

## Answers to the reviewers

First we are very glad over your positive response, both reviewers agree on the legitimacy of our study and think this addressing a very important but so far neglected parts in ecosystem modeling.

5 The intention and aim of our study was to close the knowledge gap of missing ectomycorrhizal fungi (ECM) in current ecosystem models and compare three modeling approaches of different complexity at explaining plant and soil development across a climate and N deposition gradient. Both reviewers thought the language of the paper was weak. We feel sorry for the grammatical and typographical errors in the previous submission. We have now rewritten the manuscript thoroughly and it has been further edited through professional language edition by  
10 native English speaking person who also has a PhD degree in relevant field, and hope you find this much better.

We further invite associated professor, Annemarie Reurslag Gärdenäs ([annemieke.gardenas@bioenv.gu.se](mailto:annemieke.gardenas@bioenv.gu.se)) who is specialized at ECM, soil microbes and soil  
15 biogeochemical model development as an external reviewer for this paper. She has both detailed comments on the content but also the language (her review reports attached). In the revised version of this paper, we incorporated all of her comments.

Besides, we have now rewritten the description of Bayesian calibration procedure and added more references to make it comprehensive and easier to understand. However, we would like  
20 to emphasize that the paper is to present a new model considering ECM and further compare this to two simpler approaches on explaining the observed data. We employed Bayesian calibration as a common procedure to estimate the parameter uncertainties associated with the 3 different models. Therefore the purpose is surely not only demonstrating how reduction of statistical uncertainty by Bayesian calibration can be made. Also other statistical methods can  
25 be used to demonstrate the same phenomena with respect to the link between parameter uncertainty and model structure uncertainty, providing the same data are used to constrain the model. Of course we agree with the reviewer 2 that a thoroughly and detailed description of what has been used is needed and is now added.

Last but not least, we have thoroughly improved all the figures and tables to make it easier to follow. We added Table 1 for better model comparison. More importantly, we changed  
30 previous Table 4 and Figure 5 into a new Figure 5 to compare our modelling results better with the measured data. Previously only soil C balance was compared, we have now added all the major C cycling variables: GPP, ecosystem respiration, soil respiration, NEE, and also soil N balance (see new Fig. 5). This gives a much more comprehensive comparison of the modeling  
35 approaches and data, which also additionally show large difference of litter addition and soil respiration between the “explicit”/“implicit” approach. These are all additionally included in the result section. We also have made a more thorough discussions on the modeling approaches, uncertainties and possible explanations. The abstract and conclusions are also improved. The information-rich parameter correlation tables (Tables A2, A3 and A4.) are  
40 further moved into supplementary to make the paper more concise and easier to read.

We will here answer the comments raised by the Reviewers, and how these were met by changes in the text and figures (answers are marked with blue).

#### Comments by Reviewer 1

#### 45 **Specific comments**

Introduction Consider putting in a table showing clearly what the different models described on line 48 through 69 do.

50 Table 1 is now added to make it clear and also added Coup-MYCOFON according to external reviewer's comment

Line 128 ECM growth is driven by sink strength of what?

55 Now rewrite the whole section to make it clear.

Line 141 I had to read this sentence twice as I thought the authors were comparing the approach for ECM and root respiration to the approach of something else. however I think they have just treated ECM respiration the same way they have  
60 treated root respiration. Perhaps it would be clearer to say that there are two components (maintenance and growth) for both ECM and root respiration.

Changed accordingly.

65 Line 159 Is  $\text{NUPT}_{\text{FRACMAX}}$  the fraction of total soil N available for uptake, or is it the fraction of mineral N available for uptake? Please clarify.

Clarified.  $\text{NUPT}_{\text{FRACMAX}}$  determines the fraction of mineral N available for uptake.

70 section 2.1.5 My first reaction was that degree of mycorrhization had not been taken into account; then I realised mycorrhization degree was covered in section 2.1.6. Consider switching these two sections.

We agree and switched accordingly

75 Line 166 Please add the scientific name for spruce. As this is Sweden it is probably *Picea abies*.

Added

80 Line 211 I see the point of spinning up the vegetation from the time of establishment over the lifetime of the trees (100 years in this study), but soil C pools may take considerably longer than that to come to equilibrium. For example, 500 years is a more typical spinup to initialise soil C pools in dynamic vegetation models (DVMs). The legacy of recalcitrant C from previous forest growth in the soil must be accounted for by  
85 the initial standing C stock and C/N initial values in Table 2 which the footnote says are calibration parameters; maybe make this clear in the text. Unlike the calibration parameters of Table 1, the initial values assumed for soil C pools shown in Table 2 do not have minimum and maximum values  
90 associated with them, and standing stock does not appear in Table 3.

First, we have now redesigned the Table 3 to show the calibration data and forcing data clearly. The model is constrained by the forest standing biomass and the soil C/N ratio, not the soil C or N pools as such. Previous 100-year simulation by Svensson et al. (2008a) considering the regions and similarly by Berggren Klejal et al. (2007) considering specific representative sites for the same regions, both showed that the model can describe a consistent pattern of C pools and C/N ratios. None of the regions are in a perfect steady state but the difference from a steady state is small and not possible to constrain from measured changes of the soil C pools. Thus adding soil C would not further constrain the parameters. Moreover, Svensson et al. (2008a) demonstrated that the soil C/N ratio showed consistent patterns of different N supply assumptions and expected turnover rate from differences in climate forcing and current C-pools. These previous applications thus provide a base for our current model designs and evaluations. We have described this in more detail and also added why these observational constraints (i.e. soil C/N) are selected in section 2.3.1.

However, we agree that initialization problem of the soil pools exist but this is mostly a general problem independent of the three model approaches. Ideally, the initialization of each soil organic C pool required a spin-up simulation over a longer-term (e.g. 500 years) to find a soil C equilibrium for undisturbed vegetation. After the spin-up using undisturbed vegetation, the reconstructed disturbance history was then used to get a close estimate of the SOC pools. But this two-step method requires informative historical data which are not available in our case. Besides, another significant uncertainty in this spin-up type runs is the initial estimate (500 yrs ago) of inert or very slowly decomposing organic C, which again we do not know.

Thus in our study, we use another approach by following Svensson et al., (2008) and Berggren Klejal et al., (2007) who use the measured current soil C and N content data from similar soils that only different with respect to climate. And they assume that the current soil was in close to equilibrium with respect to C/N ratios for the different regions. Eliasson et al. (2013) investigated the soil balance in Swedish forests over 300 years by different modeling approaches and also found the soil C balance generally reach equilibrium after 100 years. It should be noted that the intention was to evaluate how the ECM affect the C and N cycling in plant-soil over the lifetime of the trees in different regions providing basic assumptions on the carbon pools. Our investigation cannot be used to justify some new suggestion on the current rate of change of soil carbon pools in the different regions.

Line 212 I do not understand what is meant by this sentence: A minimum of specific regional data were used at input values. Does this refer to the number of driving variables input to the model (six in Table 2 plus two calibration parameters) or the amount of data used in the Bayesian analysis for each driving variable (30-year averages rather than time series or multiple values for each region)? I also don't understand at input values; does this mean as input values or does it mean something else? What is specific about the regional data?

We rewrite this to make it clear

Line 230 The data likelihood function which determines the parameter sets **being candidate of** the posterior distribution sounds odd; I assume that this sentence refers to the likelihood function determining **acceptance** of the parameter sets which will comprise the posterior distribution?

Changed accordingly

Line 235 Please make clear that  $\omega_i$  is a vector.

Changed accordingly

Line 244 Replace  $q_i + 1 = q_i + \epsilon$  " with  $\theta_i + 1 = \theta_i + \epsilon$ , using the same  $\epsilon$  on lines 244 (the equation) and 245. Also, consider numbering the equations.

Changed accordingly

150 Line 280 Surely it is just parameters that are being calibrated and not processes?

Removed “processes”

155 Line 306 Do fungi take up the same amount of organic N when there is sufficient mineral N available?

No, the uptake is both driven by demand and by N availability, and in our case, mineral N will first regulate the mineral uptake and reduce the demand of organic N. This is describe in detail in the 2.1.5

160 Line 333 should thus be and? Is the sentence referring to N mineralisation? A higher organic matter turnover should mean higher N mineralisation.

Changed accordingly

165 Section 3.2.3 Is it necessary to list all these correlations? The figures are better for this; perhaps only discuss the most interesting ones?

We have now moved this detailed correlation tables into supplementary files, Fig. 9 show the important and interesting ones.

170 Section 4 I take the authors’ point that there is a dearth of comparison data, especially related to ECM, but are there really zero data? is there not one observation that can be compared with the model results? What about the Lindroth et al paper cited on line 419? How does the coupling of Mycofon to CoupModel affect the simulated soil respiration, for example? it is a bit difficult to claim that the model delivers “accurate” results (line 464) without any comparison to observations. Table 4 shows the Svensson et al model results so consistency with this other model could be worth showing in a figure.

180 We have now added a new figure 5 to make this clear. The simulated regions do not have detailed measured data. In previous papers the carbon balance have been compared with the data of eddy flux measurements from some few years of each site (Svensson et al., 2008a, b). We compare the nearby sites that have been intensively measured by eddy covariant technology. We both compare the soil respiration, total respiration, GPP, NEE and also change of soil C and soil N.

## 185 2 General remarks on figures

Please include units and self-explanatory axes labels in all figures. Many readers will look at the abstract and figures before deciding to read the text; don’t make readers go searching through the text for basic information. Where possible, don’t even make readers read the captions carefully. In general, don’t make readers do more work than absolutely necessary to understand what is being shown in the figures.

We have now redesign the figures and take all of these into consideration

195 Figure 3 Is total N litter production the N released during decomposition or the N being added to the litter pool with fresh litter?

200 N total litter production is the total N in litter being added to the soil litter pool by fresh litter, we have added this in the caption to make it clear

Figure 4 There is room to add implicit model and explicit model to the right of the figure so that readers can see immediately what the upper and lower graphs mean.

Added accordingly

Figure 5 Is GPP in this figure simulated or measured? Any possibility of showing both simulated and measured GPP?

This is modelled GPP and now the new Figure 5 also include measured GPP data from nearby sites with similar conditions which are comparable to our study. We further also compare our results to Svensson et al., (2008a)

Figure 6 There is room to add nonlim, implicit and explicit to the right of the three panels, and to show the north-south gradient to the left of the Y axis of each panel.

Added accordingly

Figure 7 Show the N-S gradient to the left of the panels (ie N next to Ly, and an arrow leading to S next to Lj). Thanks for adding implicit and explicit; please also add the meanings of the parameters on the X axis (eg KH is the humus decomp. coeff.) so that readers can see at a glance what is going on without having to search the text and tables.

Added accordingly

Figure 8 Please give the units, especially for the rates. What is fungal litter rate, the rate of uptake from litter, or the rate at which hyphae die and contribute to the litter pool?

Added now and the Fungal litter rate refers to the rate at which hyphae die and add to the soil litter pool, we have added this in caption to make it clear

Figure 9 Does C assimilates mean NPP? Please make clear what parameters are being shown, so readers don't have to go searching (they probably won't have read the paper and won't realise the information is in one of the tables). Is the colour scheme here the same as in previous figures?

C assimilates refers to the C taken up by the plant, so GPP, although of course respiratory losses and litter are subtracted later so that only net growth remains. We have now added this in figure caption and also show the meaning of the parameters briefly in the figure to make the reader easier to follow. The color scheme are the same in all the figures. We rephrase the figure caption to make this clear.

### 3 Tables

Table 2 Can it be made clearer that soil C/N and standing stock of C are calibration parameters and the other data are all driving data?

Now redesigned to make it clearer

Table 3 Why are there no mean and uncertainty columns for soil C standing stock?

according to Table 2 it's a calibration parameter.

You misunderstood here and only soil C/N is a calibration parameter, not soil C stock. See above and also details are added in section 2.3.1

Table 4 The Lindroth et al data shown here are means of the highest and lowest estimates, but the full ranges are shown for the Mycofon results. Would it not be better to show ranges for both?

We have added these and integrate previous figure 5 and Table 4 into the new Figure 5 which show more detailed ecosystem C processes, model data comparison.

Table 4 The Svensson et al data generally fall within the Mycofon model ranges; are these the ranges from the posterior distributions? why is the implicit approach shown for one site and the explicit approach for the other, and what are the results for the mean of the posterior? Could this material (Svensson et al. vs. model approaches) be presented as a figure? If Lindroth et al. measured respiration, surely that is a CoupModel output which could be compared to those measurements?

We have included a new Figure 5 which includes the measured data from Lindroth et al. (2008) also Svensson et al. (2008a). However it should be noted that Svensson et al. (2008a) data are results from single model simulation established by subjective calibration and not from an ensemble approach, like in our study.

#### 4 Grammatical or typographical errors

Here is a partial list of lines with errors, including suggested corrections.  
In some cases I suggest rewordings of awkward clauses, in others I try to show the grammatical/typographical error and how to fix it. Original text is to the left of the arrow, and the replacement text to the right of the arrow. Actual changes (deletions to the left of the arrow, additions to the right of the arrow) are in **boldface**. I have tried to include enough text to make it clear why the change is necessary, such as where a grammatically plural noun is coupled with a grammatically singular verb. Generally, models and approaches are preceded by the, which is omitted repeatedly throughout the text. A global change is not possible because there are a few occasions where the is present, or where it is OK to leave it out.

We have now rewritten the paper thoroughly and include all the following into revision. The language was also improved by British language editing companies. We believe the language of paper is now significantly improved.

27 ... Coup-Mycofon model provide ! Coup-Mycofon model provides  
43 **known as ! which are**  
46 **the** ecosystem ! ecosystem  
48 research show ! research shows  
56 Moore ! **the** Moore  
60 ANAFORE ! **the** ANAFORE  
68 ECM models ... simulates ! ECM models ... simulate  
70 that coupled ! that **is** coupled  
78 approach which ! approaches **es** which  
79 The "ECM implicit" does not ! The "ECM implicit" **approach** does not  
79 incorporating ! incorporates  
80 Plants ... **does** not ! Plants ... **do** not  
100 **in** Meyer ! **by** Meyer (NB this is my personal preference but check the journal's

policy: are citations considered to be the name of the paper, in which case in is fine, or do they refer to the authors who wrote the paper, in which case by makes more sense?)

315 118 **are distinguished** between ! **distinguish** between  
 131 follow ! follows  
 132 **to prevent fungi to die ! preventing fungal death**  
 159 ... as fungi **have** are more efficient ! as fungi are more efficient  
 193 plant **uptaking of** organic N ! plant **taking up** organic N  
 320 203,213 **Tab. 2!Table 2** (likewise Table 3 in section 3.1.1; check the journal's policy, but in any case be consistent as Table is spelled out earlier in the manuscript)  
 205 managements ! management  
 206, see 100 in Svensson ! by Svensson  
 212 effects **is** not ! effects **are** not  
 325 237 both, the ! both the  
 239 **for a better constrain of** posterior ! **to better constrain** posterior  
 241 using the **Markov chain Monte Carlo method, also the Metropolis-Hastings walk ! using the Metropolis-Hastings random walk Markov Chain Monte Carlo algorithm** (and please cite van Oijen et al 2005 here too)  
 330 245 The **random numbers are generated normally distributed having** a mean of zero ! The **normally distributed random numbers \_ have** a mean of zero  
 252 **parameter: ConstantNsupply for the spruce tree, is selected as calibration parameters ! parameter ConstantNsupply for the spruce tree is a calibration parameter**  
 335 257 (C/Nmyc), ! (C/Nmyc)  
 268 The posterior model ... show ! The posterior model ... shows  
 272 than that **of** using the “implicit” **and** “explicit” approach ! than that using the “implicit” **or** “explicit” approach  
 275 generally **N more** limited ! generally **more N** limited  
 340 277 southern site, Ljungbyhed than ! southern site, Ljungbyhed, than  
 278 **show overestimation** by “implicit” approach but **change to underestimation** when “explicit” approach is used ! **is overestimated** by **the** “implicit” approach but **coloredunderestimated** when **the** “explicit” approach is used  
 345 281 **the more processes and parameters included for calibration, less likely of finding an accept combination of parameter sets ! as more parameters are included for calibration, acceptable combinations of parameter sets become less likely**  
 286 approach **show a much larger uncertainties than that of ECM “implicit” and “explicit” approaches ! approach shows much larger uncertainties than either the “implicit” or “explicit” approaches**  
 350 287 approach **simulate soil sequestration of N up to 2 g N m<sup>-2</sup> y<sup>-1</sup> ! approach simulates up to 2 g soil N m<sup>-2</sup> y<sup>-1</sup>**  
 292 **Besides the simulated soil C balance by “nonlim” approach ! The simulated soil C balance by the “nonlim” approach**  
 355 293 the soil **sequesterate C at most north site, Lycksele** but ! the soil **sequesters C at the most northerly site, Lycksele**, but  
 294 and decoupled ! and **are** decoupled  
 297 and “implicit” approach ! and the “implicit” approach  
 297 **sites overall loss soil C by 6 and 5 g C m<sup>-2</sup> y<sup>-1</sup> ! soils lose 6 and 5 g C m<sup>-2</sup> y<sup>-1</sup>, respectively**  
 360 298 **sites gain soil C by 3 and 13 g C m<sup>-2</sup> y<sup>-1</sup> ! soils gain 3 and 13 g C m<sup>-2</sup> y<sup>-1</sup>, respectively**  
 299 For “explicit” approach ! For **the** “explicit” approach  
 300 in “implicit” approach ! in **the** “implicit” approach  
 365 301 show **an** overall minor C and N losses ! show overall minor C and N losses  
 305,306 in “explicit” model ! in **the** “explicit” model

309 using “implicit” approaches ! using the “implicit” approaches  
 310 favour climate ! favour~~able~~ climate  
 311 but “explicit” approach show a ! but ~~the~~ “explicit” approach shows a  
 370 312 in “explicit” approach ! in the “explicit” approach  
 314 ~~show explicitly account~~ for ECM!~~shows that explicitly accounting~~ for ECM  
 326 except a larger uncertainties in the “explicit” . ! except ~~for~~ larger uncertainties  
 in the “explicit” ~~approach~~.  
 327 than ~~that of~~ the southern ! than ~~for~~ the southern  
 375 The rest of the manuscript is riddled with errors like the ones above; please go through  
 and fix them.

380 Anonymous Referee #2

The authors coupled an ectomycorrhizal fungi (ECM) model MYCOFON with a terrestrial  
 biogeochemistry model to show that it is important to consider the plant-ECM  
 interaction to properly model the ecosystem nitrogen dynamics. While I could see the  
 385 legitimacy of their statement, I agree with the other reviewer that the paper seems submitted  
 in a hurry: there are too many problems with grammars, syntaxes and formats,  
 making it unreadable to some extent. Thus a thorough rewritten is needed before it  
 can be better judged.

390 [The paper has been rewritten thoroughly. The language was also edited by British language edition  
 services. Please also see answers above.](#)

The language problem becomes more severe as the paper goes closer to the end. For  
 instance, the description of 2.3.2 is pretty much a mess. I guess it is really awkward that  
 395 a paper would use “Bayesian calibration procedure” as section title. Personally, I think  
 “Bayesian calibration of models” would be much more appropriate. The use of “data  
 likelihood function” is also not accordant with the general terminology in data assimilation  
 or Bayesian inference based model calibration. I strongly suggest the authors to  
 read more relevant papers and revise the description to make it more readable.  
 400 As for the description of MCMC method, there are many excellent papers on this topic,  
 however, the authors barely mention them and the description is again very poor.  
 As the paper reaches the results section, there are many more language/presentation  
 problems. Many of the sentences are incomplete, such as missing verbs or wrong use  
 of juxtapositions. The other reviewer has listed many of those issues and I won’t add  
 405 more to the list.

