

General:

Thank you again for your review. Please find our comments below in red.

* Include commas where appropriate after "i.e." (treat it as "that is")

Changes made throughout

* Always include commas after "e.g." (treat it as "for example")

Changes made throughout

* Consider noting somewhere that form drag is not included (as a general note, this may be parameterized without the bedforms being explicitly modeled)

Note added at Line 205

* Consider making a more specific comment about the grain-size range over which each of these two formulas was developed, and possibly (though only if you find this useful) even including this as horizontal bars or shaded areas in Fig. 2.

Included Lines 374-375

* As a general note, I am not sure how clear it is that "Long-term landscape evolution is disproportionately influenced by successive extreme events." Since this is a point in your response, and not in your text, let's leave this as an item for possible future discussion.

Very happy to have this discussion!

Line-by-line comments (line numbers from the "track changes" version

82 (affects 282-284, 491, and more). You change more than just input parameters, right? You also change the functional form of the sediment transport formula.

Later I see that you refer to the sediment transport formula as a parameter. While you use a binary switch to do this, practitioners of data--model intercomparison typically are quite careful to distinguish changes in parameters from changes in functional form.

I suggest that you make it clear up front that you use this as a "parameter" even though it really is different.

Agreed – there's a wider question here on whether C-L using Einstein is indeed the same model as C-L using Wilcock and Crowe. However, this is somewhat of a philosophical point in relation to the purposes of this study and the choices presented to the user applying the model. For clarity we have included notes at Lines 182 and 370-373

142-143. Herman et al. outside parentheses

173. I intuitively think I know what you mean by "second-generation LEMs" (i.e. newer than SIBERIA), but I am not sure about this. Could you help the reader out?

Have removed 'second-generation' reference.

250-251. Grammar

Changed.

253. "chosen" redundant with "selected"

Changed

274. You usually have an article with "CAESAR-Lisflood"; why not here?

Changed.

311. "which" to "that" or ", which". These have subtly different but occasionally important differences in meaning; I suggest you review this.

Changed

210. lowercase "de"

Changed

352-354. Maybe you want to be implicit that the step length is 2?

Included on line 288

403-406. Start of sentence / transition?

Changed

409. by which parameter 9 was altered (end-of-sentence preposition)

Changed

447. (I should have caught this earlier. Use actual greater-than-or-equal-to symbols. My pickiness also notes that each values should have a range, though in practicality it is clear what you mean.

Changed

690. space after colon

Changed

749-750. An alternate/additional set of reasons for this include that (1) the sources of the nonlinearity (e.g., channel width response) are not adequately included in CAESAR-Lisflood, and (2) even if they are/were, these parameters are not tunable in a simple way because their effects depend on other settings.

Both of these are indeed possible – (though in CL width response is included but your point is still valid) in that we may not be representing NL processes completely and yes – there is a significant degree of dependability on other model settings (as the SA should and does show).

However, in this case the exponential increase in sediment is due to how sediment transport rules calculate how much is entrained/moved. This is very broadly the v^2 to v^3 of the velocity so modest increases in velocity e.g., due to a slightly higher rainfall – will lead to much greater amounts of sediment being moved. That and sediment transport processes being threshold based in many sediment transport rules. We do discuss this in the 2012 paper – but for clarity we've added a sentence to that effect here (563) and tightened up a little of the text in this section to make it clearer.

1 **Global Sensitivity Analysis of Parameter Uncertainty in Landscape Evolution Models**

2 **Christopher J. Skinner¹, Tom J. Coulthard¹, Wolfgang Schwanghart², Marco J. Van De Wiel³, and**

3 **Greg Hancock⁴**

4 ¹School of Environmental Sciences, University of Hull, Hull, UK

5 ²Institute of Earth and Environmental Science, Potsdam University, Potsdam-Golm, Germany

6 ³Centre for Agroecology, Water and Resilience, Coventry University, Coventry, UK

7 ⁴University of Newcastle, Callaghan, Australia

8

9 Corresponding Author: C. J. Skinner (c.skinner@hull.ac.uk)

10

11 **Abstract**

12

13 The evaluation and verification of Landscape Evolution Models (LEMs) has long been limited by a lack
14 of suitable observational data and statistical measures which can fully capture the complexity of
15 landscape changes. This lack of data limits the use of objective function based evaluation prolific in
16 other modelling fields, and restricts the application of sensitivity analyses in the models and
17 consequential the assessment of model uncertainties. To overcome this deficiency, a novel model
18 function approach has been developed, with each model function representing an aspect of model
19 behaviour, which allows for the application of sensitivity analyses. The model function approach is
20 used to assess the relative sensitivity of the CAESAR-Lisflood LEM to a set of model parameters by
21 applying the Morris Method sensitivity analysis for two contrasting catchments. The test revealed that
22 for both catchments the model was most sensitive to the choice of the sediment transport formula,
23 and that each parameter influenced model behaviours differently, with model functions relating to
24 internal geomorphic changes responding in a different way to those relating to the sediment yields
25 from the catchment outlet. The model functions proved useful for providing a way of evaluating the
26 sensitivity of LEMs in the absence of data and methods for an objective function approach.

27

28 **1. Introduction**

29

30 Landscape Evolution Models (LEMs) investigate how the Earth's surface evolves over timescales
31 ranging from hundreds to millions of years (Coulthard and Van De Wiel, 2012; Martin and Church,
32 2004; Pazzaglia, 2003; Tucker and Hancock, 2010; Van De Wiel et al., 2011). They represent the earth's
33 surface with a regular or irregular mesh and simulate how the surface evolves over time as a function
34 of tectonic processes, and erosion and deposition from Earth surface processes. LEMs have proved to
35 be very useful scientific tools to understand how Earth surface processes interact to shape the
36 landscape.

37 More recently, LEMs have improved considerably in their ability to simulate the physical environment,
38 and this has developed in parallel with improvements in computational efficiency and power. This
39 allows LEMs to go beyond highly simplified models of landform development and to also incorporate
40 increasingly complex processes such as pedogenesis (Vanwalleghem et al., 2013; Welivitiya et al.,
41 2016) and periglacial processes (Andersen et al., 2015; Egholm et al., 2015). Other processes are now
42 being handled in more detail such as hydrodynamic flow models and aeolian processes (Adams et al.,
43 2017; Coulthard et al., 2013; Liu and Coulthard, 2017). These developments led to Coulthard et al.
44 (2013) describing them as 'second generation' LEMs that extend previously explanatory and
45 explorative models to be used for prediction of future changes in landscapes, such as for the mining
46 industry (e.g., Hancock et al., 2017; Saynor et al., 2012).

47 However, more detailed physical representations of the processes that shape the Earth's surface
48 involve a larger number of parameters that are typically estimated from proxy data or theoretical
49 considerations, or are completely unknown (Oreskes et al., 1994; Petersen, 2012). If LEMs are to be
50 operationally used for prediction or as decision-making tools in the future, their outputs must be
51 evaluated against the uncertainty in input parameters – a task that is increasingly difficult for a large

52 number of parameters. Through sensitivity analysis (SA) investigates how variations in the output of a
53 numerical model can be attributed to its input factors (Pianosi et al., 2016). This is useful for identifying
54 key parameters for later calibration but this has rarely been conducted for LEMs. The aim of this study
55 is thus to conduct a SA of the widely used and highly parameterized LEM CAESAR-Lisflood (Coulthard
56 et al., 2013) - in particular, we wish to be able to detect the parameters that have the greatest
57 influence on the model's simulation output. As model sensitivity may be influenced by different
58 landscapes, we run the SA in two individual and distinct catchments.

59

60 ***1.1 Sensitivity Analysis and Landscape Evolution Models***

61

62 The application of SA in environmental modelling has a history spanning four decades (Norton, 2008)
63 and forms an important component of using models for decision-making, including model
64 development, calibration and uncertainty analysis (Yang, 2011). SA addresses five key questions
65 (Cariboni et al., 2007; Neumann, 2012; Song et al., 2012, 2015):

66

- 67 1. Which parameters have the greatest influence on the model?
- 68 2. If additional data could be used to reduce the uncertainty in a parameter, which would most
69 reduce the model output variance?
- 70 3. Are there parameters with such low influence that their values could be fixed without impact
71 on the model outputs?
- 72 4. If parameter values emerge as incorrect, how will they influence model outputs?
- 73 5. Which parameters influence model outputs in different regions (parameter space)?

74

75 Clearly, based on the above, an appraisal of model sensitivity is important to fully understand and
76 apply model results. In a review of applications of SA in environmental models, Yang (2011) identified
77 two common approaches to SA – local and global. Local SA are limited, considering only the impacts

78 of factors on model outputs locally, i.e., within a restricted region of the model's parameter space,
79 whilst global SA typically utilise Monte-Carlo methods to assess the sensitivity of impacts across the
80 whole parameter space (Yang, 2011). For complex models with non-linear behaviours, the use of Local
81 SA can be highly biased as they neglect the non-linear interactions between parameters (Oakley and
82 O'Hagan, 2004; Pappenberger et al., 2006; Yang, 2011). Global SA are more computationally
83 expensive, but as the methods are more reliable, they are attractive to modellers (Yang, 2011).

84
85 The use of SA as a routine component of model assessment and calibration is common place in
86 climatic, meteorological, hydrological, hydraulic and many other modelling fields. However, for LEMs
87 there are surprisingly few examples of SA being carried out. This can be explained by three inter-
88 related issues: (i) LEMs typically have a large number of model parameters; (ii) long model run times
89 can make multiple simulations for SA impractical; and (iii) model behaviour can be highly non-linear
90 (e.g., Coulthard and Van De Wiel, 2007; Larsen et al., 2014; Van De Wiel and Coulthard, 2010), leading
91 to potentially complex SA interpretations. Large numbers of model parameters and long run times, in
92 particular, make Monte-Carlo methods extremely time consuming – and therefore often unviable.

93
94 There are several studies on how LEMs respond to variable forcing, process changes and model
95 parameters, including changes in climate variability and precipitation resolution (Armitage et al., 2017;
96 Coulthard and Skinner, 2016a; Ijjasz-Vasquez et al., 1992; Tucker and Bras, 2000), channel widths
97 (Attal et al., 2008), vegetation (Collins, 2004; Istanbuluoglu and Bras, 2005), and variations in initial
98 conditions (Hancock, 2006; Hancock et al., 2016; Ijjasz-Vasquez et al., 1992; Willgoose et al., 2003).

99 Campforts et al. (2017~~5~~), investigated how different numerical solvers affect LEM simulation. Yet few
100 studies explicitly perform SA and most of the applications described above are exploring LEM
101 sensitivity to processes, or changes in environmental conditions, and are more correctly referred to
102 as exploratory tests (Larsen et al., 2014). On the other hand, investigations to ascertain the model's

Field Code Changed

Formatted: English (United Kingdom)

Field Code Changed

Formatted: English (United Kingdom)

Formatted: English (United Kingdom)

103 response to potential uncertainties (e.g from model parameterisation) can be deemed as true SA (eg,
104 Armitage et al., 2017; Coulthard and Skinner, 2016a; Hancock et al., 2016).

105
106 Hydrological models faced similar issues to LEMs in the past, i.e., model complexity and long
107 processing times when applying SA. To overcome them, hydrologists have used the Morris Method
108 (MM; Morris, 1991). The MM can be regarded as a global SA, although it actually performs multiple
109 local SAs sampled from across the full parameter space – this produces a series of local evaluations,
110 the mean of which is an approximation of the global variance (van Griensven et al., 2006; Norton,
111 2009; Saltelli et al., 2000). The main strength of the MM is its computational efficiency. (Herman et al.
112 (2013) showed that the MM could estimate similar variance in model outputs to the Sobol' Variance-
113 based global SA method (Sobol', 2001), yet required 300 times less evaluations, and significant less
114 data storage for an application to a distributed catchment hydrological model. The robustness of this
115 approach has been further shown by numerous workers (e.g., Brockmann and Morgenroth, 2007;
116 Pappenberger et al., 2008; Yang, 2011). However, the MM cannot provide a full quantitative
117 assessment of parameter sensitivity and is dependent upon the user-defined bounds to the parameter
118 space. It can successfully rank parameters between the least and most influential to model outputs,
119 but cannot determine parameters' exact relative influence (Brockmann and Morgenroth, 2007). These
120 advantages and limitations entail that MM has primarily been used during the pre-screening stage of
121 models, isolating the most influential parameters for further SA with quantitative, yet more
122 computationally expensive, methods (e.g., Ratto et al., 2007; Song et al., 2015; Yang, 2011; Ziliani et
123 al., 2013).

