow water equations (or approximation thereof) for detailed inundation studies.

Precip Evap Q_{channel}

S1

Q_{bf}

Reservoirs

Figure 1. Schematic overview of a PCR-GLOBWB 2 cell and its modelled states and fluxes. S_1 , S_2 (soil moisture storage), S_3 (groundwater storage), Q_{dr} (surface runoff – from rainfall and snowmelt), Q_{sf} (interflow or stormflow), Q_{bf} (baseflow or groundwater discharge), Inf (riverbed infiltration from to groundwater). The thin red lines indicate surface water withdrawal, the thin blue lines groundwater abstraction, the thin red dashed lines return flows from surface water use and the thin dashed blue lines return flows from groundwater use surface. For each sector: withdrawal - return flow = consumption. Water consumption adds to total evaporation.

Q_{channel}

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2.2 Model structure and flexibility

PCR-GLOBWB 2 has a flexible modular structure in which the exchange of water between a series of interconnected stores is easily performed (Figure 1). The modular structure of PCR-GLOBWB 2, both in terms of model concepts and implementation (separate modules are called from a main program), makes it easy to modify or replace components according to specific objectives of the model application, to introduce new modules or components within the modelling system and to couple it to existing codes.

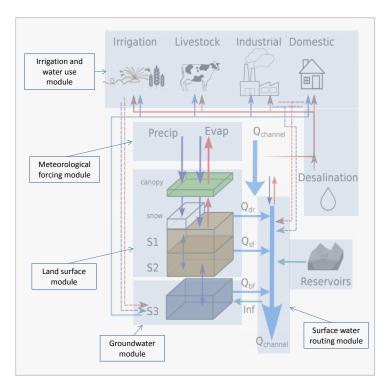


Figure 2. The five modules that make up PCR-GLOBWB 2 portrayed on the model components of Figure 1.

There are currently five main hydrological modules in PCR-GLOBWB 2 as illustrated in Figure 2 and briefly described in Section 2.3: Meteorological forcing; Land surface; Groundwater; Surface water routing; Irrigation and water use. For an extensive description of the underlying equations and methods used in each of these modules we refer to the following sources:

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196	Meteorological forcing module: van Beek (2008) http://vanbeek.geo.uu.nl/suppinfo/vanbeek2008.pdf
197	• Land surface module, groundwater module and surface water routing module: van Beek and Bierkens
198	(2009) http://vanbeek.geo.uu.nl/suppinfo/vanbeekbierkens2009.pdf; van Beek et al. (2011)
199	http://dx.doi.org/10.1029/2010WR009791
200	• Irrigation and water use module:
201	o Calculation of water demand: Wada et al., (2014) https://doi.org/10.5194/esd-5-15-2014
202	 Calculation of water withdrawal, consumption and return flows: de Graaf et al. (2014)
203	https://doi.org/10.1016/j.advwatres.2013.12.002; Wada et al. (2014) https://doi.org/10.5194/esd-
204	5-15-2014; Erkens and Sutanudjaja (2015) https://doi.org/10.5194/piahs-372-83-2015
205	
206	Furthermore: for details about coupling to MOFLOW we refer to:
207 208	One-way coupling: Sutanudjaja et al. (2011) https://doi.org/10.5194/hess-15-2913-2011 ; De Graaf et
209	al. (2017) https://doi.org/10.1016/j.advwatres.2017.01.011
210	• Two-way coupling: Sutanudjaja et al. (2014) http://dx.doi.org/10.1002/2013WR013807
211	1 no may coup.mg, building to the (2011) intermediate 1011002 2010 120100
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213	2.3 Description of the modules
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215	Hereafter, we briefly describe the main features of the five modules. Additionally, a (non-exhaustive) list of
216	the model state and flux variables is provided in Table A1, whereas Table A2 lists the model inputs and
217	parameters, including their sources.
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220	2.3.1 Meteorological forcing module
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222	Meteorological forcing of PCR-GLOBWB 2 uses time series of spatial fields of precipitation, temperature and
223	reference evaporation. Reference potential evaporation can be prescribed or calculated within the model, and is
224	used in the land surface module to calculate crop-specific potential evaporation based on crop factors of the
225	various land cover types according to the FAO guidelines (Allen et al., 1998). There are two options for
226	calculating reference potential evaporation: 1) using Hamon (1963) in case only daily mean temperature is
227	available; 2) using Penman-Monteith following the FAO guidelines (Allen et al., 1998) if net radiation, wind
228	speed and vapour pressure deficit are additionally available. See van Beek et al. (2008) for details. The
229	resulting crop specific potential evaporation is subsequently used to compute the actual evaporation for
230	different land cover types in each cell. Apart from the calculation of evaporation, temperature is also used to

partition precipitation into snow and rain and to drive snowmelt.

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2.3.2. Land surface module

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This core module of PCR-GLOBWB 2 covers the land-atmosphere exchange, the vertical flow between soil compartments and the eventual groundwater recharge, snow and interception storage and the runoff generation mechanisms. Information is organized per land cover type and the fraction it occupies within a cell. The number of land cover types is configurable; the standard parameterization of PCR-GLOBWB 2 carries: tall natural vegetation, short natural vegetation, irrigated crops (non-paddy) and paddy-irrigation (i.e., wet rice). For each land cover type, separate soil conditions can be specified. It should be noted that the soil and vegetation conditions are in any case fully spatially distributed. Thus, vegetation properties (e.g., crop factor, Leaf Area Index) and soil properties (depth, saturated hydraulic conductivity, etc.) vary not only between land cover types, but may also vary from cell-to-cell (e.g., per climate zone). In the standard parameterization vegetation properties vary over the year using a monthly climatology of phenology and crop calendars (i.e. for the crop factor and LAI). The application of irrigation water for paddy and non-paddy irrigation is done by the irrigation and water use module. It is based on the FAO guidelines of Allen et al. (1998) and is dependent on the actual soil water storage (S_1, S_2) or paddy-open water storages. All fluxes, from and to the land surface module in Figure 2, are thus calculated separately per land cover type. The resulting vertical fluxes for each land cover type are: interception evaporation, bare soil evaporation, snow sublimation, vegetation-specific transpiration. In the soil column, vertical fluxes are based on Darcian flow and interact with the underlying groundwater store, S_3 . Surface runoff (Q_{dr} , from precipitation and snowmelt) consists of infiltration excess runoff and saturation excess runoff following a sub-grid approach that mimics variable source areas, i.e. the improved Arno Scheme (Todini, 1996; Hagemann and Gates, 2003). Interflow or stormflow (Qsf), mostly occurring in regolith soils on hillslopes, is also handled with a sub-grid approach based on a runoff parameterization by Sloan and Moore (1984). All fluxes are computed per land cover type and balanced with the available storage to arrive at the net flux that is used to update the storages for the next time step. Also, to report the overall fluxes per cell, and to pass these to other modules, the land cover specific fluxes are subsequently averaged (weighted by land cover type fractions).

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For the standard parameterization of the land surface module the following data sets are combined (see Table A2): the cell fractions of various non-irrigation land cover types are based on the map of Global Land Cover Characteristics Data (GLCC) Base Version 2.0 (Loveland et al., 2000) with the land cover classification following Olson (1994a; b) and the parameter sets from Hagemann et al. (1999) and Hagemann (2002). Irrigation land cover types (i.e. paddy and non-paddy), including their crop calendars and growing season lengths, are parameterized based on the data set of MIRCA2000 (Portmann et al., 2010) and the Global Crop Water Model of Siebert and Döll (2010). We refer to van Beek et al. (2011) for detailed descriptions.