[We have now rewritten the entire description of Bayesian calibration \(section 2.3\) and added more  
 references to make it comprehensive and easier to understand. Specifically, we describe the  
 observational constraints more clearly and justify the reason for select plant biomass and soil C/N as  
 410 the accepted criteria. We add more text explaining the defined measured error. The sub-sections:  
 parameters chosen for calibration and Bayesian calibration of models are switched in position in order  
 to be followed easily by the reader. Besides, we have described the MCMC algorithm and Bayesian  
 method in much more detail, also more literature for comparison.](#)

415 [Again, we would like to emphasize that the paper is to present a new model considering ECM and  
 further compare this to two previous simpler approaches of explaining the observed data. We employed  
 Bayesian calibration to estimate the parameter uncertainties but the purpose is surely not only  
 demonstrating the reduction of statistical uncertainty by Bayesian Calibration as such.](#)

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Further, I don't know why Appendix is shown in the middle of the paper. Have the authors carefully checked their submission? Is the wrong version uploaded?  
Overall I suggest rejecting the paper for a resubmission.

We do agree that the long appendix might reduce the readability and also consider comments from reviewer 1 that parameter correlation tables (Table A2, A3 and A4) might not be needed in that detail. So we now move those into supplementary and substantially reduce the appendix. However, the equations and explanation of the parameters can be helpful for the reader to get into details of the model buildup.

Our paper overall presents a new ecosystem model (version) that can explicitly include ECM, where so far the other models cannot. Modeling comparison also clearly demonstrates the importance and legitimacy of incorporating ECM into ecosystem models. Of this, both reviewers agree on. Again we are sorry for the language issue but we believe the language has been largely improved in the revised version.

## **Interactive comment on “Simulating ectomycorrhiza in boreal forests: implementing ectomycorrhizal fungi model MYCOFON into CoupModel (V5)” by Hongxing He et al.**

**L. Gross**

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As explained in [https://www.geoscientific-model-development.net/about/manuscript\\_types.html](https://www.geoscientific-model-development.net/about/manuscript_types.html) GMD is expecting that authors upload the program code of models and the used data sets as a supplement or make the code and data available at a data repository preferable with an associated DOI (digital object identifier) for the exact model version described in the paper. If for some reason your code and/or data for the MYCOFON model cannot be made available in this form as the code availability section in your paper suggests you need to state the reasons why the code is not available or why access is restricted. Please note that in the code accessibility section you can still point the reader to your web site for updates even if you provide the code as supplement or use a DOI for a release.

All the best Lutz Gross GMD Executive Editor

Dear editor,

Thanks for the comment, we will now add the model software and version. Mycofon-CoupModel is derived from CoupModel with the implementation presented in section 2 here. The general code of CoupModel will be made available from [www.coupmodel.com](http://www.coupmodel.com), which means freely available for everyone after registration. The general information about how to install CoupModel and its different branches and tutorials are also made available from [www.coupmodel.com](http://www.coupmodel.com). The manuscript has been updated with these information and the new version used would also be specified in the section code availability.

475

**References mentioned**

480

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# Simulating ectomycorrhiza in boreal forests: implementing ectomycorrhizal fungi model MYCOFON into CoupModel (V5)

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## Abstract

Ectomycorrhizal fungi (ECM), the symbiosis between a host-plants and Ectomycorrhizal fungi (ECM) mycorrhizal fungi, has been shown to considerably influence the carbon (C) and nitrogen (N) fluxes between the soil, the rhizosphere, and plants in boreal forest ecosystems. However, ECM is either neglected or presented as an implicit, non-dynamic term in most ecosystem models which can potentially reduce the predictive power of models.

In order to investigate the necessity of an explicit consideration of ECM in ecosystem models, we implemented the previously developed MYCOFON model into a detail process-based soil-plant-atmosphere model, CoupModel-MYCOFON, which explicitly describes the C and N fluxes between ECM and roots. This new Coup-Mycofon-MYCOFON model approach (ECM explicit) is compared to two simpler model approaches, of which one contains ECM implicitly as an non-dynamic N uptake function a dynamic uptake of organic N considering the plant roots to represent the ECM (ECM implicit), and the other represents a static N approach version w where plant growth has a constant N availability is limited to a fixed N level (nonlim). Parameter uncertainties are quantified by using Bayesian calibration where the model outputs are constrained to current forest growth and soil C/N ratio conditions for four forest sites along a climate and N deposition gradient in Sweden and simulated over a 100-100-year period.

Our results show that the “nonlim” approach could not describe the soil C/N ratio, due to largely overestimation of soil N sequestration but simulate could does not describe both the forest growth reasonably well and soil C and N conditions properly. The ECM “implicit”/“explicit” approaches are both able to describe current conditions with acceptable uncertainty the soil C/N ratio well but slightly underestimate the forest growth. The “implicit” approach simulated lower litter production and soil respiration than the “explicit” approach. Meanwhile, the ECM “explicit” Coup-Mycofon model provides a more detailed description of internal ecosystems fluxes and feedbacks of C and N fluxes between plants, soil and ECM. Our modelling highlights the need of incorporating ECM and organic N uptake into current ecosystem models, and the “nonlim” approach is not recommended for future long-term soil C and N predictions. We also provide a key set of posterior fungal parameters which can be further investigated and evaluated in future ECM studies.

## 1. Introduction

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Boreal forests cover large areas on the Earth's surface and are generally considered as substantial carbon (C) sinks (Dixon et al., 1994; [Pan et al., 2011](#)). The sink strength is determined through the balance between the major C uptake and release processes, i.e., plant photosynthesis and both autotrophic and heterotrophic respiration, and is largely controlled by nitrogen (N) availability (Magnani et al., 2007; [Högberg et al., 2017](#)). For instance, numerous studies have shown that soil nitrogen-N availability is the main driver for plant and microbial growth dynamics ([Vitousek and Howarth, 1991](#); [Klemetsson et al., 2005](#); [Lindroth et al., 2008](#); [Luo et al., 2012](#); [Mäkiranta et al., 2007](#); [Martikainen et al., 1995](#)). Thus, a proper description of N dynamics in ecosystem models is the prerequisite for precisely simulating plant-soil C dynamics and greenhouse gas (GHG) balance ([Maljanen et al., 2010](#); [Schulze et al., 2009](#); [Huang et al., 2011](#)). Ecosystem models, however, vary considerably in their representation of N fluxes: from a very simplified presentation (e.g., the LPJguess model: [Sitch et al., 2003](#); [Smith et al., 2011](#)) to very complex approaches which aim to capture the whole N cycle (e.g., LandscapeDNDC: [Haas et al., 2012](#); CoupModel: [Jansson and Karlberg, 2011](#)).

Ectomycorrhizal fungi (ECM) are common symbionts of the trees in boreal forests. ECM, which are more efficient than roots in taking up different N sources from the soil ([Plassard et al., 1991](#)), as well as store vast amounts of N in their tissues ([Bååth and Söderström, 1979](#)), and can cover a large fraction of their host plants' N demand ([Leake, 2007](#); [van der Heijden et al., 2008](#)). Further, ECM are shown to respond sensitively to ecosystem N availability and are generally considered as an adaptation measure to limited N conditions ([Wallenda and Kottke, 1998](#); [Read and Perez Moreno, 2003](#); [Kjoller et al., 2012](#); [Bahr et al., 2013](#); [Choma et al., 2017](#)). Previous research showed that ECM can receive between 1 and 25% of the plants' photosynthates and constitute as much as 70% of the total soil microbial biomass, thus having a major impact on the soil C sequestration in boreal forests ([Staddon et al., 2003](#); [Clemmensen et al., 2013](#)). Overall, the functions and abundance of ECM fungi constitute numerous pathways for N turnover in the ecosystem and considerably influence the magnitude and dynamics of C and N fluxes.

However, ECM are so far rarely been considered in ecosystem models (for an overview about modelling ectomycorrhizal traits see [Deckmyn et al., 2014](#)). To our knowledge, only five ecosystem models have implemented ECM to a various degree: The ANAFORE model ([Deckmyn et al., 2008](#)), the MoBiLE environment ([Meyer et al., 2012](#)), the MyScan model ([Orwin et al., 2011](#)) and more recently the Moore et al. (2015) and Baskaran et al. (2016) ECM models ([Table 1](#)). In the ANAFORE model, ECM are described as a separate C and N pools. However, this model does not distinguish between mycorrhizal hyphae and mantle. The C allocated from the host tree to ECM is simulated as a zero order function, further regulated by nutrient and water availability. ECM can also facilitate organic matter decomposition in the ANAFORE model. The MyScan model uses a similar approach for ECM C uptake and dynamics but does not, to our knowledge, include the influence of water availability on ECM. In both models, ECM transfer of N to the host is regulated by the C/N ratios of the plant and fungi. In the MoBiLE model, C allocation to ECM is more complex than that in ANAFORE and MyScan models, and the N allocation to the host by the ECM can feed back into their C gains. Although, the N allocation to the host plant is described similarly as to the other two models. In MoBiLE, mycorrhiza are further distinguished between hyphae and mantle, but cannot neither degrade organic matter nor take up organic N forms. Hyphae and mantle differ in their capacity to take up N, and the mantle has a slower litter production rate than that of hyphae. Both Moore et al. (2015) and Baskaran et al. (2016) ECM models

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represent the ECM as a separate model pool and ~~also~~ explicitly simulate ECM decomposition, but with much ~~more~~ simpler process descriptions, and the interaction with environmental functions are neglected (Table 1).

The overall aim of this study is to improve understanding of ecosystem internal C and N flows related to symbiosis between ECM and host tree, in order to improve the model predictive power in assessment of C sequestration and climate change. This is done by presenting a new version of the CoupModel, that is coupled with an explicit description of ECM, and also to investigate how the explicit consideration of ECM affects the overall model performance and model uncertainty. We Thus, wSpecifically, we thus implement ed the previously developed MYCOFON model (Meyer et al., 2010); ~~Meyer et al., 2012~~ into the well-established soil-plant-atmosphere model, CoupModel (Jansson, 2012). The implementedWe choose the MYCOFON model because; first, it contains a very detailed description of ECM fungal C and N pools, and all major C and N ECM exchange processes (i.e., litter production, respiration, C uptake, N uptake), and second, ECM can also additionally responses to the soil N availability (Table 1). Fungal-Therefore, ECM growth and N uptake, both mineral and organic N forms, respond dynamically to environmental functions and plant C supply in the new Coup-MYCOFON model (Fig. 1). This detailed ECM explicit modelling approach (~~hereafter hereafter~~ called “ECM explicit”) is further compared with two simpler modelling approaches: ~~the~~ “ECM implicit” and “nonlim” approaches ~~which already exist in CoupModel. The “ECM implicit” approach does not represent the ECM as a separate pool but incorporates ECM into the roots implicitly. Plants are thus allowed to take up additional organic N sources, statieally from soil organic pools, and do not respond to environmental functions. SimilarThe “ECM implicit” approach was usedhas been previously used in a similar way in by Kirschbaum and Paul, (2002) and Svensson et al. (2008a). The “nonlim” approach assumes an “open” N cycle and plant growth areis limited by a constant N availability thus to a static fixed level (e.g., in Franklin et al., (2014)). These three ECM modelling approaches represent-constitute most of the current ECM representations in ecosystem models, and are tested by four forest sites situated along a climate~~ and N fertility gradient across Sweden (Fig. 2). Bayesian calibration is used to quantify the uncertainty of model parameters and identify key parameter sets.

## 2. Data and Methodologies

### 2.1 Model description

The CoupModel (“Coupled heat and mass transfer model for soil-plant-atmosphere systems”, Jansson and Karlberg, 2011) is a one-dimensional process-orientated model, simulating all the major abiotic and biotic processes (mainly C and N) in ~~the soil-plant-atmosphere-systemterrestrial ecosystem~~. The basic structure is a depth profile of the soil for which water and heat flows are calculated based on defined soil properties. Plants can be distinguished between ~~understoreyunderstory~~ and overstorey vegetation, which allows simulating competition for light, water, and N between plants. The model is driven by ~~measured~~ climate data: ~~precipitation, air temperature, relative humidity, wind speed, and global radiation~~, and can simulate ecosystem dynamics in hourly/daily/yearly resolutions. A general structural and technical overview of the CoupModel can be found in Jansson and Moon (2001) and Jansson and Karlberg (2011), ~~and~~ ~~A~~ recent overview of the model was also given by Jansson (2012). The model is freely available at [www.coupmodel.com](http://www.coupmodel.com). ~~The CoupModel (V5) was-is~~ complemented with an ectomycorrhizal module (MYCOFON, Meyer et al., 2010) which allows ~~to directlythe direct simulate-simulation ofthe the~~ C and N uptake processes of ECM. The MYCOFON model is described in detail by Meyer et al. (2010).

and here only the key processes of plant and ECM fungal growth, N uptake as well as litterfalling and respiration are described.

### 2.1.1 Plant growth in CoupModel

An overview of model functions is given in Table A.1 in the Appendix Table A.1. Plant growth is simulated according to a “radiation use efficiency approach” where the rate of photosynthesis is assumed to be proportional to the global radiation absorbed by the canopy, but limited by temperature, water conditions, and N availability (eq. 1, Table A.1(a)). Assimilated C is allocated into five main-different plant C compartments:  $C_{\text{root}}$ ,  $C_{\text{leaf}}$ ,  $C_{\text{stem}}$ ,  $C_{\text{grain}}$ , and  $C_{\text{mobile}}$ . The same compartments also represent the corresponding N amounts. The “mobile” pool ( $C_{\text{mobile}}$ ,  $N_{\text{mobile}}$ ) contains embedded reserves which are reallocated during certain time periods of the year, e.g., during leafing. Respiration is distinguished between maintenance and growth respiration, where a  $Q_{10}$  function response was used, respectively for maintenance respiration (eqs. 2.1, 2.2, Table A.1(a)). Plant litter is calculated as fractions of standing biomass (eq. 3, Table A.1(a)).

### 2.1.2 ECM Fungal C and N pools

The ECM are closely linked to the trees' fine roots and consist of a C and N pools. The C pool is distinguished between the mycelia, which are responsible for N uptake, and the fungal mantle, which covers the fine roots tips. The C pool is the difference between C gains by supply from the plant supply and C losses due to respiration and litter production (eq. 8.1, Table A.1(b)). Accordingly, the fungal N pool is the result of the difference between N gains by uptake, and N losses by litter production, and N transfer to the plant (eq. 8.2, Table A.1(b)). ECM Fungal C and N pools distinguish between mycelia and mantle which is of importance for when simulating N uptake (only the mycelia is able to take up N), and also when simulating litter production if the more complex approach for simulating fungal litter production is chosen (see section 2.1.4). The ratio between mycelia and mantle is determined by the parameter  $FRAC_{\text{MYC}}$  which defines the fraction of mycelia C in total ECM fungal C. For all other N and C exchange processes (growth, respiration, and N transfer to plant), the separation between mycelia and mantle is disregarded.

### 2.1.3 Growth of ectomycorrhizal fungi

ECM growth is limited by a defined maximum, i.e., only a certain amount of tree host assimilates will be directed to the ECM. This maximum ECM growth is determined by a potential C supply from the plant, and limited by N availability (eq. 5.1, Table A.1 (b)). The C supply is defined by a constant fraction of the root C gain and is leveled off by the function  $f(C_{\text{fungiavail}})$  as soon as a defined value of soil available total N is exceeded; i.e., in the model the potential ECM growth declines with rising soil N. This scaling function is based on observations from field and laboratory experiments, which showed that the ECM biomass of mycelia and mantle can be as much as 30-50% of fine root biomass, and the majority of ECM decreases in abundance and functioning when the soil N levels are high (e.g., Wallander, 2005; Wallenda and Kottke, 1989; Höglberg et al., 2010). This is defined according to the results from field and laboratory studies that the ECM biomass of mycelia and mantle can be as much as 30-50% of fine root biomass. Besides, ECM growth is driven by sink strength (see overview by Smith and Read, 2008). The actual ECM growth is limited by the maximum growth and calculated by a pre-defined fraction of assimilated root C, assuming that the production of an optimum mycorrhization degree requires

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650 a certain amount of ECM biomass (eq. 5.2, Table A.1 (b):  $FRAC_{OPT} * C_{IT}$ ). Therefore, we use these observations  
as the bases for the calculation of the actual ECM growth: i.e., the model aims to grow ECM biomass to a  
certain fraction of fine root biomass (eq. 5.2, Table A.1 (b):  $FRAC_{OPT} * C_{frr}$ ). This is further dependent on the N  
supply from the ECM to the roots,  $f(n_{supply})$ . The model thus follows the assumption that plants feed the ECM with  
C as long as their investment is outweighed by the their benefits obtained (Nehls et al., 2008). A minimum C  
supply to preventing ECM fungi death during C shortage is guaranteed by the term during time periods when plant  
photosynthesis is limited, and belowground C supply to root and ECM becomes zero (eq. 5.3, Table A.1 (b)). The  
655 C supply is defined by a constant fraction of the root C gain and reduced by the function  $f(e_{fungiavail})$  as soon as a  
defined value of soil available N is exceeded; i.e., in the model the potential ECM growth declines with rising  
soil N. This scaling function is based on observations from field and laboratory experiments, which showed that  
the majority of ECM decreases in abundance and functioning when the soil N levels are high (e.g., Wallander,  
2005; Wallenda and Kottke, 1989; Höglberg et al., 2010).

#### 660 2.1.4 Respiration and litter production of ectomycorrhizal fungi

Respiration is separated into two components (maintenance and growth) for both ECM and root respiration. The  
Same approach is used here for ECM and root respiration simulation, and is distinguished between  
maintenance and growth respiration, respectively (see eq. 2 and eq. 6, Table A.1). Two approaches are available  
to simulate ECM fungal litter production, which differ in complexity. The simple approach (eqs. 7.1, 7.2, Table  
665 A.1) uses one common litter rate  $L$  for both, the fungal mantle and the fungal mycelia. Consequently, possible  
specific effects of the mantle and mycelia tissue on litter production are neglected. The alternative “detailed”  
approach (eqs. 7.3, 7.4, Table A.1) has specific litter rates for ECM mantle and mycelia ( $L_M$ ,  $L_{MYC}$ ). This set-up is  
recommended when investigating different biomass ratios between mycelia and mantle and their effects on overall  
litter production. The fraction between ECM mantle and mycelia is determined by the parameter  $FRAC_{MYC}$ .  
670 Irrespective of the approach used for litter production, ECM have the capability to retain a defined amount of N  
during senescence (eqs. 7.2, 7.5, Table A.1 (b):  $nret_{fungi}$ ). In this study, the simple approach was applied.

#### 675 2.1.5 Plant mycorrhization degree, plant N uptake, and ECM fungal N transfer to plant

According to field investigations, the mycorrhization degree can vary considerably between species. For spruce  
(*Picea abies*), typical mycorrhization degrees of over 90% have been reported (Fransson et al., 2010; Leuschner,  
2004). The impact of the ECM mantle on fine root nutrient uptake has been controversially discussed, but the  
majority of studies indicate that the root is isolated from the soil solution; i.e., the nutrient uptake is hampered so  
that the plant is highly dependent on ECM supplies (Taylor and Alexander, 2005). Therefore, the mycorrhization  
degree is of major importance when plant-ECM-soil N exchange and plant nutrition are of interest. In the explicit  
680 Coup-MYCOFON model, mycorrhization degree is calculated as the ratio between ECM C pool and the defined  
optimum ECM C pool, divided by the defined optimum mycorrhization degree (eq. 9, Table A1 (b)). It should be  
noted that the optimum mycorrhization degree needs to be defined with care as there is often a discrepancy  
between the applied root diameter in experimental studies and models: in experiments, mycorrhization degrees  
usually refer to fine roots  $\leq 1$  mm, whereas models often consider fine roots as roots with a diameter of up to 2  
685 mm.

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690 The mycorrhizal mantle has an impact on the mineral plant N uptake. This is because plant ammonium and nitrate uptake is largely driven by the plant N demand (eqs. 4.1, 4.2, Table A.1), but also regulated by the N availability function (eqs. 15, 16, 17, Table A.1:  $f(n_{avail})$ ,  $f(n_{humavail})$ ) based on the assumption that only a certain fraction of soil ammonium and nitrate is available for plant uptake. The ECM fungal mantle reduces this availability in such a way that reduction is highest at maximum biomass. In a balanced symbiosis, the fungus provides nutrients to the plant in exchange for the plant's C supply. In the Coup-MYCOFON model, the amount of ECM fungal N transferred to the plant is determined by either the plant N demand or, if the plant N demands exceeds the ECM fungal capacity, the available fungal N (eqs. 10.1, 10.2, Table A.1). This is the amount of "excess" N which is available after the ECM have fulfilled their defined minimum demand as calculated by the fungal C/N ratio (eq. 10.2, Table A.1). This relation is again based on the assumption that the ECM fungi will only supply the plant with N as long as its own demand is fulfilled (Nehls et al., 2008).