124
125 (Ziliani et al., 2013) performed a two-stage SA for the CAESAR LEM, utilising the MM (as adapted by
126 Campolongo et al., 2007). Whilst this study demonstrated the feasibility of applying the MM as a global
127 SA to a reach-scale LEM, it was applied as a pre-screening stage to identify the most relevant

128 parameters for model calibration. In contrast, our study focuses on SA as a tool to investigate
129 parameter influence on model behaviour.

130

131 **1.2 Metrics for Landscape Evolution Model Assessment**

132

133 Evaluating LEMs is challenged by the paucity of comprehensive field data against which they can be
134 assessed and the lack of measures for calibration and validation (Hancock et al., 2016; Hancock and
135 Willgoose, 2001; Tucker and Hancock, 2010). Moreover, some ~~second-generation~~ LEMs (e.g., CAESAR-
136 Lisflood) simulate short (annual to decadal) and long-term (millennial time scales and longer)
137 landscape changes, necessitating data and methods to assess them across variable time scales. Thus,
138 while SA of environmental models often rely on objective functions (e.g., the Nash-Sutcliffe score
139 between observed and simulated values; Nash and Sutcliffe, 1970), this approach is generally not
140 practical for LEMs. With few exceptions (e.g., Ziliani et al., 2013), results from LEMs are therefore
141 frequently assessed qualitatively, relying on visual interpretation of the simulated landforms or cross-
142 section profiles (e.g., Coulthard and Skinner, 2016b; Hancock et al., 2010, 2015; Hancock and
143 Coulthard, 2012).

144

145 Catchment outlet statistics, such as sediment yield time series, allow for comparison between
146 simulations to indicate a catchment's response to perturbations (e.g., Coulthard et al., 2012; Coulthard
147 and Skinner, 2016b; Hancock and Coulthard, 2012). However, sediment yield time series rarely provide
148 a sufficiently complete picture of a catchment's geomorphic response. For example, Coulthard and
149 Skinner (2016b) showed that simulations calibrated to provide equivalent sediment yields produced
150 different landforms. For planning purposes these internal catchment changes are likely to be more
151 useful than catchment sediment yields. Moreover, changing topography potentially instigates a
152 feedback process that leads to complex, often non-linear catchment behaviour (Coulthard and Van De
153 Wiel, 2007, 2013; Hancock et al., 2016; Jerolmack and Paola, 2010; Van De Wiel and Coulthard, 2010).

154 Finally, the spatially and temporally heterogeneous response of erosion and deposition patterns in
155 LEMs also makes “pixel-to-pixel” comparisons difficult. For example, in a valley reach, gross patterns
156 of erosion and deposition may be identical but with the channel on the other side of the valley –
157 yielding a poor pixel-to-pixel comparison.

158

159 Few studies have tested metrics to compare topographic data or physical experiments to simulated
160 elevation changes by LEM (Hancock et al., 2010, 2011; Hancock and Willgoose, 2001; Ibbitt et al.,
161 1999). However, although the metrics often suggested a good agreement, visual analysis of the final
162 DEMs indicated clear differences between the physical models and the simulations (Hancock and
163 Willgoose, 2001). There is, therefore, a clear need for better statistical methods for critically evaluating
164 and comparing landscapes that can also be used for evaluating the accuracy (or otherwise) of LEMs.

165

166 The paucity of observational data and the lack of measures that amalgamate the complexity of spatio-
167 temporal landscape change into a single metric have prevented the objective function approach to be
168 common in modelling landscape evolution. Instead, LEMs can be evaluated by observing the changes
169 in model outputs reflective of model behaviour – these model functions can be used in lieu of objective
170 functions to allow the sensitivity of LEMs to be assessed. Model functions would be best used as a set
171 in combination to allow assessment across a range of model behaviours, and would also be
172 transferable across a range of catchments. Such an approach formalises existing methods of
173 evaluating LEM outputs and provides a framework from which multi-criteria objective function
174 approaches can be applied when suitable observation become available.

175

176 ***1.3 A Global SA for a catchment LEM***

177

178 This study uses MM to assess the sensitivity of CAESAR-Lisflood to a range of user-defined parameters,
179 and therefore demonstrates the first application a global SA to a catchment LEM. This study

180 ~~demonstrates the first application of a global SA illustrate parameter influence on model behaviour in~~
181 ~~a catchment LEM (CAESAR-Lisflood), using the MM to assess the model's sensitivity to user defined~~
182 ~~parameters.~~We selected 15 model parameters (here we consider the choice of sediment transport
183 formula as parameter)~~chosen~~ either because of their known importance to the model or because the
184 model's response to the parameter is presently poorly understood. Although not all the 15 model
185 parameters are universal between LEMs, many LEMs have equivalents. Moreover, we developed a set
186 of 15 model functions -that reflect core behavioural responses of the model. These will indicate
187 whether the same parameters influence all behaviours, or whether the different behaviours respond
188 to different parameters. The choice of 15 model parameters and 15 model functions is coincidental.
189 We conducted the SA in two catchments with contrasting environmental settings to assess how
190 transferable an individual SA is to different conditions.

191

192 It is important to state that this study is an illustration of the potential for using the MM to inform an
193 operator of how model parameter choices can impact the performance and behaviour of their model.
194 It is not an attempt to reproduce or calibrate the CAESAR-Lisflood model to real-world observations,
195 although the model has been applied to each catchment previously.

196

197 **2. Methods**

198

199 We apply the MM to perform a global SA on the CAESAR-Lisflood model for two contrasting
200 catchments (more detail in Section 2.3): the Upper Swale, UK (181 km², temperate, perennial), and
201 Tin Camp Creek, Australia (0.5 km², tropical, ephemeral). Each individual simulation runs for a 30 year
202 period, where the first 10 years are used as a spin-up to reduce the impacts of transient model
203 behaviour and therefore output analysis starts after year 10 of the simulation. The CAESAR-Lisflood
204 model is used in catchment mode, the simulations have no representation of suspended sediments
205 and bed rock, and the dune and soil evolution modules are not used. Form drag is not directly

206 considered within the model but is reflected within the setting of the Manning's n Roughness
207 Coefficient. For each catchment, we assess the 15 user-defined parameters against a set of 15 model
208 functions. Finally, we also assess the changes in elevations across different sections of the catchments.

209

210 For clarity, we here define some terms used frequently throughout this manuscript:

211

- 212 • Parameter – Adjustable value within a model. The value is determined during model
213 set-up and remains constant throughout a given simulation. The value is often based
214 on recorded values or adjusted during calibration.
- 215 • Objective function – an error score between model outputs and observations used to
216 evaluate model performance.
- 217 • Model function – a measure derived from model outputs used to evaluate model
218 behaviour in lieu of an adequate objective function.
- 219 • Elementary effect (EE) – a value used as part of the Morris Method, indicating the
220 change in function value (objective or model) resulting from a change of parameter
221 value during a single repeat.
- 222 • Main effect (ME) – the mean of the elementary effects from all repeats, for a specified
223 parameter and a specified function.

224

225 **2.1 CAESAR-Lisflood**

226

227 The LEM used is the CAESAR-Lisflood model (Coulthard et al., 2013). CAESAR-Lisflood is a second
228 generation LEM, capable of simulations with greater physical realism than first generation models but
229 also with increased complexity – the model features a large number of fixed, physically-based, or user-
230 defined parameters. This additional complexity may result in an increased non-linearity and sensitivity

231 to model parameters. We used CAESAR-Lisflood v1.8, without any additional modifications to the
232 model's functionality from the version freely available online.

233

234 A full description of the CAESAR-Lisflood model can be found in Coulthard et al. (2013), and its core
235 functionality is only summarised here. The model utilises an initial DEM built from a regular grid of
236 cells, and in the catchment mode (as used in this model set up) is driven by a rainfall timeseries, which
237 can be lumped or spatially distributed (Coulthard and Skinner, 2016b). At each timestep the rainfall
238 input is converted to surface runoff using TOPMODEL (Beven and Kirkby, 1979), and distributed across
239 the catchment and routed using the Lisflood-FP component (Bates et al., 2010). The CAESAR
240 component of the model drives the landscape development using sediment transport formulae based
241 on flow depths and velocities derived from the Lisflood-FP component. Bed load is distributed to
242 neighbouring cells proportionally based on relative bed elevations. This study has not used the
243 suspended sediment processes in the model. The model can handle nine different grain sizes, and
244 information is stored in surface and sub-surface layers where only the top surface layer is 'active' for
245 erosion and deposition. A comprehensive description of this process can be found in Van De Wiel et
246 al., 2007).

247

248 CAESAR-Lisflood is freely available and since 1996 there have been over 60 published studies using
249 the model over a wide range of temporal and spatial scales (Skinner and Coulthard, 2017). These
250 previous studies provide useful background into model parameter interactions helping to inform the
251 choice of the user-defined parameters used for the SA as described in Section 2.4. Some studies have
252 also investigated the model's sensitivities to external factors - for example, Coulthard and Skinner
253 (2016) investigated the sensitivity of the CAESAR-Lisflood model to the spatial and temporal resolution
254 of precipitation. Other studies have investigated the influence of individual processes or forcings. For
255 example, Coulthard and Van De Wiel (2017) examined how land-use influences the outputs of the
256 model.

257

258 **2.2 Morris Method**

259

260 Our study used the MM described in Ziliani et al. (2013), i.e. the original MM of Morris (1991), as
261 extended by Campolongo et al. (2007), and applied the “sensitivity” package in the R Statistical
262 Environment (Pujol, 2009) to generate the parameter sets for the SA.

263

264 To set up the MM we selected a number of parameters to be assessed, specifying a minimum and
265 maximum range for each, plus a number of iterative steps. The parameter values are equally spaced
266 based on the range and number of steps – for example, a parameter with a range of 2 to 10 and 5
267 iterative steps would have available values of 2, 4, 6, 8, and 10. This is done for each parameter and,
268 where possible, the same number of iterative steps was used for each.

269

270 The MM samples the global parameter space by performing multiple local SAs referred to as repeats.

271 The first simulation in each repeat is made up of a randomly assigned selection of parameter values

272 from the available values. To set up the second simulation in the repeat a single parameter is randomly

273 selected and its value changed by a random number of iterative steps – if we use the example above,

274 if simulation 1 used the value 4, changing this to 2 or 6 would be one iterative step change (where one

275 step is a change in value of 2), to 8 would be two steps, and using 10 would be three steps. For

276 simulation 3 in the repeat another randomly selected parameter is changed although previously

277 changed parameters are no longer available to be selected. This is continued until no further

278 parameters are available to be changed, therefore in our study each repeat contains 16 tests – 1

279 starting set of parameters, plus 15 parameter changes. In this study we have used 100 repeats, for a

280 total of 1600 individual simulations – for comparison, the implementation of the MM by Ziliani et al.

281 (2013) used 10 repeats.

