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Contribution of emissions to concentrations: The TAGGING 1.0 submodel based on the Modular Earth Submodel System (MESSy 2.52)

Volker Grewe[1,2], Eleni Tsati[1], Mariano Mertens[1], Christine Frömming[1], and Patrick Jöckel[1]

[1]Deutsches Zentrum für Luft- und Raumfahrt, Institut für Physik der Atmosphäre, Oberpfaffenhofen, Germany
[2]also at: Delft University of Technology, Aerospace Engineering, Section Aircraft Noise and Climate Effects, Delft, Netherlands

Correspondence to: Volker Grewe (volker.grewe@dlr.de)

Abstract. Questions such as "What is the contribution of road traffic emissions to climate change?" or "What is the impact of shipping emissions on local air quality?" requires a quantification of the contribution of specific emissions sectors to the concentration of radiatively active species and air quality related species, respectively. Here, we present a diagnostics, implemented in the Modular
5   Earth-System Model MESSy, which keeps track of the contribution of source categories (mainly emission sectors) to various concentrations. The diagnostics is implemented as a submodel (TAGGING) of EMAC (European Centre for Medium-Range Weather Forecasts - Hamburg (ECHAM)/Modular Earth Submodel System (MESSy) Atmospheric Chemistry). It determines the contributions of 10 different source categories to the concentration of ozone, nitrogen oxides, peroxyacytyl nitrate, car-
10   bon monoxide, non-methane hydrocarbons, hydroxyl and hydroperoxyl radicals (=tagged tracers). The source categories are mainly emission sectors and some other sources for completeness. As emission sectors, road traffic, shipping, air traffic, anthropogenic non-traffic, biogenic, biomass burning, and lightning are considered. The submodel obtains information on the chemical reaction rates, online emissions such as lightning, and wash-out rates. It then solves differential equations for the
15   contribution of a source category to each of the seven tracers. This diagnostics does not feed back to any other part of the model. For the first time, it takes into account chemically competing effects: For example the competition between NO_x, CO, and NMHCs in the production and destruction of ozone. We show that the results are in-line with results from other tagging schemes and provide plausibility checks for concentrations of trace gases such as OH and HO_2, which have not previously
20   been tagged. The budgets of the tagged tracers, i.e. the contribution from individual source categories (mainly emission sectors) to, e.g., ozone, are only marginally sensitive to changes in model resolu-

tion, though the level of detail increases. A reduction in road traffic emissions by 5% shows that road traffic global tropospheric ozone is reduced by 4% only, because the net ozone productivity increases. This 4% reduction in road traffic tropospheric ozone corresponds to a reduction in total

25 tropospheric ozone by ≈0.3%, which is compensated by an increase in tropospheric ozone from other sources by 0.1%, resulting in a reduction in total tropospheric ozone of ≈0.2%. This compensating effect compares well previous findings. The computational costs of the TAGGING submodel are low with respect to computing time, but a large number of additional tracers are required. The advantage of the tagging scheme is that in one simulation and at every time step and grid point, infor-

30 mation is available on the contribution of different emission sectors to the ozone budget, which then can be further used in upcoming studies to calculate the respective radiative forcing simultaneously.

1 Introduction

Nitrogen oxides (NO_x), carbon monoxide (CO), methane (CH_4) and non-methane hydrocarbons (NMHC) are precursors of tropospheric ozone (O_3). The assessment of the contribution of indi-

35 vidual emissions of these precursors on air quality and climate requires a detailed analysis of the chemical conversion, transport and deposition of these species in numerical atmosphere-chemistry simulations. A frequently used method is called 'tagging' (Horowitz and Jacob, 1999; Lelieveld and Dentener, 2000; Meijer et al., 2000; Grewe, 2004; Gromov et al., 2010; Grewe et al., 2012). Technically, this method adds a set of diagnostic tracers for each chemical species considered, i.e. one addi-

40 tional tracer per source category for each chemical species considered. For example, for the species NO_x a set of tagged tracers NO_x^{ant}, NO_x^{rt}, NO_x^{shp}, NO_x^{air}, NO_x^{bio}, NO_x^{bb}, NO_x^{lig}, NO_x^{ch4}, NO_x^{n2o}, and NO_x^{str} is added, which describes the NO_x concentration from anthropogenic non-traffic (e.g. industry, households), road traffic, ships, air traffic, biogenic, biomass burning, lightning, methane and nitrous oxide decomposition and stratospheric ozone production. The idea is that these tagged tracers

45 experience the same chemical conversions and loss processes (such as deposition), as the simulated tracer NO_x. If all emissions of NO_x are considered and tagged, the sum of all tagged diagnostic NO_x tracers equals the simulated NO_x tracer in this approach. A full partition of the simulated tracer concentration with respect to emission sectors can be achieved. Thus, the contribution of an emission sector, such as industry, road traffic, etc. to a concentration is provided by the tagging method.

50 A different approach, the perturbation approach (e.g. Hoor et al., 2009; Grewe et al, 2007), where results from two simulations are compared which differ in the strength of an individual emission source, identify the impact of changes in emissions (e.g. by mitigation options) on the atmospheric composition. It is important not to confuse both approaches. For example, the change in ozone due to a 100% reduction in road traffic emissions is smaller by a factor of 5 than the contribution of

55 the road traffic emissions to ozone (Grewe et al., 2012). Emmons et al. (2012) showed that similar results (factor of 3) are obtained for biomass burning NO_x emissions and the impact on ozone.

Clearly, the non-linearity in the ozone chemistry leads to these large differences. Any reduction in NO_x emission leads mostly to a larger ozone production efficiency. Grewe et al. (2012) showed that in the simulation without road traffic NO_x emissions, the obvious large reduction in ozone
60   from the reduced road traffic contribution to ozone is compensated by larger contributions from other emission sectors, not because these emissions are changed, but because the ozone production efficiency is increased.

These two different approaches answer two different questions. The perturbation approach quantifies how much a concentration changes if emissions are changed, whereas tagging addresses the
65   contribution of an emission to the concentration. The combination of both approaches leads to much better insights in the reasons how emission changes lead to concentration changes (Grewe et al., 2012). Note also that the perturbation approach often requires the identical meteorology in either simulation to enhance the signal-to-noise ratio enabling a robust signal. However, this is not feasible in fully coupled chemistry-climate models unless run in a so-called "QCTM-mode", which replaces
70   instantaneous chemical feedbacks by climatological values (Deckert et al., 2011, see also below).

Most tagging approaches address a straight process chain from the emission of e.g. NO_x to a concentration of e.g. ozone. Grewe et al. (2010), as well as Grewe (2013a) and Tsati (2014) proposed a more general tagging approach, where competing mechanisms in the production of ozone can be taken into account, e.g. both NO_x and carbon compounds (CO, CH_4, NMHCs) are precursors of
75   ozone. This more general tagging approach allows the contribution of road traffic NO_x, CO, and NMHC emissions to ozone, for example, to be determined. This generalised method has also been successfully applied to a non-chemical application, namely temperature in an energy balance model (Grewe, 2013b).

Here, we present a submodel (TAGGING) of an Earth-System-Model (EMAC), which applies this
80   general tagging approach to allow the contribution of NO_x, CO and NMHC emissions from a variety of emission sectors to ozone and HO_x chemistry to be quantified. In Section 2 we present the basic equations of the tagging scheme, whereas in Section 3 we present what emissions are addressed and how the tagging method is implemented. In Section 4 we show results of a base simulation and compare them with other modelling studies. Since no measurements are available for contributions
85   of emissions to ozone concentrations, a direct comparison with observational data is not possible. Instead, we show that the results are in agreement with other studies. Since the tagging of HO_x components is new, we discuss those results in more detail, especially with the focus on aviation and shipping emissions. Finally, we address sensitivities of the methodology (Sec. 5), with respect to the resolution and emission changes, and provide a comparison of the perturbation and tagging method.

2 Basics on tagging

The tagging approach, which we adopt here is based on Grewe et al. (2010) and Grewe (2013a). We first describe the basic mechanism and describe in Sec. 3 how this mechanism is applied in the submodel TAGGING. Exemplarily, we concentrate on the main reaction for tropospheric ozone production (rate limiting step; NO_2 is photolysed and recombines with O_2 to form ozone):

$$NO + HO_2 \longrightarrow NO_2 + OH. \tag{R1}$$

The ozone production rate P_{R1} depends on the abundance of NO and HO_2, and the reaction rate coefficient k_{R1} (Reaction R1). The NO concentration in turn depends on emissions of NO from different emission sectors (here N in total), such as industry and road traffic with the respective concentration NO_x^{ind} and NO_x^{rt}. Thus, the ozone production rate P_{R1} has to be distributed to the sectors industry, road traffic, etc. This is achieved by a combinatoric redistribution according to the concentrations of the tagged species of NO_x and HO_2. This means that all possible combinations between a tagged NO_x species and another tagged HO_2 species are evaluated and its probability calculated consistently with the calculation of the chemical production rate P_{R1}. This is just a full partitioning of the production rate P_{R1}:

$$P_{R1} = k_{R1} NO \, HO_2 \tag{1}$$

$$= k_{R1} \sum_{i=1}^{N} NO^i \sum_{j=1}^{N} HO_2^j \tag{2}$$

Here i and j represent a counter for all N source categories; Here, we have chosen $N = 10$ source categories (see Sec. 3) With this combinatorial approach, Grewe et al. (2010) showed (here as an example for industry) that the reaction rate P_{R1}^{ind}, that is the ozone production due to NO_x from industry (NO_x^{ind}) and due to HO_2^{ind} from industry equals:

$$P_{R1}^{ind} = P_{R1} \frac{1}{2} \left(\frac{NO^{ind}}{NO} + \frac{HO_2^{ind}}{HO_2} \right). \tag{3}$$