695 ECM can take up both, mineral and organic N. For both N forms, a the potential fungal uptake is first defined at first. This, which is determined by the size of fungal C pool, the fraction of fungal C which is capable to of take up N uptake (the mycelia,  $FRAC_{MJC}$ ), and an uptake rate ( $NO_3RATE$ ,  $NH_4RATE$ ,  $NORG_{RATE}$  (eqs. 11.1, 11.3, 11.4, 11.6, Table A.1 (b)). This function is based on the assumption that only the fungal mycelia can take up N. Values for  $NO_3RATE$ ,  $NH_4RATE$ , and  $NORG_{RATE}$  were are derived from published values (Table 1). The actual N uptake is dependent on the available soil N as well as the fungal N demand (eq. 11.2, Table A.1). The N availability function  $f(n_{avfungi})$  determines the fraction of soil N which is available for fungal uptake, and is controlled by the parameters  $NUPT_{ORGFACMAX}$  and  $NUPT_{FRACMAX}$ . N availability for fungi corresponds to the plant available N (eq. 16, Table A.1), but as fungi are more efficient in the uptake of nutrients, the availability is enhanced for both mineral and organic N (eqs. 17.1, 17.2, 17.3, Table A.1). To prevent the fungal N demand to be being covered by only one N form only, the parameters  $f_{NO_3}$ ,  $f_{NH_4}$ ,  $f_{LIT}$ , and  $f_{HUM}$  are included, and corresponding to the ratio of nitrate and ammonium in total available soil N. If the potential N uptake exceeds the available soil N, the actual uptake corresponds to the available N (eq. 11.2 and eq. 11.5, Table A.1 (b)).

## 710 2.1.6 Ectomycorrhizal fungal N uptake Plant mycorrhization degree, plant N uptake, and fungal N transfer to plant

According to field investigations, the mycorrhization degree can vary considerably between species. For spruce, typical mycorrhization degrees of over 90% have been reported (Fransson et al., 2010; Leuschner, 2004). The impact of the ECM mantle on fine root nutrient uptake has been controversially discussed, but the majority of studies indicate that the root is isolated from the soil solution, i.e., the nutrient uptake is hampered so that the plant is highly dependent on ECM supplies (Taylor and Alexander, 2005). Therefore, the mycorrhization degree is of major importance when plant-ECM soil N exchange and plant nutrition are of interest. In the explicit Coup-MYCOFON model, the mycorrhization degree is calculated as the ratio between ECM C pool and the defined optimum ECM C pool, multiplied by the defined optimum mycorrhization degree (eq. 9, Table A.1 (b)). However, the optimum mycorrhization degree needs to be defined with care, as there is often a discrepancy between the applied root diameter in experimental studies and models: in experiments, mycorrhization degrees usually refer to fine roots  $\leq 1$  mm, whereas models often consider fine roots as roots with a diameter of up to 2 mm. The mycorrhizal mantle has an impact on the mineral plant N uptake. Generally, plant ammonium and nitrate uptake is regulated by the plant N demand (eqs. 4.1, 4.2, Table A.1). The actual uptake is estimated by the N

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725 availability function (eqs. 15, 16, 17, Table A.1:  $f(n_{avai}), f(n_{minhumavai})$ ) based on the assumption that only a certain  
 730 fraction of soil ammonium and nitrate is available for plant uptake. The fungal mantle reduces this availability in  
 such a way that reduction is highest at maximum biomass. In a balanced symbiosis, the fungus provides nutrients  
 to the plant in exchange for the plant's C supply. In the Coup-MYCOFON model, the amount of fungal N  
 transferred to the plant is determined by either the plant N demand or, if the plant N demands exceeds the fungal  
 capacity, the available fungal N (eqs. 10.1, 10.2, Table A.1). This is the amount of "excess" N which is available  
 after the ECM have fulfilled their defined minimum demand as defined by the fungal C/N ratio (eq. 10.2, Table  
 A.1). This relation is based on the theory that the fungi will only supply the plant with N as long as its own demand  
 is fulfilled (Nehls et al., 2008). In the Coup-MYCOFON model, ECM can take up both mineral and organic N.  
 For both N forms, the potential ECM uptake is first defined. This is determined by the size of ECM C pool, the  
 735 fraction of ECM C which is capable of N uptake (the mycelia,  $FRAC_{MYC}$ ), and an uptake rate ( $NO3_{RATE}$ ,  $NH4_{RATE}$ ,  
 $NORG_{RATE}$  (eqs. 11.1, 11.3, 11.4, 11.6, Table A.1 (b)). This function is based on the assumption that only the ECM  
 fungal mycelia can take up N. Values for  $NO3_{RATE}$ ,  $NH4_{RATE}$ , and  $NORG_{RATE}$  are derived from published values  
 but with wide ranges (Table 2). The actual N uptake is dependent on the available soil N as well as the ECM N  
 demand (eq. 11.2, Table A.1). The N availability function  $f(n_{avfungi})$  determines the fraction of soil N which is  
 740 available for ECM fungal uptake, and is controlled by the parameters  $NUPT_{ORGFACMAX}$  (the fraction of organic N  
 available for uptake) and  $NUPT_{FRACMAX}$  (the fraction of mineral N available for uptake). N availability for ECM  
 corresponds to the plant available N (eq. 16, Table A.1), but as ECM are more efficient in the uptake of nutrients,  
 the availability is enhanced for both mineral and organic N (eqs. 17.1, 17.2, 17.3, Table A.1). To prevent the ECM  
 N demand being covered by only one N form, the parameters  $r_{NO3}$ ,  $r_{NH4}$ ,  $r_{LIT}$ , and  $r_{HUM}$  are included, corresponding  
 745 to the ratio of nitrate and ammonium in total available soil N (litter and humus). If the potential N uptake exceeds  
 the available soil N, the actual uptake corresponds to the available N (eq. 11.2 and eq. 11.5, Table A.1 (b)).

## 2.2 Transect modelling approach

### 2.2.1 Three ECM modelling approaches

750 Three modelling approaches of different complexity were applied in this study differing in their complexity.  
 The basic "nonlim" approach is was conducted to test if a plant N uptake can be described as proportional to the  
 C demand of the plants of the respective sites. In this case, the plant N uptake is not regulated by the the actual  
 soil N availability, and N is used from a virtual source potentially exceeding the soil N availability, thus as an  
 "open" N cycle. The "ECM implicit" approach simulates the plant uptake of organic N which is assumed  
 755 to be of via ECM origin, i.e. ECM are considered implicitly as being responsible for the N source uptake, but  
 they do are not physically exist represented in the model. The rate of the organic N uptake is determined by the  
 plant N demand and restricted by the availability of organic N in the soil humus pools (eqs. 4.4, 4.5, Table A.1).  
 Plants can also additionally take up ammonium and nitrate (eqs. 4.1, 4.2, Table A.1). In the "ECM explicit"  
 approach, ECM fungi are fully physically considered as described above. Fungal-ECM growth interacts  
 760 dynamically with plant growth and responds to changes in soil N availability and soil temperature. ECM fungi  
 can take up both mineral and organic N forms.

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## 2.2.2 Simulated regions and database

Simulations were performed for four forests sites: Lycksele, Mora, Nässjö, and Lungbyhed, situated along a climate and N deposition gradient in Sweden (Fig. 2). Climate and site information is given in Table 23, and the climate data were taken from the Swedish Meteorological and Hydrological Institute (SMHI). Data on forest standing stock volumes and as well as forest management were derived from the database and practical guidelines of the Swedish Forest Agency (2005), and were applied as previously described by Svensson et al. (2008a). Soil C content as well as soil C/N ratio, were previously determined by Berggren Kleja et al. (2008) and Olsson et al. (2007), were and used to describe soil properties in the initial model setting up. For all simulated sites and for all model modeling approaches, the development of managed Norway spruce forests was simulated in daily step over a 100-year period from a newly established to a closed mature forest. The measured 100-year old forest standing biomass and as well as soil C/N ratio were used for model calibration. Climate input data were quadrupled in order to cover the entire period, and thus climatic warming effects are not considered here. A minimum of specific regional data including the meteorological data, N deposition and soil data, were used as input values (Table 23). Otherwise, model parameters were kept identical between model modeling approaches in order to evaluate the general model applicability. An overview of the parameter values is shown in Table A.1 (d) in the Appendix. For a more detailed site description and CoupModel setup, see Svensson et al. (2008a).

## 2.3 Brief description of Bayesian calibration

### 2.3.1 Overview Observational constraints

We performed a Bayesian calibration for all model modeling approaches and sites. In this study, we emphasize the models' predictability in precisely describing the long term plant and soil developments, also aiming at maximized model flexibility. This allows us to compare the different model approaches in terms of explaining the measured data, and also to investigate distributions and uncertainty of key parameters. In this study, we emphasize the models' predictability in precisely describing the long term C and N in the soils and standing stock in the forest stock, also aiming at a maximized model flexibility. Measured data including tree biomass and the C/N ratio of soil organic matter are thus used as accepted criteria. The previous modeling study by Svensson et al. (2008a) demonstrated that the changes of soil C in these sites were rather small over a 100-year period while the soil C/N ratio showed large variabilities with different N supply assumptions. Therefore, in this study the measured C/N ratio of soil organic matter and standing stock biomass were used as observational constraints. Measured data including tree biomass and the C/N ratio of soil organic matter are thus used as accepted criteria. The measured error (also called relative uncertainty in Table 4) for both the soil C/N ratio and the standing stock biomass were difficult to assume due to lack of information. An uncertainty estimate of 30% was generally recommended under such conditions (van Oijen et al., 2005). In order to reduce the weight of values close to zero on behalf of large peaks, a minimum measured error that is 10% of the measured value was defined in this study (Klemedtsson et al., 2008). This is also because our intention was to force the model to simulate tree biomass and soil C/N ratio precisely, to better constrain posterior parameter distributions for the respective model approach and site. This allowed allows us to investigate the distributions and uncertainty of key parameters of the respective ECM model modeling approaches ("nonlim", "implicit", and "explicit"), as well as analyze model uncertainties and dependencies between parameters. Uncertainties in parameter values are expressed as probability distributions. The posterior probability distributions of parameters are estimated by considering the prior distribution and the

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likelihood function in the calibration procedure. The likelihood function is determined by the measured data on output variables and the respective error estimates of the simulated model output. The Bayesian calibration as applied in this study is briefly described below, however for a detailed description of the general methodology see e.g., van Oijen et al. (2005) or Klemedtsson et al. (2008) and van Oijen et al. (2005).

### 2.3.2 Model parameters chosen for calibration

The different ECM modeling approaches were calibrated for a comprehensive set of key parameters which are chosen according to their function as regulating factors of the C and N fluxes in the plant-soil-mycorrhiza continuum (Table 2). In the “nonlim” approach, the constant N supply parameter *ConstantNsupply* for the spruce tree was a calibration parameter. In the “implicit” approach, the fraction of organic N available for plant uptake (*NUPT<sub>ORGFRACMAX</sub>*) was included in the calibration based on Svensson et al. (2008a). For the ECM “explicit” approach, all ECM fungal parameters in MYCOFON including ECM growth (C and N assimilation and uptake, C and N losses), overall N uptake and plant N supply, respiration, and littering were calibrated. For all three approaches, the humus decomposition rate (*K<sub>H</sub>*), the C/N ratio of microbes (*CN<sub>mic</sub>*) regulating soil mineralization thus soil N availability, and the fraction of plant C assimilates allocated to the rooting zone (*F<sub>ROOT</sub>*) regulating ECM fungal growth were additionally calibrated.

### 2.3.3 Bayesian calibration of models

The prior distributions of the parameters were chosen as uniform and non-correlated, with wide ranges of possible values (Table 2). Bayesian calibration combines the prior information about the parameters, and the observational constraints on model outputs to obtain a revised probability distribution or called posterior distribution (Yeluripati et al., 2009). The posterior probability of any parameter vector is proportional to the product of its prior probability and its corresponding data likelihood (eq. (1)). The data likelihood function which determines acceptance of the parameter sets as the posterior distributions, is based on the assumption that the model errors (the differences between simulated and observed values) are normally distributed and uncorrelated (van Oijen et al., 2005). Furthermore, model errors are assumed to be additive so that the log-likelihood function reads:

The data likelihood function which determines the parameter sets being candidate of the posterior distributions is based on the assumption that the model errors, i.e., the differences between simulated and observed values, are normally distributed and uncorrelated (van Oijen et al., 2005). Furthermore, model errors are assumed to be additive so that the log-likelihood function reads:

$$\log L = \sum_{i=1}^n \left( -0.5 \left( \frac{y_i - f(\omega_i \cdot \theta_i)}{\sigma_i} \right)^2 - 0.5 \cdot \log(2\pi) \right) - \log(\sigma_i) \quad (1)$$

where  $y_i$  = observed values,  $f(\omega_i \cdot \theta_i)$  = simulated values for a given model input vector  $\omega_i$  and parameter set  $\theta_i$ ,  $\sigma_i$  = standard deviation across the measured replicates, and  $n$  = number of variables measured.

In this study, a measured uncertainty of 10% for both the soil C/N ratio and the standing stock biomass data is used. The uncertainty estimate is low (van Oijen et al., 2005), as our intention was to force the model to simulate tree biomass and soil C/N ratio precisely, to better constrain posterior parameter distributions for the respective model approaches and sites.

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To construct the posterior parameter distribution, many sets of parameter  $\theta$  were sampled. In this study, candidate parameter sets were generated by investigating the parameter space using the Metropolis-Hastings random walk Markov Chain Monte Carlo (MCMC) algorithm (van Oijen et al., 2005; Vrugt, 2016). Briefly, a parameter ensemble of “walkers” move around randomly and the integrand value at each step was calculated. A few number of tentative steps may further be made to find a parameter space with high contribution to the integral. MCMC thus increases the sampling efficiency by using information about the shape of the likelihood function to preferentially sample in regions where the posterior probability is high (Rubinstein and Kroese, 2016). For each simulation, the model’s likelihood is evaluated for a certain parameter set. After each run, a new parameter set was generated by adding a vector of random numbers  $\epsilon$  to the previous parameter vector:

$$q_{i+1} = q_i + \epsilon \quad (2)$$

where  $q_i$  = previous parameter vector,  $q_{i+1}$  = new parameter vector, and  $\epsilon$  = random numbers.

The normally distributed random numbers  $\epsilon$  have a mean of zero and a step length of 0.05, i.e., 5% of the prior parameter range as proposed by van Oijen et al. (2005). After a sufficiently long iteration (referred to as the “burn-in” period), the Markov chain reaches a stationary distribution that converges to the joint parameter posterior (Ricciuto et al., 2008). Van Oijen et al. (2005) recommended chain lengths in the order of  $10^4$ – $10^5$  for modelling forest ecosystems with many observational constraints. In this trial study, we performed  $10^4$  runs for each ECM modelling approach and site to ensure posterior convergence. This is because a length of  $10^4$  model runs with a burn-in length of around  $10^3$  runs results in numerically stable results for our current considered problem. The step sizes used in this study result in acceptance rates between 25 to 50% (Table 4), which is also generally the most efficient range for the MCMC algorithm (Harmon and Challenor, 1997).

### 2.3.3 Model parameters chosen for calibration

The different ECM modelling approaches are calibrated for a comprehensive set of key parameters which are chosen according to their function as regulating factors of the C and N fluxes in the plant-soil-mycorrhiza continuum (Table 3). In the “nonlim” approach, the constant N supply parameter *ConstantNsupply* for the spruce tree is a calibration parameter. In the “implicit” approach, the fraction of organic N available for plant uptake (*NUPT<sub>FORGERACMAX</sub>*) is included in the calibration based on Svensson et al. (2008a). For the ECM “explicit” approach, all fungal parameters in MYCOFON including: fungal growth (C and N assimilation and uptake, C and N losses), overall N uptake and plant N supply, respiration, and littering are calibrated. For all three approaches, the humus decomposition rate (*K<sub>H</sub>*), the C/N ratio of microbes (*CN<sub>mic</sub>*) regulating soil mineralization, and the fraction of plant C assimilates allocated to the rooting zone (*F<sub>ROOT</sub>*), regulating fungal growth are also calibrated. Overall, we include a rather generous number of parameters for Bayesian calibration following Klemetsson et al., (2008) which who emphasized the importance of a holistic perspective when considering model parameters. Prior distributions of parameters are assumed to be uniform, i.e., each value is equally probable, with a given minimum and maximum values (Table 3). Values were chosen based on either previous modelling applications (e.g., plant parameters determined by Svensson et al. 2008a, b), or literature data (Table 3).

## 3. Results

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### 3.1 Comparison of the three model modeling approaches

#### 3.1.1 General ability to reproduce tree growth and soil C/N

The three model modeling approaches show different abilities accuracies in reproducing current plant growth and soil C/N ratio after calibration (Table 3B4B). The posterior model in the “implicit” and “explicit” approaches shows a clear better performance of simulating soil C and N, as indicated by the soil C/N ratio, than the “nonlim” approach. The latter tends to simulate a lower soil C/N ratio, indicated by the negative mean errors (ME, difference between the simulated and measured values) in the posterior model (ME is the difference between the simulated and measured values) (Table 3B4B). The ME by the “nonlim” approach is also two to five times higher than that when using the “implicit” or “explicit” approach (Table 3B4B). The “nonlim” approach tends to overestimate the plant growth as the posterior mean of ME for plant C is always positive, while the “implicit” and “explicit” approaches tend to show an underestimation (Table 3B4B).

All posterior models underestimate soil C/N for the northern sites which are generally more N limited, but gradually switch to overestimation at the southern sites. The model with the “nonlim” approach simulates a better plant growth for the most southernmost site, Ljungbyhed, than the other sites. Further, Model modeled plant growth at Ljungbyhed is overestimated by the “implicit” approach, but underestimated when the “explicit” approach is used (Table 3B4B). The acceptance of model runs in posterior is higher for the “nonlim” approach (25 to 48%); and the “implicit” approaches (42 to 50%), followed by the “explicit” approaches (30 to 33%) which can be explained by the model complexity, i.e. as more parameters are included for calibration, accepted combinations of parameter sets become less likely. No major differences could be found for the summed log-likelihood for both calibration variables (Table 3B4B).

#### 3.1.2 Ecosystem C-N and N-C fluxes and comparison to measured data budget

Model modeled major ecosystem N fluxes, and soil C, and N balance in the posterior are shown in Figure 3. The modeled N litterfall, uptake and leaching fluxes differ significantly from one modeling approach to another where the “nonlim” approach always gives the highest fluxes. The “explicit” and “implicit” approaches show similar modeled N fluxes for the northernmost site, Lycksele. However, the differences between these two approaches become larger when moving towards south where higher fluxes are simulated by the “explicit” approach (Fig. 3). For instance, modeled N litter production in “explicit” approach increases by 1 to 30% compared to the “implicit” approach, but N losses due to uptake and leaching also increase by 10 to 50% for Lycksele and Ljungbyhed, respectively (Fig. 3). The modeled N pool sizes for these two sites also differ where the “explicit” approach shows a larger mineral N in the soil and a smaller organic N pool compare to the “implicit” approach (Fig. 4).

In general, the “nonlim” approach shows much larger greater uncertainties in the model modeled N fluxes than either the ECM “implicit” or “explicit” approaches. The “nonlim” approach simulates soil sequestration of N up to 2 g N m<sup>-2</sup> yr<sup>-1</sup> for all the sites, but much lower or close to zero values are found when using other two model modeling approaches (Fig. 3). Soil N is expected to reach a steady state over a period of 100 years (Svensson et al., 2008a). Therefore, the “nonlim” approach largely overestimates the soil N sequestration which. This can be attributed to the assumed “virtual” constant N uptake from the unlimited source. According to our model predictions, this “virtual” N fraction accounts for 20 to 30% of the total plant N uptake. The simulated soil C balance by the “nonlim” approach also contrasts with that of soil N, where the soil sequesters C at the most northern site, Lycksele, but losses C at a rate of 6 to 17 g C m<sup>-2</sup> yr<sup>-1</sup> for the other three sites (Fig. 3). Therefore,

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soil C and N are not in a steady states and are decoupled in the “nonlim” approach over the simulated 100-year period.