282

283 The sensitivity of the model to changes in parameter values is evaluated by the changes of objective
 284 function values between sequential tests within repeats relative to the number of incremental steps
 285 the parameter value has been changed by. The change in objective function score between two
 286 sequential tests divided by the number of incremental step changes is an elementary effect (EE) of
 287 that objective function and the parameter changed, as shown by (Equation 1 -). ~~After all 1600 tests~~
 288 ~~have been performed, the main effect (ME) for each objective function and parameter is calculated~~
 289 ~~from the mean of the relevant EEs – the higher the ME the greater the model’s sensitivity. Alongside~~
 290 ~~the ME, the standard deviation of the EEs is also calculated as this provides an indication of the non-~~
 291 ~~linearity within the model.~~

292 **Equation 1**

293
$$d_{ij} = \left| \frac{y(x_1, x_2, \dots, x_{i-1}, x_i + \Delta_i, x_{i+1}, \dots, x_k) - y(x_1, x_2, \dots, x_{i-1}, x_i, x_{i+1}, \dots, x_k)}{\Delta_i} \right|$$

294
 295 ~~where~~ Here d_{ij} is the value of the j^{th} EE ($j = 1, \dots, r$; where r is the number of repetitions (here $r =$
 296 100)) of the i^{th} parameter (e.g. $i = 1$ refers to sediment transport formula, see Table 1), x_i is the value
 297 of the i^{th} parameter, k is the number of parameters investigated (here 15), $y(x_1, x_2, \dots, x_k)$ is the value
 298 of the selected objective function, and Δ_i is the change in incremental steps parameter i was altered
 299 ~~by~~.

301 After all 1600 tests have been performed, the main effect (ME) for each objective function and
 302 parameter is calculated from the mean of the relevant EEs – the higher the ME the greater the model’s
 303 sensitivity. Alongside the ME, the standard deviation of the EEs is also calculated as this provides an
 304 indication of the non-linearity within the model.

306 **2.3 Study Basins**

307

308 **2.3.1. Upper Swale, UK**

309

310 The Swale catchment, UK, is a medium sized basin (181 km²) with 500 m of relief (Figure 1). It has been
311 used extensively in previous CAESAR/CAESAR-Lisflood applications (Coulthard et al., 2012; Coulthard
312 and Macklin, 2001; Coulthard and Skinner, 2016a; Coulthard and Van De Wiel, 2013). For this SA, it
313 represents a medium basin in a temperate climate. All simulations on the Swale are use a 50 m
314 resolution DEM based on airborne LiDAR. Precipitation inputs are 10 years of NIMROD composite
315 RADAR rainfall estimates (Met Office, 2003), applied at a 1 h temporal and 5 km spatial resolution,
316 and repeated three times for a 30 year timeseries.

317

318 **2.3.2. Tin Camp Creek, Australia**

319

320 The Tin Camp Creek catchment is a small sub-catchment (0.5 km²) of the full Tin Camp Creek system
321 (Hancock et al., 2010; Hancock, 2006) (Figure 1). The basin has 45 m of relief and is in the tropical
322 region of the Northern Territory, Australia. In contrast to the Swale, Tin Camp Creek is a small basin
323 and the region has pronounced wet and dry seasons, with short intense rainstorms a feature of wet
324 season precipitation. The DEM is at 10 m grid cell resolution produced from high resolution digital
325 photogrammetry (Hancock, 2012).The rainfall input is taken from observations from a single raingauge
326 at Jabiru Airport, providing a 1 h – lumped (single catchment-average) resolution timeseries for 23
327 years, with the first 7 years repeated to produce a continuous 30 year timeseries..

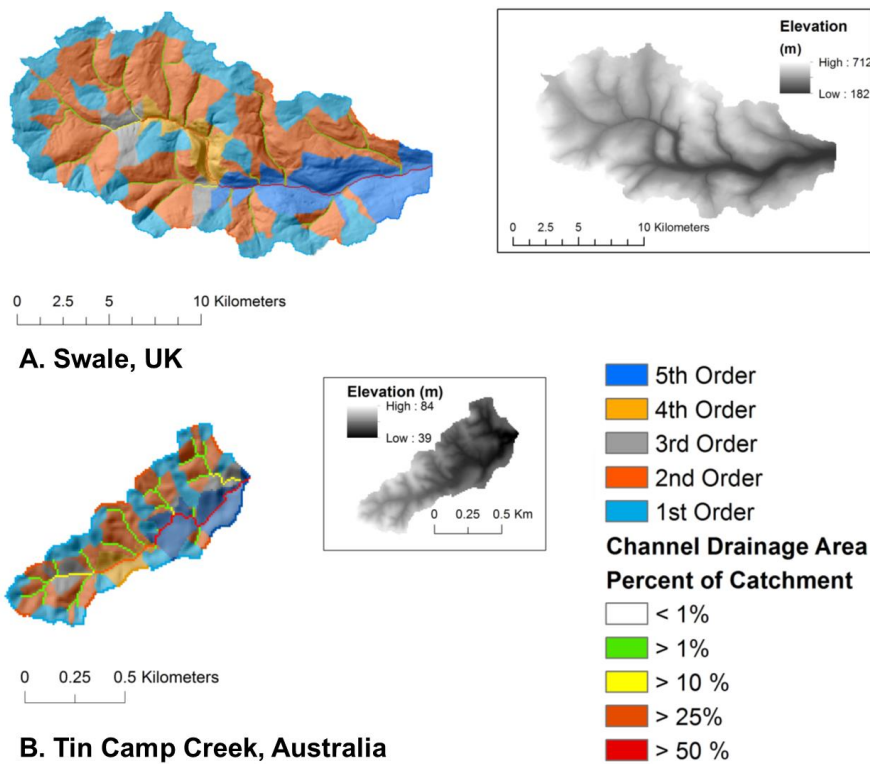
328

329 **2.3.2 Stream Orders**

330

331 The changes in the mean elevation across different areas of the catchments were assessed as an
332 illustration of spatial differences in geomorphic change. Each basin was sub-divided into regions
333 corresponding to the watersheds of five stream orders based on the proportion of the catchment

334 drained in the initial DEM – 1st \leq 1 %; 2nd \geq 1 %; 3rd \geq 10 %; 4th \geq 25 %; 5th \geq 50 % (see
 335 Figure 1). This method is novel and was developed to provide a consistent method of sub-dividing
 336 both catchments independent of factors such as connectivity and DEM resolution.
 337



338
 339 **Figure 1 – Elevation map for the Upper Swale catchment, UK (top), and Tin Camp Creek catchment, Australia**
 340 **(bottom) – note the differences in scale between catchments. Each catchment is sub-divided into watersheds**
 341 **of five stream orders based on the proportion of the catchment drained.**

343 **2.4 User-Defined Parameters**

344

345 The MM implemented here used 15 user-defined parameters, each with 5 iterative step values (as
 346 described in Section 2.2). The only exception was the choice of sediment transport formula parameter
 347 (SED, Table 1) where only two options are available. The parameters, their ranges, and available values
 348 are shown in Table 1.

349

350 **Table 1 – User-defined parameters used and the min-max values for the two study catchments.**

Code	Parameter	Steps	Upper Swale	Tin Camp Creek
(1) SED	Sediment Transport Formula	2	1 Wilcock & Crowe / 2 Einstein	1 Wilcock & Crowe / 2 Einstein
(2) MEL	Max Erode Limit (m)	5	0.01; 0.015; 0.02; 0.025; 0.03	0.001; 0.0015; 0.002; 0.0025; 0.003
(3) CLR	In Channel Lateral Erosion Rate	5	10; 15; 20; 25; 30	10; 15; 20; 25; 30
(4) LAT	Lateral Erosion Rate	5	2.5e ⁻⁶ ; 3.75e ⁻⁶ ; 5e ⁻⁶ ; 6.25e ⁻⁶ ; 7.5e ⁻⁶	1.5e ⁻⁶ ; 2.25e ⁻⁶ ; 3e ⁻⁶ ; 3.75e ⁻⁶ ; 4.5e ⁻⁶
(5) VEG	Vegetation Critical Shear Stress (Pa)	5	10; 15; 20; 25; 30	2; 3.25; 4.5; 5.75; 7
(6) MAT	Grass Maturity Rate (yr)	5	0.5; 0.75; 1; 1.25; 1.5	0.5; 0.875; 1.25; 1.625; 2
(7) SCR	Soil Creep Rate (m/yr)	5	0.00125; 0.001875; 0.0025; 0.003125; 0.00375	0.00125; 0.001875; 0.0025; 0.003125; 0.00375
(8) SFT	Slope Failure Threshold (°)	5	40; 42.5; 45; 47.5; 50	40; 42.5; 45; 47.5; 50
(9) IOD	In/Out Difference (m ³ .s ⁻¹)	5	2.5; 3.75; 5; 6.25; 7.5	0.1; 0.175; 0.25; 0.325; 0.4
(10) MinQ	Min Q Value (m)	5	0.25; 0.375; 0.5; 0.625; 0.75	0.025; 0.0375; 0.05; 0.0625; 0.075
(11) MaxQ	Max Q Value (m)	5	2.5; 3.75; 5; 6.25; 7.5	2.5; 3.75; 5; 6.25; 7.5
(12) SEC	Slope for Edge Cells	5	0.0025; 0.00375; 0.005; 0.00625; 0.0075	0.0025; 0.00375; 0.005; 0.00625; 0.0075
(13) EVR	Evaporation Rate (m/d)	5	0.00067; 0.001005; 0.00134; 0.001675; 0.00201	0.0025; 0.004375; 0.00625; 0.008125; 0.01
(14) MNR	Manning's n Roughness	5	0.03; 0.035; 0.04; 0.045; 0.05	0.03; 0.0325; 0.035; 0.0375; 0.04
(15) GSS	Grain Size Set	5	Set 1; Set 2; Set 3; Set 4; Set 5	Set 1; Set 2; Set 3; Set 4; Set 5

351

352 The MM varies the value of each parameter tested once per repeat, and here we use 100 repeats.
 353 Therefore, careful consideration was required in the selection of parameters as each parameter tested
 354 added 100 model runs to the test – there are 49 user-defined parameters in the version of CAESAR-

355 Lisflood model used (v1.8), and even excluding parameters associated with dune and soil
356 development, there are still 35 user-defined parameters. To test each would require 3600 model runs
357 for each catchment, yet the inclusion of some parameters is likely to add little value. Thus this was
358 narrowed to a set of 15 user-defined parameters (Table 1) with the selection based largely on prior
359 knowledge of the importance of these parameters, or due to a lack of previous knowledge of the
360 influence of the parameters on the model – full justification of the selection of parameters, and
361 descriptions of their purpose within the model, can be found in Supplementary Material S1.

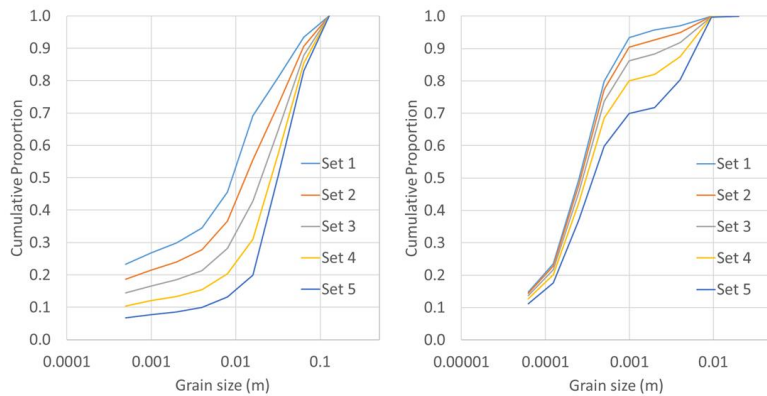
362

363 The MM is subjective in that the relative sensitivities shown depend on the minimum and maximum
364 range values set by the user. Therefore, it is necessary to set each parameter's range to be broadly
365 equal to the others in order to obtain useful information. To be consistent, where possible we have
366 used a default value taken from past calibrations and varied this by +/- 25 % and +/- 50 %. There are
367 some instances where this was not appropriate and a minimum and maximum bound was set instead,
368 with 5 iterative steps of equal distance determined (for example, the Manning's n Roughness for Tin
369 Camp Creek where +/- 50 % would have resulted in obviously physically unrealistic values – see Table
370 1 for values used).