Note that this includes the reactions of NO_x^{ind} with all other HO_2 molecules and vice versa HO_2^{ind} with other NO_x molecules without any double counting. The relevant differential equation for the tagged species is then

$$\frac{d}{dt} O_3^{ind} = P^{ind} - D^{ind}, \tag{4}$$

where P^{ind} and D^{ind} are the sum of all relevant production and loss terms. With this approach, Grewe et al. (2010) showed that the sum of all emissions contributions adds up to the total concentration of the respective species. For example, the ozone field is completely partitioned into emission sectors contributions, if all emission sectors are included, leading to

$$\sum_{i=1}^{N} O_3^i = O_3. \tag{5}$$

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This approach is identical to a different formulation, which describes the right hand side of the differential equation more generally as the relative sensitivity of the individual production and loss terms with respect to the emission sector considered (Grewe, 2013a):

$$P_{R1}^{ind} \quad = \quad P_{R1} \frac{S^{ind^T} \nabla_S P_{R1}}{S^T \nabla_S P_{R1}}, \tag{6}$$

125 where S is the vector of all chemical compounds, e.g.,

$$S^T \quad = \quad (NO_x, CO, NMHC, O_3, ...)^T, \text{ and} \tag{7}$$

$$S^{ind^T} \quad = \quad (NO_x^{ind}, CO^{ind}, NMHC^{ind}, O_3^{ind}, ...)^T, \tag{8}$$

and

$$\nabla_S P_{R1} \quad = \quad \frac{d}{dS} P_{R1} \tag{9}$$

130 providing two different interpretations of the differential equation (4).

To summarise, this tagging approach fully partitions individual chemical fields into the contribution of individual emission sectors. There is no linearisation required and the approach utilises the identical chemical parameterisation as the underlying chemical scheme, with respect to the probability that a reaction occurs. Note that the new aspect of this tagging approach compared to other tagging approaches (Grewe, 2007; Lelieveld and Dentener, 2000; Emmons et al., 2012) is the competing effect of NO_x and carbon compounds in producing ozone. Since the differential equation for the tagging scheme (4) fully relies on the reaction rates and concentrations, the tagging scheme can be implemented independently from the main chemical solver. However, details on many reaction rates have to be transferred from the chemical solver to the tagging scheme.

3 Implementation in EMAC/MECO(n)

The objective of the implementation of this tagging scheme is to be able to monitor online, i.e. at every model's timestep, the contribution of individual emission sectors to ozone and OH, allowing a competition between ozone precursors, linearisation to be avoided, and applicable in decadal simulations. The tagging approach requires to quantify all sources of the species considered. Therefore, in 145 addition to the emission sectors considered, there are additional source categories considered, such as ozone produced by photolysis of oxygen in the stratosphere. In the following the base models are described for which the tagging scheme is developed, an overview on the tagging scheme is given, and the tagging chemistry is described.

3.1 Description of MESSy, EMAC and MECO(n)

150 The TAGGING model described here (see also Tsati, 2014) is written as a submodel of the Modular Earth Submodel System (MESSy), which comprises a standard interface to couple different processes, a simple coding standard and a set of different submodels (Jöckel et al., 2005). The TAG-

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Table 1. Brief description of the submodels used together with the TAGGING submodel. A complete list can be found in the supplement of Mertens et al. (2016).

Submodel	Description	Reference
CLOUD	large scale cloud/rain properties	based on Roeckner et al. (2003) see also Jöckel et al. (2006)
CONVECT	convective cloud/rain properties and related transport)	Tost et al. (2006a)
DDEP	dry deposition of trace gases	Kerkweg et al. (2006a)
JVAL	photolysis rates	Landgraf and Crutzen (1998), see also Jöckel et al. (2006)
LNOX	lightning NO_x emissions	Tost et al. (2007b) and Grewe et al. (2001)
MECCA	tropospheric and stratospheric gas-phase chemistry	Sander et al. (2011)
OFFEMIS	prescribed emissions of trace gases	Kerkweg et al. (2006b) (named OFFLEM therein)
ONEMIS	online calculated emissions of trace gases	Kerkweg et al. (2006b) (named ONLEM therein)
SCAV	wet deposition and scavenging of trace gases	Tost et al. (2006b)

GING submodel is implemented in MESSy2 (Jöckel et al., 2010) and consists of two parts, the Sub-Model Interface Layer (SMIL) and the SubModel Core Layer (SMCL). The SMIL part is mainly

155 important for data management, defining and handling the tracers (using the TRACER submodel described in Jöckel et al. (2008)) and the diagnostic output fields using the CHANNEL submodel (Jöckel et al., 2010). The coupling for the necessary input fields are also handled via the CHANNEL submodel. These input fields comprise, for example, lightning NO_x emissions and chemical production/loss-rates from the chemical solver MECCA (Module Efficiently Calculating the Chem-

160 istry of the Atmosphere, Sander et al. (2011)).

The TAGGING submodel is implemented in EMAC (ECHAM5/MESSy Atmospheric Chemistry) and MECO(n) (MESSyfied ECHAM and the Consortium for Small-Scale Modelling model COSMO nested n-times). While EMAC uses ECHAM5 as a global circulation model, MECO(n) consists of COSMO/MESSy as a regional-scale model with EMAC as the driving model (Kerkweg and Jöckel,

165 2012a), which are coupled online. The SMCL of the TAGGING submodel is independent of the base model and consists mainly of the code needed to solve the relevant equations. A detailed description of the TAGGING submodel, including individual subroutines of the SMIL and the SMCL, are provided in the supplement. The model set-up is identical to that of Mertens et al. (2016). A detailed list of applied submodels can be found in the supplement of Mertens et al. (2016) (page 42, therein).

170 Table 1 describes only those submodels, which are of direct relevance for the TAGGING submodel. An evaluation of the model configurations of EMAC and MECO(n) with respect to the chemical composition of the atmosphere can be found in Jöckel et al. (2016) and Mertens et al. (2016).

3.2 TAGGING overview: families, emission sectors and workflow

The objective of the tagging scheme is to determine the contribution of emissions from various

175 sectors. Here, we discriminate between ten different sources, four anthropogenic: non-traffic an-

Table 2. Submodels which provide the source terms (emissions or production terms) for the individual emission sectors (first column) and tagged species (columns 2-4).

Sector	Tagged species with emissions and other sources			
	NO_y	CO	NMHC	O_3
—— Anthropogenic ——				
Non-traffic	OFFEMIS	OFFEMIS	OFFEMIS	-
Road Traffic	OFFEMIS	OFFEMIS	OFFEMIS	-
Ships	OFFEMIS	OFFEMIS	OFFEMIS	-
Air Traffic	OFFEMIS	OFFEMIS	OFFEMIS	-
—— Natural ——				
Lightning	LNOX	-	-	-
Biogenic	ON-/OFFEMIS	ON-/OFFEMIS	ON-/OFFEMIS	-
N_2O	MECCA	-	-	-
CH_4	-	-	MECCA	-
Strat-O_3	-	-	-	MECCA
—— Mixed ——				
Biomass Burning	OFFEMIS	OFFEMIS	OFFEMIS	-

thropogenic (industry, energy, households), road traffic, ships, and air traffic, five natural sources: lightning, emissions from biogenic sources including soils, decomposition of N_2O, decomposition of CH_4, stratospheric ozone production by photolysis of O_2, and a mixed class: biomass burning (see Table 2).

180   We use a configuration of the chemical scheme MECCA (Sander et al., 2011), which consists of 72 species. We only tag a reduced set of species, which resemble the main species and families for tropospheric chemistry, in order to limit the required memory. Besides CO, O_3, peroxyacytyl nitrate (PAN), HO_2, and OH, 2 families are considered: NO_y and NMHC which include all chemically active nitrogen compounds (15) and hydrocarbons (42). All together, the tagging scheme consists of 185   7 species times 10 emission sectors, thus 70 tagged tracers. For each tracer initialisation, transport (except for OH and HO_2), emissions, dry and wet deposition, and chemical conversion has to be deduced from the base model (Figure 1). The tagging scheme utilises the EMAC submodels, e.g. for tracer transport, for emissions computed online during the simulation, and for emissions prescribed by inventories (Table 1; for details see supplement), such as industry, road traffic, etc. (Figure 1, 190   middle column). It further obtains information on online emissions (lightning, soils), dry and wet deposition, background tracers and reaction rates (left column). This information is processed in tagging core routines (right).

    Here, we concentrate on the TAGGING submodel (Figure 1, right column). For the initialisation of the tagged tracers two options are available. First, the variables can be initialised from files, or

195 second the tagged tracers can be initialised according to their key characteristics. In this case, the tagged stratospheric ozone is initialised by the ozone field above the tropopause and all other tagged ozone fields are zero above the tropopause and vice versa. Below the tropopause, all but the tagged stratospheric ozone tracer, obtain one ninth of the tropospheric ozone concentration.

At each time step during the simulation, the online emissions (soil emissions) are added to the
200 respectively tagged tracer (Table 2). The emission rate is obtained by recording the concentration of NO_x before and after the calculation of online emissions. The tagged lighting NO_y tracer obtains the same lightning emissions as the chemical NO tracer, which is provided by the lightning submodel LNOX (Tost et al., 2007b; Grewe et al., 2001). Dry and wet deposition is treated as a bulk process. Changes in the concentration of all relevant chemical species are calculated in a practical and simple
205 manner, by the difference in the respective concentrations before and after dry and wet deposition is calculated. This tendency of the concentration is provided to the tagging submodel and distributed among the tagged species according to their relative contribution to the total concentration.