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2.3.3. Groundwater module

The groundwater module calculates groundwater storage dynamics subject to recharge and capillary rise (calculated by the land surface module), groundwater discharge ($Q_{\rm bf}$; in case of a positive groundwater storage) and riverbed infiltration (Inf). Groundwater discharge (assumed the same as groundwater baseflow here) depends on a linear storage-outflow relationship ($Q_{\rm bf} = S_3/J$) where the proportionality constant J is calculated following drainage theory of Kraijenhoff-van de Leur (1958) based on drainage network density and aquifer properties. Riverbed infiltration occurs only in case $Q_{\rm bf}$ becomes 0 by groundwater withdrawal, and only in areas where under natural conditions (without groundwater withdrawal) significant groundwater discharge occurs. Under persistent groundwater withdrawal (calculated with the Irrigation and Water use module) that is larger than the sum of recharge and riverbed infiltration, the groundwater storage S_3 is allowed to become negative. In this case, the part of the withdrawn groundwater in excess of the input (recharge and riverbed infiltration) is seen as non-renewable groundwater withdrawal leading to groundwater depletion (permanent loss of groundwater from storage). In case withdrawal becomes smaller than the input, the remaining input is used to first fill the negative storage to zero, before baseflow $Q_{\rm bf}$ commences again. Alternatively, an initial estimate of a fossil, i.e. a non-actively replenished, groundwater store can be imposed that provides a similar functionality.

It is possible to use a full-fledged groundwater flow model based on MODFLOW: Harbaugh et al., 2000) coupled to PCR-GLOBWB 2 in order to calculate groundwater heads and flow paths. This can be done as a one-way coupling where PCR-GLOWB 2 is first run with the standard groundwater module (reservoir S_3 with only vertical fluxes) to yield time series of net groundwater recharge (recharge – capillary rise) and surface water levels. These fluxes/inputs are subsequently used to force the groundwater flow model (see e.g. Sutanudjaja et al., 2011; de Graaf et al., 2017). Another possibility is to use a two-way coupling where the groundwater module of PCR-GLOBWB 2 is replaced by the groundwater flow model. In this case, at each time step fluxes are exchanged between the groundwater model and the land surface module, and the groundwater model and the surface water routing module (Sutanudjaja et al. 2014).

2.3.4 Surface water routing module

Following an 8-point steepest gradient algorithm across the terrain surface (local drainage direction or LDD), all cells of the modelled domain are connected to a strictly convergent drainage network that together make up the river basins and sub-basins of the model domain. The lowermost cell is either connected to the ocean or to an endorheic basin. Per cell, the sum of the three daily runoff fluxes (Figure 1) is aggregated and routed along the drainage network. Routing can be done in three ways of increasing complexity: 1) accumulation of the fluxes over the drainage network, typically aggregated over longer time steps (e.g. month or year) that are

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larger than the travel times of water along the longest river length; 2) a travel-time characteristic solution (Karssenberg et al., 2007). Here, for each cell flow velocity is calculated in advance based on bankfull discharge and Manning's equation (assuming the energy slope to be equal to the bed slope). Next, this velocity is used to move the volume of water in the channel of a cell the corresponding distance within one time step along the drainage network. This method works reasonably well for relatively steep rivers in humid climates where the friction slope is close to the bed slope and the rivers are equally filled with water throughout the year; 3) the kinematic wave approximation of the Saint Venant equations with flow described by Manning's equation. Also, here, it is assumed that friction slope and bed slope are equal, which makes it valid for rivers without backwater effects. The kinematic wave is solved using a time-explicit variable subtime stepping scheme based on the minimum Courant number. Of these methods, the kinematic wave solution simulates the propagation of the flood wave more realistically while the others provide an expedient means to approximate discharge over longer periods. Using the kinematic wave method, it is possible to model floodplain inundation which occurs if the discharge exceeds the bankfull capacity of a channel. The excess discharge volume is spread over the entire cell from the lowest part of the cell (based on a higher resolution sub-grid DEM) yielding a flooded area with an approximated flood depth. In case of flooding, the simulated river flow is impacted by adjusting the wetted area and wetted perimeter and calculating a weighted Manning coefficient from the individual Manning coefficients of the floodplains and the channel. Lakes and reservoirs are part of the drainage network. Lakes and reservoirs can extend over multiple cells, in which case the storage is subdivided by area such as to ensure that lake and reservoir levels are the same across their extent. The active storage of lakes and the actual storage of reservoirs are dynamically updated; for the lake outflow a standard storage-outflow relationship is used based on a rectangular cross-section over a broad-crested weir, while reservoirs follow a release strategy. This strategy is, by default, aimed at passing the average discharge, while maintaining levels between a minimum and maximum storage, but more elaborate strategies that take account of downstream water demand are possible (e.g. Van Beek et al., 2011). Lakes and reservoir areas change based on global volume-area relationships. All surface water areas, i.e. the river channel, inundated floodplains, lakes and reservoirs, are subject to open water evaporation calculated from reference potential evaporation multiplied with a factor depending on water type and water depth. Moreover, surface waters are subject to surface water withdrawal calculated with the Irrigation and Water Use module. If the kinematic wave approach is used, it can be also augmented with an energy routing scheme to simulate surface water temperature (Van Beek et al., 2012). Finally, it should be noted that it is possible to run the routing routine from PCR-GLOBWB 2 as a stand-alone routine, which allows it to be fed with the specific discharge from other land surface models.

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343 The routing methods that are available in PCR-GLOBWB 2 will yield significant errors for wide lowland 344 rivers where backwater effects are important. In this case, it is possible to replace the surface water module for part of the modelling domain with hydrodynamic models solving the shallow water equations (Hoch et al., 345 346 2017a). Hoch et al. (2017b) developed a generic coupler for this purpose that enables coupling to multiple hydrodynamic modelling codes (https://doi.org/10.5281/zenodo.597107). 347 348 349 Although any data set can be used to define the drainage network and locate the lakes and reservoirs, the 350 standard parameterization of PCR-GLOBWB 2 that runs globally uses the drainage network derived from HydroSHEDS (Lehner et al., 2008) combined with GTOPO30 (Gesch et al., 1999) and Hydro1k (Verdin and 351 352 Greenlee, 1996; USGS EROS Data Center, 2006); lakes taken from GLWD1 (Lehner and Döll, 2004) and 353 reservoirs obtained from GranD (Lehner et al., 2011). 354 355 356 2.3.5 Irrigation and water use module 357 358 In PCR-GLOWB 1 water demand was calculated separately from the hydrology and water availability calculated as a post-processing step by subtracting upstream demand (Wada et al., 2011a,b). In PCR-359 360 GLOBWB 2 water use (withdrawal and consumption) is fully integrated. Hereafter, the main features of the 361 irrigation and water use module are described in the following order: water demand, water withdrawal, water 362 consumption and return flows. 363 364 Water demand 365 Irrigation water demand is calculated based on the crop composition (which changes per month and includes 366 multi-cropping) and the irrigated area per cell. As stated above, these are obtained from MIRCA2000 (Portmann et al., 2010) and the Global Crop Water Model (Siebert and Döll, 2010). In the standard PCR-367 GLOBWB 2 parameterization the irrigated areas change over time. In want of detailed data, fractions of paddy 368 and non-paddy irrigation, as well as the crop composition per month stay fixed (as obtained from 369 370 MIRCA2000), while the total irrigated area per cell changes over time and is based on the FAOSTAT (FAO, 371 2012) reported irrigated areas. Irrigation water demand is computed using the FAO guidelines (Doorenbos and 372 Pruit, 1977; Allen et al., 1998): in case of non-paddy irrigation, water is applied whenever soil moisture falls 373 below a pre-set value and then the soil column is replenished up to field capacity. In case of paddy irrigation, 374 the water level is kept at a water depth of 5 cm above the surface until the late crop development stage (~ 20 375 days) before the harvest. After that, no irrigation is applied anymore such that the water level is allowed to 376 drop to zero under infiltration and evaporation (Wada et al., 2014). The net irrigation demand is augmented to account for limited irrigation efficiency and losses. In the standard parameterization of PCR-GLOBWB the 377 378 irrigation water demand is increased by 40% to obtain gross irrigation water demand (meaning an irrigation 379 efficiency of $(1/1.4) \times 100 = 71\%$.) However, it is possible to use spatio-temporal varying irrigation 380 efficiencies if needed, which is the case for all other variables.