371

372 Here we have considered the selection of sediment transport formula as a parameter despite doing
373 so is to change the functional form of the model. For clarity, and in line with how the choice is
374 presented within the Graphical User Interface of the model, we will henceforth consider this choice in
375 the same way as a parameter. The sediment transport formulae employed for SED were Einstein
376 (derived for sand-bed rivers) (Einstein, 1950) and Wilcock & Crowe (formulated on sediment ranges
377 between 0.5 and 64 mm) (Wilcock and Crowe, 2003). These were not selected as representing the
378 best fit for the catchments simulated but because they are the formulae available in the unmodified
379 version of CAESAR-Lisflood. The sediment transport formulae parameter was applied as a binary
380 choice, with the model switching from one formula to the other once per repeat (no other parameter

381 values were varied when this occurs, as per the description of the MM in Section 2.2). It was assumed
382 that this change constituted a single iterative step change for calculating related EEs.



383
384 **Figure 2 – Sediment grain size distribution sets for the Upper Swale (left) and Tin Camp Creek (right),**
385 **showing the cumulative proportions.**

386
387 Grain size distribution has been shown to influence erosion patterns and erosion rate(Hancock and
388 Coulthard, 2012)). It is more difficult to define iterative steps for the sediment grain size sets which
389 include 9 different grain sizes and proportions in each. Instead, these were skewed by altering the
390 proportions of the five smallest grain sizes +/- 25 % and 50 %, and the opposite to the four largest
391 grain sizes, before adjusting the final proportions to equal one based on the relative values. This
392 produces two sets biased for smaller grain sizes (Sets 1 and 2), and two sets biased for larger grain
393 sizes (Sets 4 and 5), as well as the default grain size set (Set 3) (Figure2). Note, that the grain size sets
394 presented in Figure 2 contain non-cohesive silts and this requires an extrapolation of the sediment
395 transport formulae (Van De Wiel et al., 2007).

396
397 **2.5 Model Functions**

398

399 The common method of assessing a model's sensitivity to parameters values via SA, and the method
400 employed by the MM, is to observe the variations to objective function measures. However, the
401 difficulties in applying an objective function approach to LEMs were highlighted in Section 1.2, and in
402 order to apply an SA a novel approach is required. The method we have developed eschews the
403 objective function approach and instead assesses the model against a series of model functions
404 designed to reflect some of the core behaviours displayed in the model – these can be seen in Table
405 2. This represents a philosophical difference to traditional applications of SA – here we are not testing
406 the model against its skill in simulating the physical environment, but rather how the model responds
407 behaviourally to changes in the user-defined parameters detailed in Section 2.4. The 15 model
408 functions (Table 2) are simple, scalable and transferable between different catchment types, and can
409 be applied to simulations of different timeframes. The model functions are based on outputs which
410 are not unique to CAESAR-Lisflood, so can be applied to other LEM and geomorphic models

411 **Table 2 – Model Functions and the associated core behaviours.**

Model Function	Core Behaviour
Total Sediment Yield (m ³)	Catchment Sediment Yield
Mean Daily Sediment Yield (m ³)	
Peak Daily Sediment Yield (m ³)	
Time to Peak Sediment Yield (s)	
Days when Sediment Yield > Baseline (d)	
Total Net Erosion (m ³)	Internal Geomorphology
Total Net Deposition (m ³)	
Area with > 0.02 m Erosion (m ²)	
Area with > 0.02 m Deposition (m ²)	
Total Discharge (m ³)	Catchment Discharge
Mean Daily Discharge (m ³)	
Peak Daily Discharge (m ³)	
Time to Peak Discharge (s)	
Days when Discharge > Baseline (d)	
Total Model Iterations (calculations)	Model Efficiency

412

413 The model functions were applied to the MM as described in Section 2.2, substituting the model
414 functions in place of the objective functions with no further changes to the method. Model function
415 values were calculated at the end of each simulation.

416

417 To summarise the large amount of information produced, the ME of each parameter and model
418 function combination was normalised based on the proportion of the ME for highest ranking
419 parameter for that model function – therefore the highest ranked parameter for each model function
420 always scored 1. The scores for each parameter were aggregated for across all model functions based
421 on the mean of the scores. The model functions were sub-divided into core behaviour groups (Table
422 2), and the scores aggregated again for each core behaviour. The same was also done, separately, for
423 the standard deviations of each parameter and model function.

424

425 **3. Results**

426

427 ***3.1 All Model Functions***

428

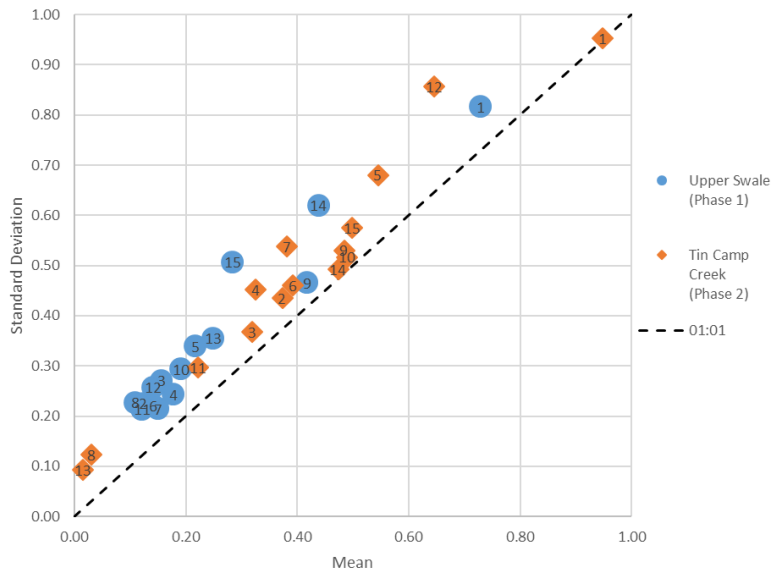
429 Figure 3 shows the spread of parameter influence for both catchments, where a higher mean of the
430 aggregated MEs indicates greater sensitivity in the model to that parameter, and a higher standard
431 deviation shows greater non-linearity when interacting with other parameters. Table 3 shows the
432 parameters ranked for both catchments, based on the aggregated mean ME values. The most
433 influential parameter is SED (see Table 1 for full description of parameter abbreviations), ranked top
434 for both catchments and also being most influential by a reasonable margin, having an aggregated
435 mean of at least 0.2 higher than the 2nd ranked parameter. Other parameters, such as VEG, IOD, MNR,
436 MinQ and GSS, rank highly or mid-range. There is a visually close correlation between the most
437 influential parameters and those that display the most non-linearity (Figure 3).

438

439 Table 3 – Parameters ranked by means for each catchment from the aggregated scores for all Elementary
 440 Effects. SED = sediment transport formula; MEL = maximum erode limit; CLR = in channel lateral erosion rate;
 441 LAT = lateral erosion rate; VEG = vegetation critical shear stress; MAT = grass maturity rate; SCR = soil creep
 442 rate; SFT = slope failure threshold; IOD in/out difference; MinQ = minimum Q value; MaxQ = maximum Q
 443 value; SEC = slope for edge cells; EVR = evaporation rate; MNR = Manning’s n roughness coefficient; and GSS
 444 = grain size set.

Rank (by mean: 1 = most influential)	Upper Swale	Tin Camp Creek
1	SED	SED
2	MNR	SEC
3	IOD	VEG
4	GSS	GSS
5	EVR	MinQ
6	VEG	IOD
7	MinQ	MNR
8	LAT	MAT
9	CLR	SCR
10	SCR	MEL
11	SEC	LAT
12	MAT	CLR
13	MEL	MaxQ
14	MaxQ	SFT
15	SFT	EVR

445



446
 447 **Figure 3 – Aggregated scores for all Elementary Effects where: 1 = sediment transport formula (SED); 2 =**
 448 **maximum erode limit (MEL); 3 = in channel lateral erosion rate (CLR); 4 = lateral erosion rate (LAT); 5 =**
 449 **critical vegetation shear stress (VEG); 6 = grass maturity rate (MAT); 7 = soil creep rate (SCR); 8 = slope**
 450 **failure threshold (SFT); 9 = in/out difference (IOD); 10 = minimum Q value (MinQ); 11 = maximum Q value**
 451 **(MaxQ); 12 = slope for edge cells (SEC); 13 = evaporation rate (EVR); 14 = Manning’s n roughness coefficient**
 452 **(MNR); and 15 = grain size set (GSS).**

453
 454 **3.2 Catchment Sediment Yield Vs Internal Geomorphology**

455
 456 The core behaviours of catchment sediment yield and internal geomorphology show a different
 457 response to the changes in parameter values, as can be seen in Figure 4, and also the rankings in Table
 458 4. For both catchments, SED is ranked as most influential for catchment sediment yields. For influence
 459 on the internal geomorphology, SEC ranks higher in the Tin Camp Creek catchment. The Upper Swale
 460 catchment displays a similar response with both behaviours, with SED and MNR most influential and
 461 by similar amounts, although GSS has less influence on internal geomorphology. The change in

462 response for Tin Camp Creek is more varied – SED is less influential on internal geomorphology, and
 463 SEC is the most influential with a higher aggregated mean. GSS is slightly less influential, and MNR
 464 slightly more, and VEG is more influential on the internal geomorphology than it is on catchment
 465 sediment yield. For both model functions, there again is a strong visually correlation between those
 466 parameters showing the most influence and those showing the most non-linear behaviour.

467

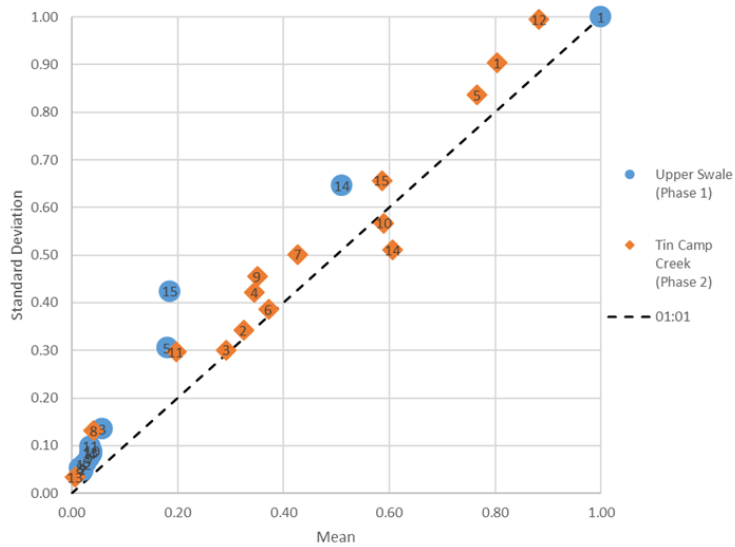
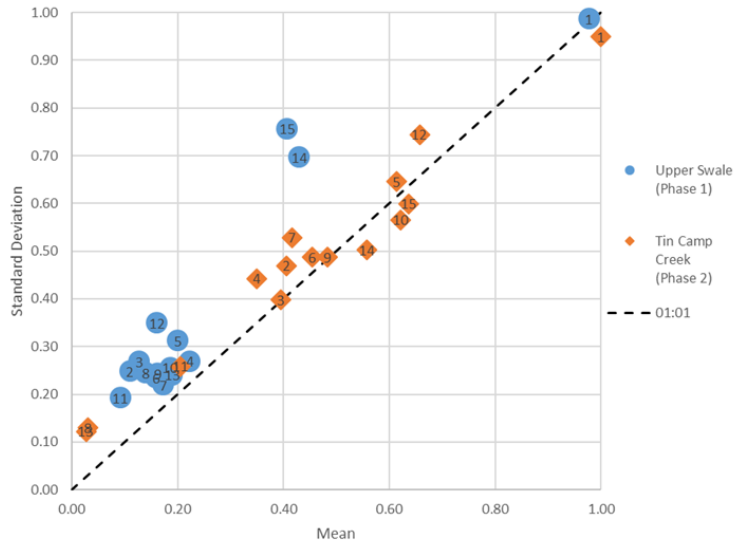
468 **Table 4 – Parameters ranked by means for each catchment from the aggregated scores for catchment**
 469 **sediment yields (SY) and internal geomorphology (IG) elementary effects. SED = sediment transport formula;**
 470 **MEL = maximum erode limit; CLR = in channel lateral erosion rate; LAT = lateral erosion rate; VEG = vegetation**
 471 **critical shear stress; MAT = grass maturity rate; SCR = soil creep rate; SFT = slope failure threshold; IOD in/out**
 472 **difference; MinQ = minimum Q value; MaxQ = maximum Q value; SEC = slope for edge cells; EVR = evaporation**
 473 **rate; MNR = Manning’s n roughness coefficient; and GSS = grain size set.**