3.3   TAGGING chemistry

The core of the tagging submodel is the distribution of the chemical tendencies to the tagged tracers
210 as introduced in Sec. 2. Therefore, the individual production and loss terms have to be determined adequately to calculate concentration changes via Eq. (4). Here, we consider effective ozone production and loss terms according to Crutzen and Schmaizl (1983). This implies that a family is considered for ozone (see supplementary material for more details), which includes all fast exchanges between ozone and other chemical species. The ozone production basically requires splitting up an oxygen
215 molecule. For the identification of ozone production and loss reactions, we apply the tool *ProdLoss* (see supplementary material for more detailed information), which identifies the effective production and loss reactions for a family in the selected chemical mechanism. This family for effective ozone is hereafter referred to as ozone for simplicity. This results in two ozone production terms, which are applied to any tagged ozone field with the exception of stratospheric ozone. This is reaction (R1) and
220 the combination of reactions of the type (see supplementary material for more detailed information)

$$NO + RO_2 \longrightarrow NO2 + RO, \tag{R2}$$

with reaction rate P_{R2}. The production and loss terms of these tagged ozone fields are then

$$PO_3^{tag} = \frac{1}{2}P_{R1}\left(\frac{NO_y^{tag}}{NO_y} + \frac{HO_2^{tag}}{HO_2}\right) \tag{10}$$

225 $$+\frac{1}{2}P_{R2}\left(\frac{NO_y^{tag}}{NO_y} + \frac{NMHC^{tag}}{NMHC}\right)$$
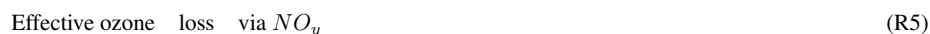
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$$
\begin{aligned}
DO_3^{tag} \;=\; & \tfrac{1}{2}P_{R3}\left(\frac{OH^{tag}}{OH} + \frac{O_3^{tag}}{O_3}\right) \\[2mm]
& + \tfrac{1}{2}P_{R4}\left(\frac{HO_2^{tag}}{HO_2} + \frac{O_3^{tag}}{O_3}\right) \\[2mm]
& + \tfrac{1}{2}P_{R5}\left(\frac{NO_y^{tag}}{NO_y} + \frac{O_3^{tag}}{O_3}\right) \\[2mm]
& + \tfrac{1}{2}P_{R6}\left(\frac{NMHC^{tag}}{NMHC} + \frac{O_3^{tag}}{O_3}\right) \\[2mm]
& + P_{R7}\frac{O_3^{tag}}{O_3},
\end{aligned}
\tag{11}
$$

with tag denoting one of the ten source tags and with the reaction rates P_{R3}, P_{R4}, P_{R5}, P_{R6}, P_{R7} referring to the reactions

$$OH + O_3 \;\longrightarrow\; HO_2 + O_2 \tag{R3}$$

$$HO_2 + O_3 \;\longrightarrow\; OH + 2\,O_2 \tag{R4}$$

$$\text{Effective ozone} \quad \text{loss} \quad \text{via } NO_y \tag{R5}$$

$$RO_2 + O_3 \;\longrightarrow\; RO + 2\,O_2 \tag{R6}$$

$$OH + O_3 \;\longrightarrow\; HO_2 + O_2 \tag{R7}$$

The tagged species NO_y, CO, NMHC, and PAN are treated similarly and will be discussed here only briefly, while more detailed information is provided in the supplement. Figure 2 sketches the principal relations between the tagged species. Methane (not tagged) is depleted and the chemical products are then tagged as "NMHC from methane". The species in the NMHC family are eventually transformed into CO and further into CO_2. The decomposition of N_2O (not tagged) constitutes a source for "stratospheric NO_y". Reactions between NO_y and NMHCs form PAN (not included in NO_y). PAN is an important species which can be transported over long distances before it thermally decomposes (Roberts, 2007).

HO_x chemistry (Figure 3 and Table 3) and the calculation of the individual contributions to the concentrations of OH and HO_2 is much more complex, hence we discuss it here in more detail. The main source of OH is the reaction of H_2O with $O(^1D)$. The chemical reactions between OH and HO_2 involve species such as CO, CH_4, NO_y, and NMHC. Losses of HO_x are the formation of H_2O_2 and HNO_3, which are soluble and can be easily rained out.

Since the lifetime of both OH and HO_2 is short, we assume steady-state for the contributions. We regard the main HO_x reactions, for which the production and loss rates are calculated in and provided by the MECCA submodel (see also Table 2).

Table 3. Reactions and reaction rates used for the calculation of OH and HO_2 contributions.

Reaction			OH Prod	OH Loss	HO_2 Prod	HO_2 Loss
$H_2O + O(^1D)$	$\longrightarrow$	$2\ OH$	$0.5\ P_1^{OH}$			
$HO_2 + O_3$	$\longrightarrow$	$OH + 2\ O_2$	P_2^{OH}			$L_1^{HO_2}$
$NO + HO_2$	$\longrightarrow$	$NO_2 + OH$	P_3^{OH}			$L_2^{HO_2}$
$OH + CO$	$\xrightarrow{O_2}$	$HO_2 + CO_2$		L_1^{OH}	$P_1^{HO_2}$	
$OH + CH_4$	$\xrightarrow{O_2}$	$NMHC + H_2O$		L_2^{OH}		
$OH + O_3$	$\longrightarrow$	$HO_2 + O_2$		L_3^{OH}	$P_2^{HO_2}$	
$OH + NMHC$	$\xrightarrow{O_2}$	$NMHC + H_2O$		L_4^{OH}		
$OH + HO_2$	$\longrightarrow$	$H_2O + O_2$		L_5^{OH}		$L_3^{HO_2}$
$OH + NO_2$	$\longrightarrow$	HNO_3		L_6^{OH}		
$NMHC + NO$	$\longrightarrow$	$NMHC + HO_2 + NO_2$			$P_3^{HO_2}$	
$NMHC + HO_2$	$\longrightarrow$	$NMHC + O_2$				$L_4^{HO_2}$
$HO_2 + HO_2$	$\longrightarrow$	$H_2O_2 + O_2$				$L_5^{HO_2}$

The steady-state assumption for the contributions to the OH and HO_2 concentrations, i.e., OH^{tag} and HO_2^{tag} implies that the individual production terms equal the individual loss terms:

$$P_{OH}^{tag} = L_{OH}^{tag} \tag{12}$$

$$P_{HO_2}^{tag} = L_{HO_2}^{tag}. \tag{13}$$

Again the more complex part of the tagging chemistry is to derive the production and loss terms. Using the reactions in Table 3 and the approach from Grewe et al. (2010), we obtain for the production and loss of OH^{tag}:

$$P_{OH}^{tag} = P_1^{OH} \frac{O_3^{tag}}{O_3} + \frac{1}{2} P_2^{OH} \left(\frac{HO_2^{tag}}{HO_2} + \frac{O_3^{tag}}{O_3} \right)$$
$$+ \frac{1}{2} P_3^{OH} \left(\frac{NO_y^{tag}}{NO_y} + \frac{HO_2^{tag}}{HO_2} \right) \tag{14}$$

$$L_{OH}^{tag} = L_1^{OH} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{CO^{tag}}{CO} \right) \tag{15}$$

$$+ L_2^{OH} \left(\frac{OH^{tag}}{OH} \right)$$

$$+ L_3^{OH} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{O_3^{tag}}{O_3} \right)$$

$$+ L_4^{OH} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{NMHC^{tag}}{NMHC} \right)$$

$$+ L_5^{OH} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{HO_2^{tag}}{HO_2} \right)$$

$$+ L_6^{OH} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{NO_y^{tag}}{NO_y} \right)$$

This set of equations includes the assumption that exchanges within a family are fast enough to achieve equally distributed tags among family members. For example, concerning P_1^{OH}, the contribution of one source to $O(^1D)$ equals that of O_3, i.e. $\frac{O(^1D)^{tag}}{O(^1D)} = \frac{O_3^{tag}}{O_3}$.

Similarly, we derive the individual production and loss terms for HO_2:

$$P_{HO_2}^{tag} = P_1^{HO_2} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{CO^{tag}}{CO} \right) \tag{16}$$

$$+ P_2^{HO_2} \frac{1}{2} \left(\frac{OH^{tag}}{OH} + \frac{O_3^{tag}}{O_3} \right)$$

$$+ P_3^{HO_2} \frac{1}{2} \left(\frac{NMHC^{tag}}{NMHC} + \frac{NO_y^{tag}}{NO_y} \right)$$

$$L_{HO_2}^{tag} = L_1^{HO_2} \frac{1}{2} \left(\frac{HO_2^{tag}}{HO_2} + \frac{O_3^{tag}}{O_3} \right) \tag{17}$$

$$+ L_2^{HO_2} \frac{1}{2} \left(\frac{HO_2^{tag}}{HO_2} + \frac{NO_y^{tag}}{NO_y} \right)$$

$$+ L_3^{HO_2} \frac{1}{2} \left(\frac{HO_2^{tag}}{HO_2} + \frac{OH^{tag}}{OH} \right)$$

$$+ L_4^{HO_2} \frac{1}{2} \left(\frac{HO_2^{tag}}{HO_2} + \frac{NMHC^{tag}}{NMHC} \right)$$

$$+ L_5^{HO_2} \frac{HO_2^{tag}}{HO_2}.$$

Now the Eqs. (12) and (13) can be written as

$$0 = A^{tag} - L^{OH} \frac{OH^{tag}}{OH} + P^{OH} \frac{HO_2^{tag}}{HO_2} \tag{18}$$

$$0 = B^{tag} + P^{HO_2} \frac{OH^{tag}}{OH} - L^{HO_2} \frac{HO_2^{tag}}{HO_2}, \tag{19}$$

with

$$A^{tag} = P_1^{OH}\frac{O_3^{tag}}{O_3} + \frac{1}{2}P_2^{OH}\frac{O_3^{tag}}{O_3} \tag{20}$$