However, the “implicit” and “explicit” approaches show a strong coupling between soil C and N (Fig. 3). That is, in for the “implicit” approach, Lycksele and Mora soils lose 6 and 5 g C m<sup>-2</sup> yr<sup>-1</sup>, respectively, while Nässjö and Ljungbyhed soils gain 3 and 13 g C m<sup>-2</sup> yr<sup>-1</sup>, respectively. Similarly, Lycksele and Mora loses N by 0.2 and 0.1 g N m<sup>-2</sup> yr<sup>-1</sup>, while Nässjö and Ljungbyhed gain N by 0.3 and 0.6 g N m<sup>-2</sup> yr<sup>-1</sup>. For the “explicit” approach, soil C and N losses at the two northern sites are slightly higher than that in the “implicit” approach (Fig. 3). In contrast with to the “implicit” approach, the two southern sites also show overall minor C and N losses with large standard deviations (Fig. 3). Model Modeled N litter production increases by 1 to 30% compared to the “implicit” approach, but N losses due to uptake and leaching also increase by 10 to 50 % (for Lycksele and Ljungbyhed, respectively, (Fig. 3). The increased litter addition of easily degradable C and N stimulates microbial activity, thus leading to a higher microbial respiration, which explains the minor losses of C and N in the southern sites in the “explicit” model. The higher N leaching in the “explicit” model can be attributed to a higher uptake from organic N (eqs. 11.5, 11.6, Table 1.B) and also a stimulated microbial growth thus increases net mineralization, both of which leaves more mineral N in the soil (Fig. 4).

Figure 5 shows the modeled major ecosystem C fluxes and comparison with previous results by Svensson et al. (2008a) and measured data from three other Swedish sites (Flakaliden, Knottåsen and Asa, Fig. 2) at comparable latitudes and on comparable soils by Lindroth et al. (2008). The simulated plant gross primary production (GPP) using three approaches all show an increasing trend from the northern sites to the southern sites, due to a more favorable climates and N availability for spruce forest growth. For the studied four sites, the “nonlim” approach simulates the highest GPP followed by the “explicit” and lastly the “implicit” approach. The variation of modeled GPP between the “explicit” and “implicit” approach ranges from 12% in northernmost Lycksele site to 7% in the southernmost Ljungbyhed site (Fig. 5). Simulated GPP in this study are generally higher than that by Svensson et al. (2008a) but comparable with the measured data from Lindroth et al. (2008). It should be noted that the GPP at the southern site, Asa was only measured for one year thus can associated with large uncertainties due to annual variations. Modeled ecosystem respiration generally follows the pattern of GPP. The net ecosystem exchange (NEE) predicted by the three approaches all show an overall atmospheric C uptake for all the sites where the “explicit” approach seems to have a higher uptake strength than the others (Fig. 5). Current estimates of NEE are again within the measured range by Lindroth et al., (2008), although a small net release of C was measured at Knottåsen, likely caused by the abnormal high temperature during those measured years. In addition, explicitly including ECM also increase the soil respiration for the four sites except the northernmost Lycksele site. The simulated ranges however are somehow smaller than that by Svensson et al. (2008a).

The “nonlim” approach generally shows much higher uncertainties in the modeled N fluxes than either the “implicit” or “explicit” approaches. The “nonlim” approach simulated soil N sequestration up to 2 g N m<sup>-2</sup> yr<sup>-1</sup> for all the sites, but much lower or close to zero values were found when using the other two modeling approaches (Fig. 5). The simulated soil C balance by the “nonlim” approach also contrasts with that of soil N, where the soil sequesters C at the northernmost site, Lycksele, but loses C at a rate of 6 to 17 g C m<sup>-2</sup> yr<sup>-1</sup> for the other three sites (Fig. 5). Therefore, soil C and N are not in steady state and are decoupled in the “nonlim” approach over the simulated 100-year period. The “implicit” and “explicit” approaches, however, show a strong coupling between soil C and N (Fig. 5). That is, for the “implicit” approach, Lycksele and Mora soils lose 6 and 5 g C m<sup>-2</sup> yr<sup>-1</sup>

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distributions for the three northern sites. However, but much lower values are obtained for the south most southernmost Ljungbyhed site (Fig. 7), reflecting a more soil N rich environment. Overall, parameters are less constrained and only minor differences between sites are found when the “explicit” approach is used (Fig. 7).

### 3.2.2 ECM fungal specific parameters:

The posterior distributions of all ECM fungal specific parameters are all constrained to log-normal or normal distributions (data not shown). The mean values of the N uptake parameters ( $NORG_{RATE}$ ,  $NH4_{RATE}$ ,  $NO3_{RATE}$ ) show a decreasing trend from the northern to southern sites (Fig. 8). This again means an higher-enhanced ECM fungal N uptake rate is necessary to explain the observed soil and plant data at the more N-limited northern sites. Similarly, lower values for the northern and higher values for the southern regions are also found for the minimum ECM fungal C/N ratio parameter ( $CN_{FMIN}$ ). The optimum ratio between fungal-ECM and root C content,  $FRAC_{OPT}$ , tends to be higher at the northern sites and lower at the southern sites, also implying that a model modeled higher ECM biomass at the northern sites (Fig. 8).  $MIN_{SUP}$ , the minimum supply of N from fungi-ECM to the host plant parameter, does not show a clear trend. Further, differences of the other ECM parameters for the four regions sites are minor (Fig. 8).

### 3.2.3 Correlation between parameters

An overview of correlations for all posterior model parameters can be found in the Appendix supplementary in Tables A2, A3, and A4. Key parameter sets Parameters showing correlation with each other (defined here as a Pearson correlation coefficient  $r \geq 0.3$  or  $\leq -0.3$ ) are identified as the key parameter sets, shown in Figure 9. When the “implicit” approach is used, a significant positive correlation is obtained between the humus decomposition rate,  $K_H$ , and the fraction of C allocated to rooting zone,  $F_{ROOT}$ . The organic N uptake parameter,  $NUPT_{ORGFACMAX}$  and microbial C/N ratio,  $CN_{MIC}$  are significantly negative correlated, except for a weak correlation for Ljungbyhed (Fig. 9). A weak correlation between  $NUPT_{ORGFACMAX}$  and  $F_{ROOT}$  is also found for the Näs sjö site only (see Table A2-Appendix). For the “explicit” approach, the correlation coefficients between  $K_H$  and  $F_{ROOT}$  are all decreased, and there is also also a weaker correlation between  $NUPT_{ORGFACMAX}$  and  $CN_{MIC}$  for all sites comparing compared to that in the “implicit” approach (Fig. 9). No clear correlation between common and ECM fungal parameters is obtained. Further, a negative correlation occurred between microbial C/N ratio,  $CN_{MIC}$ , and the fungal N uptake rates ( $NORG_{RATE}$ ,  $NH4_{RATE}$ ,  $NO3_{RATE}$ ), but only for the Northern sites Lycksele and Mora (Table A4). A moderate correlation is found for  $K_H$  and the fungal litter rate,  $L$  for Ljungbyhed. Among fungal parameters, the N uptake rates correlated moderately correlate to the litter production rate,  $L$  at the northern sites, but correlations at Näs sjö and Ljungbyhed are either weak non-existent or non-existent weak (Table A4). Our identified inter-connections and correlations between the parameters in general reflect the complex and interrelated nature of ECM, soil, and plant interactions (He et al., 2016; Klemetsson et al., 2008). But more importantly, they also highlight the need to calibrate a number of parameters simultaneously rather than calibrating just one single parameter when applying the different process interactions and explanations provided by the applied modeling approaches, for the observational constraints such detailed ecosystem models (He et al., 2016; Klemetsson et al., 2008).

## 4. Discussion

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Our new version of the CoupModel provides a detailed model-predictive model framework to explicitly account for ECM in the plant-soil-ECM continuum. Model comparison to two earlier-simpler ECM model-modeling approaches show that large variations in N dynamic simulations, and that ECM and organic N uptake have to be included in ecosystem models to be able to describe the long-term plant and soil C and N development. Our results confirm that ECM have a substantial effect on soil C and N storage, and can also impact forest plant growth. But more importantly, including them into ecosystem models is both important and feasible.

ECM have to be included in ecosystem models (“implicitly or explicitly”) to be able to describe the long long term plant, and soil C and N development. Overall, the models perform similarly in the “implicitly or explicitly” approaches, while the “nonlim” approach significantly overestimates soil N uptake. Our results thus confirm that ECM have a substantial effect on soil C and N storage, and can also impact on forest plant growth. But more importantly, including them into ecosystem models is both important and feasible.

#### 4.1 Comparison of the three ECM modeling approach

The “nonlim” approach in this study shows an overestimation of plant growth and also larger biases in soil N than the “implicit” and “explicit” approaches even after calibration (Table 4). Soil N is expected to reach a steady state over a period of 100 years (Svensson et al., 2008a). Therefore, the “nonlim” approach largely overestimates soil N sequestration which can be attributed to the assumed “virtual” constant N uptake from the unlimited source. According to our model predictions, this “virtual” N fraction accounts for 20 to 30% of the total plant N uptake. A previous CoupModel application by Wu et al. (2012) demonstrated that the “nonlim” approach could possibly describe short-term C and water dynamics for a Finnish forest site. The same “nonlim” approach was also used in Franklin et al. (2014) to simulate Swedish forest biomass growth and its competition with ECM. These seem to suggest that plant growth and the C cycle can be simulated reasonably with the “nonlim” approach, although a slight trend of overestimation is exhibited. However, our modeling exercise further indicates that in this simplified approach soil C and N are uncoupled (Fig. 5) and therefore this approach is not recommended for future long-term soil C and N predictions. This is also reflected in the posterior model parameter distributions where the *constantNSupply* rate parameter shows primary control on the modeled plant growth and soil conditions. Other parameters have minor or no importance for the model results, reflecting an oversimplified model structure of N. Thus, the following discussion focuses on the other two modeling approaches.

Moore et al. (2015) demonstrated that accounting ECM in ecosystem models would substantially affect soil C storage, and that the impact is largely dependent on plant growth. Our study additionally shows that ECM representation in ecosystem models could further feedback into the predicted plant growth through N. When ECM are implicitly included, the model simulates a 48 g C m<sup>-2</sup> (average of four sites,  $\pm$ std: 86) lower plant biomass compared to the measured data. When they are explicitly included, the difference becomes even larger, 185 ( $\pm$ 35) g C m<sup>-2</sup> (Table 4). Including ECM explicitly in the model therefore results in decreased plant growth. This somehow differs from the general assumption that growth should be higher in mycorrhized plants, i.e., boreal forest trees, due to optimized nutrient supply (Pritsch et al., 2004; Finlay et al., 2008, see also review by Smith and Read, 2008). This discrepancy can be possibly due to: 1) an enhanced root litterfall due to a higher turnover of ECM mycelia. Simulated litter production is 50 to 110 g C m<sup>-2</sup> yr<sup>-1</sup> higher by the “explicit” approach compared to the “implicit” approach. This could be explained by the conceptually considering the ECM implicitly into the roots where the litterfall rate of roots is c.a. three times lower than that of ECM (calibrated litter rate of ECM is

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0.0075 d<sup>-1</sup>, Fig. 8, whereas the litter rate of roots is 0.0027 d<sup>-1</sup>, Table A1(d)). These two approaches thus show large differences in simulating litter production. The discrepancy could also be due to: 2) an enhanced N immobilization in ECM under N-limited conditions based on the assumption that ECM retain more N in their own biomass in response to plant allocation of newly assimilated C (Nehls et al., 2008). The increasing trend towards the northern sites shown by the constrained optimum ECM fungi C allocation fraction parameter (Fig. 8) also indicates a higher proportional C “investment” by the forest plants in ECM in northern, N limited conditions. The resulting ECM-plant competition for N could then potentially result in decreased plant N uptake, and thus plant growth (Näsholm et al., 2013). Finally, the discrepancy could be due to 3) biases in simulating ECM N uptake due to model/parameter uncertainties caused by high variability among ECM species and the scarcity of direct measurements in the field (Smith and Read, 2008; Clemmensen et al., 2013). The current “explicit” approach implements many biotic interactions and internal feedbacks within the plant-soil-ECM continuum. However, increasing the number of processes and interactions in an already complex ecosystem model will not necessarily generate more reliable model predictions; as shown here, the parameters in the “explicit” approach have a larger uncertainty range even after calibration. This is also shown by the smaller accepted ratio in the calibration (Table 4) which can be explained by model complexity; i.e. as more parameters are included for calibration, accepted combinations of parameter sets become less likely.

It should also be noted that the “explicit” and “implicit” approaches show considerable difference in estimating soil respiration. Compared to the “implicit” approach, the “explicit” approach simulates a 15% higher soil respiration for the northernmost site and 40% for the southernmost site. The measured soil respiration at Flakaliden is 400 to 590 g C m<sup>-2</sup> yr<sup>-1</sup> (Coucheney et al., 2013) and 460 to 520 g C m<sup>-2</sup> yr<sup>-1</sup> at Asa (Von Arnold et al., 2005) and these data generally align better with the modeled results by the “explicit” approach (Fig. 5). The estimated higher soil respiration is partly due to the higher litter production and consequently soil respiration in the “explicit” approach, but also due to a higher decomposition of the old organic matter (humus) as shown by the constrained higher humus decomposition coefficient,  $K_H$  in the “explicit” approach (Fig. 7). This collaborates well with findings from field measurements and recent modeling studies that ECM are able to degrade complex N polymers in humus layers, thus enhancing soil N transformation under low N conditions (Hartley et al., 2012; Moore et al., 2015; Lindahl and Tunlid, 2015; Parker et al., 2015; Baskaran et al., 2016). The modeled higher soil respiration further explains the minor losses of soil C and N in the southern sites, and also a higher mineral N pool thus higher N leaching in the “explicit” approach (Fig. 3 and Fig. 4).

#### 4.2 Constrained parameters

Our constrained parameters generally indicate a shift in the role of ECM from northern to southern sites with a corresponding shift in both climate and soil conditions (Fig. 6, Fig. 7 and Fig. 8). The ECM N uptake parameters show a decreasing trend with increasing soil N availability in the “explicit” approach. This is consistent with observations that at the northern N limited sites, organic N uptake by ECM is highly important for plant growth, becoming less important as N availability increases southwards (e.g., Hyvönen et al., 2008; Näsholm et al., 2013). Shown by the “explicit” approach, the mycorrhization degree of tree roots at Lycksele and Mora (> 90%) is much higher than that of Ljungbyhed (15%), thus the majority of modeled N uptake is through fungal mycelia in northern sites. Similar trend is also found for the organic N uptake parameter in the “implicit” approach, but with a larger site to site difference, thus indicating a stronger response to soil N conditions (Fig. 6). This is expected as more

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detailed ECM processes in the “explicit” approach should result in more internal interactions and feedbacks, thus more resilience to the change of environmental conditions.

Most ECM fungal parameters in the “explicit” approach are not – or only weakly – dependent on the differing environmental conditions along the modeled transect, except for the N uptake parameters,  $NORG_{RATE}$  and ECM fungal minimum C/N ratio,  $CN_{FMIN}$ , which show different mean values (Fig. 8). As such, these parameters need to be calibrated carefully when further applying the model to other sites with different soil nutrient levels or climate conditions.

The correlation between the humus decomposition coefficient,  $K_H$  and the fraction of C that is allocated to the rooting zone,  $F_{ROOT}$ , reflects the strong connection of the root-ECM symbiosis and also soil N availability. When ECM are explicitly modeled, this becomes less important, which can be explained by a more detailed internal cycling of N supply and uptake from the ECM; i.e., plant N supply is further regulated by simulated higher root litter input and N uptake from the soil (Fig. 3, Fig. 9). The correlations between the ECM fungal litter rate and ECM fungal N uptake rates in the “explicit” model, and that between fungal N uptake rates,  $NORG_{RATE}$  and the microbial C/N ratio,  $CN_{MIC}$ , for the northern sites (Fig. 9) further indicate the close coupling between ECM fungal N uptake (N loss from the soil) and litter production (N input to the soil). Such an incorporated tight cycle is of major importance for the overall plant N supply, and thus for the C and N dynamics of plant and soil at the N limited sites in the boreal forests. One of the major difficulties of explicitly including ECM in ecosystem models is the unknown turnover of ECM mycelia (Ekblad et al., 2013). Previously reported turnover rates of newly formed mycelia vary from days to weeks, even up to 10 years (Staddon et al., 2003; Wallander et al., 2004), mostly due to the high variability in ECM species and structures (see review by Ekblad et al., 2013). Additionally, root turnover rates can also vary considerably between species, soils, and climate zones (Brunner et al., 2012). Thus far, very few studies have reported parameterization of C and N cycling for ECM in boreal forests. The present model calibration thus provides a key set of ECM parameters that can be further tested by field observations, and more importantly, can in combination with the identified model parameter correlation, act as a guideline for future ECM modeling studies. ECM alter plant soil C and N dynamics

The “nonlim” model in this study shows overestimations of plant growth and also a clear larger biases in soil N than “implicit and explicit” approaches even after calibration (Table 3). A previous CoupModel application by Wu et al. (2012) demonstrated that the “nonlim” approach could possibly describe short term carbon and water dynamics in for a Finnish forest site. The same approach with open N cycle was also used in Franklin et al. (2014) to simulate the Swedish forest biomass growth and its competition with ECM. It therefore seems that plant growth and thus the C cycle can be simulated reasonably with the “nonlim” approach, plant growth thus C cycle can be simulated reasonably, although with a slightly trend of overestimation is shown exhibited here. However, our model modeling further indicates that this simplified approach has an uncoupled soil C and N in its model structure (Fig. 3) and is thus not recommended for future long term soil C and N predictions. This is also reflected in the posterior model parameter distributions where the constant N supply rate parameter shows primary control on the modeled plant growth and soil conditions. Other parameters have minor or no importance for the model results, reflecting an oversimplified soil C and N model structure. Thus, in the following discussion we focus on the other two model modeling approaches. Soil N is expected to reach a steady state over a period of 100 years (Svensson et al., 2008a).

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Moore et al. (2015) demonstrated that ECM have a substantially affect effect on soil C storage, and the its impact is largely dependent on plant growth. The Present study additionally shows that ECM representation in ecosystem models could also feedback on into the predicted plant growth through the feedback of N. As when ECM are implicitly included, the model simulates a 78 (average of four sites, std: 102) g C m<sup>-2</sup> lower plant biomass comparing compared to the “nonlim” approach. Further, when they are and when explicitly included, the difference becomes even larger, 214 (50) g C m<sup>-2</sup> (Table 3). Including ECM in the model thus shows a decreased plant growth. This somehow differs from with what is the generally assumed assumption that growth should be higher in mycorrhized plants, i.e. boreal forests, due to optimized nutrient supply (Pritsch et al., 2004; Finlay et al., 2008, see also review by Smith and Read, 2008). This discrepancy could be possibly due to: 1) an enhanced root litterfall, due to a higher turnover of fungal mycelia, shown by a higher litter turnover rate (calibrated litter rate of ECM is 0.0075 d<sup>-1</sup>, Fig. 8, whereas the litter rate of roots is 0.0027 d<sup>-1</sup>, Table A1(d)). When ECM is explicitly considered, litter production is modeled to be higher than in the “implicit” approach (difference from 50 to 110 g C m<sup>-2</sup> yr<sup>-1</sup>, data not shown). These two model modeling approaches thus show large differences in simulating litter production. Field data are further needed to clarify this. The discrepancy could also be due to: out; 2) an enhanced N immobilization in ECM under N-limited conditions, due to the because ECM retains more N in its their own biomass in response to plant allocation of newly assimilated C (Nehls et al., 2008). The constrained optimum fungi C allocation fraction parameter shows an increasing trend towards the more northern sites (Fig. 8). This indicating indicates a higher proportional C “investment” by the forest plants on ECM in more northern N-limited conditions. The resulting ECM-plant competition for N could then potentially result in a decreased plant N uptake, and thus plant growth (Näsholm et al., 2013). Finally, and the discrepancy could be due to 2) biases in simulating ECM N uptake due to model/parameter uncertainties caused by high variability among fungal species and the scarcity of direct measurements in the field (Smith and Read, 2008, Clemmensen et al., 2013). The Current “explicit” approach implemented many biotic interactions and internal feedbacks within the plant-soil-ECM continuum. However, increasing the number of processes and interactions in an already complex ecosystem model will not necessarily generate more reliable model predictions, as shown here, the parameters in the “explicit” approach have a larger uncertainty range even after calibration. Thus, future model evaluation, together with more detailed ECM data, are needed is of need to better understand the tightly coupled soil-ECM-plant continuum.