474

Rank (by mean: 1 = most influential)	Upper Swale		Tin Camp Creek	
	SY	IG	SY	IG
1	SED	SED	SED	SEC
2	MNR	MNR	SEC	SED
3	GSS	GSS	GSS	VEG
4	LAT	VEG	MinQ	MNR
5	VEG	CLR	VEG	MinQ
6	EVR	LAT	MNR	GSS
7	MinQ	MinQ	IOD	SCR
8	SCR	MaxQ	MAT	MAT
9	IOD	EVR	SCR	IOD
10	SEC	IOD	MEL	LAT

11	MAT	MAT	CLR	MEL
12	SFT	SEC	LAT	CLR
13	CLR	SCR	MaxQ	MaxQ
14	MEL	MEL	SFT	SFT
15	MaxQ	SFT	EVR	EVR

475



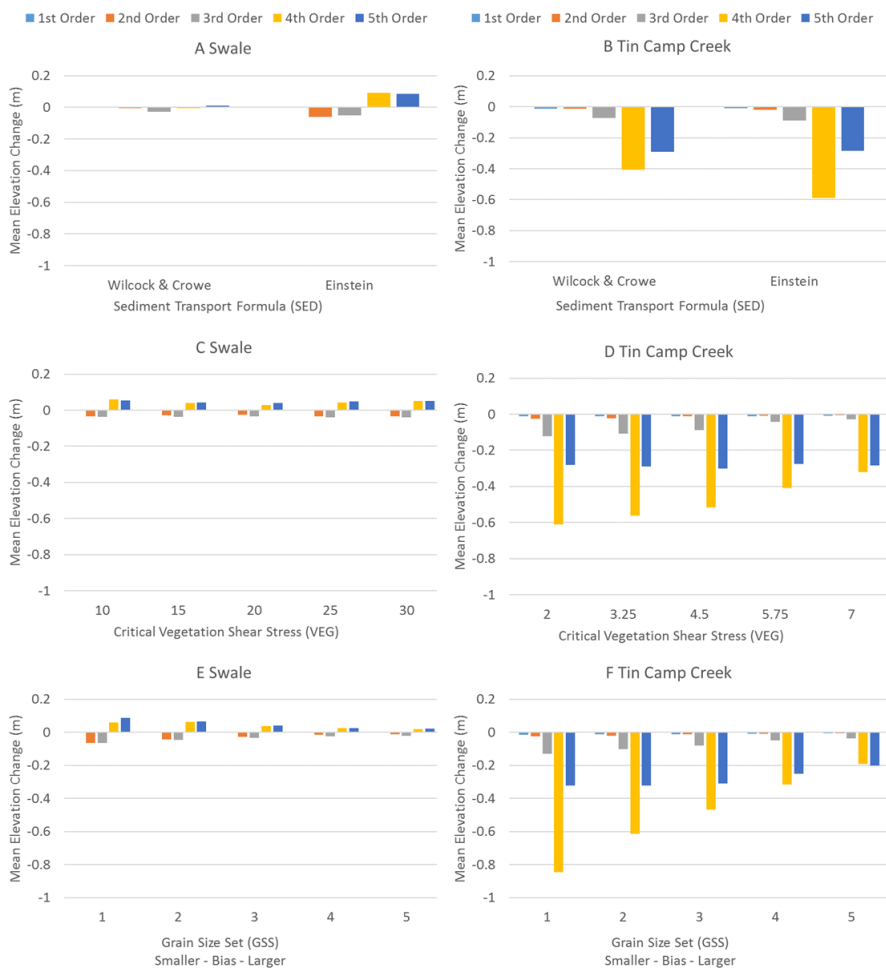
476
 477 **Figure 4 – Aggregated scores for sediment yield (top) and internal geomorphology (bottom) where: 1 =**
 478 **sediment transport formula (SED); 2 = maximum erode limit (MEL); 3 = in channel lateral erosion rate (CLR); 4**
 479 **= lateral erosion rate (LAT); 5 = critical vegetation shear stress (VEG); 6 = grass maturity rate (MAT); 7 = soil**
 480 **creep rate (SCR); 8 = slope failure threshold (SFT); 9 = in/out difference (IOD); 10 = minimum Q value (MinQ);**

481 11 = maximum Q value (MaxQ); 12 = slope for edge cells (SEC); 13 = evaporation rate (EVR); 14 = Manning's n
 482 roughness coefficient (MNR); and 15 = grain size set (GSS).

483

484 **3.3 Changes in the Mean Elevations**

485



486

487 **Figure 5 – Illustration of changes in the mean elevations for Upper Swale (A, C and E), and Tin Camp Creek (B,**

488 **D and F) for the tests split by SED (A and B), VEG (C and D), and GSS (E and F) where 1 and 2 are biased smaller,**

489 and 4 and 5 are biased larger. The catchment is sub-divided into watersheds of five stream orders, based on
490 proportion of catchment drained.

491

492 The test results were binned by the parameter values used, and the mean changes in the mean
493 elevations across the 5 stream orders calculated – Figure 5 illustrates how changes in parameter values
494 might influence the spatial patterns of landscape change using SED, VEG and GSS as examples. For SED
495 (Fig 5.A and 5.B), the most obvious difference is the scale of changes seen using each formula with
496 Einstein generally showing greater change. For Tin Camp Creek (Fig 5.B) the spatial changes are
497 similar, but for the larger Swale (Fig 5.A) there are differences in relative rates in 2nd and 4th order
498 areas. In the Swale, VEG (Fig 5.C) appears to have little impact on the patterns and scale of changes,
499 yet in Tin Camp Creek (Fig 5.D) there is reduction in the rates of erosion across the catchment with
500 higher values, except in the 5th order areas which remain at a similar level. Finally, both catchments
501 show a reduction in rates of erosion with a greater proportion of larger grain sizes, yet this is more
502 pronounced 4th order areas in Tin Camp Creek (Fig 5.F).

503

504 **4. Discussion**

505

506 The results reveal some important insights into the application of the SA to LEMs generally, and also
507 on specific behaviours of the CAESAR-Lisflood model. Here we discuss model functions (Section 4.1),
508 sediment transport formulae (Section 4.2), implications for calibrating LEMs (Section 4.3), full
509 uncertainty analyses of LEMs (Section 4.4), and limitations of this study (Section 4.5).

510

511 **4.1 Model Functions**

512

513 Our findings show that different model functions provide us with different indications of model
514 sensitivity. This has important implications for how to measure LEM performance – and more widely

515 how to quantify and assess geomorphic change within a basin. For example, Figure 4 and Table 4 show
516 how any LEM assessment must depend on the applied metric for comparison. Model functions that
517 quantify sediment yield (derived at the catchment outlet) indicate different sensitivities compared to
518 model functions that quantify the internal landform response (based on spatial measures from within
519 the catchment). Whilst at-a-point sediment yields are straightforward to extract from simulation data
520 and easily related to field measurements (e.g., gauges, although these have their own associated
521 uncertainties), similar or identical yields may conceal very different behaviours within the basin. This
522 highlights an important aspect of LEM calibration: changes in sediment yields from a catchment outlet
523 only provide partial information of what is changing internally. We therefore argue that metrics
524 incorporating *spatial* changes in the basin (as well as bulk figures) are vital for assessing LEM
525 performance. (i.e., time series of high resolution DEM data from LiDAR/photogrammetry) This is
526 especially important as the shape of the landscape – where material has been eroded and deposited
527 – is effectively the basins geomorphic memory and will directly influence subsequent model
528 performance. For other basin scale models (e.g., hydrological models) this aspect is possibly not so
529 important over longer-terms given the limited temporal extent memory of basin antecedence. Some
530 of the challenges of LEM output comparison are similar to those of meteorology/climatology and may
531 require a shift in expectation from end users as to what is possible. For example, predicting detailed
532 patterns of local erosion and deposition is akin to predicting weather (low comparability especially
533 over longer time scales) but more general (spatial and temporal) patterns of basin change are similar
534 to climate predictions (better comparability especially for longer time scales).

535

536 **4.2. Sediment Transport Formulae**

537

538 Our SA shows that the choice of sediment transport formula (SED) had a very strong impact on the
539 model functions. As sediment transport formulae are also integrated into other LEMs and geomorphic
540 models they will affect their outcomes too. Looking at sediment transport formulae themselves,

541 Gomez and Church (1989) tested 11 different sediment transport formulae to the same data sets and
542 showed widespread variation in predictions – in some cases over orders of magnitude. The variation
543 in the model performance can be explained by the derivation of the sediment transport formulae
544 themselves, that are often theory-based but fitted to limited laboratory and field data, sometimes
545 representing temporal averages over equilibrium conditions (Gomez and Church, 1989). The formulae
546 do not, and were likely never intended to, represent the full variation of actual flow conditions in
547 natural river. As LEMs commonly amalgamate a set of geomorphic models or transport formulae, their
548 performance hinges ~~on the~~ a number of individual model components. Therefore, when applied to
549 different situations, they may not be appropriate (Coulthard et al., 2007a).

550

551 **4.3 Implications for Calibrating LEMs**

552

553 This, however, presents a challenge, as it is highly likely that the sediment transport formula to be
554 used was neither designed nor calibrated for a particular model application. The SIBERIA model
555 (Hancock et al., 2010, 2016, 2017; Hancock and Willgoose, 2001; Willgoose et al., 2003) overcomes
556 this issue by having a version of the Einstein sediment transport formula (Einstein, 1950) that is
557 calibrated or tuned to field data on erosion rates. However, even when calibrated, LEMs (and their
558 sediment transport formulae) face another hurdle with the non-stationarity of basin sediment yields.
559 For example, a calibrated LEM will be adjusted to perform for a set of observed sediment outputs or
560 erosion and deposition patterns. If, due to climate change for example, [sediment supply, rainfall and](#)
561 [or channel flows significantly increase outside of the range of the initial calibration](#) then ~~that~~ initial
562 calibration may no longer be valid (Coulthard et al., 2007b). This is similar to issues faced by calibrating
563 hydrological models (e.g., Li et al., 2012) though the non-linear sediment response of LEMs like
564 CAESAR-Lisflood (Coulthard et al., 2012) may make LEMs more sensitive to this. [Such a non-linear](#)
565 [sediment response to hydrological increases can be traced to the calculation of sediment transport as](#)
566 [a square or cubic function of flow velocity](#). Furthermore, this analysis suggests that detailed

567 justification and calibration of model choices around sediment transport will lead to the most effective
568 gains in model skill.

569

570 **4.4 Full Uncertainty Analysis**

571

572 It is important to note that the MM does not provide an absolute value of sensitivity, but ranks each
573 factor based on its relative influence on the model. This means it can be used to assess the main
574 sources of uncertainty on a particular model set up. The next step is then to establish how the
575 uncertainty caused by model parameters (e.g., the choice of sediment transport formula) compares
576 to other identified sources of uncertainty, such as rainfall input uncertainty, DEM observation and
577 resolution uncertainty, and length of spin-up period. For example, it may be that the choice of
578 sediment transport formula may only be a minor source of uncertainty compared to the DEM
579 resolution, or equally, it might be the most significant source of uncertainty in a LEM's output.