$$+\frac{1}{2}P_3^{OH}\frac{NO_y^{tag}}{NO_y}$$

$$-\frac{1}{2}L_1^{OH}\frac{CO^{tag}}{CO} - \frac{1}{2}L_3^{OH}\frac{O_3^{tag}}{O_3}$$

$$-\frac{1}{2}L_4^{OH}\frac{NMHC^{tag}}{NMHC} - \frac{1}{2}L_6^{OH}\frac{NO_y^{tag}}{NO_y}$$

$$B^{tag} = \frac{1}{2}P_1^{HO_2}\frac{CO^{tag}}{CO} + \frac{1}{2}P_2^{HO_2}\frac{O_3^{tag}}{O_3} \tag{21}$$

$$+\frac{1}{2}P_3^{HO_2}\frac{NMHC^{tag}}{NMHC} + \frac{1}{2}P_3^{HO_2}\frac{NO_y^{tag}}{NO_y}$$

$$-\frac{1}{2}L_1^{HO_2}\frac{O_3^{tag}}{O_3} - \frac{1}{2}L_2^{HO_2}\frac{NO_y^{tag}}{NO_y}$$

$$-\frac{1}{2}L_4^{HO_2}\frac{NMHC^{tag}}{NMHC}$$

$$P^{OH} = \frac{1}{2}\left(P_2^{OH} + P_3^{OH} - L_5^{OH}\right) \tag{22}$$

$$L^{OH} = \frac{1}{2}\left(L_1^{OH} + 2L_2^{OH} + L_3^{OH}\right. \tag{23}$$

$$\left. + L_4^{OH} + L_5^{OH} + L_6^{OH}\right)$$

$$P^{HO_2} = \frac{1}{2}\left(P_1^{HO_2} + P_2^{HO_2}\right) \tag{24}$$

$$L^{HO_2} = \frac{1}{2}\left(L_1^{HO_2} + L_2^{HO_2} + L_3^{HO_2}\right. \tag{25}$$

$$\left. + L_4^{HO_2} + 2L_5^{HO_2}\right)$$

The Eqs. (18) and (19) can easily be solved resulting in

$$OH^{tag} = \frac{A^{tag}L^{HO_2} + B^{tag}P^{OH}}{L^{OH}L^{HO_2} - P^{OH}P^{HO_2}}OH \tag{26}$$

$$HO_2^{tag} = \frac{A^{tag}P^{HO_2} + B^{tag}L^{OH}}{L^{OH}L^{HO_2} - P^{OH}P^{HO_2}}HO_2. \tag{27}$$

The quantity A^{tag} (B^{tag}) represents the contribution of chemical tracers (tagged and non-tagged, other than OH and HO$_2$) to the net OH production (net HO$_2$ production). The terms L^{OH} and P^{OH} are primarily contributions to OH loss and production rates, which depend on the contribution to OH (OH^{tag}) and HO$_2$ (HO_2^{tag}), respectively. Only the reaction of OH with HO$_2$ forming water vapour and molecular oxygen constitutes an exception, since the loss of OH is dependent from both OH and HO$_2$ (L_5^{OH}). Therefore, it also contributes to P^{OH} (see Eq. (22)), the last term in equation (18), which depends on HO$_2$. Note that in this case it does not lead to a production but destruction of OH.

4 Present-day simulation and comparison to other studies

In this Section, we present results of a present day simulation. An actual validation of the tagging method is not feasible, since only the full quantities can be measured, e.g. the ozone concentration, but not the contribution from individual sources. Therefore, we concentrate on a comparison to earlier studies. In the following sections we present the simulation set-up and give a plausibility check for contributions to the HO_x concentration based on shipping and aviation, focussing on the ozone concentration.

4.1 Simulation set-ups

4.1.1 EMAC

To evaluate the TAGGING submodel we conduct two different simulations, one base simulation with all emissions and a second simulation where we reduced all road traffic emissions by five percent. The set-up follows the "Specified Dynamics Reference Simulation" for the Chemistry Climate Model Initiative, and is identical to the RC1SD-base10a set-up described and evaluated by Jöckel et al. (2016), however, extended by the TAGGING module, which we described above.

The simulation is performed with a spectral resolution of T42 and a vertical resolution of 90 levels (up to 0.01 hPa). For the anthropogenic emissions we use the MACCity emissions dataset with a resolution of $0.5°$ described by Granier et al. (2011). The lightning emissions are calculated online using the parameterisation described by Grewe et al. (2001). Emissions of NO from soil and biogenic origin as well as biogenic isoprene (C_5H_8) are calculated online by the MESSy submodel ONEMIS (described as ONLEM in Kerkweg et al. (2006b)). The submodel ONEMIS uses an algorithm based on Yienger and Levy (1995) for NO and Guenther et al. (1995) for isoprene. The dynamic state of the atmosphere is relaxed towards ERA-Interim reanalysis data (Dee et al., 2011) using a weak Newtonian relaxation ("nudging") of the four prognostic variables temperature, divergence, vorticity and the logarithm of surface pressure (Jöckel et al., 2006). Sea-surface temperature and sea-ice concentration are taken from ERA-Interim as well.

One important difference of our simulation to the RC1SD-base10a set-up is the use of the QCTM-mode of EMAC (Deckert et al., 2011). This QCTM-mode decouples the chemistry and the dynamics by using monthly climatologies (here derived from the RC1SD-base10a simulation) in the radiation code and for the heterogeneous stratospheric reactions. The application of the QCTM mode is important for overcoming the problem of a low signal to noise ratio in the case of a direct comparison of a base case simulation with one with a small chemical perturbation, which would be present with a fully coupled system. The dynamical and chemical differences between the RC1SD-base10a and our base simulation are shown in the supplement. The simulation covers the period 2004-2010 and is initialized from the RC1SD-base10a simulation. The first year was used as a spin-up period, resulting in an evaluation period from 2005-2010.

4.1.2 MECO(n)

The COSMO/MESSy simulation shown in Sec. 5.1 covers the European domain, including parts of the Eastern Atlantic and North Africa, with a resolution of $0.44°$ (≈ 50 km). Simulated is the period from July 2007 until December 2008, with the six months of 2007 used as spin-up phase. The driving EMAC model is applied at a resolution of T42 with 31 horizontal levels and is relaxed towards ERA-Interim reanalysis as well. The same QCTM mode as described above is applied for EMAC and COSMO/MESSy. Both model instances use the anthropogenic MACCity emissions, as well as online calculated soil/biogenic emissions as described above. The simulation differs, however, from the EMAC simulation described above, by using the lightning parameterisation after Price and Rind (1992) to simulate the lightning NO_x emissions on the global scale. In COSMO/MESSy we use the same emissions as on the global scale by regridding the corresponding emissions from EMAC. We have chosen this approach to have emissions as comparable as possible in both model instances. More detailed information about this simulation, including an evaluation of chemical tracer concentrations, is provided by Mertens et al. (2016).

4.2 Contributions of emission sectors to NO_y, CO, NMHCs, and O_3

The six year annual average contributions of the ten emission sectors to the ozone concentration are shown in Figure 4. We compare these results with an earlier model version, which only tags NO_y and ozone (Grewe, 2007, Figure 5b therein), and to earlier similar studies by Lelieveld and Dentener (2000) and Emmons et al. (2012). This comparison aims at verifying that the implementation of the TAGGING mechanism is correct by comparing contribution patterns and magnitudes. We have to keep in mind that the approach is conceptually different from earlier studies and takes into account all ozone precursor emissions and not only NO_y. Hence the individual contributions have to be smaller and no agreement can be expected, except for pattern and magnitude. The only direct intercomparison can be performed for the stratospheric ozone mixed into the troposphere, since this process is independent from any precursors. Here Lelieveld and Dentener (2000); Lamarque et al. (2005); Grewe (2007); Emmons et al. (2012) (hereafter denoted as LD00, L05, G07, and E12, respectively) estimated a 5 to 40% contribution from stratospheric ozone to tropospheric ozone in the Southern Hemisphere and mostly systematically lower values of 10% to 25% in the Northern Hemisphere, while tropical values are below 10% (Table 4). Our simulation shows also a minimum in the tropics and lower values in the Southern Hemisphere compared to the Northern Hemisphere. The mixing ratios for January and July are very similar to those of Emmons et al. (2012, not shown).

Ozone formed from lightning NO_x (Figure 4) shows a maximum in the tropics and upper troposphere and larger contributions in the Southern Hemisphere than in the Northern Hemisphere, which is in agreement with G07 and LD00. The maximum contribution from lightning is around 25-30%

385 and thus lower than G07 (40%) and LD00 (50%), because here we regard the ozone production of
all precursors, whereas in G07 and LD00 only NO_x as a precursor is considered (see above).

Agreement between the studies LD00, G07, E12, and our work can also be found with respect to
the contribution of anthropogenic emissions to tropospheric ozone. These emissions (here: anthro-
pogenic non-traffic, road traffic, shipping and aviation) predominantly contribute by 30 to 50% in

390 the Northern Hemisphere. The ozone contribution from biomass burning peaks in the lower tropical
troposphere with values of around 10% to 15% which compares well with G07 and LD00 (20%).
Around 15% of the tropospheric ozone originates from methane, which reacts with OH and con-
tributes to NMHC compounds and eventually to CO and CO_2.

Figure 5 shows the contribution of the individual emission sectors to the tropospheric budgets

395 of NO_y, CO, NMHC, PAN, and O_3. Lightning and non-traffic anthropogenic emissions show the
largest contributions to NO_y. The emitted NO_y from lightning and aviation remains much longer in
the atmosphere compared to a surface source such as non-traffic anthropogenic NO_y, since lightning
and aviation emit mainly in the upper troposphere. Aviation, shipping, and biomass burning have
approximately the same contribution.