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Non-irrigation water demand covers three sectors; industry, households and livestock. For each of these sectors, the gross demand and net demand are prescribed to the model and calculated using separate scripts. The calculation of net non-irrigation water demand, which varies with time, follows methods developed by Wada et al (2014). We refer to Wada et al. (2014) for an extensive description. Trends in water demand are prescribed on an annual basis as a function of population, electricity demand and gross domestic product (GDP) per capita. In addition, domestic water demand exhibits a seasonal variation on the basis of temperature. Domestic and industrial gross water demand is calculated from net water demand using a country-specific recycling ratio RC (based on development stage or GDP per capita and additionally access to domestic water demand): gross = net/(1-RC). This takes into account that much of the domestic and industrial water is not consumed but returned as surface water. For livestock, the return flow is assumed to be zero, meaning all water is consumed.

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Water withdrawal

The water withdrawal estimation is based on the work by de Graaf et al. (2014) and Wada et al. (2014). In PCR-GLOBWB 2 water withdrawal is set equal to gross water demand (summed over all the sectors) unless sufficient water is not available. In that case, water withdrawal is scaled down to the available water and then allocated proportionally to gross water demand per sector. Thus, no allocation preference is available in the standard parameterization of PCR-GLOBWB 2, but it would be rather straightforward to change this.

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Water can be abstracted from three sources: surface water, groundwater (fossil and non-fossil) and desalinated water. The latter is prescribed (Wada et al., 2011a), while the fractions of the other two sources are determined as function of their relative abundance. Groundwater and surface water availability are determined based on two-year running means of groundwater recharge and river discharge respectively, thus keeping track of the prevalence of local resources and their temporal change (de Graaf et al., 2014). These fractions determine on a monthly basis from which source water is abstracted. Surface water withdrawal is ceased if river discharge falls below 10% of the long-term average yearly discharge under naturalized flow conditions (determined by running the model without withdrawal). If, for some reason, the surface water amount is insufficient, the model falls back on groundwater to meet the resulting gap. Groundwater is first abstracted from the renewable groundwater storage, and if not this is not present, non-renewable groundwater is abstracted. The amount of groundwater that can be abstracted is, however, capped by the groundwater pumping capacity which is based on data by IGRAC GGIS database. The described dynamic allocation scheme is not always in line with local preferences or the infrastructure. However, there is a possibility to use literature fractions of groundwater withdrawal and surface water withdrawal. For urban areas, we rely on the data set of McDonald et al. (2014) that states whether a surface water distribution infrastructure is available. If this is the case, industrial and domestic water withdrawals are mainly taken from surface water before abstracting groundwater. If surface water infrastructure is limited, groundwater source is prioritized (see e.g. Erkens and Sutanudjaja, 2015). For

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118	urban areas that are not in the McDonald (2014) data set, we give preference to the dynamic allocation
119	scheme. For irrigation, we use the ratios supplied by Siebert et al. (2010) in regions where they are said to be
120	reliable. In regions where they are not fully reliable, we take the average ratio provided by Siebert et al. (2010)
121	and the one provided by the dynamic allocation scheme. For regions where the data of Siebert (2010) are not
122	reliable (i.e., extrapolated data), we give preference to the dynamic allocation scheme.
123	
124	Moreover, we cannot assume that all the water demand is supplied from surface water and groundwater
125	resources in the same cell. Ideally, the local water redistribution network should be used to define a surface
126	water service area. Unfortunately, this information is not available at the global scale. Therefore, in our current
127	parameterization of PCR-GLOBWB 2, we pool water availability of desalinated and surface water over zones
128	of approximately 1 arc-degree around each 5 arc-minute cell that are truncated by country and basin borders if
129	applicable. For groundwater, 0.5 arc-degree zones are used. The downside of the current scheme is that a cell
130	does not always have access to its nearest water resource if this lies outside its prescribed service area.
131	Available surface water for abstraction is stored in channels, lakes and reservoirs within each cell and service
132	area. Groundwater availability is also limited by the pumping capacity in the service area.
133	
134	Water consumption and return flows
135	In case of irrigation, all the withdrawn water is applied to the soil (non-paddy) or the water level on the field
136	(paddy). Part of that water is lost by transpiration and part by soil and open water evaporation. Transpiration
137	and evaporation together make up the irrigation water consumption. The remaining part of irrigated water is
138	lost by percolation and contributes to groundwater recharge as return flow. Irrigation efficiency (not including
139	conveyance losses) could also be calculated after the fact by the difference between withdrawal and
140	transpiration. In case of domestic and industrial water use, water consumption depends on the recycling ratio
141	RC and equals withdrawal×(1-RC), while withdrawal×RC constitutes return flow. All return flow is added to
142	the surface water. For livestock, the consumption is set equal to the withdrawal and no return flow is assumed.
143	
144	
145	2.4 Differences between PCR-GLOBWB 1 and 2
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147	PCR-GLOBWB 2 has the following new capabilities compared to PCR-GLOBWB 1 (cf. Van Beek et al.,
148	2011; Wada et al, 2011):
149	• the model was completely rewritten in PCRaster Python and now has a modular structure;
150	• the inputs and outputs are in the form of NetCDF files and output can be reported for daily monthly and
1 51	yearly time steps;

parameterizations are available at 30 arc-minute and 5 arc-minute resolution;

water use (demand, withdrawal, consumption and return flow) is fully integrated;

distinction is made between paddy and non-paddy irrigation and irrigation follows FAO guidelines;