580

581 Importantly, whilst the simulation of long-term development of landscapes may be somewhat
582 resilient to some uncertainties, e.g., initial conditions (Hancock et al., 2016), any attempt to reproduce,
583 predict or forecast physical changes should have the same appreciation of uncertainty and rigorous
584 testing that is applied to models in other fields (e.g., hydrology and hydraulics). There are many
585 methods available, but when discussing CAESAR-Lisflood the applications applied to Lisflood-FP seem
586 a reasonable place to start.- Lisflood-FP has been rigorously tested and benchmarked for decision-
587 making purposes (Hunter et al., 2005; Neelz & Pender, 2013), and the use of SA to assess model
588 response and uncertainty is standard practise (Di Baldassarre et al., 2009; Fewtrell et al., 2008, 2011;
589 Hall et al., 2005; Horritt and Bates, 2001, 2002; Hunter et al., 2008; Neal et al., 2011; Sampson et al.,
590 2012), often as a stage of calibration using the GLUE method (Aronica et al., 2002; Bates et al., 2004;
591 Horritt et al., 2006; Hunter et al., 2005; Pappenberger et al., 2007; Wong et al., 2015). Uncertainty in
592 model predictions can be accounted for by utilising probabilistic measures and uncertainty cascades

Field Code Changed

593 (for example, Pappenberger et al., 2005; Stephens et al., 2012). This is not considered unique to
594 CAESAR-Lisflood, and any application of an LEM or other geomorphic model for operational, decision-
595 making or forecasting applications should make full consideration of all associated uncertainties.

596

597 **4.5. Limitations**

598

599 The main limitation of the MM is the subjectivity in selection of parameter values and ranges. Here,
600 this has been mitigated by consistently selected ranges of +/- 50 % of a default value obtained from
601 previous calibrations (where feasible). An issue emerges with categorical parameters, such as SED,
602 where multiple values cannot be placed in spectrum across a range between minimum and maximum
603 values. The MM has no formal method for dealing with such categorical parameters, so here it has
604 been assumed that switching from one formula to another is a single iterative step change, and this
605 would be the same even with more choices available. This reflects the purpose of the MM, which is to
606 inform about the relative importance of choices of parameter values on the performance/behaviour
607 of the model. However, to assess the impact of this single step-change assumption, we performed a
608 further analysis, where it was assumed that switching formula was a change of four iterative steps.
609 This analysis shows that the relative sensitivity of the model to the sediment transport formula choice
610 becomes less important, with other parameters such as Manning's n Roughness and grain size sets
611 increasing in relative influence (see Supplementary Material S2 for full results of this analysis).

612

613

614 An obvious limitation to this exercise is computational resource. This study incorporated 1600
615 individual model runs to test the behavioural response of the model to 15 parameters, in just two
616 catchments, and this partly influenced the choice to limit simulation periods to 20 years. The bulk of
617 simulations used Intel i7-5960X processors and using Solid State Drives (SSD), yet the run times varied
618 considerably depending on the parameter sets chosen. As an indication, the mean simulation run time

619 for the first repeat in each catchment was 11 hours and 23 minutes for the Swale and 21 minutes for
620 Tin Camp Creek. We used a batch mode functionality of CAESAR-Lisflood to run simulations of each
621 repeat (16 model runs each) consecutively, and distributed batches across different machines – this is
622 feasible for the model set ups described. However, for long-term simulations for catchments the size
623 of the Upper Swale, individual model runs can take several weeks and running several runs
624 consecutively becomes prohibitive. One solution would be to distribute the jobs on High Performance
625 Computing (HPC) facilities, where the time for a single model run would not significantly decrease, but
626 several hundred, even thousands, of individual model runs can be performed coincidentally.

627

628 Here, the methodology has only been applied to the CAESAR-Lisflood model, and although some of
629 the findings will be unique to CAESAR-Lisflood and the model set ups presented, they have
630 implications for all LEMs. Importantly, the methodology can serve as a highly useful tool for users to
631 determine the behaviour of any LEM model set up prior to calibration and/or simulation.

632

633 **5. Conclusions**

634

635 The feasibility of performing global SA to a highly parameterised catchment LEM has been
636 demonstrated through the application of the MM to the CAESAR-Lisflood model. The analysis was
637 repeated over two different catchments suggesting some model behaviours are universal, and others
638 vary depending on the catchment characteristics providing crucial information to inform future model
639 developments. This analysis confirms that the sediment transport formulae are a significant source of
640 uncertainty in LEMs, and in the CAESAR-Lisflood model the use of one formula over another can result
641 in an order of magnitude differences in sediment yields when all other factors are kept constant.

642 Another finding with relevance to SA and calibration of LEMs was the influence of parameters on each
643 model function, showing that one aspect of model behaviour (e.g., catchment sediment yield) is not
644 fully reflective of other, albeit related, model behaviours (e.g., internal geomorphology).

645

646 In addition to the above, the results reveal the parameters in CAESAR-Lisflood which exert the greatest
647 influence, and whilst we can only apply this to the CAESAR-Lisflood model itself, it is likely that LEMs
648 with comparable parameters will display similar behaviours. Some of the most influential parameters,
649 like Manning's n roughness coefficient, grain size distributions, and vegetation critical shear stress are
650 physically-based, so any uncertainty can be reduced by more detailed field measurements. We also
651 show that parameters that determine the numerical efficiency of CAESAR-Lisflood exert a medium
652 influence on the simulation results. Although some parameters exerted less influence on model
653 behaviour relative to others, there were no parameters which did not influence the model in some
654 way.

655

656 The application of a global SA should become a vital step in any investigation using LEMs. This paper
657 has demonstrated that the use of the MM is efficient for this purpose and yielded some valuable
658 insights into model behaviour that can ultimately feed back into model set up, as well as future model
659 development.

660

661 **Model and Data Availability**

662

663 The data produced by this study is made available on request from the corresponding author. The
664 CAESAR-Lisflood model used in this study is freely available under a GNU licence from

665 <http://www.coulthard.org.uk>

666

667 **Competing Interests**

668 The authors declare that they have no conflict of interest.

669

670 **Acknowledgements**

671

672 The authors wish to thank the two reviewers, Andy Wickert and Daniel Hobley, for their insightful and
673 helpful comments which have improved this manuscript. The Landscape Evolution Model Sensitivity
674 Investigation Project (LEMSIP) has emerged from the Field and Computer Simulation in Landscape
675 Evolution (FACSIMILE) network. The aims of the project are to collate and generate knowledge
676 pertaining to the sensitivities and uncertainties associated with Landscape Evolution Models, and how
677 these influence the simulation of landscape development. The authors wish to thank the Young
678 Geomorphologists group who donated computational resource. This work was supported by the NERC
679 Flooding from Intense Rainfall (FFIR) project, Susceptibility of Basins to Intense Rainfall and Flooding
680 (SINATRA) NE/K008668/1. The CAESAR-Lisflood model used in this study is freely available under a
681 GNU licence from <http://www.coulthard.org.uk>

682

683 **References**

- 684 Adams, J. M., Gasparini, N. M., Hobbey, D. E. J., Tucker, G. E., Hutton, E. W. H., Nudurupati, S. S. and
685 Istanbuluoglu, E.: The Landlab v1.0 OverlandFlow component: a Python tool for computing shallow-
686 water flow across watersheds, *Geosci. Model Dev*, 10, 1645–1663, doi:10.5194/gmd-10-1645-2017,
687 2017.
- 688 Andersen, J. L., Egholm, D. L., Knudsen, M. F., Jansen, J. D. and Nielsen, S. B.: The periglacial engine
689 of mountain erosion - Part 1: Rates of frost cracking and frost creep, *Earth Surf. Dyn.*, 3(4), 447–462,
690 doi:10.5194/esurf-3-447-2015, 2015.
- 691 Armitage, J. J., Whittaker, A. C., Zakari, M. and Campforts, B.: Numerical modelling landscape and
692 sediment flux response to precipitation rate change, *Earth Surf. Dyn. Discuss.*, (May), 1–31,
693 doi:10.5194/esurf-2017-34, 2017.
- 694 Aronica, G., Bates, P. D. and Horritt, M. S.: Assessing the uncertainty in distributed model predictions
695 using observed binary pattern information within GLUE, *Hydrol. Process.*, 16(10), 2001–2016,
696 doi:10.1002/hyp.398, 2002.
- 697 Attal, M., Tucker, G. E., Whittaker, A. C., Cowie, P. A. and Roberts, G. P.: Modelling fluvial incision
698 and transient landscape evolution: Influence of dynamic Channel adjustment, *J. Geophys. Res. Earth*
699 *Surf.*, 113(3), 1–16, doi:10.1029/2007JF000893, 2008.
- 700 Di Baldassarre, G., Schumann, G. and Bates, P. D.: A technique for the calibration of hydraulic models
701 using uncertain satellite observations of flood extent, *J. Hydrol.*, 367(3), 276–282,
702 doi:10.1016/j.jhydrol.2009.01.020, 2009.
- 703 Bates, P. D., Horritt, M. S., Aronica, G. and Beven, K.: Bayesian updating of flood inundation
704 likelihoods conditioned on flood extent data, *Hydrol. Process.*, 18(17), 3347–3370,
705 doi:10.1002/hyp.1499, 2004.
- 706 Bates, P. D., Horritt, M. S. and Fewtrell, T. J.: A simple inertial formulation of the shallow water
707 equations for efficient two-dimensional flood inundation modelling, *J. Hydrol.*, 387(1–2), 33–45,
708 doi:10.1016/j.jhydrol.2010.03.027, 2010.
- 709 Beven, K. and Kirkby, M.: A physically based, variable contributing area model of basin hydrology/Un
710 modèle à base physique de zone d'appel variable de l'hydrologie du bassin versant, *Hydrol. Sci. J.*,
711 24(1), 37–41 [online] Available from:
712 <http://www.tandfonline.com/doi/abs/10.1080/02626667909491834> (Accessed 8 May 2014), 1979.
- 713 Brockmann, D. and Morgenroth, E.: Comparing global sensitivity analysis for a biofilm model for two-
714 step nitrification using the qualitative screening method of Morris or the quantitative variance-based
715 Fourier Amplitude Sensitivity Test (FAST), *Water Sci. Technol.*, 56(8), 85–93,
716 doi:10.2166/wst.2007.600, 2007.
- 717 [Campforts, B., Schwanghart, W., Govers, G.: Accurate simulation of transient landscape evolution by](#)
718 [eliminating numerical diffusion: the TTLEM 1.0 model. *Earth Surface Dynamics*, 5, 47-66, doi:](#)
719 [10.5194/esurf-5-47-2017, 2017.](#) [Campforts, B., Vanacker, V., Vanderborght, J., Baken, S., Smolders, E.](#)
720 [and Govers, G.: Simulating the mobility of meteoric ¹⁰Be in the landscape through a coupled soil-](#)
721 [hillslope model \(Be2D\), *Earth Planet. Sci. Lett.*, 439, 143–157, doi:10.1016/j.epsl.2016.01.017, 2016.](#)
- 722 Campolongo, F., Cariboni, J. and Saltelli, A.: An effective screening design for sensitivity analysis of
723 large models, *Environ. Model. Softw.*, 22(10), 1509–1518, doi:10.1016/j.envsoft.2006.10.004, 2007.
- 724 Cariboni, J., Gatelli, D., Liska, R. and Saltelli, A.: The role of sensitivity analysis in ecological modelling,
725 *Ecol. Modell.*, 203(1–2), 167–182, doi:10.1016/j.ecolmodel.2005.10.045, 2007.