400 The different emission sectors have very different emission characteristics. Some are only emitting
NO_y, such as lightning, or NO_y and NMHCs, such as most anthropogenic sources. This is well re-
flected in the budgets (Figure 5). Since NO_y is required to form PAN, the decomposition of PAN also
produces NO_y and NMHCs with the original tag, e.g. the lightning tag. This is fully consistent with
the chosen tagging approach and leads to minor contributions of non-CO and non-NMHC emitting

405 emission sectors to the CO and NMHC budgets (lightning, stratosphere, aviation). The formation of
PAN and hence contributions to PAN (Figure 5, second row) requires both NO_y and NMHCs. None
of the ten emission sectors has a large contribution from both. And hence, the contributions of each
of the ten sectors to PAN are almost equally distributed around 10%. One exception is methane,
which contributes largely to NMHC concentrations but not to NO_y. In addition, the NMHCs from

410 methane are predominantly occurring in areas with low NO_y background, which reduces the impact
on PAN. The contribution to tropospheric ozone (Figure 5, second row) reflects the distribution pre-
sented in Figure 4, with major contributions from lightning, stratosphere, anthropogenic non-traffic
emissions and methane.

4.3 Contribution of emission sectors to HO_x concentrations

415 In this Section, we present the effects of a surface source (shipping) and a higher altitude atmospheric
source (aviation) on their contribution to the HO_x concentrations. We have chosen the Mediterranean
Sea for shipping, since it includes pristine areas in the middle of the Sea on the one hand, as well
as areas which are largely affected by other sources e.g. in southern France (Marseilles) and Italy
(harbour areas such as Genoa, Figure 6), on the other hand.

420　　We have identified four areas (A-D) with different chemical characteristics (Table 5, see also
Figure 6): Highly polluted areas with high concentrations of NO_2 (A and B) and with large (some)
impact from shipping in region A (B); a pristine area with some impact from shipping on NO_x and
O_3 (C), and a pristine area with a larger values of shipping ozone (D).

　　Large NO_x concentrations in the background (A and B) impact the chemistry and net production
425　　efficiencies, i.e. the ozone enhancement per NO_x is decreasing with increasing NO_x concentrations
(e.g. Dahlmann et al., 2011). The reaction (R1), which transforms HO_2 into OH, in principle in-
creases (decreases) the OH (HO_2) concentration in the region where large amounts of shipping NO_x
is present. However, this reaction only dominates the OH to HO_2 ratio if enough ozone is available
for the HO_x production. In region A, the very low ozone concentration due to ozone titration by
430　　NO_x limits the availability of OH and the contribution of shipping NO_x to OH is even negative.
Region B is less polluted than region A and has lower values of shipping NO_x and therefore reaction
(R1) dominates the OH and HO_2 contributions from shipping leading to positive contributions to
OH and negative to HO_2. The tagged shipping ozone is larger in area D compared to A and B (not
shown). This leads to a larger contribution to OH via the main production reaction of H_2O with
435　　$O(^1D)$, where the $O(^1D)$ originates from the tagged ozone (see also Table 3). The close coupling
of OH with HO_2 also enhances the tagged HO_2 especially in region D. These processes then lead
to a complex picture. It shows negative contributions to OH in region A, mainly due to low ozone
concentration limiting the OH availability which is even more pronounced by shipping emissions.
The shipping contribution to HO_2 in the polluted areas A and B are negative mainly driven by the
440　　reaction (R1). Large positive contributions of shipping to OH and moderate negative contributions
to HO_2 are found in region C, resulting from a combination of effects from reaction (R1) and the
main OH production resulting from tagged shipping ozone. Whereas in region D moderate positive
contributions of shipping to OH and large negative contributions to HO_2 are found. Overall, the
contributions from shipping emissions to the OH and HO_2 concentrations show a complex picture,
445　　which results from variations in both the background concentrations and shipping concentrations.
The impact of an enhanced horizontal resolution is discussed for the same situation in Sec. 5.

　　Figure 7 shows annual mean contributions of aviation NO_x emissions to OH (left) and HO_2 (right).
The air traffic contribution to OH peaks at around 10-20 fmol/mol at the main flight altitude. At the
surface, there are other secondary peaks, basically at the locations of the airports. Lee et al. (2010)
450　　summarised the work of Grewe et al. (2002) and Köhler et al. (2008) in their Figure 10 and showed
4 atmospheric regions, which are affected differently by air traffic. In the first region (RNOy in their
paper), which is mainly the air traffic corridor, the reaction (R1) controls the chemical impact from
air traffic emissions. This implies that air traffic largely contributes to OH and negatively contributes
to RHO_2 as shown in Figure 7. The region north of RNOy is called RHO2 and the aviation impact
455　　is largely controlled by the reaction of O_3 with HO_2 (see Table 3). Hence, this reaction leads to
a reduction in HO_2 without affecting the OH concentration in a similar manner. The region RO3

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is located in the lower troposphere and away from the major flight corridor. Here, a significant contribution from air traffic to ozone is found, but not so much to NO_y (not shown). The region is controlled by an increase in ozone. Hence it leads to a general increase in HO_x via the reaction of

460　H_2O with $O(^1D)$ (see Table 3). This comparison shows that the OH and HO_2 contributions form aviation, calculated here, are consistent with the chemical regimes identified in previous studies.

　　A more detailed view on this tagging mechanism is feasible by applying it to a Lagrangian framework (Grewe et al., 2014). Within the EU-Project REACT4C (Reducing Emissions from Aviation by Changing Trajectories for the benefit of Climate), the HO_x tagging mechanism was implemented

465　in the same EMAC model version, including a Lagrangian transport algorithm. Aviation-like pulse emissions of NO_x were released at selected points in the atmosphere, and trajectories with these emissions were tagged so that reactions with the background can be determined in detail. Note that aviation is not emitting CO and NMHCs in our simulation, hence the equations look simplified in Grewe et al. (2014) as the values for CO^{tag} and $NMHC^{tag}$ are zero (see Sec. 3.3). Figure 8 shows the

470　temporal development of several NO_x related species (top and mid) as well as ozone production and loss terms (bottom) for a pulse emission at 45°W, 50°N and 300 hPa. The NO_x emission induces net production of O_3^{tag} (see Eq. 10 in Grewe et al., 2014), mainly via Reaction (R1) and enhanced HO_x^{tag} as calculated via Eqs. (26) and (27). NO_x reacts with OH and forms HNO_3, which eventually leads to washout and a reduction of NO_y^{tag} within a few weeks. When NO_x^{tag} is no longer available for

475　O_3-production, O_3^{tag} is subsequently depleted. We denote the chemical regime, where enough NO_x is available to produce larger amounts of ozone with RegNOx and the following regime as RegO3 (see also Figure8). Regarding the destruction of CH_4^{tag}, these two regimes are also characterising the two different depletion pathways. First, as long as sufficient NO_x^{tag} is available, CH_4^{tag} is reduced because of an increase of OH^{tag} via Reaction (R1) (NO_x driven CH_4 destruction). Second, when

480　NO_x^{tag} is removed, OH^{tag} is mainly produced via photolysis of O_3^{tag} and the subsequent Reaction P_1^{OH} ($H_2O + O^1D \rightarrow 2\,OH$). The tagged OH and HO_2 are far lower in the RegO3 regime compared to the RegNOx regime (Figure 8, mid). And consequently, the gradient in the O_3 driven CH_4 destruction is not as steep. However, due to the longer time period, it dominates the total amount of methane destruction in this case, which can be seen from a budget analysis for the chemical regimes

485　RegNOx (blue bars) and RegO3 (red bars) given in Figure 9. Note that the trajectory is transported into polar night around day 5, which leads to a reduction of OH and HO_2 and a reduction of the photochemical activity. This example shows a reasonable temporal behaviour of the tagged species and it further shows how combining the tagging methodology and a Lagrangian transport algorithm results in a powerful tool, facilitating a detailed analysis of particular processes.

5 Sensitivities

In this Section, we investigate if our tagging scheme responds reasonably to changes in resolution (Sec. 5.1) and emissions (Sec. 5.2). In general, there are no strict verification tests other than checking for plausibility and stability.

5.1 Higher resolution: MECO(n)

By applying the MECO(n) system (Kerkweg and Jöckel, 2012a, b; Hofmann et al., 2012; Mertens et al., 2016), we have increased the horizontal resolution over Europe by roughly a factor of 5, from a resolution of roughly 200 km times 300 km in EMAC to 50 km times 50 km in the nested grid. Figure 10 shows the contributions of the individual emission sectors to the tropospheric ozone column as a mean over Europe for the coarse resolution (top) and the finer resolution (bottom). Clearly, the individual contributions are very similar in terms of mean values and the seasonal cycle. The finer resolution simulation shows finer resolved structures in the horizontal (not shown), which however do not largely affect the large-scale budgets.

As an example of the effects of finer resolution, we present OH and HO_2 contributions from shipping over the Mediterranean Sea, as discussed in Sec. 4.3 and Figure 6. The OH enhancement along shipping routes is much more visible in the finer resolved case (Figure 6, lower left) compared to the lower resolution (upper left). The structures for the OH and HO_2 contributions are again similar: A positive contribution to the OH concentration in the area of shipping emissions (B-D) and a decrease in the contribution to OH and HO_2 where background NO_x is largely enhanced (region A).

The comparison of the coarser and finer resolution clearly shows that the tagging scheme is stable in its behaviour. Naturally, the finer resolution enables more detailed and finer resolved chemical changes due to emissions to be quantified, but basic structures are reproduced in either resolution. This implies that, depending on the underlying research question, either model can be used.