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455 three different options for surface water routing are available and a surface water temperature module is 456 fully integrated with the routing scheme; 457 it is possible to run surface water routines separately with specific discharge from other sources (e.g. other 458 land surface models); 459 PCR-GLOBWB 2 can be coupled to a two-layer transient groundwater model (Sutanudjaja et al., 2014; De Graaf et al., 2017) and to the hydrodynamic models Delft3D Flexible Mesh (Kernkamp et al., 2011) or 460 LISFLOOD-FP (Bates et al., 2010, see Hoch et al., 2017b). 461 462 463 464 2.5 Model code 465 The original PCR-GLOBWB version 1 (van Beek et al., 2011) was written in the PCRaster scripting language. 466 467 PCRaster (Wesseling et al., 1996) is a high-level programming language that started as a dynamic raster-based 468 Geographical Information System (GIS) and is tailored to spatiotemporal modelling for environmental and earth science applications. The generic nature of PCRaster with its many tailor-made built-in hydrological 469 470 functions and its syntax that reads like pseudo-code, generally results in short and readable model codes, short 471 development times and limited programming errors. Karssenberg et al. (2010) developed a PCRaster Python package such that PCRaster functions, implemented in C++, can also be called via Python 472 473 (http://www.python.org/). Using PCRaster Python also makes it possible for students and beginner modellers 474 to contribute to the model quickly, while it allows experts to be more productive and focus on the science 475 rather than on the programming language syntax. Realising the aforementioned advantages, PCR-GLOBWB, particularly starting from this version 2, has been rewritten in the Python scripting language. 476 477 478 To run the model a so-called initialization file or configuration file is used (with extension .ini). In this file the following aspects are defined: the spatial and temporal domain, the time step, the settings of the different 479 480 modules (e.g. which surface water routing, human water use or not etc.) and the locations and names of the 481 parameter files and forcing files. As mentioned above, PCR-GLOBWB 2 uses NetCDF files for most input 482 and all output, thus making it easier to exchange data with other scientists and use existing tools to analyse its 483 output. 484 485 PCR-GLOBWB 2 generally runs under Linux. It is also possible to run it under Windows, but Windows 486 memory constraints limit domain size and time steps simulated. In order to run PCR-GLOBWB the following 487 additional software needs to be installed: PCRaster version 4, Python versions 2.7 with Python packages 488 numPy and netCDF4 and gdal version 1.8 or higher. 489

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490 491	3. Model demonstration and evaluation
492	To test and evaluate the performance of PCR-GLOBWB 2, we ran the model at both 30 arc-minute and 5 arc-
493	minute resolution over the period 1958-2015. We compared the results of both simulations with discharge data
494	from the GRDC Global Runoff Data Centre (GRDC, 2014), with total basin water storage estimates
495	from GRACE (Gravity Recovery and Climate Experiment; Wiese, 2015) and with water withdrawal
496	data from the FAO AQUASTAT database (FAO, 2016).
497	
498	
499	3.1 Model run setup
500	
501	3.1.1 Parameterization
502	
503	We used the standard parameterization (parameters, forcing and their sources in Table A2) of PCR-
504	GLOBWB 2 at 30 arc-minute and 5 arc-minute spatial resolutions to simulate global hydrology at daily
505	resolution over 1958-2015. Outputs were reported as monthly averages. Note that parameterizations
506	were derived directly following their source data sets using hydrological concepts described in Van
507	Beek and Bierkens (2009). We stress that no calibration was performed. We ran the model with human
508	water use options on and used the travel-time characteristic solution routing option.
509	
510	3.1.2 Forcing
511	
512	The forcing data set is based on time series of monthly precipitation, temperature and reference
513	evaporation from the CRU TS 3.2 data set of Harris et al. (2014) downscaled to daily values with
514	ERA40 (1958-1978; Uppala et al., 2005) and ERA-Interim (1979-2015; Dee et al., 2011). CRU is
515	specified at 30 arc-minute spatial resolution and directly usable. We used ERA40 and ERA-I results
516	that had been resampled by ECMWFs resampling scheme from their original resolutions (~1.2° and
517	~0.7°) to 30 arc-minutes first. Precipitation was temporally downscaled by first applying a threshold of
518	0.1 mm/day to the ERA daily time series to estimate the number of rain days for ERA. The amount of
519	rainfall below this threshold was proportionally allocated to the rain days. Next, the daily rainfall totals
520	were scaled in order to reproduce the CRU monthly precipitation total using multiplicative scaling.
521	Equally, monthly reference potential evaporation, computed with Penman-Monteith from the CRU data
522	set was scaled using multiplicative scaling and downscaled to daily data proportional to Hamon (1967)
523	evaporation calculated from daily ERA temperatures. For the air temperature, an additive scaling factor

was used. To better simulate snow-dynamics for the 5-arc-minute model, the temperature values from

CRU were further spatially downscaled to 5 arc-minutes using a temperature lapse-rate derived from

the higher-resolution CRU V1.0 climatology (New et al., 2002). For areas where the number of stations

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527 underlying the CRU data set was found to be small, preference was given to using directly the 528 meteorological data from ERA. The method used to create the forcing data set is described more 529 extensively in Van Beek (2008). 530 531 3.1.3 Spin-up 532 533 The large groundwater response times for certain regions (e.g. Niger and Amazon) requires substantial 534 spin-up for the groundwater volumes to be in equilibrium with the current climate. To reach this equilibrium, the model was spun-up using the average climatological forcing over the years 1958-2000 535 536 back-to-back for 150 years to reach a dynamic steady state. This spin-up was executed under 537 naturalized condition which means no reservoirs and no human water use. 538 539 3.1.4 Computation time and parallelization 540 541 The models were run on Cartesius, the Dutch national supercomputer (https://userinfo.surfsara.nl/systems/cartesius). Without parallelization, the wall clock time for a one-542 543 year global simulation run of the 30 arc-minute model was about one hour. This entails that a one-year 544 global simulation run with the 5 arc-minute model, might result in wall clock times at least 36 hours. 545 Hence, to speed-up computation, the 5 arc-minute model domain was divided into 53 groups of river 546 basins such that it could be run as 53 separate processes. With this simple parallelization technique, the 547 wall clock time for a one-year simulation run of the 5 arc-minute model reduced to about one hour 548 again. Note that these computation times were obtained for simulations with the travel-time characteristic 549 routing option. Calculation times would have been significantly longer if the kinematic wave routing had been 550 used (e.g. about 6 hours for a one-year 5 arc-minutes global run including parallelization). 551 552 553 554 3.2 Data used for comparison 555 556 3.2.1 River discharge 557 558 We used discharge stations from GRDC (2014) to compare simulated discharge from PCR-GLOBWB 559 2 with monthly reported discharge. From all the globally available stations in the database, we selected 560 a subset of stations using the following criteria: 1) allowing a not more than 15% difference in catchment area between PCR-GLOBWB 2 and the area reported with the GRDC discharge station; 2) 561 562 not more than 1 cell distance between the station location and the nearby location of a main river in PCR-GLOBWB 2; 3) at least 1 year of discharge data. This yielded 5363 stations for the 5 arc-minute 563

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564 simulation, 3910 stations for the 30 arc-minute simulation and 3597 stations fulfilling the criteria for both resolutions. As we jointly compared the performance of both simulations, we used the set of 3597 565 locations throughout. The average time series length of these stations is equal to 36 years. 566 567 568 3.2.2 Total water storage 569 570 We compared total water storage (TWS) as simulated by PCR-GLOBWB 2 with the TWS estimated 571 from GRACE (Gravity Recovery and Climate Experiment) gravity anomalies. We used the GRACE 572 JPL Mascon product PL-RL05M (Wiese, 2015; Watkins et al., 2015; Wiese et al., 2016). Scanlon et al. 573 (2016) suggest that recent developments in mascon (mass concentration) solutions for GRACE have 574 significantly increased the spatial localization and amplitude of recovered terrestrial TWS signals. They 575 also claim that one of the advantages of using the mascon solutions relative to traditional SH (spherical 576 harmonic) solutions is that it makes it much easier for non-geodesists to apply GRACE data to 577 hydrologic problems. Note that although the data of PL-RL05M are represented on a 30 arc-minutes 578 lat-lon grid, they represent the 3x3 arc-degree equal-area zones, which is the actual resolution of JPL-579 RL05M. We compared trends on a pixel-by-pixel basis. Given the coarse resolution of GRACE 580 products of about 300 km by 300 km we compared cross-correlations only for major river basins with an area of 900,000 km² and up. 581 582 583 3.2.3 Water withdrawal 584 585 The water withdrawal for a large number of countries is taken from FAO's AQUASTAT database 586 (FAO, 2016) This data is on average reported in every 5 years. We compared simulated water 587 withdrawal per sector and per water source (surface water and groundwater) with reported values per 588 country and per reporting period, whenever available. 589 590 591 592 593 3.3 The global water balance simulated at 30 and 5 arc-minutes 594 595 We calculated the main global water balance components from the 30 arc-minute and 5 arc-minute simulations over the period 2000-2015. The results in Table 1 show that there are some differences 596 597 between the two model runs, but values are in the same order of magnitude. The small difference in precipitation is due to the fact that the area of the land cells is slightly different at the two resolutions. 598 Differences in evaporation and runoff show that the runoff and evaporation parameterization of PCR-599 GLOBWB 2 is not entirely scale-consistent. Differences in evaporation may also be causing the 600