Both approaches simulate the soil C and N stock well (Table 3). The respective net change in the soil C pools of the “implicit” approach corresponds well to the results by Svensson et al. (2008a) who also suggesting a small loss of soil C in the north while a gain in the south. However, when the “explicit” approach is used, the soils in the south are also predicted to loss C and N, mostly due to an enhanced soil respiration (see section 3.1.2). It is difficult to evaluate which approach gives a more realistic prediction as, since field data are not available. However, Lindroth et al. (2008), who measured C fluxes at three sites in Sweden, which are situated at comparable latitudes and on comparable soils, found. They also found a similar trend in the soil net C change as simulated by the “explicit” approach here, but with a higher loss rate of between 24 to and 133 g C m<sup>-2</sup> year<sup>-1</sup> (Table 4).

#### 4.2 Parameter and model responses to different environmental conditions

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Our modelling results show a consistent pattern with observations (e.g., Hyvönen et al., 2008; Näsholm et al., 2013) that at the northern N-limited sites, organic N uptake by ECM is highly important for plant growth, and it becomes less important as N-availability increases southwards. As indicated by the “explicit” approach, the mycorrhization degree of tree roots at Lycksele and Mora (>90%) is much higher than that of Ljungbyhed (15%), thus the majority of modelled N uptake is through fungal mycelia in northern sites. The constrained fungal organic and mineral N uptake parameters also show a decreasing trend (Fig. 8). Similarly, the organic N uptake parameter,  $NUPT_{ORGFRACMAX}$ , in the “implicit” approach decreases from north to south, but with a more clear site-to-site difference, thus indicating a stronger response to environmental conditions (Fig. 6). This is expected as more detailed ECM processes in the “explicit” approach should result in more internal interactions and feedbacks, thus damping the direct environmental regulations. Current modelling also indicates a higher mineralisation, shown by the humus decomposition coefficient,  $K_H$ , in the northern sites. However, the decomposition of also mineralization is also enhanced when ECM is “explicitly” considered (Fig. 7). This collaborates well with findings from the field measurements and recently modelling studies, that ECM is able to degrade complex N-polymers in humus layers, thus enhance enhancing soil N transformation under low N conditions (Moore et al., 2015; Lindahl and Tunlid, 2015; Baskaran et al., 2016).

In the “implicit” approach, the humus decomposition coefficient,  $K_H$ , was found to correlate with the fraction of C that allocates to the rooting zone,  $F_{ROOT}$ . Since As ECM is implicitly included in the roots, this correlation thus therefore indirectly indicates a strong connection of the root-ECM symbiosis and soil N-availability. But when ECM are explicitly considered, this becomes less important, again due to a more detailed internal cycling of N supply and uptake from the fungi, i.e., plant N supply is further regulated by simulated higher litter input and N uptake from the soil in the “explicit” model (Fig. 3, Fig. 9). Our modelling shows the that fungal litter rates correlate to fungal N uptake rates in the “explicit” model, and the that fungal N uptake rates have significant correlations to the microbial C/N ratio,  $CN_{MIC}$ , for the northern sites (Fig. 9). This indicates the close coupling between fungal N uptake (N-loss from the soil) and fungal litter production (N input to the soil). Such an incorporated tight cycle is of major importance for the overall plant N supply, and thus C and N dynamics of plant and soil at the N-limited sites in the boreal forests.

Most fungal parameters in the “explicit” approach are not — or only weakly — dependent on the differing environmental conditions along the modelled transect, except for the N uptake parameters and fungal minimum C/N ratio,  $CN_{FMIN}$ , which show different mean values (Fig. 8). Thus As such, these parameters need to be calibrated carefully when further applied applying the model to other sites with different soil nutrient levels or climate conditions.

One of the major difficulties of the explicitly inclusion of ECM in ecosystem models is the unknown turnover of fungal mycelia (Ekblad et al., 2013). Previously reported turnover rates of newly formed mycelia vary from days to weeks, even up to 10 years (Staddon et al., 2003; Wallander et al., 2004), mostly due to the high variability in ECM species and structures (see review by Ekblad et al., 2013). Besides, root turnover rates can also vary considerably between species, soils, and climate zones (Brunner et al., 2012). So Thus far, very few studies have reported parameterization of C and N cycling for ECM in boreal forests. Our calibration study thus provides a key

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set of ECM parameters that can be further tested through field observation, and more importantly, together with the identified correlations with the variables, can act as a guidelines for future ECM model modeling studies.

## 5. Conclusions

The key components and features of the Coup-MYCOFON<sub>veefon</sub> model have been described. The new version of CoupModel simulates C and N fluxes and pools, with the capacity of explicitly accounting for the links and feedbacks between the ECM, soil, and the plant. The comparison of three commonly ECM modeling approaches which differing in complexity demonstrates that the simple “nonlim” approach cannot describe the measured soil C/N ratio, and also overestimates measured the forest growth. When including ECM either implicitly or explicitly, both models deliver accurate long-term quantitative predictions on forest C and N cycling with simultaneous considerations of the impact of ECM fungi on ecosystem dynamics. However, the “implicit” approach shows a much lower litter production and soil respiration than the “explicit” approach, and both approaches they but slightly underestimate forest growth. The ECM explicit Coup-MYCOFON<sub>veefon</sub> model provides a more detailed description of internal ecosystems fluxes and feedbacks of C and N fluxes. The constrained ECM parameter distributions presented in this study can be used as guidelines for future model applications. Overall, Our model implementation and comparison overall suggest that ecosystem models need to incorporate ECM fungi into their model structure for a better prediction of ecosystem C and N dynamics, and the new version of CoupModel now provides such an alternative option.

## 6. Code and data availability

The model and extensive documentation with tutorial excises are freely available from the CoupModel home page <http://www.coupmodel.com/> (CoupModel, 2015). The source code can be requested for non-commercial purposes from Per-Erik Jansson (pej@kth.se). CoupModel is written in the C programming language (code also available in Fortran) and run mainly under Windows/Linux systems. Inputs and outputs are in binary format. The version used as the basis for the present development was version 5 from 12 April 2017. The simulation files including the model and calibration set-up, the used parameterization, and corresponding input and validation files can be requested from Hongxing He (hongxing.he@gu.se). However, the majority of the input and output data used for the current model modeling is public available publicly, i.e. through SMHI or previous publications, i.e. Svensson et al. (2008). Please contact the first author of this publication or Per-Erik Jansson if you plan an application of the model and further collaboration. The model and extensive documentation with tutorial excises are freely available from the CoupModel home page <http://www.coupmodel.com/> (CoupModel, 2015). The source code will be available to download from the home page and a link to a repository for MS Visual studio can also be provided. CoupModel is written in the C programming language and runs mainly under Windows systems. The version used as the basis for the present development was version 5 from 12 April 2017. The simulation files including the model and calibration set-up, the used parameterization, and corresponding input and validation files can be requested from Hongxing He (hongxing.he@gu.se).

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services in a Changing Climate, [www.cec.lu.se/research/becc](http://www.cec.lu.se/research/becc)), and the Linnaeus Centre LUCCI (Lund University Centre for studies of Carbon Cycle and Climate Interactions).

APPENDIX:

Table A.1 Model functions describing plant growth, ECM fungal growth, model parameters, and response functions of plant and ECM. Parameters are always entitled with capital letters

Table A.1 (a) Description of plant model functions. (i = fine roots, coarse roots, stem, leaves, grain, mobile)

No.	Equation
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Plant photosynthesis (g C m<sup>-2</sup> d<sup>-1</sup>):

$$1 \quad c_{atm \rightarrow plant} = \varepsilon_L \times f(T_l) \times f(CN_l) \times f\left(\frac{E_{ta}}{E_{tp}}\right) \times r_s$$

$\varepsilon_L$  = coefficient for radiation use efficiency,  $f(T_l)$ ,  $f(CN_l)$ ,  $f(E_{ta}/E_{tp})$  = response functions to leaf temperature, leaf CN, and air moisture (see Table A.1 (c)),  $r_s$  = global radiation absorbed by canopy.

Plant maintenance respiration (g C m<sup>-2</sup> d<sup>-1</sup>):

$$2.1 \quad c_{plantM \rightarrow atm} = c_i \times K_{RMi} \times f(T_l)$$

$c_i$  = C content of each respective plant compartment i (g C m<sup>-2</sup>) and  $K_{RMi}$  is a coefficient.

Plant growth respiration (g C m<sup>-2</sup> d<sup>-1</sup>):

$$2.2 \quad c_{plantG \rightarrow atm} = c_{m \rightarrow i} \times K_{RGi}$$

$c_{m \rightarrow i}$  = C gain (growth) of each plant compartment i (g C m<sup>-2</sup> d<sup>-1</sup>) and  $K_{RGi}$  is a coefficient.

Plant litter production (g C m<sup>-2</sup> d<sup>-1</sup>):

$$3 \quad c_{i \rightarrow lit} = c_i \times L_i$$

where  $C_i$  is the C content of each plant compartment i (g C m<sup>-2</sup>) and  $L_i$  (= 0.0027 d<sup>-1</sup>) is a coefficient.

Plant nitrate and ammonium uptake (g N m<sup>-2</sup> d<sup>-1</sup>) (only shown for nitrate, equivalent for ammonium):

$$4.1 \quad n_{NO3 \rightarrow plant} = dem_{Nplant} \times r_{NO3} \quad \text{if } f(n_{minavail}) \times n_{NO3soil} \geq dem_{Nplant} \times r_{NO3}$$

$$4.2 \quad n_{NO3 \rightarrow plant} = f(n_{minavail}) \times n_{NO3soil} \times dem_{Nplant} \quad \text{if } f(n_{minavail}) \times n_{NO3soil} \leq dem_{Nplant} \times r_{NO3}$$

and where

$$4.3 \quad dem_{Nplant} = \sum \frac{c_{a \rightarrow i} - c_{i \rightarrow atm}}{CN_{iMIN}}$$

$f(n_{NO3avail})$  = fraction of soil NO<sub>3</sub> available for plant uptake (see response functions Table A.1 (d)),  $n_{NO3soil}$  = soil NO<sub>3</sub>-N content (g N m<sup>-2</sup>),  $dem_{Nplant}$  = plant N demand (g N m<sup>-2</sup> d<sup>-1</sup>),  $r_{NO3}$  = fraction of soil NO<sub>3</sub>-N in total mineral soil N,  $c_{a \rightarrow i}$  = plant C gain (g C m<sup>-2</sup> d<sup>-1</sup>),  $c_{i \rightarrow atm}$  = respiration of respective plant compartment i (g C m<sup>-2</sup> d<sup>-1</sup>),  $CN_{iMIN}$  = defined minimum C:N ratio of each plant compartment i.

Plant organic N uptake (g N m<sup>-2</sup> d<sup>-1</sup>) from the humus layer:

$$4.4 \quad n_{hum \rightarrow plant} = dem_{Nplant} \times r_{hum} \quad \text{if } f(n_{humavail}) \times n_{humsoil} \geq dem_{Nplant} \times r_{hum}$$

$$4.5 \quad n_{hum \rightarrow plant} = f(n_{humavail}) \times n_{humsoil} \quad \text{if } f(n_{humavail}) \times n_{humsoil} < dem_{Nplant} \times r_{hum}$$



$f(n_{humavail})$  = response function for plant available N from the humus layer,  $n_{humsoil}$  = soil N content in humus layer (g N m<sup>-2</sup>).

**Table A.1 (b) Functions describing processes related to ECM fungal growth and N exchange to plant**

Fungal-ECM fungal maximum C supply (g C m<sup>-2</sup> d<sup>-1</sup>):

$$5.1 \quad c_{a \rightarrow fungi} = c_{a \rightarrow root} \times FRAC_{FMAX} \times f(c_{fungiavail})$$

Fungal-ECM fungal actual growth (g C m<sup>-2</sup> d<sup>-1</sup>):

$$5.2 \quad c_{a \rightarrow fungi} = ((c_{frt} \times FRAC_{OPT}) - c_{fungi}) \times f(n_{supply})$$

$c_{a \rightarrow root}$  = C available for root and mycorrhiza growth (g C m<sup>-2</sup> d<sup>-1</sup>),  $FRAC_{FMAX}$  = maximum fraction of total root and mycorrhiza available C which is available for ECM,  $f(c_{fungiavail})$  = response function which relates fungal-ECM growth to N availability,  $c_{frt}$  = total root C content (g C m<sup>-2</sup>),  $FRAC_{OPT}$  = optimum ratio between root and fungal-ECM C content,  $c_{fungi}$  = total ECM C content (g C m<sup>-2</sup>),  $f(n_{supply})$  = response function of fungal growth to the amount of N (both mineral and organic N) which is transferred from ECM to plant.

Minimum ECM fungal C supply (g C m<sup>-2</sup> d<sup>-1</sup>):

$$5.3 \quad c_{a \rightarrow fungi} = c_{fungi \rightarrow atm} \quad \text{if } c_{a \rightarrow root} \leq 0$$

Total ECM fungal respiration (g C m<sup>-2</sup> d<sup>-1</sup>):

$$6.1 \quad c_{fungi \rightarrow atm} = c_{mfungi \rightarrow a} + c_{gfungi \rightarrow a}$$

where  $c_{mfungi \rightarrow a}$  = ECM fungal maintenance respiration and  $c_{gfungi \rightarrow a}$  = ECM fungal growth respiration (all in g C m<sup>-2</sup> d<sup>-1</sup>).

Fungal-ECM fungal maintenance respiration (g C m<sup>-2</sup> d<sup>-1</sup>):

$$6.2 \quad c_{mfungi \rightarrow a} = c_{fungi} \times K_{RM} \times f(T_l)$$

$c_{fungi}$  = total ECM C content (g C m<sup>-2</sup>),  $K_{RM}$  = maintenance respiration coefficient,  $f(T_l)$  = temperature response function.

Fungal-ECM fungal growth respiration (g C m<sup>-2</sup> d<sup>-1</sup>):

$$6.3 \quad c_{gfungi \rightarrow a} = c_{a \rightarrow fungi} \times K_{RG}$$

$c_{a \rightarrow fungi}$  = ECM fungal growth (g C m<sup>-2</sup> d<sup>-1</sup>),  $K_{RG}$  = growth respiration coefficient.

Fungal-ECM fungal C and N litter production ( $c_{fungi \rightarrow lit}$ : g C m<sup>-2</sup> d<sup>-1</sup>,  $n_{fungi \rightarrow lit}$ : g N m<sup>-2</sup> d<sup>-1</sup>):

If ECM fungal growth = simple

$$7.1 \quad c_{fungi \rightarrow lit} = c_{fungi} \times L$$

$$7.2 \quad n_{fungi \rightarrow lit} = n_{fungi} \times L - n_{ret\_fungi}$$

$$7.3 \quad n_{ret\_fungi} = n_{fungi} \times L \times (1 - N_{RET})$$

$c_{fungi}$  = ECM C content (g C m<sup>-2</sup>),  $n_{fungi}$  = ECM fungal N content (g N m<sup>-2</sup>),  $L$  = litter rate,  $nret_{fungi}$ : ECM fungal N which is retained in fungal tissue,  $N_{RET}$  = fraction of N retained in fungal tissue from senescence.

If ECM fungal growth = detailed

$$7.4 \quad c_{fungi \rightarrow lit} = c_{fungi} \times (FRAC_{MYC} \times L_{MYC} + ((1 - FRAC_{MYC}) \times L_M))$$

$$7.5 \quad n_{fungi \rightarrow lit} = n_{fungi} \times (FRAC_{MYC} \times L_{MYC} + ((1 - FRAC_{MYC}) \times L_M)) - nret_{fungi}$$

7.6  $FRAC_{MYC}$  = fraction of mycorrhizal hyphaemycelia in total fungal biomass,  $L_{MYC}$  = litter rate of mycorrhizal hyphaemycelia,  $L_M$  = litter rate of ECM fungal mantle tissue.

Fungal-ECM fungal biomass (g C m<sup>-2</sup> g N m<sup>-2</sup>)

$$8.1 \quad c_{fungi} = c_{a \rightarrow fungi} - c_{fungi \rightarrow litter} - c_{fungi \rightarrow a}$$

$$8.2 \quad n_{fungi} = n_{N \rightarrow fungi} - n_{fungi \rightarrow litter} - n_{fungi \rightarrow plant}$$

Mycorrhization degree

$$9 \quad m = \frac{c_{fungi}}{c_{fri} \times FRAC_{OPT} \times M_{OPT}}$$

$c_{fri}$  = fine root biomass (g C m<sup>-2</sup>),  $FRAC_{OPT}$  = coefficient defining optimum ratio between ECM fungal and fine root biomass,  $M_{OPT}$  = optimum mycorrhization degree, and  $m=1$ , when  $\frac{c_{fungi}}{c_{fri} \times FRAC_{OPT}} \geq M_{opt}$ .

Uptake and transfer processes of ECM and plant

N transfer from ECM to plant (g N m<sup>-2</sup> d<sup>-1</sup>)

$$10.1 \quad n_{fungi \rightarrow plant} = dem_{Nplant} \quad \text{if } dem_{Nplant} \leq n_{fungiavail}$$

$$n_{fungi \rightarrow plant} = n_{fungiavail} \quad \text{if } dem_{Nplant} > n_{fungiavail}$$

$dem_{Nplant}$  = plant N demand,  $n_{fungiavail}$  = fungal available N for transfer to plant (all g N m<sup>-2</sup> d<sup>-1</sup>)

$$10.2 \quad n_{fungiavail} = n_{fungi} - \frac{c_{fungi}}{CN_{FMAX}}$$

$c_{fungi}$  = ECM biomass (g C m<sup>-2</sup>),  $CN_{FMAX}$  = maximum C:N ratio of fungal tissue, which allows N transfer to plant.

Fungal-ECM fungal nitrate and ammonium uptake (given for nitrate, equivalent for ammonium with ammonium specific parameter)

$$11.1 \quad n_{NO3 \rightarrow fungi} = n_{NO3pot \rightarrow fungi} \times r_{NO3} \times f(n_{demfungi}) \quad \text{if } N_{NO3pot \rightarrow fungi} < n_{NO3soil} \times f(n_{avfungi})$$

$$11.2 \quad n_{NO3 \rightarrow fungi} = n_{NO3soil} \times f(n_{avfungi}) \quad \text{if } N_{NO3pot \rightarrow fungi} > n_{NO3soil} \times f(n_{avfungi})$$

$$11.3 \quad n_{NO3pot \rightarrow fungi} = NO3_{RATE} \times c_{fungi} \times FRAC_{MYC}$$

1365  $n_{NO_3pot \rightarrow fungi}$  = potential ECM nitrate uptake ( $g\ N\ m^{-2}\ d^{-1}$ ),  $r_N$  = fraction of ammonium-N and total mineral-N in the soil,  $f(n_{demand})$  = N uptake response to N demand,  $n_{NO_3soil}$  = soil nitrate content ( $g\ N\ m^{-2}$ ),  $f(n_{avfungi})$  = N uptake response to soil availability,  $NO_3RATE$  = nitrate specific uptake rate ( $g\ N\ m^{-2}\ d^{-1}$ ),  $c_{fungi}$  = [ECM](#) fungal biomass ( $g\ C\ m^{-2}$ ),  $FRAC_{MYC}$  = fraction of mycorrhizal mycelia in total [fungal-ECM](#) biomass.