5.2 Emission changes

We performed an additional global simulation with EMAC where we reduced the road traffic emissions by 5%. The simulation set-up hence follows Hoor et al. (2009). This means that the chemical composition and the ozone productivity is different from the base simulation, which leads to roughly 2-3% reduction in the tagged road traffic ozone (not shown). Generally, a reduction of surface NO_x emissions is increasing the ozone productivity (Emmons et al., 2012; Grewe et al., 2012) and consequently a 5% reduction in emissions is expected to lead to significantly less than a 5% reduction in road traffic ozone, which is consistent with our results. Figure 11 shows the relative change in tropospheric ozone induced by the road traffic emission reduction of 5%. The total ozone change of 0.08% (black bar) is a consequence of the reduction of the contribution of road traffic to the tro-

pospheric total ozone by 0.16% and other compensating effects. In total, this leads to a factor of 2

525 difference between the total ozone change and the road traffic ozone change. The compensating effects are resulting from larger net ozone production rates for the shipping emission sector and other anthropogenic non-traffic emission sectors. This leads to a larger contribution of the anthropogenic non-traffic (dark blue bar) and the (other than road traffic) traffic emission sector (green bar) to total ozone by 0.04% and 0.06%, respectively. Other non-anthropogenic, i.e. natural emission sectors (red

530 bar) reduce this compensation.

The ratio of 2 between the reduction in total ozone and road traffic indicates that a calculation of the road traffic contribution to tropospheric ozone using the perturbation method, i.e. difference between two simulations with changing emissions, underestimates this contribution by exactly this factor of 2. Other studies have shown slightly larger factors, e.g. a factor of 3 for biomass burn-

535 ing NO_x emissions (Emmons et al., 2012) and a factor of 5 for road traffic NO_x emissions. Here, a smaller factor can be expected, since emissions other than NO_x and their impact on ozone are tagged, which reduces the effects from road traffic emission changes. Further, this factor largely depends on the chemical state of the atmosphere, which differ between the simulations. Hence, a direct intercomparison is not possible, however, the results are plausible.

540 ## 6 Conclusions

We present a submodel for the Earth-System Model EMAC, which diagnoses online the contributions of individual source categories (mainly emission sectors) to the concentrations of various trace species. For the first time, we take into account the competition of nitrogen oxides, carbon monoxide and non-methane hydrocarbons for the production and destruction of ozone. We concentrated on 10

545 source categories and 7 species and families, which are tagged. As a result, we introduced 70 new tracers. The physical and chemical tendencies for these tracers are obtained from other submodels of EMAC, such as the chemistry (MECCA), scavenging (SCAV), etc. The tagging mechanism is distributing the calculated physical and chemical tendencies into the tagged tracer fields. Therefore, the computing time increase by the TAGGING submodel is small, around 10%.

550 We performed a present-day simulation and showed that the TAGGING submodel provides contributions of individual emission sectors to the concentration of ozone, which roughly agree with previous estimates. A detailed analysis of the calculated contribution of aviation and shipping to OH and HO_2 shows reasonable results in different chemical regimes. Changes in the model's resolution shows a stable performance of the TAGGING submodel. Changes in the strength of road traffic

555 emissions yields a decrease in ozone, which is partly compensated for by an increase in ozone from other source categories, since the ozone production efficiency increases, which is in agreement with earlier findings (Grewe et al., 2012; Emmons et al., 2012).

The advantage of this specific tagging scheme is that (1) the effect of ten source categories on ozone and other trace species can be monitored online in one simulation, (2) the competition be-

560 tween ozone precursors is included, (3) no linearisation is required, and (4) the scheme is applicable for long-term simulations, e.g. over a century. On the other hand, the disadvantage is that (1) the family concept is not flexible and fixed in this specific way, and consequently (2) any change in the set of chemical species requires an adaptation of the TAGGING scheme, and (3) due to memory limitations, a restriction to the main chemical species and families is required.

565 To summarise, the TAGGING submodel provides a powerful tool to identify the contribution of individual emission sectors to main atmospheric constituents at every grid point and timestep of the simulation and can be further used to derive, for instance, radiative forcings or contribution to air quality information for individual emission sectors.

7 Code availability

570 The Modular Earth Submodel System (MESSy) is continuously further developed and applied by a consortium of institutions. The usage of MESSy and access to the source code is licensed to all affiliates of institutions which are members of the MESSy Consortium. Institutions can become a member of the MESSy Consortium by signing the MESSy Memorandum of Understanding. More information can be found on the MESSy Consortium Website (http://www.messy-interface.org).

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Table 4. Comparison of different studies with respect to the contribution (%) of stratospheric ozone to the tropospheric ozone concentration. Numbers are rough estimates, only, as taken from published figures. Note that values for L05 are surface values only and percentage values from E12 are estimated from mixing ratios, however a mean value of 17% is given therein. See text for more explanations. SH and NH abbreviates Southern and Northern Hemisphere, respectively.

Reference	SH	Tropics	NH
LD00	40	10	25
L05	20	<10	10
G07	5-10	10	15
E12	20	<5	15
This work	10-15	5-10	10-20

Table 5. Qualitative characterisation of four different regions (A-D) in the Mediterranean Sea. A: Southern France; B: Strait of Gibraltar; C: Central Mediterranean Sea; D: Tunesian Coast. See Figure 6 (top row) for the location of the regions. The signs '++', '+', '○', and '-' indicate a qualitative estimate of the respective characteristics, 'very strong/very large', 'strong/large', 'moderate', 'negative'.

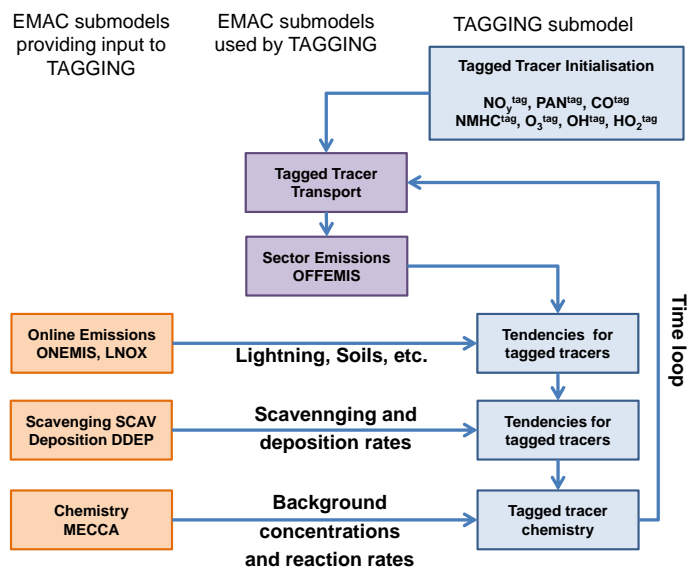|	A	B	C	D
Region has polluted background	++	+	○	○
Region is impacted by shipping NO_x	++	+	+	++
Region is impacted by shipping ozone	+	+	+	++
Shipping emissions are converting HO_2 into OH via $NO+HO_2 \longrightarrow OH + NO_2$	++	++	+	+
Shipping ozone produces OH via $O_3 \longrightarrow O(^1D) \xrightarrow{H_2O} OH$	+	+	+	++
Contribution of shipping emissions to OH	-	+	++	+
Contribution of shipping emissions to HO_2	-	-	○	-

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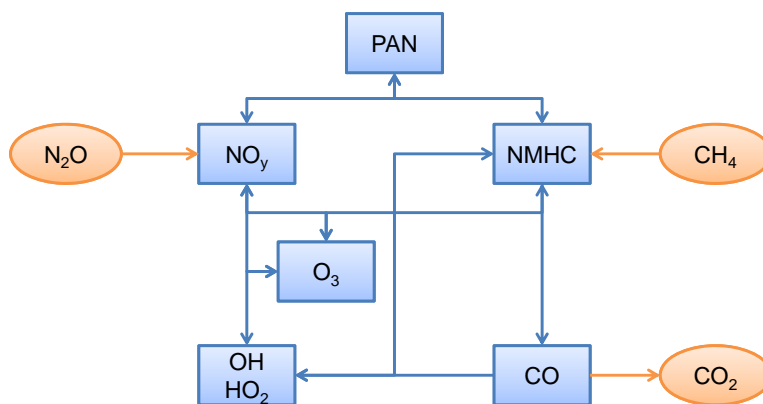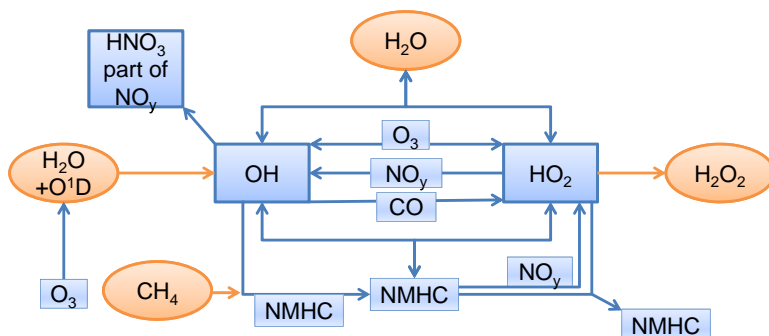**Figure 1.** Sketch of the tagging algorithm.