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 differences in irrigation water demand which in turn may explain the differences in water withdrawal. Recently, Samaniego et al. (2017) applied their multiscale parameter regionalization (MPR) technique to PCR-GLOBWB 2 for the Rhine basin, showing that scale-consistent flux-preserving parameterisation is possible. However, a global application of this method to all PCR-GLOWB 2 parameters is not possible yet. Nonetheless, when comparing the results of both model runs with data reported in the literature, it shows that the global water balance components are similar to recent assessments (e.g. by Rodell et al., 2015) and groundwater withdrawal and total withdrawal estimates match those of previous studies (see Table 2).

From Table 1, it can also be seen that there is a negative change in total terrestrial water storage in both model runs. Table 1 shows that this can only be partly explained by groundwater depletion, which is localized to certain regions (see also Sect. 3.4.2). Further analysis shows that this change can also be attributed to the trends in precipitation forcing used, particularly over the tropics.

 Table 1. Global Water balance components and human water withdrawal (in km³/year) over the period 2000-2015 as obtained from the 30 arc-minutes and the 5 arc-minute simulations. The numbers are shown to high significance to show the water balance closure. This does not mean that we pretend to know e.g. global discharge with a km³ accuracy (actual accuracy of the large fluxes is more in the order of 10³ km³)

		30 arc-min	5 arc-min
		(km ³ /year)	(km ³ /year)
Global water	Precipitation	107452	107495
balance	Desalinated water use	3	2
	Runoff	42393	43978
	Evaporation*	65754	63974
	Change in total water storage	-693	-455
Groundwater	Groundwater recharge	27756	25521
budget	Groundwater withdrawal	737	632
	Non-renewable groundwater withdrawal (groundwater depletion)	173	171
	Renewable groundwater withdrawal	564	460
Withdrawal	Agricultural water withdrawal (irrigation + livestock)	2735	2309
by sector	Domestic water withdrawal	380	314
	Industrial water withdrawal	798	707
Withdrawal	Total water withdrawal	3912	3330
by source	Surface water withdrawal	3172	2697
	Desalinated water use	3	2
	Groundwater withdrawal	737	632

^{*} Includes consumptive water use for livestock, domestic and industrial sectors

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Table 2. Groundwater withdrawal (a) and total water withdrawal (b) as compared to other studies

a) Groundwater withdrawal

a) Grodiawater withdrawar		
Source	Year	Value (km ³ /year)
Döll et al. (2012)	1998-2002	571
Döll et al. (2014) (their Table 2).	2003-2009	690-888
Döll et al. (2014) (their Table 6).	2000-2009	665
Pokhrel et al. (2015)	1998-2002	570 (±61)
Wada et al. (2010) (from the IGRAC database)	2000	734 (±87)
This study (5 arc-minutes)	2000-2015	632

b) Total water withdrawal

Source	Year	Value (km ³ /year)
Döll et al. (2012)	1998-2002	4340
Döll et al. (2014) (their Table 2)	2003-2009	3000-3700
FAO (2016)	2010	3583
Oki and Kanae (2006): 3800 km3/year (contemporary)	2006	3800
Vörösmarty et al. (2005)	1995-2000	3560
This study (5 arcminutes)	2000-2015	3330

3.4 Evaluation of the 30 and 5 arc-minute simulations

3.4.1 Discharge

When evaluating the simulated discharge with discharge observations from GRDC we calculated three different measures: 1) the cross-correlation coefficient between simulated and observed GRDC time series, which is a measure of reproducing correct timing of high and low discharge. A correlation coefficient of 1 is perfect timing; 2) the Kling-Gupta efficiency coefficient or KGE (Gupta et al., 2009) which equally measures bias, differences in amplitude and differences in timing. The KGE varies between 1 and minus infinity, where 1 means a perfect fit in terms of bias, amplitude and timing; 3) anomaly cross-correlation, i.e. the cross-correlation between time series after the seasonal signal (climatology) has been removed. This statistic measures the ability of the model to correctly simulate timing of seasonal and the inter-annual anomalies from the yearly climatology. It shows if the model is capable of capturing hydrological extremes and is not only driven by the climatology. An anomaly correlation of 1 indicates perfect characterization of the extremes and values below 0 indicates a lack thereof.

 Figure 3 shows maps of the cross-correlation coefficients for the GRDC stations considered and Figure 4 shows histograms of cross-correlation and KGE values. Both figures show that the validation results of the 5 arc-minute simulation are notably better than those of the 30 arc-minute simulation. For the 30

when moving from a coarser to higher resolution.

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arc-minute model the number of catchments with KGE > 0, 0.3 and 0.6 are equal to 48%, 26% and 7% of the total catchments respectively. For the 5 arc-minute model these values are respectively equal to 63%, 40% and 12% of the total catchments. Note that for both runs the standard parameterization was used. Possible explanations for the better performance of the 5 arc-minute run are: a better delineation of the outline of the basins, particularly the smaller ones, a better characterization of basin relief and the drainage network, more accurate sub-grid parameterization of soil and land cover due to a smaller scale-gap that needs to be overcome, better estimates of the basin storage and better snow dynamics due to the downscaling of temperature to 5 arc-minute resolution. The KGE values are less favourable than the cross-correlation coefficients. This is mostly due to biases in runoff caused by to incorrect meteorological forcing. The maps of cross-correlations (Figure 3) show the best results in Europe and North America where the meteorological forcing is generally more accurate as a result of more data used in the re-analysis products and higher station availability in the CRU data set. Also, monsoon-dominated basins are well simulated due to the strong seasonal nature of both forcing and related discharge. The improvement of the 5 arc-minute simulation over the 30 arc-minute simulation in Europe is mostly seen in the Alps and the Norwegian mountains. This reflects the fact that topography and thus snow dynamics is better represented at higher resolution. The least accurate results are obtained for some of the African rivers, in particular the Niger where the groundwater recession constants are probably over-estimated and inland delta evaporation is under-estimated, and for some rivers in the Rocky Mountains, which may be the result of errors in snow dynamics. Although results are generally better, the spatial distribution of results is similar to those found by Van Beek et al. (2011) for PCR-GLOBWB 1. The histograms of validation results in Figure 4 do not show a strong relationship between catchment size and validation statistics. This suggests that the improvements of model results equally apply to all catchment sizes

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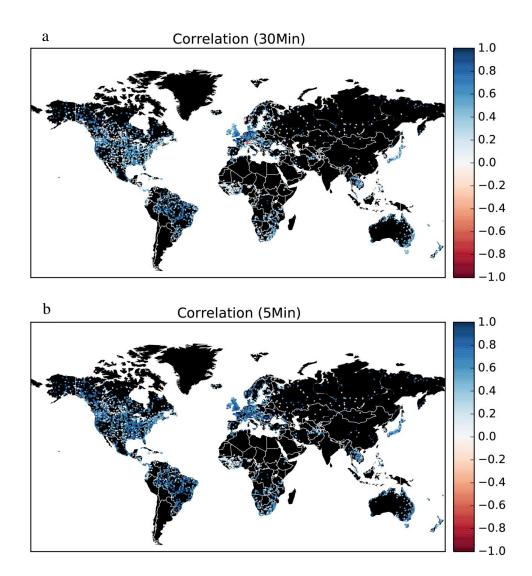


Figure 3 Maps of cross-correlation between simulated and observed discharge time series for 3597 GRDC discharge stations; a. results for the 30 arc-minute simulation; b. results for 5 arc-minute simulation.