- 726 Collins, D. B. G.: Modeling the effects of vegetation-erosion coupling on landscape evolution, *J. Geophys. Res.*, 109(F3), 1–11, doi:10.1029/2003JF000028, 2004.
- 727
- 728 Coulthard, T., Hicks, D. and Wiel, M. Van De: Cellular modelling of river catchments and reaches: Advantages, limitations and prospects, *Geomorphology*, 90(3–4), 192–207, doi:10.1016/j.geomorph.2006.10.030, 2007a.
- 729
- 730
- 731 Coulthard, T., Neal, J., Bates, P., Ramirez, J., de Almeida, G. and Hancock, G.: Integrating the LISFLOOD-FP 2D hydrodynamic model with the CAESAR model: implications for modelling landscape evolution, *Earth Surf. ...*, 38(15), 1897–1906, doi:10.1002/esp.3478, 2013.
- 732
- 733
- 734 Coulthard, T. J. and Macklin, M. G.: How sensitive are river systems to climate and land-use changes? A model-based evaluation, *J. Quat. Sci.*, 16(4), 347–351, doi:10.1002/jqs.604, 2001.
- 735
- 736 Coulthard, T. J. and Skinner, C. J.: The sensitivity of landscape evolution models to spatial and temporal rainfall resolution, *Earth Surf. Dyn.*, 4(3), 757–771, doi:10.5194/esurf-4-757-2016, 2016a.
- 737
- 738 Coulthard, T. J. and Skinner, C. J.: The sensitivity of landscape evolution models to spatial and temporal rainfall resolution, *Earth Surf. Dyn. Discuss.*, 1–28, doi:10.5194/esurf-2016-2, 2016b.
- 739
- 740 Coulthard, T. J. and Van De Wiel, M. J.: Quantifying fluvial non linearity and finding self organized criticality? Insights from simulations of river basin evolution, *Geomorphology*, 91(3–4), 216–235, doi:10.1016/j.geomorph.2007.04.011, 2007.
- 741
- 742
- 743 Coulthard, T. J. and Van De Wiel, M. J.: Modelling river history and evolution, *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, 370(1966), 2123–2142, doi:10.1098/rsta.2011.0597, 2012.
- 744
- 745 Coulthard, T. J. and Van De Wiel, M. J.: Climate, tectonics or morphology: What signals can we see in drainage basin sediment yields?, *Earth Surf. Dyn.*, 1(1), 13–27, doi:10.5194/esurf-1-13-2013, 2013.
- 746
- 747 Coulthard, T. J. and Van De Wiel, M. J.: Modelling long term basin scale sediment connectivity, driven by spatial land use changes, *Geomorphology*, 277, 265–281, doi:10.1016/j.geomorph.2016.05.027, 2017.
- 748
- 749
- 750 Coulthard, T. J., Lewin, J. and Macklin, M. G.: 12 Non-stationarity of basin scale sediment delivery in response to climate change, *Dev. Earth Surf. Process.*, 11(07), 315–331, doi:10.1016/S0928-2025(07)11131-7, 2007b.
- 751
- 752
- 753 Coulthard, T. J., Ramirez, J., Fowler, H. J. and Glenis, V.: Using the UKCP09 probabilistic scenarios to model the amplified impact of climate change on drainage basin sediment yield, *Hydrol. Earth Syst. Sci.*, 16(11), 4401–4416, doi:10.5194/hess-16-4401-2012, 2012.
- 754
- 755
- 756 Egholm, D. L., Andersen, J. L., Knudsen, M. F., Jansen, J. D. and Nielsen, S. B.: The periglacial engine of mountain erosion - Part 2: Modelling large-scale landscape evolution, *Earth Surf. Dyn.*, 3(4), 463–482, doi:10.5194/esurf-3-463-2015, 2015.
- 757
- 758
- 759 Einstein, H. A.: The Bed-Load Function for Sediment Transportation in Open Channel Flows, *Soil Conserv. Serv.*, (1026), 1–31 [online] Available from: https://ponce.sdsu.edu/einstein_bedload_function.pdf (Accessed 4 July 2018), 1950.
- 760
- 761
- 762 Fewtrell, T. J., Bates, P. D., Horritt, M. and Hunter, N. M.: Evaluating the effect of scale in flood inundation modelling in urban environments, *Hydrol. Process.*, 22(26), 5107–5118, doi:10.1002/hyp.7148, 2008.
- 763
- 764
- 765 Fewtrell, T. J., Duncan, A., Sampson, C. C., Neal, J. C. and Bates, P. D.: Benchmarking urban flood models of varying complexity and scale using high resolution terrestrial LiDAR data, *Phys. Chem. Earth, Parts A/B/C*, 36(7), 281–291, doi:10.1016/j.pce.2010.12.011, 2011.
- 766
- 767

Formatted: French (France)

768 Gomez, B. and Church, M.: An Assessment of Bedload Sediment transport Formulae for Gravel Bed
769 Rivers, *Water Resour. Res.*, 25(6), 1161–1186, 1989.

770 van Griensven, A., Meixner, T., Grunwald, S., Bishop, T., Diluzio, M. and Srinivasan, R.: A global
771 sensitivity analysis tool for the parameters of multi-variable catchment models, *J. Hydrol.*, 324(1–4),
772 10–23, doi:10.1016/j.jhydrol.2005.09.008, 2006.

773 Hall, J. W., Tarantola, S., Bates, P. D. and Horritt, M. S.: Distributed Sensitivity Analysis of Flood
774 Inundation Model Calibration, *J. Hydraul. Eng.*, 131(2), 117–126, doi:10.1061/(ASCE)0733-
775 9429(2005)131:2(117), 2005.

776 Hancock, G. and Willgoose, G.: Use of a landscape simulator in the validation of the SIBERIA
777 catchment evolution model: Declining equilibrium landforms, *Water Resour. Res.*, 37(7), 1981–1992,
778 doi:10.1029/2001WR900002, 2001.

779 Hancock, G. R.: The impact of different gridding methods on catchment geomorphology and soil
780 erosion over long timescales using a landscape evolution model, *Earth Surf. Process. Landforms*,
781 31(8), 1035–1050, doi:10.1002/esp.1306, 2006.

782 Hancock, G. R. and Coulthard, T. J.: Channel movement and erosion response to rainfall variability in
783 southeast Australia, *Hydrol. Process.*, 26(5), 663–673, doi:10.1002/hyp.8166, 2012.

784 Hancock, G. R., Lowry, J. B. C., Coulthard, T. J., Evans, K. G. and Moliere, D. R.: A catchment scale
785 evaluation of the SIBERIA and CAESAR landscape evolution models, *Earth Surf. Process. Landforms*,
786 35(8), 863–875, doi:10.1002/esp.1863, 2010.

787 Hancock, G. R., Coulthard, T. J., Martinez, C. and Kalma, J. D.: An evaluation of landscape evolution
788 models to simulate decadal and centennial scale soil erosion in grassland catchments, *J. Hydrol.*,
789 398(3–4), 171–183, doi:10.1016/j.jhydrol.2010.12.002, 2011.

790 Hancock, G. R., Lowry, J. B. C. and Coulthard, T. J.: Catchment reconstruction - erosional stability at
791 millennial time scales using landscape evolution models, *Geomorphology*, 231, 15–27,
792 doi:10.1016/j.geomorph.2014.10.034, 2015.

793 Hancock, G. R., Coulthard, T. J. and Lowry, J. B. C.: Predicting uncertainty in sediment transport and
794 landscape evolution - the influence of initial surface conditions, *Comput. Geosci.*, 90, 117–130,
795 doi:10.1016/j.cageo.2015.08.014, 2016.

796 Hancock, G. R., Verdon-Kidd, D. and Lowry, J. B. C.: Sediment output from a post-mining catchment -
797 Centennial impacts using stochastically generated rainfall, *J. Hydrol.*, 544, 180–194,
798 doi:10.1016/j.jhydrol.2016.11.027, 2017.

799 Herman, J. D., Kollat, J. B., Reed, P. M. and Wagener, T.: Technical Note: Method of Morris
800 effectively reduces the computational demands of global sensitivity analysis for distributed
801 watershed models, *Hydrol. Earth Syst. Sci.*, 17(7), 2893–2903, doi:10.5194/hess-17-2893-2013, 2013.

802 Horritt, M., Bates, P. and Mattinson, M.: Effects of mesh resolution and topographic representation
803 in 2D finite volume models of shallow water fluvial flow, *J. Hydrol.*, 329(1–2), 306–314,
804 doi:10.1016/j.jhydrol.2006.02.016, 2006.

805 Horritt, M. S. and Bates, P. D.: Effects of spatial resolution on a raster based model of flood flow, *J.*
806 *Hydrol.*, 253(1–4), 239–249, doi:10.1016/S0022-1694(01)00490-5, 2001.

807 Horritt, M. S. and Bates, P. D.: Evaluation of 1D and 2D numerical models for predicting river flood
808 inundation, *J. Hydrol.*, 268(1), 87–99, doi:10.1016/S0022-1694(02)00121-X, 2002.

809 Hunter, N. M., Horritt, M. S., Bates, P. D., Wilson, M. D. and Werner, M. G. F.: An adaptive time step
810 solution for raster-based storage cell modelling of floodplain inundation, *Adv. Water Resour.*, 28(9),

811 975–991, doi:10.1016/j.advwatres.2005.03.007, 2005.

812 Hunter, N. M., Bates, P. D., Neelz, S., Pender, G., Villanueva, I., Wright, N. G., Liang, D., Falconer, R.
813 A., Lin, B., Waller, S., Crossley, A. J. and Mason, D. C.: Benchmarking 2D hydraulic models for urban
814 flooding, *Proc. Inst. Civ. Eng. - Water Manag.*, 161(1), 13–30, doi:10.1680/wama.2008.161.1.13,
815 2008.

816 Ibbitt, R. P., Willgoose, G. R. and Duncan, M. J.: Channel network simulation models compared with
817 data from the Ashley River, New Zealand, *Water Resour. Res.*, 35(12), 3875–3890,
818 doi:10.1029/1999WR900245, 1999.

819 Ijjasz-Vasquez, E. J., Bras, R. L. and Moglen, G. E.: Sensitivity of a basin evolution model to the nature
820 of runoff production and to initial conditions, *Water Resour. Res.*, 28(10), 2733–2741,
821 doi:10.1029/92WR01561, 1992.

822 Istanbuluoglu, E. and Bras, R. L.: Vegetation-modulated landscape evolution: Effects of vegetation
823 on landscape processes, drainage density, and topography, *J. Geophys. Res. Earth Surf.*, 110(2), 1–
824 19, doi:10.1029/2004JF000249, 2005.

825 Jerolmack, D. J. and Paola, C.: Shredding of environmental signals by sediment transport, *Geophys.*
826 *Res. Lett.*, 37(19), 1–5, doi:10.1029/2010GL044638, 2010.

827 Larsen, L., Thomas, C., Eppinga, M. and Coulthard, T.: Exploratory modeling: Extracting causality
828 from complexity, *Eos (Washington, DC)*, 95(32), 285–286, doi:10.1002/2014EO320001, 2014.

829 Li, C., Zhang, L., Wang, H., Zhang, Y., Yu, F. and Yan, D.: The transferability of hydrological models
830 under nonstationary climatic conditions., *Hydrol. Earth ...*, 16(4), 1239–1254, doi:10.5194/hess-16-
831 1239-2012, 2012.

832 Liu, B. and Coulthard, T. J.: Modelling the interaction of aeolian and fluvial processes with a
833 combined cellular model of sand dunes and river systems, *Comput. Geosci.*, 106, 1–9,
834 doi:10.1016/j.cageo.2017.05.003, 2017.

835 Martin, Y. and Church, M.: Numerical modelling of landscape evolution: geomorphological
836 perspectives, *Prog. Phys. Geogr.*, 28(3), 317–339, doi:10.1191/0309133304pp412ra, 2004.

837 Met Office: 5km UK Composite Rainfall Data from the Met Office NIMROD System, NCAS Br. Atmos.
838 Data Centre, available at : <http://catalogue.ceda.ac.uk/uuid/82adec1f896af6169112d09cc1174499>
839 (last access: 20 September 2016), 2003.

840 Morris, M. D.: Factorial Sampling Plans for Preliminary Computational Experiments, *Technometrics*,
841 33(2), 161–174, doi:10.2307/1269043, 1991.

842 Nash, J. and Sutcliffe, J.: River flow forecasting through conceptual models part I—A discussion of
843 principles, *J. Hydrol.*, 10, 282–290 [online] Available from:
844 <http://www.sciencedirect.com/science/article/pii/0022169470902556> (Accessed 8 May 2014), 1970.