Figure 2. Sketch of the chemistry of tagged species (blue) and key relations to other species (orange). Note that stratospheric ozone is not included here. For HO_x chemistry see also Figure 3

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Figure 3. Atmospheric HO_x chemistry used in the TAGGING scheme. Blue boxes indicate tagged species and families and orange circles non-tagged species. Arrows indicate reactions.
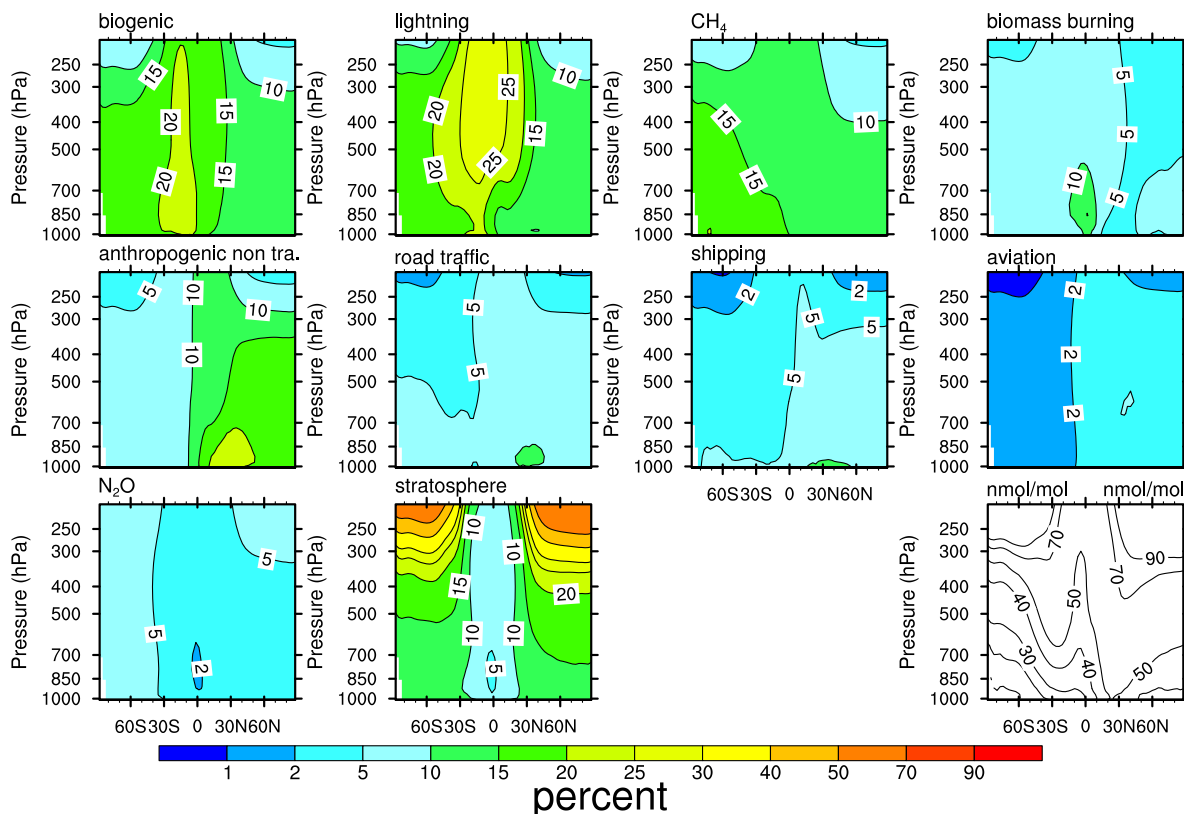
Figure 4. Annual mean contributions [%] of 10 emission sectors to the simulated ozone volume mixing ratios. The simulated ozone volume mixing ratio is shown in the lower right panel.
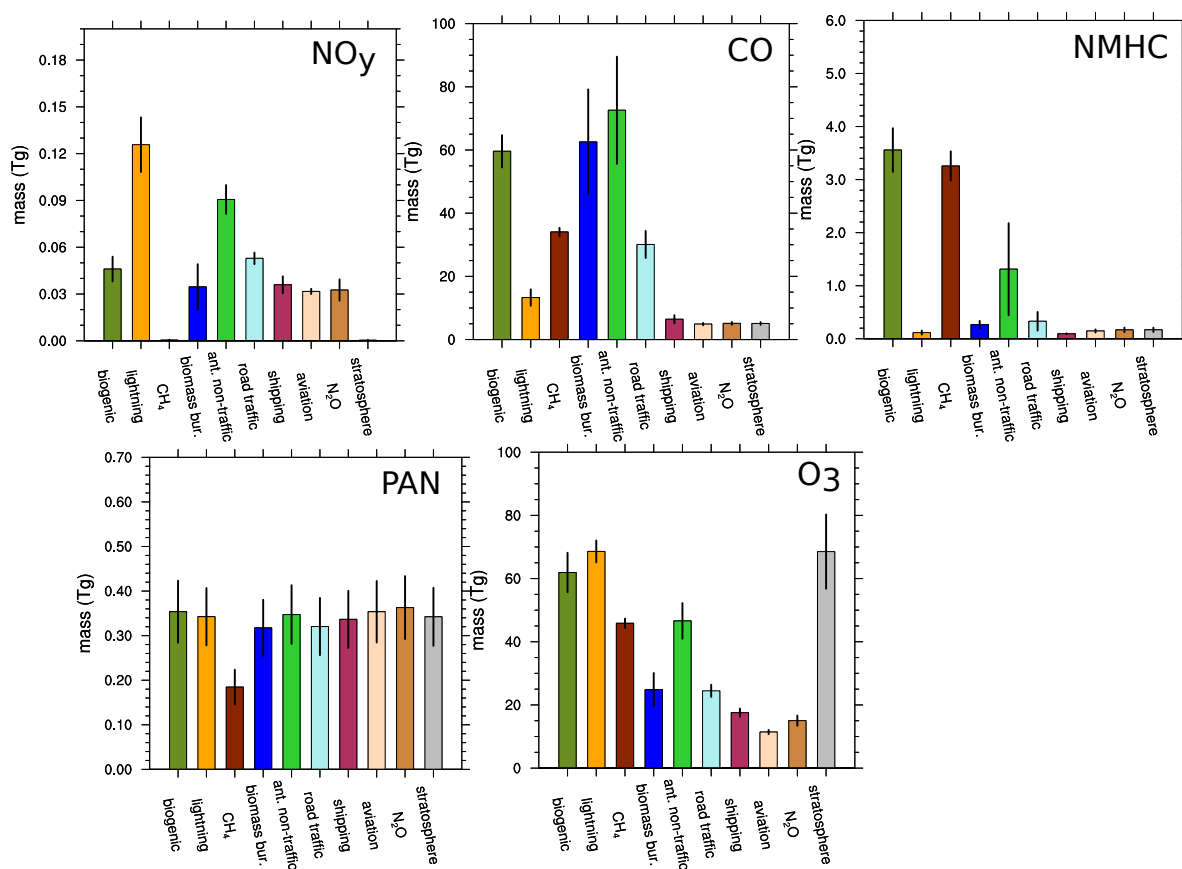
Figure 5. Contributions to the annual mean tropospheric budgets [Tg] of 10 emission sectors. Top row: NO_y, CO, and NMHCs; Bottom row: PAN and O_3. Errorbars indicate the interannual variability.

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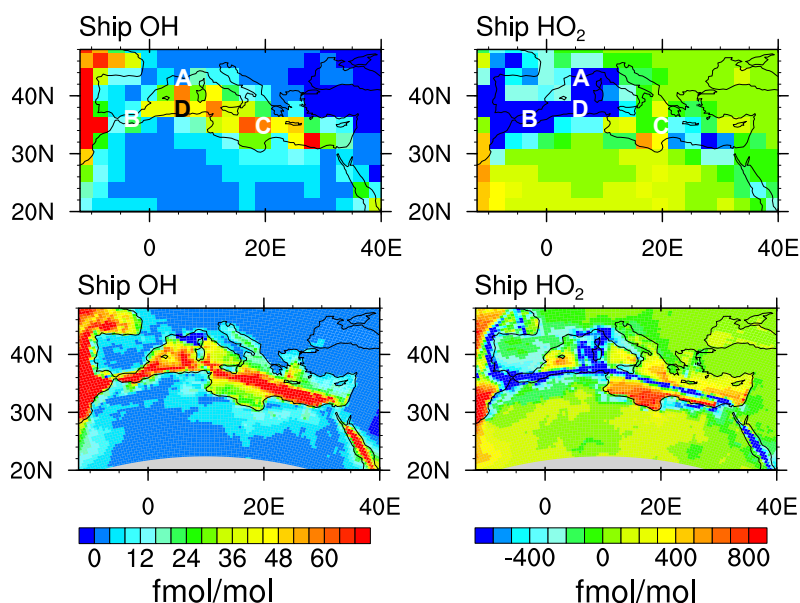
Figure 6. Absolute contribution of shipping to the simulated OH (left) and HO$_2$ (right) volume mixing ratios (in fmol/mol) for August 2007. Top row: EMAC; Bottom row: MECO(n). Regions A-D are characterised by different chemical situations. A: Southern France; B: Strait of Gibraltar; C: Central Mediterranean Sea; D: Tunesian Coast; See text for more details.
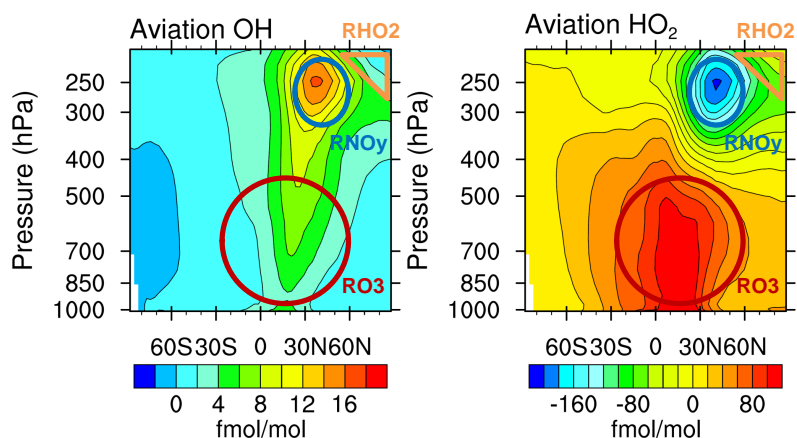