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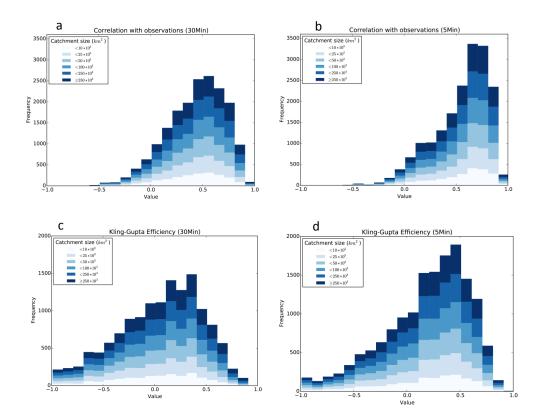


Figure 4. Histograms of validation statistics showing the cross-correlation and Kling-Gupta efficiency (KGE) values for the simulated discharge for the 30 arc-minutes and the 5 arc-minute simulations based on 3597 GRDC discharge stations; a. cross-correlation 30 arc-minute simulation; b. cross-correlation 5 arc-minute simulation; c. KGE 30 arc-minute simulation; d. KGE 5 arc-minute simulation; note: the percentage catchments with KGE < -1 are 21% and 12% for 30 and 5 arc-minutes respectively.

The histograms of the anomaly cross-correlation are shown in Figure 5 The anomaly cross-correlations are generally lower than the cross-correlations, showing that seasonality explains part of the skill in many regions where seasonal variation is dominant when compared to intra-annual or inter-annual variability. Clearly, the 5 arc-minute results are much better than those of the half-degree simulation, indicating a higher skill with regard to capturing extremes and anomalies. Figure 6 shows a map of the difference between the anomaly cross-correlation and the cross-correlation for the 5 arc-minute case. This map shows that there are some regions where the anomaly cross-correlation is better than the cross-correlation (blue colours), e.g. snow-dominated regions in Canada and the Niger basin. These are catchments where the model has difficulty

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reproducing the correct seasonality as a result of errors in snow dynamics (Canada) or groundwater dynamics (Niger). Also, in case of the Niger River, not representing the inner delta flooding and resulting high evaporation may be the cause of poor seasonal timing of discharge.

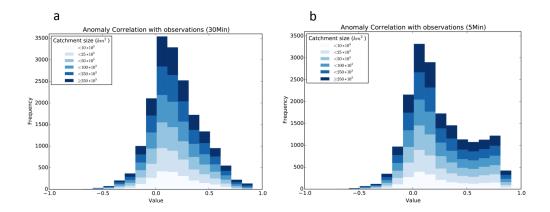


Figure 5. Histograms of validation statistics showing the anomaly cross-correlation for the simulated discharge for the 30 arc-minutes and the 5 arcminute simulations based on 3597 GRDC discharge stations; a. anomaly cross-correlation half arc-degree simulation; b. anomaly cross-correlation 5 arcminute simulation.

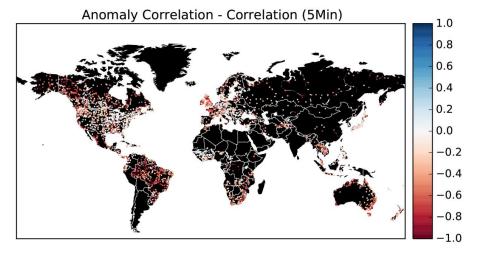


Figure 6. Map showing for the 5 arc-minutes run the difference between the cross-correlation and the anomaly cross-correlation between simulated and observed discharge time series for 3597 GRDC discharge stations; negative values mean that the cross correlation is higher than the anomaly cross-correlation.

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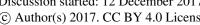
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707 708 709 3.4.2 Total water storage 710 711 Figure 7 compares the trends in 5 arc-minute simulated total water storage (TWS) with those from GRACE, 712 estimated as the average change in m/year over the period 2003-2015. Generally, the PCR-GLOBWB 2 713 simulation is able to capture major groundwater depleted regions as suggested by GRACE, such as those in the 714 Central Valley aquifer, the High Plains aquifer, the North China Plain aquifer, as well as parts of the Middle 715 East, Pakistan and India. For these regions, the absolute rates of TWS change (i.e. TWS declines) of PCR-716 GLOBWB 2 are generally larger, while the spatial pattern in the GRACE map tends to be smoother. This is 717 mainly due to the lower resolution and spatial averaging used in the GRACE product, as well as the fact that 718 the current PCR-GLOBWB 2 simulation does not include lateral groundwater flow between cells. In the polar 719 regions where GRACE estimates mass loss due to melting glaciers and ice sheets, PCR-GLOBWB 2 simulates accumulation as a result of lack of a glacier parameterization. Finally, there are some clear differences over the 720 721 Amazon and some parts of Africa. A possible explanation are errors in meteorological forcing data, which is 722 not very accurate in these parts, but also problems with the over-estimation of PCR-GLOBWB's groundwater 723 response times in these regions which therefore fail to be sufficiently sensitive to recent changes in terrestrial 724 precipitation. 725 726 Further analyses were conducted at the basin-scale resolution, where both TWS time series of PCR-GLOBWB 2 and GRACE JPL-RL05M were averaged over a river basins areas map derived from the 5 arc-minute PCR-727 GLOBWB drainage network. We identified all river basins with sizes larger than 900,000 km², which is 728 729 similar to the GRACE resolution. Smaller river basins were merged to the nearest river basins or grouped 730 together. For the remaining map of large basins, the cross-correlations between PCR-GLOBWB 2 and 731 GRACE basin-average monthly and annual TWS time series were calculated. Monthly cross-correlation 732 provides information about PCR-GLOBWB's ability to correctly time TWS seasonal variability (with a value 733 equal to 1 for perfect timing), while the cross-correlation for annual time series measures inter-annual 734 variability. 735 736 The results in Figure 8 show that PCR-GLOBWB 2 is able to capture GRACE's TWS seasonality for most 737 basins around the world, with the exception of some cold regions in high latitudes (e.g. the Yukon River basin, 738 Iceland). This shortcoming is most likely due to the lack of a proper representation of glacier and ice processes 739 in PCR-GLOBWB 2. As expected, the cross-correlation values for inter-annual time series are generally lower 740 than the ones for monthly time series. There are some areas with negative correlation values, such as the 741 Amazon, Niger and Nile river basins. Apart from the uncertainty in the GRACE signal, these deficiencies may 742 be related to errors in model forcing and structural errors such as errors in the groundwater response time and 743 the effects of wetlands that have not been represented sufficiently well.