845 Neal, J., Schumann, G., Fewtrell, T., Budimir, M., Bates, P. and Mason, D.: Evaluating a new
846 LISFLOOD-FP formulation with data from the summer 2007 floods in Tewkesbury, UK, *J. Flood Risk*
847 *Manag.*, 4(2), 88–95, doi:10.1111/j.1753-318X.2011.01093.x, 2011.

848 Neelz, S. & Pender, G.: Benchmarking the latest generation of 2D hydraulic modelling packages.
849 [online] Available from: [http://evidence.environment-](http://evidence.environment-agency.gov.uk/FCERM/Libraries/FCERM_Project_Documents/SC120002_Benchmarking_2D_hydraulic_models_Report.sflb.ashx)
850 [agency.gov.uk/FCERM/Libraries/FCERM_Project_Documents/SC120002_Benchmarking_2D_hydrauli](http://evidence.environment-agency.gov.uk/FCERM/Libraries/FCERM_Project_Documents/SC120002_Benchmarking_2D_hydraulic_models_Report.sflb.ashx)
851 [c_models_Report.sflb.ashx](http://evidence.environment-agency.gov.uk/FCERM/Libraries/FCERM_Project_Documents/SC120002_Benchmarking_2D_hydraulic_models_Report.sflb.ashx), 2013.

852 Neumann, M. B.: Comparison of sensitivity analysis methods for pollutant degradation modelling: A
853 case study from drinking water treatment, *Sci. Total Environ.*, 433(October), 530–537,

854 doi:10.1016/j.scitotenv.2012.06.026, 2012.

855 Norton, J. P.: Algebraic sensitivity analysis of environmental models, *Environ. Model. Softw.*, 23,
856 963–972, doi:10.1016/j.envsoft.2007.11.007, 2008.

857 Norton, J. P.: Selection of Morris trajectories for initial sensitivity analysis, *IFAC.*, 2009.

858 Oakley, J. E. and O’Hagan, A.: Probabilistic Sensitivity Analysis of Complex Models : A Bayesian
859 Approach Author (s): Jeremy E . Oakley and Anthony O’Hagan Published by : Wiley for the Royal
860 Statistical Society Stable URL : <http://www.jstor.org/stable/3647504> Probabilistic sensitiv, , 66(3),
861 751–769, 2004.

862 Oreskes, N., Shrader-Frechette, K. and Belitz, K.: Verification, Validation, and Confirmation of
863 Numerical Models in the Earth Sciences, *Science (80-.)*, 263, 641–646, doi:10.2307/2883078, 1994.

864 Pappenberger, F., Beven, K. J., Hunter, N. M., Bates, P. D., Gouweleeuw, B. T., Thielen, J. and Roo, A.
865 P. J. De: Cascading model uncertainty from medium range weather forecasts (10 days) through a
866 rainfall-runoff model to flood inundation predictions within the European Flood Forecasting System
867 (EFFS), *Hydrol. Earth Syst. Sci. Discuss.*, 9(4), 381–393, doi:10.5194/hess-9-381-2005, 2005.

868 Pappenberger, F., Harvey, H., Beven, K., Hall, J. and Meadowcroft, I.: Decision tree for choosing an
869 uncertainty analysis methodology : a wiki experiment, *Hydrol. Process.*, 20, 3793–3798,
870 doi:10.1002/hyp, 2006.

871 Pappenberger, F., Frodsham, K., Beven, K., Romanowicz, R. and Matgen, P.: Fuzzy set approach to
872 calibrating distributed flood inundation models using remote sensing observations, *Hydrol. Earth
873 Syst. Sci. Discuss.*, 11(2), 739–752 [online] Available from: [https://hal.archives-ouvertes.fr/hal-
874 00305049/](https://hal.archives-ouvertes.fr/hal-00305049/) (Accessed 24 May 2017), 2007.

875 Pappenberger, F., Beven, K. J., Ratto, M. and Matgen, P.: Multi-method global sensitivity analysis of
876 flood inundation models, *Adv. Water Resour.*, 31(1), 1–14, doi:10.1016/j.advwatres.2007.04.009,
877 2008.

878 Pazzaglia, F. J.: Landscape evolution models, pp. 247–274., 2003.

879 Petersen, A. C. (Arthur C.: Simulating nature : a philosophical study of computer-simulation
880 uncertainties and their role in climate science and policy advice, CRC Press. [online] Available from:
881 [https://books.google.co.uk/books?hl=en&lr=&id=I4GhNkiPv3EC&oi=fnd&pg=PP1&dq=Simulating+N
882 ature:+A+Philosophical+Study+of+Computer-
883 Simulation+Uncertainties+and+Their+Role+in+Climate+Science+and+Policy+Advice&ots=EKmUbPtt
884 VZ&sig=BisleTDNw3E0_EpozyLbxjJHudg#v=onepage&q=Simulating+Nature%3A+A+Philosophical+Study
885 of+Computer-Simulation+Uncertainties+and+Their+Role+in+Climate+Science+and+Policy+Advice&f=false
886 \(Accessed 18 August 2017\), 2012.](https://books.google.co.uk/books?hl=en&lr=&id=I4GhNkiPv3EC&oi=fnd&pg=PP1&dq=Simulating+Nature:+A+Philosophical+Study+of+Computer-Simulation+Uncertainties+and+Their+Role+in+Climate+Science+and+Policy+Advice&ots=EKmUbPttVZ&sig=BisleTDNw3E0_EpozyLbxjJHudg#v=onepage&q=Simulating+Nature%3A+A+Philosophical+Study+of+Computer-Simulation+Uncertainties+and+Their+Role+in+Climate+Science+and+Policy+Advice&f=false)

887 Pianosi, F., Beven, K., Freer, J., Hall, J. W., Rougier, J., Stephenson, D. B. and Wagener, T.: Sensitivity
888 analysis of environmental models: A systematic review with practical workflow, *Environ. Model.
889 Softw.*, 79, 214–232, doi:10.1016/j.envsoft.2016.02.008, 2016.

890 Pujol, G.: R Package “sensitivity”. Version 1.4-0, 2009.

891 R Hancock, G.: Modelling stream sediment concentration: An assessment of enhanced rainfall and
892 storm frequency, *J. Hydrol.*, 430–431, 1–12, doi:10.1016/j.jhydrol.2012.01.022, 2012.

893 Ratto, M., Pagano, A. and Young, P.: State Dependent Parameter metamodelling and sensitivity
894 analysis, *Comput. Phys. Commun.*, 177(11), 863–876, doi:10.1016/j.cpc.2007.07.011, 2007.

895 Saltelli, A., Chan, K. and Scott, E. M.: Sensitivity Analysis, John Wiley, New York, 2000.

896 Sampson, C. C., Fewtrell, T. J., Duncan, A., Shaad, K., Horritt, M. S. and Bates, P. D.: Use of terrestrial
897 laser scanning data to drive decimetric resolution urban inundation models, *Adv. Water Resour.*, 41,
898 1–17, doi:10.1016/j.advwatres.2012.02.010, 2012.

899 Saynor, M. J., Lowry, J., Erskine, W. D., Coulthard, T. and Hancock, G.: Assessing Erosion and Run-Off
900 Performance of a Trial Rehabilitated, *Proc. Life Mine Conf. July 2012*, (July), 10–12, 2012.

901 Skinner, C. and Coulthard, T.: Caesar-Lisflood Existing Applications Parameter Listings - May 2017, ,
902 doi:10.5281/ZENODO.800558, 2017.

903 Sobol', I.: Global Sensitivity Indices for Nonlinear Mathematical Models:Review, *Math. Comput.*
904 *Simul.*, 55, 271–280, doi:10.1016/S0378-4754(00)00270-6, 2001.

905 Song, X., Zhan, C., Xia, J. and Kong, F.: An efficient global sensitivity analysis approach for distributed
906 hydrological model, *J. Geogr. Sci.*, 22(2), 209–222, doi:10.1007/s11442-012-0922-5, 2012.

907 Song, X., Zhang, J., Zhan, C., Xuan, Y., Ye, M. and Xu, C.: Global sensitivity analysis in hydrological
908 modeling: Review of concepts, methods, theoretical framework, and applications, *J. Hydrol.*,
909 523(225), 739–757, doi:10.1016/j.jhydrol.2015.02.013, 2015.

910 Stephens, E. M., Bates, P. D., Freer, J. E. and Mason, D. C.: The impact of uncertainty in satellite data
911 on the assessment of flood inundation models, *J. Hydrol.*, 414–415, 162–173,
912 doi:10.1016/j.jhydrol.2011.10.040, 2012.

913 Tucker, G. E. and Bras, R. L.: A stochastic approach to modelling the role of rainfall variability in
914 drainage basin evolution, *Water Resour. Res.*, 36(7), 1953, doi:10.1029/2000WR900065, 2000.

915 Tucker, G. E. and Hancock, G. R.: Modelling landscape evolution, *Earth Surf. Process. Landforms*,
916 35(1), 28–50, doi:10.1002/esp.1952, 2010.

917 Vanwalleghem, T., Stockmann, U., Minasny, B. and McBratney, A. B.: A quantitative model for
918 integrating landscape evolution and soil formation, *J. Geophys. Res. Earth Surf.*, 118(2), 331–347,
919 doi:10.1029/2011JF002296, 2013.

920 Welivitiya, W. D. D. P., Willgoose, G. R., Hancock, G. R. and Cohen, S.: Exploring the sensitivity on a
921 soil area-slope-grading relationship to changes in process parameters using a pedogenesis model,
922 *Earth Surf. Dyn.*, 4(3), 607–625, doi:10.5194/esurf-4-607-2016, 2016.

923 Van De Wiel, M. J. and Coulthard, T. J.: Self-organized criticality in river basins: Challenging
924 sedimentary records of environmental change, *Geology*, 38(1), 87–90, doi:10.1130/G30490.1, 2010.

925 Van De Wiel, M. J., Coulthard, T. J., Macklin, M. G. and Lewin, J.: Embedding reach-scale fluvial
926 dynamics within the CAESAR cellular automaton landscape evolution model, *Geomorphology*, 90(3–
927 4), 283–301, doi:10.1016/j.geomorph.2006.10.024, 2007.

928 Van De Wiel, M. J., Coulthard, T. J., Macklin, M. G. and Lewin, J.: Modelling the response of river
929 systems to environmental change: Progress, problems and prospects for palaeo-environmental
930 reconstructions, *Earth-Science Rev.*, 104(1–3), 167–185, doi:10.1016/j.earscirev.2010.10.004, 2011.

931 Wilcock, P. R. and Crowe, J. C.: Surface-based Transport Model for Mixed-Size Sediment, *J. Hydraul.*
932 *Eng.*, 129(2), 120–128, doi:10.1061/(ASCE)0733-9429(2003)129:2(120), 2003.

933 Willgoose, G. R., Hancock, G. R. and Kuczera, G.: A Framework for the Quantitative Testing of
934 Landform Evolution Models, pp. 195–216, American Geophysical Union., 2003.

935 Wong, J. S., Freer, J. E., Bates, P. D., Sear, D. A. and Stephens, E. M.: Sensitivity of a hydraulic model
936 to channel erosion uncertainty during extreme flooding, *Hydrol. Process.*, 29(2), 261–279,
937 doi:10.1002/hyp.10148, 2015.

938 Yang, J.: Convergence and uncertainty analyses in Monte-Carlo based sensitivity analysis, *Environ.*
939 *Model. Softw.*, 26(4), 444–457, doi:10.1016/j.envsoft.2010.10.007, 2011.

940 Ziliani, L., Surian, N., Coulthard, T. J. and Tarantola, S.: Reduced-complexity modeling of braided
941 rivers: Assessing model performance by sensitivity analysis, calibration, and validation, *J. Geophys.*
942 *Res. Earth Surf.*, 118(4), 2243–2262, doi:10.1002/jgrf.20154, 2013.

943