1370 [Fungal-ECM](#) organic N uptake from litter and humus (given for litter, equivalent for humus with humus specific parameter)

$$11.4 \quad n_{lit \rightarrow fungi} = n_{litpot \rightarrow fungi} \times r_{lit} \times f(n_{demand}) \quad \text{if } n_{litpot \rightarrow fungi} \times r_{lit} < n_{litsoil} \times f(n_{litavfungi}) \times r_{lit}$$

$$11.5 \quad n_{lit \rightarrow fungi} = n_{litsoil} \times f(n_{litavfungi}) \times r_{lit} \quad \text{if } n_{litpot \rightarrow fungi} \times r_{lit} > n_{litsoil} \times f(n_{litavfungi}) \times r_{lit}$$

$$11.6 \quad n_{litpot \rightarrow fungi} = LITRATE \times c_{fungi} \times FRAC_{MYC}$$

where  $n_{litpot \rightarrow fungi}$  = potential ECM organic N uptake from litter ( $g\ N\ m^{-2}\ d^{-1}$ ),  $r_{lit}$  = fraction of litter-N in total organic-N in the soil,  $f(n_{demand})$  = N uptake response to N demand,  $n_{litsoil}$  = soil litter content ( $g\ N\ m^{-2}$ ),  $LITRATE$  = litter specific uptake rate ( $g\ N\ g\ C^{-1}\ d^{-1}$ ),  $c_{fungi}$  = [ECM](#) fungal biomass ( $g\ C\ m^{-2}$ ),  $FRAC_{MYC}$  = fraction of mycorrhizal mycelia in total [fungal-ECM](#) biomass.

**Table A1 (c) Overview of response functions of plant and [ECM](#) fungal growth and N uptake**

No. Equation

Plant response to air temperature

$$f(T_l) = \begin{cases} 0 & T_l < P_{min} \\ \frac{(T_l - P_{min})}{(P_{01} - P_{min})} & P_{min} \leq T_l \leq P_{01} \\ 1 & P_{01} < T_l < P_{02} \\ 1 - \frac{(T_l - P_{02})}{(P_{max} - P_{02})} & P_{02} < T_l < P_{max} \\ 0 & T_l > P_{max} \end{cases}$$

where  $T_l$  = leaf temperature ( $^{\circ}C$ ) and  $P_{min}$  ( $-4^{\circ}C$ ),  $P_{01}$  ( $10^{\circ}C$ ),  $P_{02}$  ( $25^{\circ}C$ ),  $P_{max}$  ( $40^{\circ}C$ ) are coefficients.

Photosynthetic response to leaf C/N ratio

$$f(CN_l) = 1 + \left( \frac{CN_l - P_{CNOPT}}{P_{COPT} - P_{CNTH}} \right) \quad \text{if } P_{CNTH} \leq CN_l \leq P_{CNOPT}$$

$$0 \quad \text{if } CN_l > P_{CNTH}$$

where  $CN_l$  = leaf C/N ratio and  $P_{CNOPT}$  (25) and  $P_{CNTH}$  (75) are parameters.

Plant response to soil moisture

$$f\left(\frac{E_{ta}}{E_{tp}}\right) = \frac{E_{ta}}{E_{tp}}$$

where  $E_{ta}$  = actual transpiration and  $E_{tp}$  = potential transpiration ( $mm\ d^{-1}$ ).

Plant mineral N uptake response to N availability and [ECM](#) fungal mantle

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$$15 \quad f(n_{\min \text{ avail}}) = NUPT_{FRACMAX} \times e^{(-FM \times m)}$$

Where  $NUPT_{FRACMAX}$ , coefficient describing fraction of soil N available, and  $FM$ , uptake reduction due to [ECM fungal mantle](#).

1405 Plant organic N uptake response to N availability and [ECM fungal mantle](#) (given for litter, equivalent for humus)

$$16 \quad f(n_{\text{litavail}}) = NUPT_{ORGFRACMAX} \times e^{(-FM \times m)}$$

Where  $NUPT_{FRACMAX}$  is the respective uptake coefficient for N from humus (included in calibration), and  $FM$  the uptake reduction due to [ECM fungal mantle](#).

1410 ECM N uptake response to N availability

$$17.1 \quad f(n_{\text{avfungi}}) = NUPT_{FRACMAX} \times UPT_{MINENHANCE} \quad \text{for nitrate}$$

$$17.2 \quad f(n_{\text{avfungi}}) = NUPT_{FRACMAX} \times UPT_{MINERAL} \times UPT_{NH4} \quad \text{for ammonium}$$

$$17.3 \quad f(n_{\text{orgavfungi}}) = NUPT_{ORGFRACMAX} \times UPT_{ORG} \quad \text{for litter/humus}$$

1415 ECM N uptake response to N demand

$$18 \quad f(n_{\text{denfungi}}) = 1 - \frac{CN_{FMIN}}{CN_{fungi}}$$

where  $CN_{FMIN}$  = minimum ECM C/N ratio.

$$19 \quad f(c_{\text{fungiavail}}) = e^{(-N_{AVAILCOEF} \times n_{\min \text{ soil}}^2)^3}$$

1420 Where  $NAVAILCOEF$  is a coefficient and  $N_{\min \text{ soil}}$  is the total soil content of ammonium and nitrate ( $\text{g N m}^{-2}$ ).

$$20.1 \quad f(n_{\text{supplyfungi}}) = 1 \quad \text{if } \min_{NPlant} < n_{\text{fungi} \rightarrow \text{plant}}$$

$$20.2 \quad f(n_{\text{supplyfungi}}) = \frac{n_{\text{fungi} \rightarrow \text{plant}}}{n_{\text{fungi} \rightarrow \text{plant}} + n_{\text{soil} \rightarrow \text{plant}}} \quad \text{if } \min_{NPlant} > n_{\text{fungi} \rightarrow \text{plant}}$$

$$20.3 \quad \min_{NPlant} = MIN_{SUPL} \cdot (n_{\text{fungi} \rightarrow \text{plant}} + n_{\text{soil} \rightarrow \text{plant}})$$

1425 Where  $\min_{NPlant}$  = defined minimum [ECM fungal N supply](#) in plant N uptake,  $n_{\text{fungi} \rightarrow \text{plant}}$  = actual ECM N supply to plant ( $\text{g N m}^{-2} \text{d}^{-1}$ ),  $n_{\text{soil} \rightarrow \text{plant}}$  = total plant N uptake from mineral and organic fraction ( $\text{g N m}^{-2} \text{d}^{-1}$ ).

1430 Table A1 (d) Overview of model parameters: previous CoupModel parameters are mostly from Svensson et al. (2008a) and ECM parameters are from literature value (Meyer et al. (2012) and references therein the paper text)

Parameter	Description	Value	Unit
$CN_{FMIN}$	Minimum <a href="#">fungal-ECM</a> C/N ratio for fungal N demand	18	$\text{gC gN}^{-1}$
$CN_{FMAX}$	Maximum <a href="#">fungal-ECM</a> C/N ratio for N transfer to plant	30	$\text{gC gN}^{-1}$
$CN_{iMIN}$	Minimum C/N ratio of fine roots,	40	$\text{gC gN}^{-1}$
	Needles/leaves	22	$\text{gC gN}^{-1}$

		Coarse roots and stem	450	gC gN <sup>-1</sup>	
	E <sub>L</sub>	<del>coefficient</del> Coefficient for radiation use efficiency	8		Formatted: English (United States)
	E <sub>NH4</sub>	<del>funeral</del> Fungal-ECM NH <sub>4</sub> uptake enhancement factor	5		Formatted: English (United States)
	FM	<del>plant</del> Plant N uptake reduction due to ECM mantle	0.5		Formatted: English (United States)
1440	FRAC <sub>FMAX</sub>	Maximum fraction of C allocated to rooting zone which is made available for ECM	0.5		Formatted: English (United States)
	FRAC <sub>MYC</sub>	Fraction of fungal-ECM mycelia in total biomass	0.5		
	FRAC <sub>OPT</sub>	Optimum fraction between root and fungal-ECM biomass	0.3		Formatted: English (United States)
	K <sub>RGF</sub>	Growth respiration coefficient of ECM	0.21	d <sup>-1</sup>	Formatted: English (United States)
1445	K <sub>RMi</sub>	Maintenance respiration coefficient of plant compartment i (i = fine roots, coarse roots, stem, leaves)	0.001	d <sup>-1</sup>	
	K <sub>RGi</sub>	Growth respiration coefficient of plant compartment i	0.21	d <sup>-1</sup>	
	L <sub>FRT</sub>	Litter rate of fine roots	0.0027	d <sup>-1</sup>	
1450	L <sub>CRT</sub>	Litter rate of coarse roots	0.000027	d <sup>-1</sup>	
	L <sub>LEAF</sub>	Litter rate of needles	0.0002	d <sup>-1</sup>	
	L <sub>STEM</sub>	Litter rate of stem	0.000027	d <sup>-1</sup>	
	L	Litter rate of ECM (if fungal growth = simple)	0.004		
	L <sub>M</sub>	Litter rate of ECM-fungal mantle (if fungal growth = detailed)	0.0014	d <sup>-1</sup>	Formatted: English (United States)
1455	L <sub>MYC</sub>	Litter rate of fungal-ECM mycelia (if fungal growth = detailed)	0.01	d <sup>-1</sup>	Formatted: English (United States)
	M <sub>OPT</sub>	Optimum mycorrhization degree of fine roots < 2 mm	0.5		
1460	N <sub>RET</sub>	N retained by ECM from senescence	0.54	d <sup>-1</sup>	
	NUPT <sub>FRACMAX</sub>	fraction of mineral N available for uptake	0.08	d <sup>-1</sup>	

Table A2 Correlation between common model parameters for all simulated sites with the “implicit” and “explicit” approaches, respectively. Correlation is given as the Pearson correlation coefficient

		implicit				explicit			
	-	K <sub>H</sub>	NUPT <sub>OEM</sub>	F <sub>ROOT</sub>	CN <sub>MIC</sub>	K <sub>H</sub>	NUPT <sub>OEM</sub>	F <sub>ROOT</sub>	CN <sub>MIC</sub>
Lyckeå	K <sub>H</sub>	1	-0.20	0.67	0.23	1	-0.08	0.28	0.21
	NUPT <sub>OEM</sub>		1	0.24	-0.57		1	0.02	-0.35
	F <sub>ROOT</sub>			1	0.18			1	0.02
	CN <sub>MIC</sub>				1				1
Mora	K <sub>H</sub>	1	-0.13	0.73	0.11	1	0.08	0.22	0.04
	NUPT <sub>OEM</sub>		1	0.18	-0.64		1	0.10	-0.46
	F <sub>ROOT</sub>			1	0.13			1	0.12
	CN <sub>MIC</sub>				1				1
Nässjö	K <sub>H</sub>	1	0.03	0.70	-0.08	1	0.13	0.29	0.16
	NUPT <sub>OEM</sub>		1	0.31	-0.60		1	0.29	-0.53
	F <sub>ROOT</sub>			1	0.02			1	0.12
	CN <sub>MIC</sub>				1				1
Ljungbyhed	K <sub>H</sub>	1	0.03	0.66	-0.18	1	0.33	0.26	-0.19
	NUPT <sub>OEM</sub>		1	0.17	-0.28		1	0.23	-0.26
	F <sub>ROOT</sub>			1	0.24			1	0.07
	CN <sub>MIC</sub>	-	-	-	1	-	-	-	1

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Table A3 Correlation between fungal and common model parameters with “explicit” approach for all sites. Correlation is given as the Pearson correlation coefficient

-	-	NO <sub>3</sub> RATE	NH <sub>4</sub> RATE	NO <sub>3</sub> RATE	K <sub>RM</sub>	L <sub>MYC</sub>	L <sub>M</sub>	CN <sub>FMIN</sub>	MIN <sub>SUPL</sub>	FRAC <sub>OPT</sub>	NAV	Formatted: English (United States)
Lyckeåsele	K <sub>H</sub>	0.17	0.16	0.16	0.01	-0.30	-0.27	0.02	0.00	-0.17	-0.13	Formatted: English (United States)
	NUPT <sub>OFM</sub>	-0.32	-0.28	-0.28	0.09	0.13	0.13	0.18	0.10	0.01	0.02	Formatted: English (United States)
	F <sub>ROOT</sub>	0.06	0.03	0.03	-0.05	0.03	0.03	0.00	0.06	-0.04	-0.15	
	CN <sub>MIC</sub>	-0.33	-0.34	-0.34	0.00	0.21	0.21	0.23	0.03	-0.12	-0.01	
Mora	K <sub>H</sub>	0.22	0.20	0.20	-0.09	-0.25	-0.21	-0.02	0.08	-0.14	-0.04	
	NUPT <sub>OFM</sub>	-0.15	-0.09	-0.09	0.08	0.02	0.02	0.05	0.11	-0.08	0.00	Formatted: English (United States)
	F <sub>ROOT</sub>	-0.11	-0.12	-0.12	0.06	0.26	0.26	0.25	-0.06	-0.01	0.01	
	CN <sub>MIC</sub>	-0.38	-0.40	-0.40	-0.03	0.29	0.29	0.33	-0.10	-0.08	-0.08	
Nässiå	K <sub>H</sub>	0.20	0.18	0.18	-0.06	-0.33	-0.32	-0.13	0.08	-0.03	-0.08	
	NUPT <sub>OFM</sub>	-0.07	-0.03	-0.03	-0.05	-0.11	-0.11	-0.12	-0.03	0.18	-0.06	Formatted: English (United States)
	F <sub>ROOT</sub>	-0.06	-0.03	-0.03	0.01	0.08	0.08	0.09	-0.08	0.14	-0.08	
	CN <sub>MIC</sub>	-0.23	-0.20	-0.20	0.05	0.11	0.11	0.15	-0.02	-0.17	0.09	
Ljungbyhed	K <sub>H</sub>	0.34	0.36	0.36	-0.08	-0.51	-0.53	-0.13	0.18	-0.22	-0.20	Formatted: English (United States)
	NUPT <sub>OFM</sub>	0.10	0.16	0.16	0.05	-0.21	-0.21	-0.24	0.06	-0.13	-0.07	
	F <sub>ROOT</sub>	-0.11	-0.07	-0.07	0.19	0.10	0.10	0.11	0.04	0.02	-0.02	
	CN <sub>MIC</sub>	-0.22	-0.21	-0.21	0.01	0.15	0.15	0.18	-0.05	0.02	0.07	Formatted: English (United States)

Table A4 Correlation between fungal model parameters with the “explicit” approach for all sites. Correlation is given as the Pearson correlation coefficient

-		NO <sub>3</sub> RATE	NH <sub>4</sub> RATE	NO <sub>3</sub> RATE	K <sub>RM</sub>	L <sub>MYC</sub>	L <sub>M</sub>	CN <sub>FMIN</sub>	MIN <sub>SUPL</sub>	FRAC <sub>OPT</sub>	NAVAH <sub>COEF</sub>
Lycksele	NO <sub>3</sub> RATE	1	0.91	0.91	0.01	-0.55	-0.59	-0.10	-0.07	0.07	-0.03
	NH <sub>4</sub> RATE		1	0.99	0.01	-0.50	-0.56	-0.07	-0.05	0.07	-0.03
	NO <sub>3</sub> RATE			1	0.01	-0.50	-0.56	-0.07	-0.05	0.07	-0.03
	K <sub>RM</sub>				1	-0.1	-0.1	-0.06	-0.07	-0.03	-0.04
	L <sub>MYC</sub>					1	0.95	0.04	0.07	-0.17	-0.03
	L <sub>M</sub>						1	0.04	0.07	-0.13	-0.02
	CN <sub>FMIN</sub>							1	0.05	0.07	0.05
	MIN <sub>SUPL</sub>								1	0	0.05
	FRAC <sub>OPT</sub>									1	0.17
	NAVAH <sub>COEF</sub>										1
Mora	NO <sub>3</sub> RATE	1	0.88	0.88	-0.09	-0.40	-0.48	0.02	-0.05	0.04	0.06
	NH <sub>4</sub> RATE		1	0.99	-0.08	-0.32	-0.43	0.01	-0.03	0.09	0.08
	NO <sub>3</sub> RATE			1	-0.08	-0.32	-0.43	0.01	-0.03	0.09	0.08
	K <sub>RM</sub>				1	-0.07	-0.06	0.01	-0.15	0.05	0.05
	L <sub>MYC</sub>					1	0.95	-0.08	0.05	-0.21	-0.02
	L <sub>M</sub>						1	-0.07	0.07	-0.19	-0.03
	CN <sub>FMIN</sub>							1	-0.08	-0.01	0.04
	MIN <sub>SUPL</sub>								1	0.06	0.13
	FRAC <sub>OPT</sub>									1	0.02
	NAVAH <sub>COEF</sub>										1

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Table A4 (continued)											
-	-	NO <sub>3</sub> RATE	NH <sub>4</sub> RATE	NO <sub>3</sub> RATE	K <sub>RM</sub>	L <sub>MYC</sub>	L <sub>M</sub>	CN <sub>FMIN</sub>	MIN <sub>SUPL</sub>	FRAC <sub>OPT</sub>	NAVAIL <sub>COEF</sub>
Nissjöfj	NO <sub>3</sub> RATE	+	0.86	0.86	0.05	-0.13	-0.20	-0.08	-0.10	0.09	-0.02
	NH <sub>4</sub> RATE		+	0.99	0.11	0.00	-0.07	-0.09	-0.07	0.15	-0.02
	NO <sub>3</sub> RATE			+	0.11	0.00	-0.07	-0.09	-0.07	0.15	-0.02
	K <sub>RM</sub>				+	-0.05	-0.06	0.01	0.05	-0.01	0.01
	L <sub>MYC</sub>					+	0.96	0.07	0.06	-0.11	-0.02
	L <sub>M</sub>						+	0.06	0.07	-0.11	-0.05
	CN <sub>FMIN</sub>							+	-0.07	-0.07	0.08
	MIN <sub>SUPL</sub>								+	-0.05	-0.04
	FRAC <sub>OPT</sub>									+	0.02
	NAVAIL <sub>COEF</sub>		-	-	-	-	-	-	-	-	+
Ljungbyhed	NO <sub>3</sub> RATE	+	0.86	0.86	-0.13	-0.32	-0.40	-0.06	0.07	0.04	-0.03
	NH <sub>4</sub> RATE		+	0.99	-0.07	-0.21	-0.28	-0.05	0.02	0.06	0.00
	NO <sub>3</sub> RATE			+	-0.07	-0.21	-0.28	-0.05	0.02	0.06	0.00
	K <sub>RM</sub>				+	-0.09	-0.08	-0.01	0.01	-0.03	-0.05
	L <sub>MYC</sub>					+	0.96	0.01	-0.08	-0.04	0.12
	L <sub>M</sub>						+	0.02	-0.10	-0.04	0.10
	CN <sub>FMIN</sub>							+	-0.03	0.16	0.04
	MIN <sub>SUPL</sub>								+	-0.07	-0.03
	FRAC <sub>OPT</sub>									+	0.01
	NAVAIL <sub>COEF</sub>		-	-	-	-	-	-	-	-	+