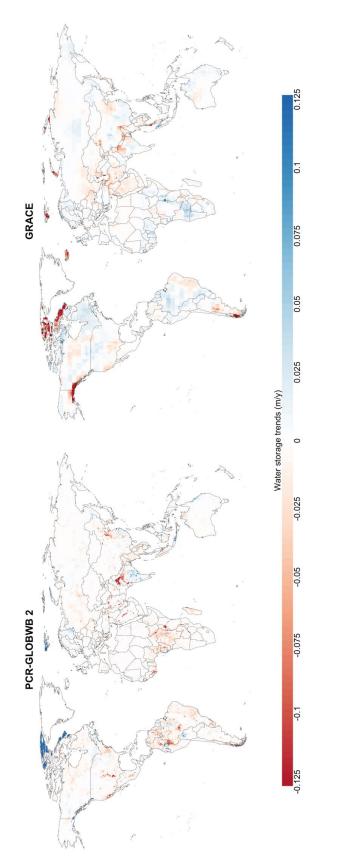
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simulated with PCR-GLOBWB 2 at 5 arc-minute resolution (~10 km at the equator). Negative values indicate declining TWS (e.g. groundwater depleted regions). b. TWS trends obtained based on the GRACE JPL PL-RL05M Mascon product. The GRACE data were resampled to the resolution of 30 arc-minutes, but they Figure 7. Comparison of PCR-GLOBWB 2 total water storage trends (m/year) with those estimated with GRACE over the period 2003-2015. a. TWS trends actually represent the 3 x 3 arc-degree (~300 km x 300 km) area, which is the native resolution of the GRACE signal.

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Annual Correlation PCR-GLOBWB 2.0 and Grace Monthly

2003-2015. b. Comparison of annual TWS series (inter-annual variability). Comparison is only done for the larger basins over 900,000 km², conform the 3x3 arcdegree resolution of GRACE. Figure 8. a. Cross-correlation between monthly TWS time series simulated PCR-GLOBWB 2 and the GRACE JPL PL-RL05M Mascon product over the period

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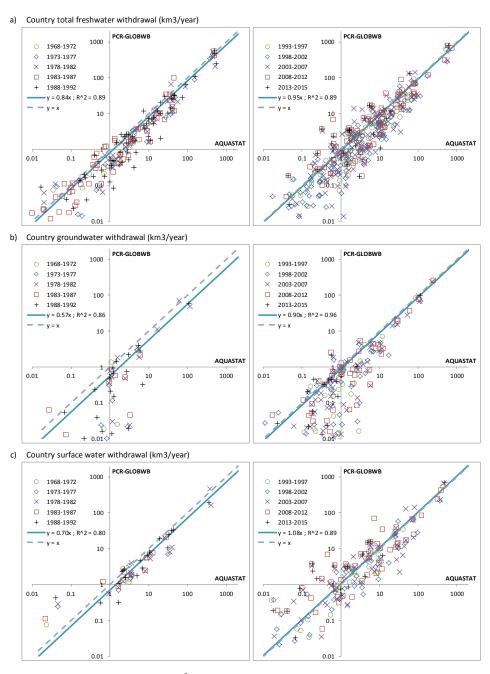


Fig. 9: Country water withdrawal (km³/year) by source; validation of simulations with PCR-GLOBWB 2 with reported values in AQUASTAT (FAO, 2016) for various periods; a) total water withdrawal; b) groundwater withdrawal; c) surface water withdrawal.

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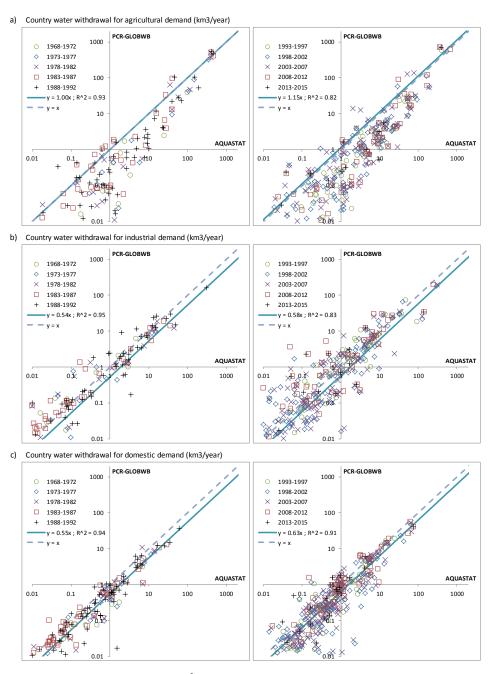


Fig. 10: Country water withdrawal (km³/year) by sector; validation of simulations with PCR-GLOBWB 2 with reported values in AQUASTAT (FAO, 2016) for various periods; a) withdrawal for agricultural demand (irrigation and livestock); b) withdrawal for industrial demand; c) withdrawal for domestic demand.

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3.4.3 Water withdrawal

We compared simulated water withdrawal data from PCR-GLOBWB 2 with reported withdrawal data per country from AQUASTAT (FAO, 2016). The results are shown subdivided per source (Figure 9) and per sector (Figure 10). These figures show that PCR-GLOBWB 2 is able to reproduce reported withdrawal values reasonably well (R² between 0.80 and 0.96 and regression slopes between 0.54 and 1.15). There is some underestimation of groundwater withdrawal for the countries with lower withdrawal values. This may be the result of not sufficiently accounting for domestic groundwater withdrawal in populated areas. Also, Figure 10 shows that agricultural water withdrawal is underestimated for countries with smaller withdrawal. A possible cause of this may be the overestimation of irrigation efficiency.

4. Conclusions and future work

 We presented the most recent version of the open source global hydrology and water resources model PCR-GLOBWB. This version, PCR-GLOBWB 2, has a global coverage at 5 arc-minute resolution. Apart from the higher resolution, the new model has an integrated water use scheme, i.e. every day sector specific water demand is calculated, resulting in groundwater and surface water withdrawal, water consumption and return flows. Dams and reservoirs from the GranD database (Lehner et al., 2011) are added progressively according to their year of construction. PCR-GLOBWB 2 has been rewritten in Python and uses PCRaster-Python functions (Karssenberg et al., 2007). It has a modular structure, which makes the replacement and maintenance of model parts easier. PCR-GLOBWB 2 can be dynamically coupled to a global 2-layer groundwater model (De Graaf et al., 2017) and a one-way coupling to hydrodynamic models for large-scale inundation modelling (Hoch et al., 2017b) is also available.

Comparing the 5 arc-minute with 30 arc-minute simulations using discharge data we clearly find an improvement in the model performance of the higher resolution model. We find a general increase in correlation, anomaly correlation and KGE, indicating that the higher resolution model is better able to capture the seasonality, hydrological extremes and the general discharge characteristics. Also, PCR-GLOBWB 2 is able to reproduce trends and seasonality in total water storage as observed by GRACE for most river basins. It simulates the hotspots of groundwater decline that around in GRACE as well. Simulated water withdrawal, by source and sector, matches reasonably well with reported water withdrawal from AQUASTAT.

Future work will concentrate on developing a full dynamic (two-way) coupling with hydrodynamic models, developing 5 km and 1 km resolution (or higher) parameterizations of PCR-GLOBWB 2 using scale-consistent parameterizations (e.g. using MPR; Samaniego et al., 2017), incorporating a crop growth model and solving the full surface energy balance. Other foreseeable developments are using the model in probabilistic settings and in data-assimilation frameworks.