Figure 7. Annual mean absolute contribution [fmol/mol] of aviation to the simulated OH (left) and HO$_2$ (right) volume mixing ratios. The regions RO3, RNOy, and RHO2 are characterised by distinct different chemical response to aviation emissions as described by Grewe et al. (2002) (see text for further details).
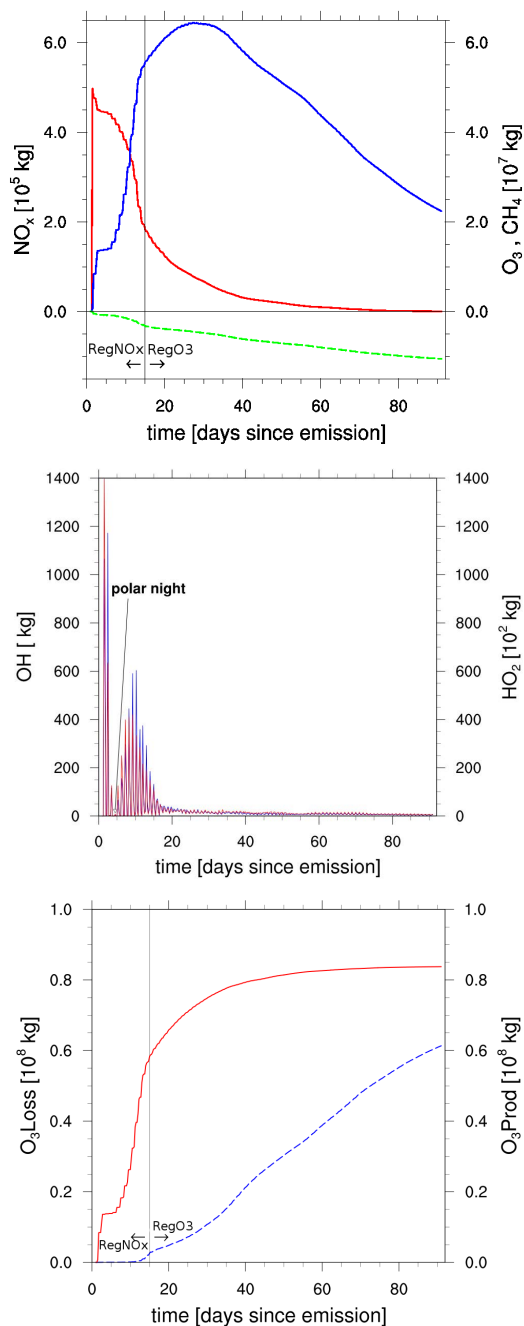
Figure 8. Temporal development of NO_x related species (top: NO_x (red), O_3 (blue), CH_4 (green), mid: OH (blue), HO_2 (red)) and production or loss terms (bottom: cumulative O_3-loss (blue) and cumulative O_3-production (red)) induced by a pulse emission at 45°W, 50°N and 300 hPa on December, 23, 2000. The discrimination between the regimes RegNOx and RegO3 refers to the NO_x dominated (days 1-15 after emission) and the O_3 dominated regime (days 16-90 after emission), respectively.
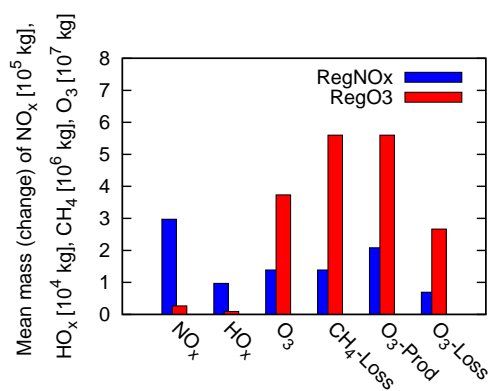
Figure 9. Mean contributions to NO_x related species and production or loss terms for the RegNOx regime (NO_x dominated, blue) and RegO3 regime (O_3 dominated, red), respectively. Values are given as temporal means over the two time periods.

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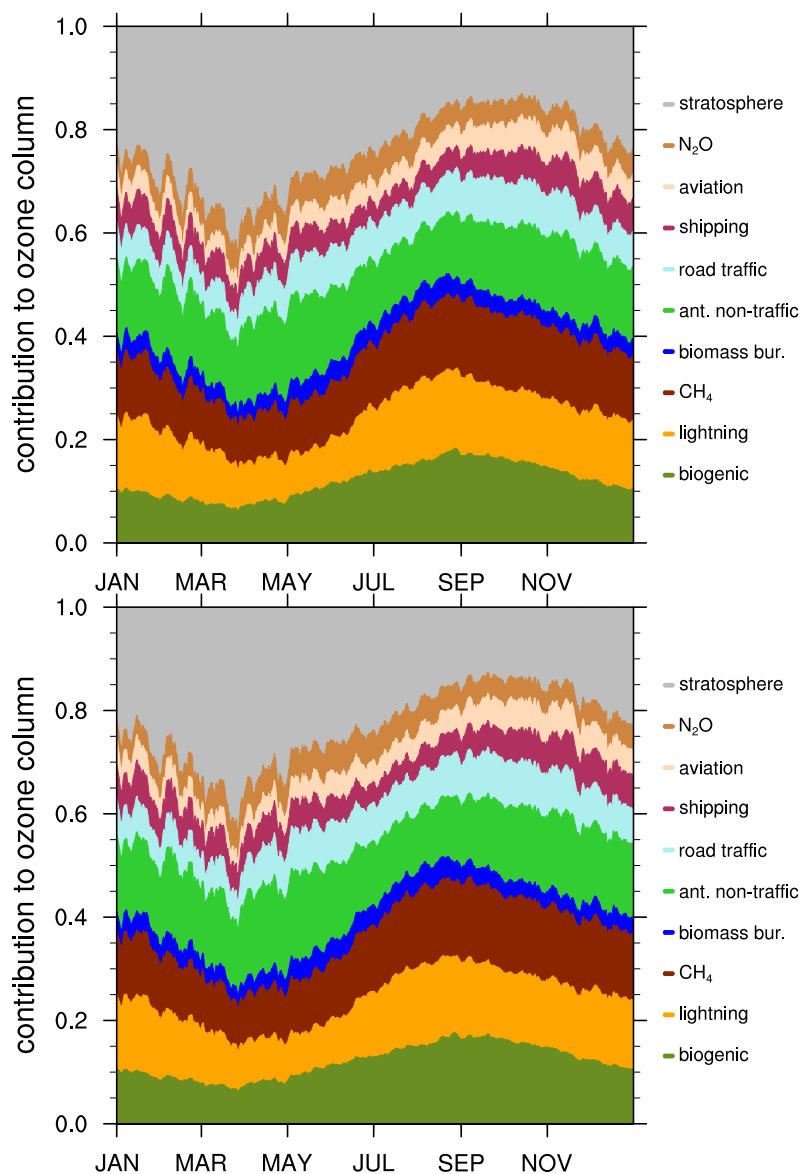
Figure 10. Contributions (fraction) of individual emission sectors to the European tropospheric ozone concentration for a coarser resolution simulation with EMAC (top) and a finer resolution with MECO(n) (bottom) for the year 2008.

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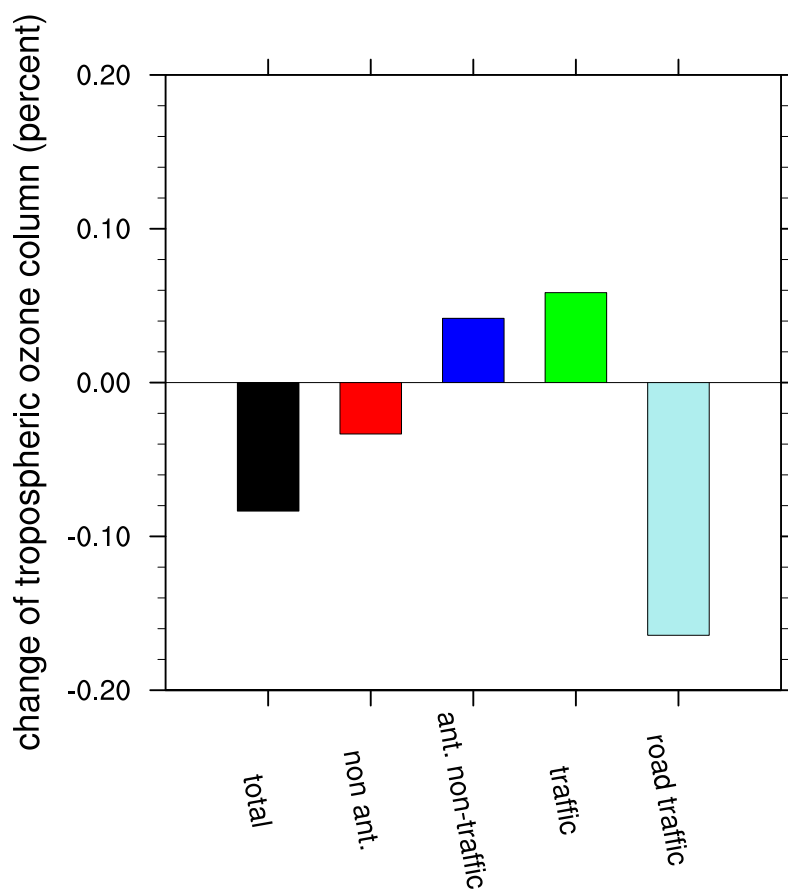
Figure 11. Changes in the global tropospheric ozone budget [%] resulting from a 5% reduction in the road traffic emissions.