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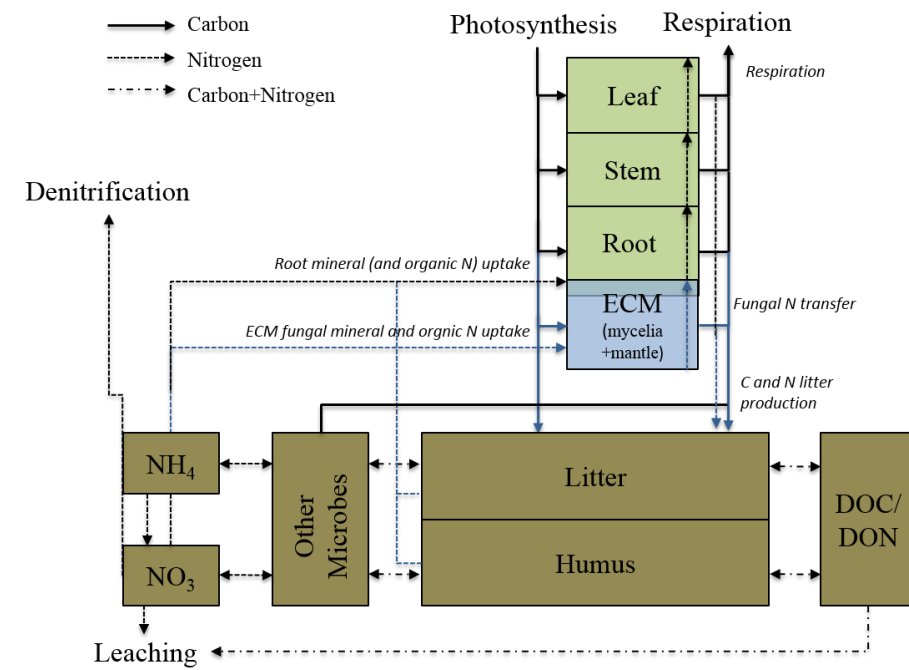
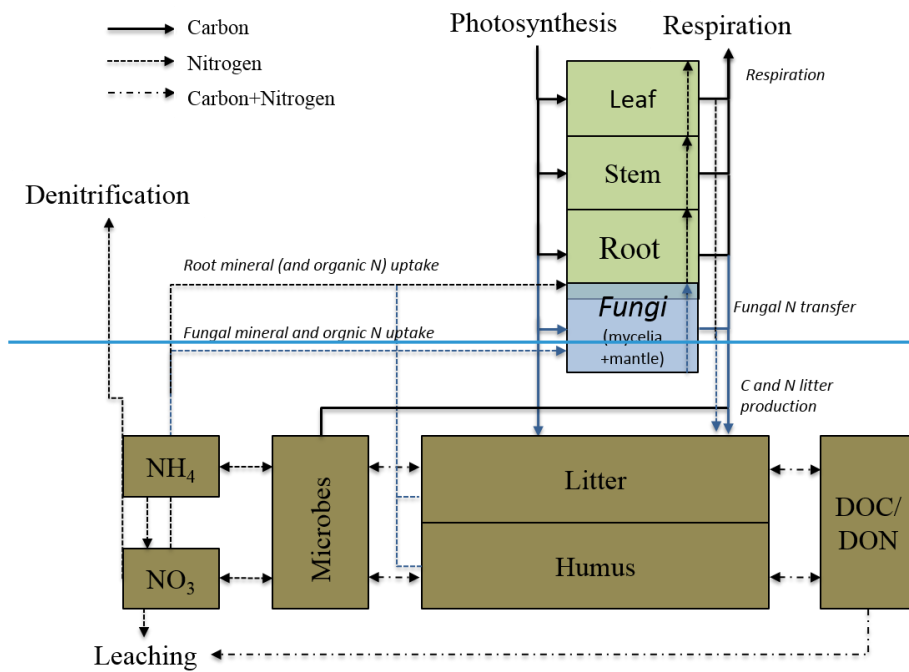
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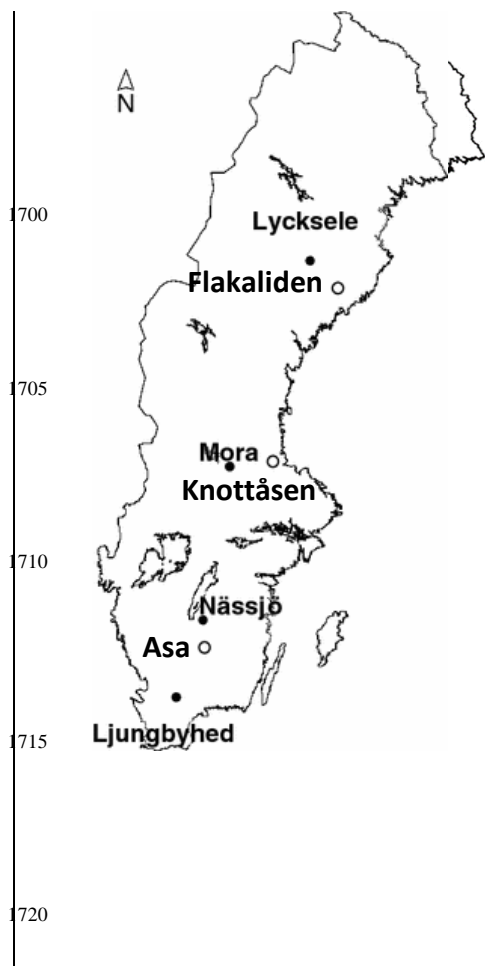




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Figure 11 A simplified overview of C and N fluxes between plants, mycorrhiza fungi, and the soil in the Coup-MYCOFON model. Light blue indicates the newly implemented MYCOFON model

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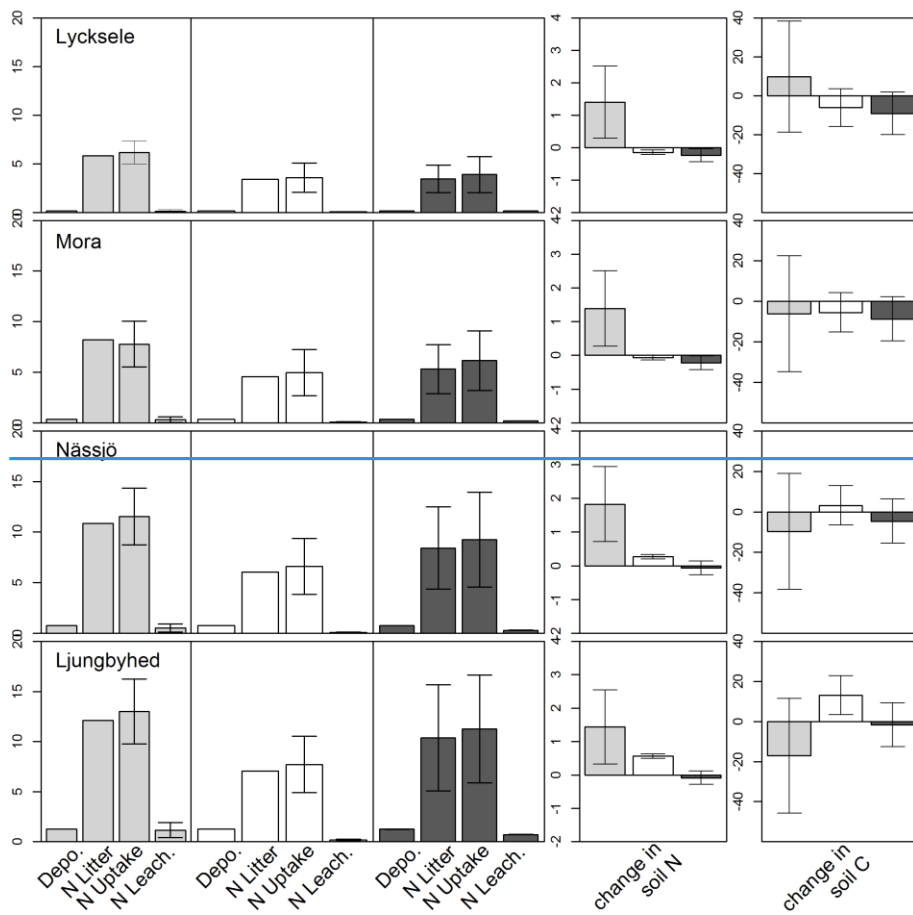


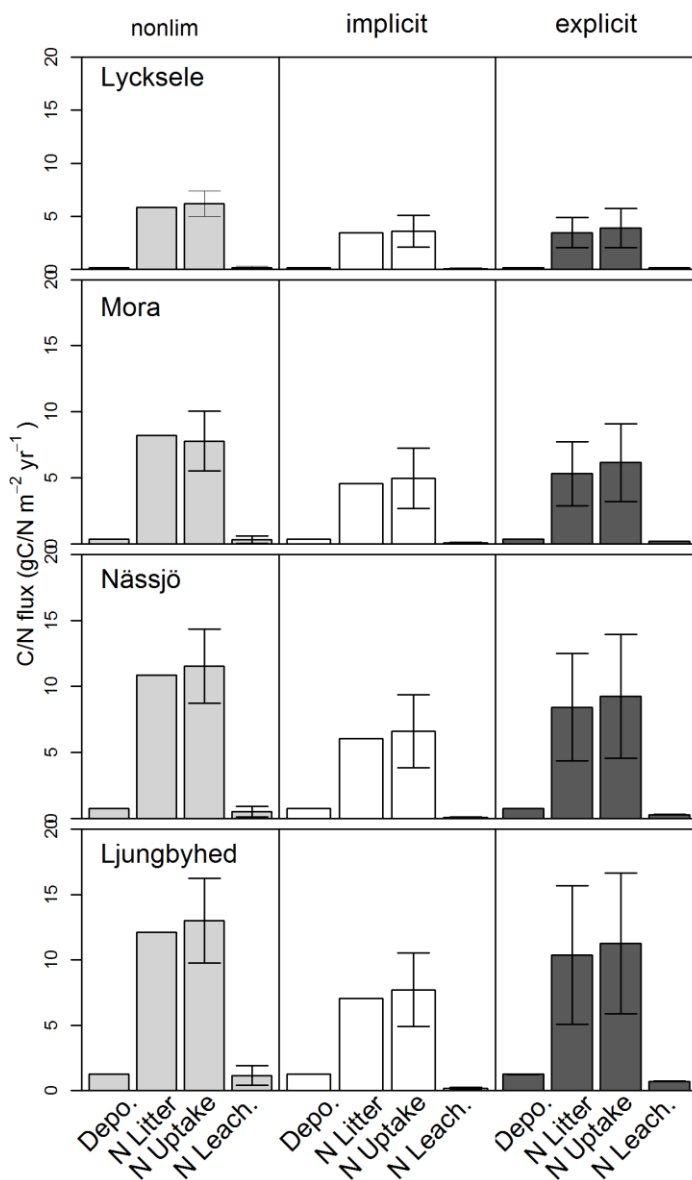
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**Figure 2 Position** Location of the four study sites in Sweden modified from Svensson et al. (2008a). Filled cycles represent the studied four sites. Open circles are the measured sites reported in Lindroth et al. (2008) used for comparison

**Figure 2** Position of the four study sites in Sweden

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**Figure 33** Soil N fluxes for the nonlim (grey columns, left), implicit (white, 2<sup>nd</sup> left column), and explicit (black, 3<sup>rd</sup> left column) model approaches, same color scheme used for the other figures. Presented are the major N inputs (N deposition, total N litter production, added to the soil litter pool by fresh litter), and outputs (N uptake from the plant/ECM fungi, N leaching), and the net change in the total soil N pool (mineral and organic). For C, the net change is presented (right column). Error bars indicate the 90<sup>th</sup> percentile of accepted model runs (posterior). Units for N are g N m<sup>-2</sup> yr<sup>-1</sup> and g C m<sup>-2</sup> yr<sup>-1</sup> for C

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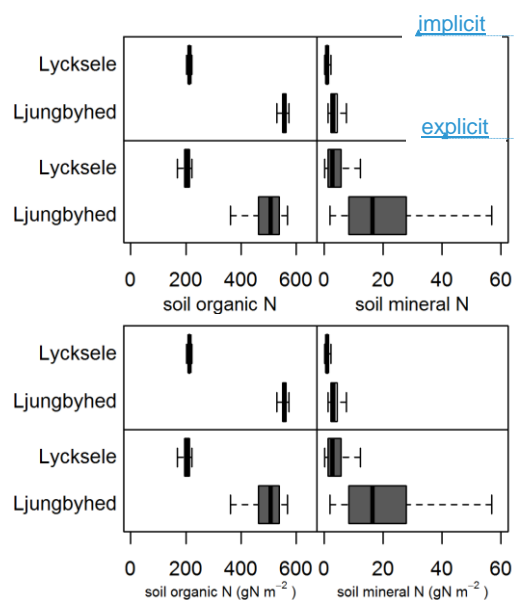
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Figure 44 Average soil organic and mineral (ammonium and nitrate) content (g N m<sup>-2</sup>) in the implicit ECM model (upper graph) and explicit ECM model (lower graph) for the two sites Lycksele and Ljungbyhed. Box plots indicate the median (bold line), the 25<sup>th</sup> and 75<sup>th</sup> percentile (bars), and the 10<sup>th</sup> and 90<sup>th</sup> percentile (whiskers)

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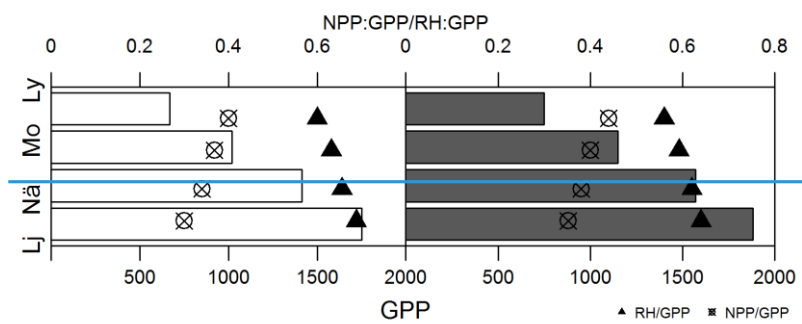
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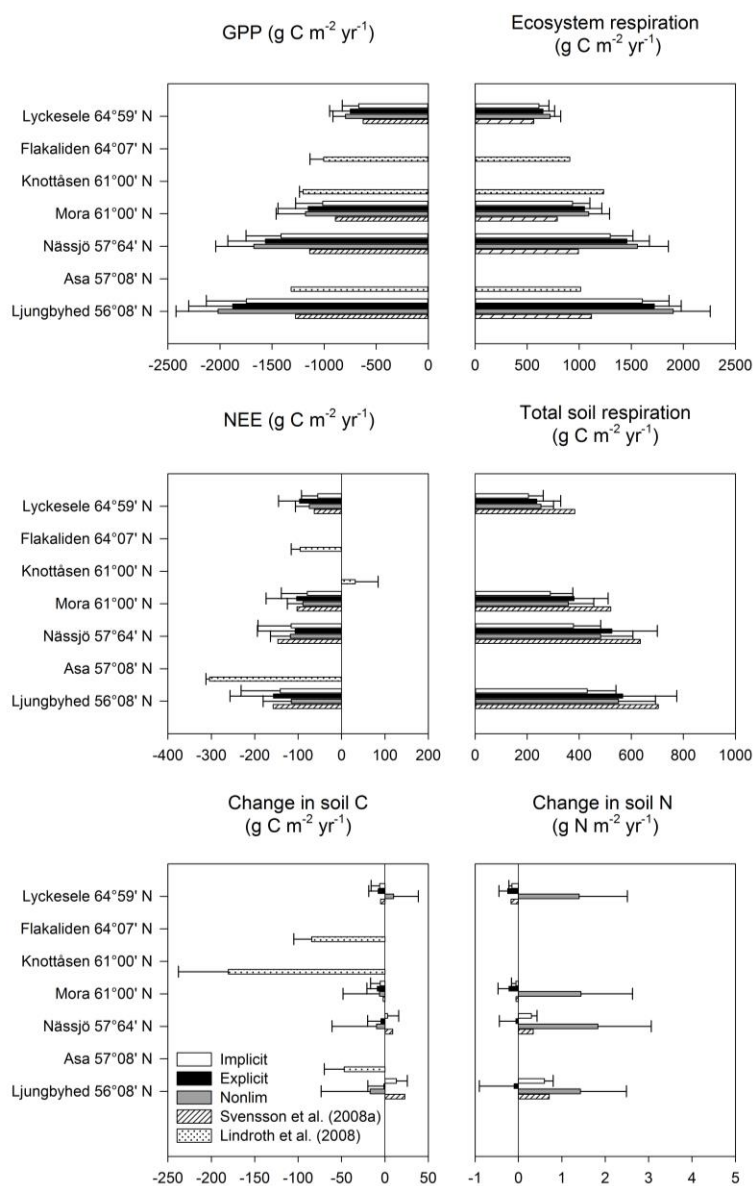
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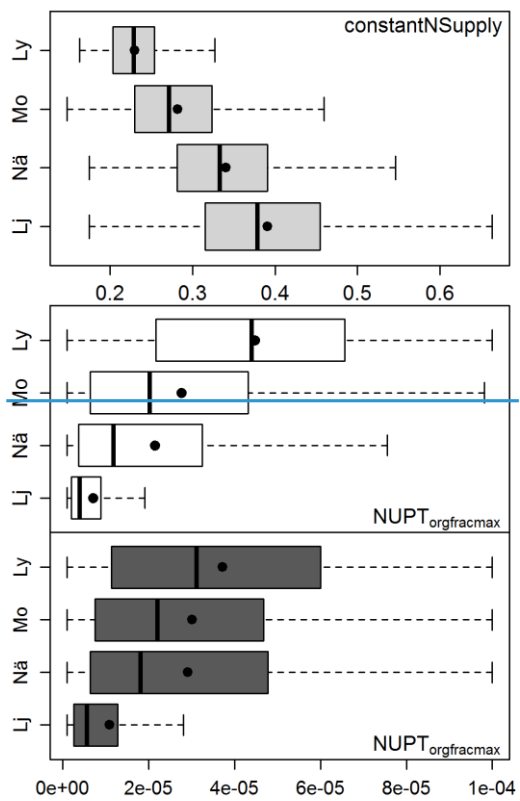


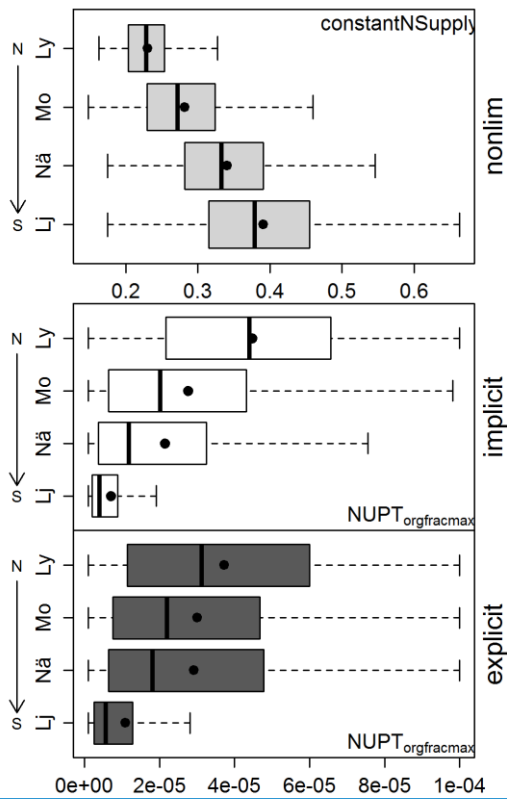




**Figure 55** GPP (bars), Rh/GPP ratio (triangle), and NPP/GPP ratio (cross circles) for all four sites simulated with the implicit (left) and explicit (right) ECM model approach. Simulated GPP, ecosystem respiration, NEE, soil respiration, change in soil C and change in soil N for all four sites with the three ECM modeling approaches and also compared with modelled data by Svensson et al. (2008a) and measurements by Lindroth et al. (2008) and the net change in the total soil N pool (mineral and organic). For C, the net change is presented (right column).