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796	5. Code and data availability
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798	PCR-GLOBWB 2 is open source and distributed under the terms of the GNU General Public License version
799	3, or any later version, as published by the Free Software Foundation. The model code is provided through a
800	Github repository: https://github.com/UU-Hydro/PCR-GLOBWB_model (Sutanudjaja et al., 2017a,
801	https://doi.org/10.5281/zenodo.595656). This keeps users and developers immediately aware of any new
802	revisions. Also, it allows developers to easily collaborate, as they can download a new version, make changes
803	and suggest and upload the newest revisions. The configuration ini-files for the global 30 arc-minutes and 5-
804	arcminute models and the associated model parameters and input files are provided on
805	https://doi.org/10.5281/zenodo.1045338 (Sutanudjaja et al., 2017b). Development and maintenance of the
806	official version (main branch) of PCR-GLOBWB 2 is conducted at the Department of Physical Geography,
807	Utrecht University. Yet, contributions from external parties are welcome and encouraged. For news on latest
808	developments and papers published based on PCR-GLOBWB 2 we refer to http://www.globalhydrology.nl
809	and for the underlying PCRaster-Python code to http://pcraster.geo.uu.nl .
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812	Acknowledgements
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814	development of PCR-GLOBWB 2. The authors are very grateful to all the contributors (as acknowledged in
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816	for Scientific Research (NWO) for the grant that enabled us to use the national super computer Cartesius with
817	the help of SURFsara Amsterdam.
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Appendix

Table A1 - List (non-exhaustive) of state and flux variables defined in PCR-GLOBWB

Description	Symbol	Unit
Interception storage	S _{int}	m
Snow cover/storage in water equivalent thickness (excluding liquid part S sla)	S swe	m
Liquid/melt water storage in the snow pack	S _{sla}	m
Upper and lower soil storages	S_1 and S_2	m
Surface water storage (lakes, reservoirs, rivers and inundated water)	S _{wat}	m
groundwater storage (renewable part)	S ₃	m
fossil groundwater storage (non-renewable)	Snow	m
total groundwater storage = $S_3 + S_{now}$	Sgwt	m
total water storage thickness = $S_{int} + S_{swe} + S_{siq} + S_1 + S_2 + S_{gwt}$	TWS	m
potential evaporation	E _{pot}	m.day
evaporation flux from the intercepted precipitation	E _{int}	m.day
evaporation from melt water stored in the snow pack	$E_{\rm slq}$	m.day
bare soil evaporation	E soil	m.day
transpiration from the upper and lower soil stores	T_1 and T_2	m.day
total land evaporation = $E_{pot} + E_{int} + E_{slq} + E_{soil} + T_1 + T_2$	E_{land}	m.day
surface water evaporation	E _{wat}	m.da
total evaporation = $E_{land} + E_{wat}$	E tot	m.da
direct runoff	Q_{dr}	m.da
interflow, shallow sub-surface flow	Q_{sf}	m.da
baseflow, groundwater discharge	Q_{bf}	m.da
specific runoff from land	Q_{loc}	m.da
local change in surface water storage	Q_{wat}	m.da
total specific runoff	Q_{tot}	m.da
routed channel (surface water) discharge	Q_{chn}	m ³ .se
net fluxes from the upper to lower soil stores	Q ₁₂	m.da
net groundwater recharge, fluxes from the lower soil to groundwater stores	$RCH = Q_{23}$	m.da
surface water infiltration to groundwater	Inf	m.da
desalinated water withdrawal	$oldsymbol{W}_{sal}$	m.da
surface water withdrawal	$W_{ m wat}$	m.da
renewable groundwater withdrawal	W_3	m.da
non-renewable groundater withdrawal (groundwater depletion)	W_{nrw}	m.da
total groundwater withdrawal = $W_3 + W_{nrw}$	W_{gwt}	m.da
water withdrawal allocated for irrigation purpose	$A_{\rm irr}$	m.da
water withdrawal allocated for livestock demand/sector	$A_{ m liv}$	m.da
water withdrawal allocated for agricultural sector = $A_{irr} + A_{liv}$	A_{agr}	m.da
domestic water withdrawal	A_{dom}	m.da
industrial water withdrawal	A_{ind}	m.da

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823 MIRCA 2000 dataset (Portmann et al., 2010), FAOSTAT (2012) Canadell et al. (1996), Hagemann et al. (1999); Hagemann HydroSHEDS (Lehner et al., 2008); Hydro1k (Verdin and (2002); van Beek (2008); van Beek and Bierkens (2009) Canadell et al. (1996); van Beek and Bierkens (2009) FAO (2007) soil map; van Beek and Bierkens (2009) GLCC v2.0 map (USGS, 1997); Olson (1994a, 1994b); van Beek (2008); van Beek and Bierkens (2009) Greenlee, 1996); GTOPO30 (Gesch et al., 1999) References/sources dimensionless dimensionless dimensionless $m^2.m^{-2}$ $\mathrm{m}^3.\mathrm{m}^{-3}$ m.day⁻¹ Unit Ε Ε Ε Ε Ε Ε Ε $f_{
m wmin}$ and $f_{
m wmax}$ $\theta_{s\text{-}1}$ and $\theta_{s\text{-}2}$ SC_1 and SC_2 $\psi_{s\text{-}1}$ and $\psi_{s\text{-}2}$ Rf upp & Rf low Z_1 and Z_2 $\theta_{r\text{-}1}$ and $\theta_{r\text{-}2}$ K_1 and K_2 β_1 and β_2 DEM_{avg} DEMfpl W_{max} $f_{
m lcov}$ DEM Arno scheme (Todini, 1999; Hagemann and Gates, 2003) exponents Land cover fraction: Land cover areas (including extent of irrigated - Saturated hydraulic conductivities of upper and lower soil stores Ratios of cell-minimum and cell-maximum soil storage to W_{max} - Soil water storage capacity per soil layer: SC = Z / (θ_s - θ_r) Total soil water storage capacities = $SC_{low} + SC_{low}$ Table A2 - List of model inputs and parameters - Exponent in the soil water retention curve defining soil water capacity distribution Upper and lower soil store parameters: - Soil matric suctions at saturation - Residual soil moisture content Soil moisture at saturation Root fractions per soil layer Topographical parameters - Flood plain elevation areas) over cell areas Cell-average DEM - Soil thickness Description

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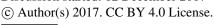
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825 Hagemann et al. (1999); Hagemann (2002); van Beek (2008); van Beek (2008); CRU (Harris et al., 2014); ERA40 (Uppala et GLWD1 (Lehner and Döll, 2004); GranD (Lehner et al., 2011) GLHYMPS map (Gleeson et al., 2014); van Beek (2008); van al., 2005); ERA-Interim (Dee et al., 2011) Wada et al., (2011a); FAO (2016) van Beek and Bierkens (2009) Beek and Bierkens (2009) References/sources Wada et al (2014) dimensionless m.day⁻¹ m.day⁻¹ m.day⁻¹ m.day⁻¹ $m^2.m^{-2}$ $\mathrm{m}^3.\mathrm{m}^{-3}$ °C or K day⁻¹ Unit S int-max Symbol $E_{
m ref,pot}$ ړ $\mathcal{T}_{\mathrm{air}}$ Q Sy - Non-irrigation sectoral water demand (i.e. livestock, dometic and industrial) Table A2 - List of model inputs and parameters (continued) Reference potential evaporation and transpiration Groundwater recession coefficient Parameters related to phenology - Atmospheric air temperature - Vegetation cover fraction **Groundwater parameters** - Aquifer transmissivity - Aquifer specific yield Meteorological forcing - Interception capacity - Lakes and reservoirs - Total precipitation - Desalinated water - Crop coefficient Description Others:

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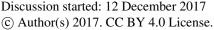
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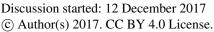
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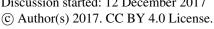
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