**Figure 6** Posterior parameter distributions for N uptake parameters: constant N supply rate in the “nonlim” approach (light grey), and organic N uptake capacity in the implicit (white) and explicit (dark grey/black) ECM model approaches. Distributions are presented as box plots over the prior range of variation (corresponding to the range in the x-axis). Box plots depict the median (bold line), the 25th and 75th percentile (bars), and the 10<sup>th</sup> and 90<sup>th</sup> percentile (whiskers)

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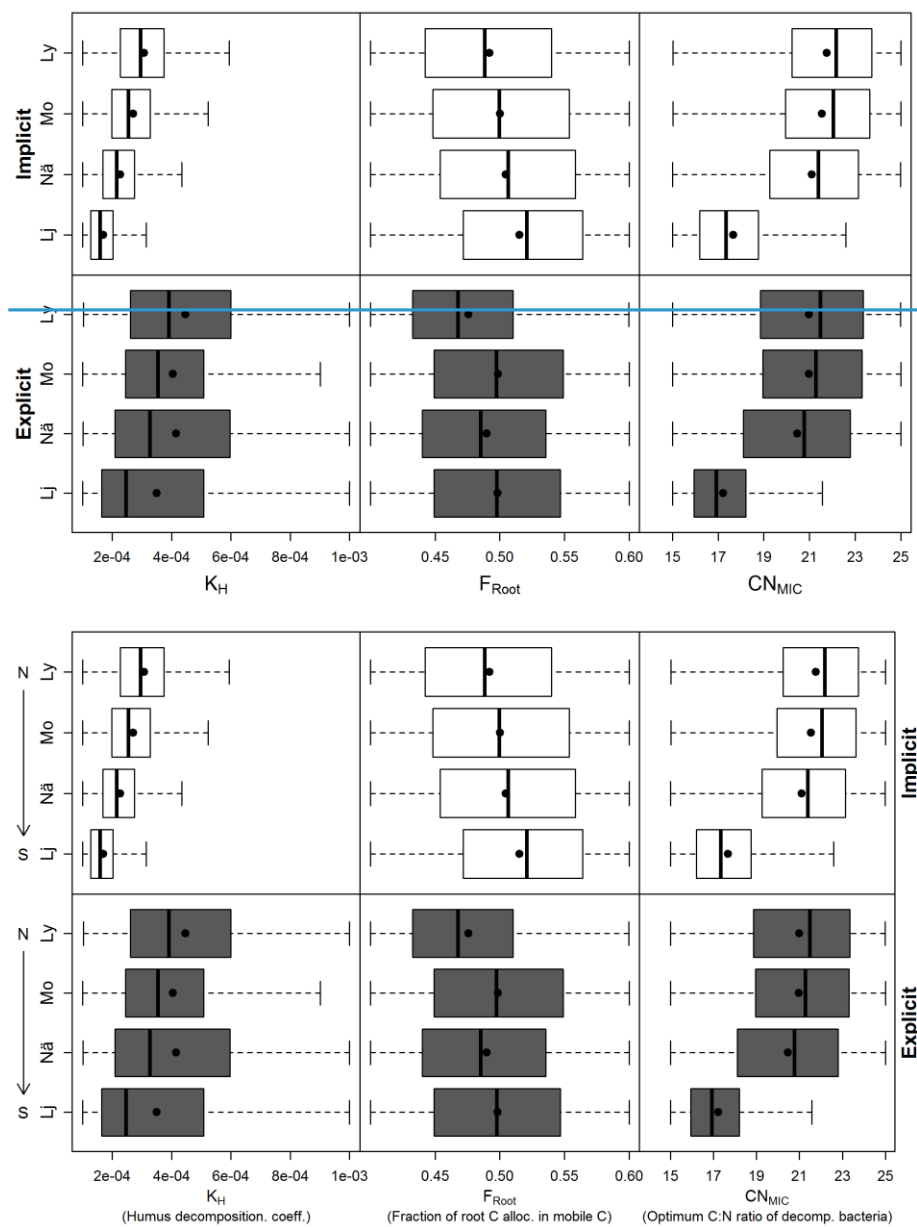


Figure 7 Posterior parameter distributions for common parameters using the implicit (top: white boxes) and explicit (bottom: dark grey boxes) ECM approaches for four different sites from North to South. Distributions are presented as box plots over the prior range of variation (corresponding to the range in the x-axis). Box plots depict the median (bold line), the 25th and 75th percentile (bars), and the 10<sup>th</sup> and 90<sup>th</sup> percentile (whiskers). The parameters

1780 shown are:  $K_H$ : the humus decomposition coefficient,  $F_{Root}$ : the fraction of C assimilates distributed to the roots, and  $E_{CM}$ ,  $CN_{mic}$ : the microbial C/N ratio

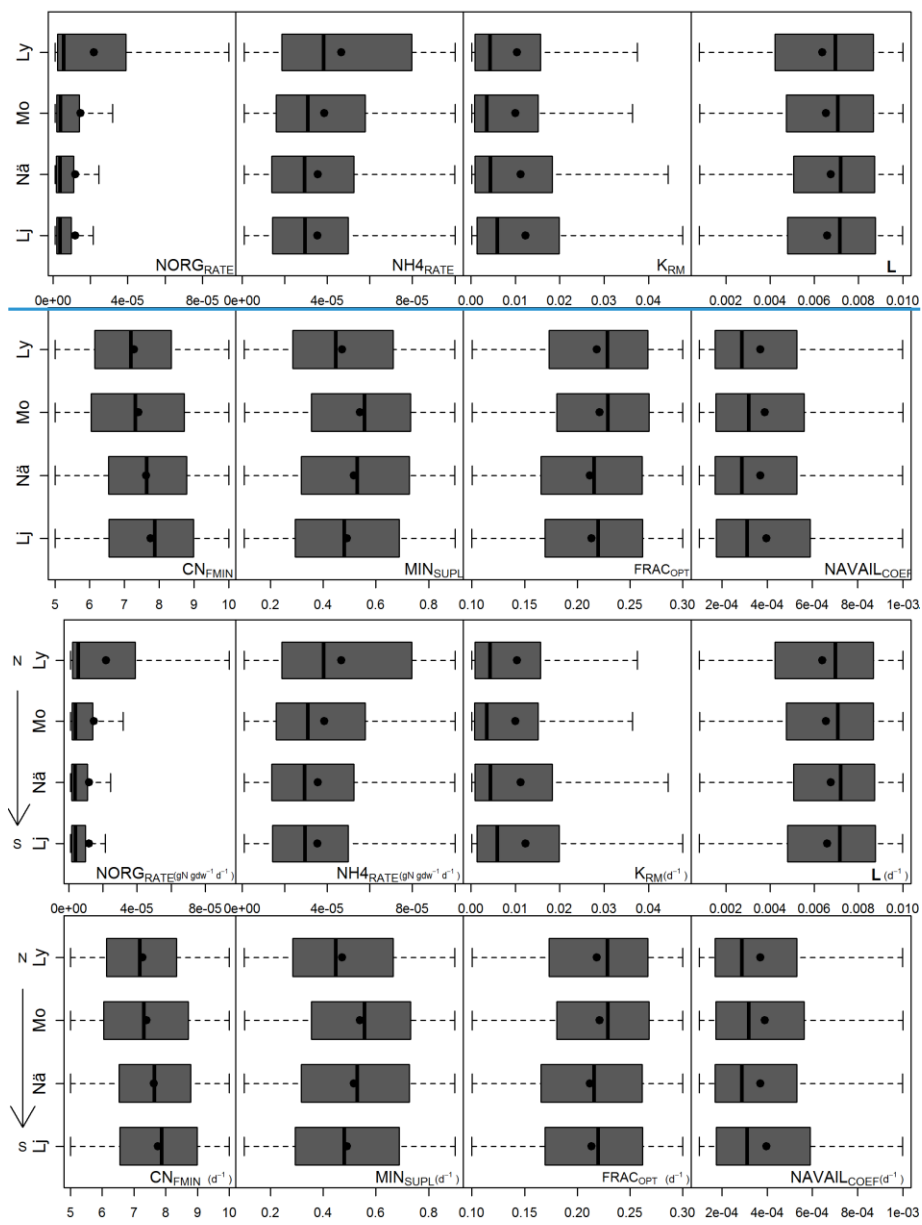


Figure 8 Posterior parameter distributions of ECM fungal specific parameters (from top left to bottom right): organic N uptake rate ( $NORGRATE$ ), ammonium uptake rate ( $NH4RATE$ ), respiration coefficient ( $K_{RM}$ ), ECM fungal litter rate coefficient (the rate at which mycelia and mantle die and add to the soil litter pool,  $L$ ), minimum ECM fungal C/N ratio ( $CN_{FMIN}$ ), fungal-ECM minimum N supply to plant ( $MIN_{SUPL}$ ), optimum ratio between fungal-ECM and root C content ( $FRAC_{OPT}$ ), and N sensitivity coefficient ( $NAVAIL_{COEF}$ ). Distributions are presented as box plots over the prior range of

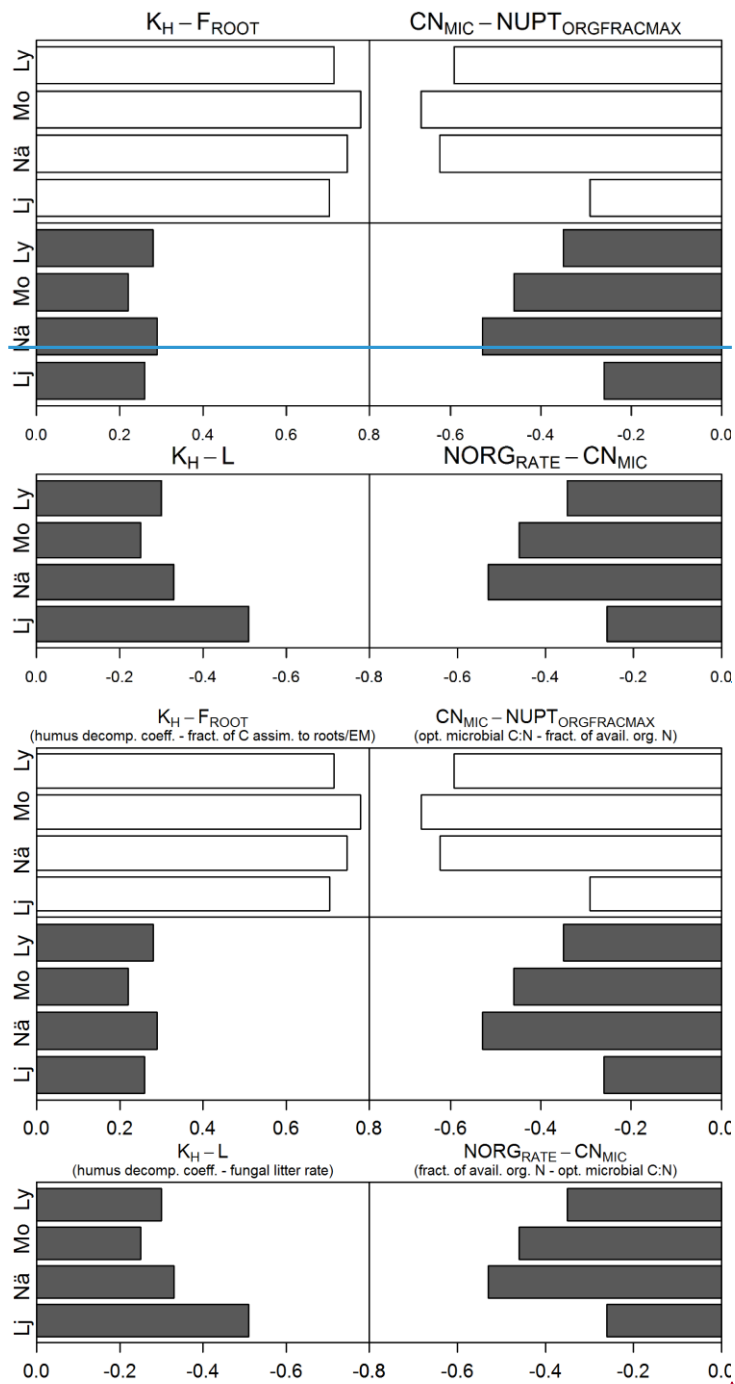
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variation (corresponding to the range in the x-axis). Box plots depict the median (bold line), the mean (black point), the 25th and 75th percentile (bars), and the 10<sup>th</sup> and 90<sup>th</sup> percentile (whiskers)

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Figure 9 Correlation between model parameters, given as the Pearson correlation coefficient, for the implicit (white) and explicit ECM (black) approaches. Top left: correlation between humus decomposition coefficient ( $K_H$ ) and the fraction of C assimilates (GPP) directed to ECM and roots ( $F_{ROOT}$ ). Top right: C/N of microbes ( $CN_{MIC}$ ) and fraction of organic N available for uptake ( $NUPT_{ORGFACMAX}$ ). Correlation between ECM fungal parameters: bottom left: humus decomposition coefficient ( $K_H$ ) and ECM fungal litter rate (L). Bottom right: fungal-ECM organic N uptake ( $NORG_{RATE}$ ) and C/N of microbes ( $CN_{MIC}$ )

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**Table 1 Main characteristics of previous ecosystem models include ECM**

Models	Time step	Elements included	Differentiation in mycelia and mantle	Organic matter decomposition	C allocation	Plant N uptake	Is sensitive to soil N
<a href="#">ANAFOR</a> , <a href="#">Deckmyn et al. (2011)</a>	hourly	<a href="#">C, N</a>	<a href="#">No</a>	<a href="#">Yes</a>	<a href="#">Fraction of C allocated to roots, regulated by water and N</a>	<a href="#">Function of the available mineral and organic N pools</a>	<a href="#">No</a>
<a href="#">MoBiLE</a> and <a href="#">Mycofon</a> , <a href="#">Meyer et al. (2010, 2012)</a>	Daily	<a href="#">C, N</a>	<a href="#">Yes</a>	<a href="#">No</a>	<a href="#">A certain ratio between root and ECM biomass exists to reach the optimum degree of mycorrhization, regulated by soil N and temperature</a>	<a href="#">Separated root and mycelia mineral N uptake and regulated by plant and ECM N demand</a>	<a href="#">Yes</a>
<a href="#">MySCaN</a> , <a href="#">Orwin et al. (2011)</a>	Daily	<a href="#">C, N, P</a>	<a href="#">No</a>	<a href="#">Yes</a>	<a href="#">Constant fraction of plant C assimilates, modified by nutrients</a>	<a href="#">Driven by C to nutrient ratios in pools</a>	<a href="#">No</a>
<a href="#">Moore et al. (2015) model</a>	Monthly	<a href="#">C</a>	<a href="#">No</a>	<a href="#">Yes</a>	<a href="#">Constant fraction of plant C assimilates</a>		<a href="#">No</a>
<a href="#">Baskaran et al. (2016) model</a>	Annual	<a href="#">C, N</a>	<a href="#">No</a>	<a href="#">No</a>	<a href="#">Constant fraction of plant C assimilates</a>	<a href="#">Root inorganic N uptake by Michaelis-Menten function and ECM N uptake by ECM C to N ratio</a>	<a href="#">No</a>
<a href="#">Coup-MYCOFON</a> (This study)	Daily	<a href="#">C, N</a>	<a href="#">Yes</a>	<a href="#">No</a>	<a href="#">Similar to MoBiLE</a>	<a href="#">Similar to MoBiLE, but allows organic N uptake for ECM</a>	<a href="#">Yes</a>

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**Table 4.2** Maximum and minimum parameters values prior to Bayesian calibration for the nonlim, implicit, and explicit model approaches

**A. Common parameters (all three approaches, including the “implicit” approach)**

Parameter	Unit	Min	Max
<i>Humus decomposition</i>			
$K_H$	$d^{-1}$	0.0001	0.001
<i>Fraction of organic N available for uptake</i>			
$NUPT_{ORGFRACMAX}$	$d^{-1}$	0.000001	0.0001
<i>Fraction of root C allocation in mobile C</i>			
$F_{ROOT}$	$d^{-1}$	0.4	0.6
<i>C/N ratio of decomposing microbes</i>			
$CN_{MIC}$	$d^{-1}$	15	25

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**B. Parameters of the “nonlim” approach**

Parameter	Unit	Min	Max
<i>Plant N Supply</i>			
$ConstantN_{Supply}$	-	0.1	0.7

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**C. Fungal-ECM fungal parameters of the “explicit” approach**

Parameter	Unit	Min	Max
<i>Fungal-ECM N uptake</i>			
$NORG_{RATE}$	$g\ N\ gdw^{-1}\ d^{-1}$	0.000001 <sup>a</sup>	0.0001
$NH4_{RATE}$	$g\ N\ gdw^{-1}\ d^{-1}$	0.000001 <sup>a</sup>	0.0001
$NO3_{RATE}$	$g\ N\ gdw^{-1}\ d^{-1}$	0.000001 <sup>a</sup>	0.0001
<i>Fungal-ECM respiration coefficient</i>			
$K_{RM}$	$d^{-1}$	0.0002 <sup>b</sup>	0.05
<i>Fungal-ECM litter rate</i>			
$L$	$d^{-1}$	0.0008 <sup>c</sup>	0.01
<i>Minimum ECM fungal C/N ratio</i>			
$CN_{FMIN}$	$d^{-1}$	5 <sup>d</sup>	10
<i>Fungal-ECM minimum N supply to plant</i>			
$MIN_{SUPL}$	$d^{-1}$	0.1 <sup>e</sup>	0.9
<i>Optimum ECM fungi C allocation fraction</i>			
$FRAC_{OPT}$	$d^{-1}$	0.1 <sup>f</sup>	0.3 <sup>f</sup>

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<i>N</i>	<i>sensitivity coefficient</i>			
NAVAIL <sub>COEF</sub>	d <sup>-1</sup>	0.0001	0.001	
<sup>a</sup> Plassard et al.; (1991);, Chalot et al.; (1995);, and Smith and Read; (2008)				
<sup>b</sup> <del>set</del> <u>Set</u> equally to trees according to Thornley and Cannell; (2000)				
<sup>c</sup> Staddon et al.; (2003);, and Ekblad et al.; (2013)				
<sup>d</sup> Högberg and Högberg; (2002) and; Wallander and Nilsson; (2003)				
<sup>e</sup> <del>estimated</del> <u>Estimated</u>				
<sup>f</sup> Leake (2007), Staddon et al.; (2003), and; Johnson et al.; (2005)				

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**Table 2.3 Climatic and soil data, and initial settings of the four study soils applied in all model approaches**

Sites	Location	Altitude (m asl)	Driving data				Calibration data			
			Air temperature <sup>a</sup> (°C)	Precipitation <sup>a</sup> (mm)	N deposition <sup>a</sup> (kg N ha <sup>-1</sup> yr <sup>-1</sup> )	Soil C (g C m <sup>-2</sup> )	Soil N (g N m <sup>-2</sup> )	Soil C/N <sup>e</sup>	Standing stock (g C m <sup>-2</sup> ) <sup>b</sup>	N deposition <sup>a</sup> (kg N ha <sup>-1</sup> yr <sup>-1</sup> )
Lycksele	64°59'N 18°66'E	223	0.7	613	1.5	7006	223	31.5	5371	1.5
Mora	61°00'N 14°59'E	161	3.3	630	3.5	8567	295	29.1	7815	3.5
Nässjö	57°64'N 14°69'E	305	5.2	712	7.5	9995	367	27.2	10443	7.5
Ljungbyhed	56°08'N 13°23'E	76	7.1	838	12.5	10666	539	19.8	11501	12.5

<sup>a</sup> 30-year (1961 to 1991) annual average

<sup>b</sup> according to Skogsdata for a 100-year-old forest (2003: <http://www.slu.se/en/webbtjanster-miljoanalys/forest-statistics/skogsdata/>)

<sup>c</sup> used as calibration parameter

<sup>d</sup> used as driving data

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▲ **Table 4 Comparison between modelled soil C and N of this study and literature value**

Reference	Site	Ecosystem type	Forest age (years)	Soil C change (g C m <sup>-2</sup> yr <sup>-1</sup> )	Soil N change (g N m <sup>-2</sup> yr <sup>-1</sup> )
▲ <u>Svensson et al. 2008a</u>	Lycksele	Coniferous on podzol	100	-5	
	Mora			-2	
	Nässjö			9	
	Ljungbyhed			23	
▲ <u>Lindroth et al. 2008</u>	Flakaliden	Coniferous on podzol	39-42 (in 2002)	-79 <sup>a</sup>	
	Knottåsen			-133 <sup>a</sup>	
	Asa			-24	
▲ <u>This study</u>	Lycksele	Coniferous on podzol	100		
	Mora			-6 to 13.1 <sup>b</sup>	-0.2 to 0.6 <sup>b</sup>
	Nässjö			-8.7 to -1.6 <sup>c</sup>	-0.2 to -0.1 <sup>c</sup>
	Ljungbyhed				

<sup>a</sup>mean Mean of the highest and lowest error estimates

<sup>b</sup>implicit Implicit approach

<sup>c</sup>explicit Explicit approach

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