



1 2 3	Modeling Regional Air Quality and Climate: Improving Organic Aerosol and Aerosol Activation Processes in WRF/Chem version 3.7.1
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10	ABSTRACT
18	Air quality and climate influence each other through the uncertain processes of aerosol formation
19	and cloud droplet activation. In this study, both processes are improved in the Weather, Research
20	and Forecasting model with Chemistry (WRF/Chem) version 3.7.1. The existing Volatility Basis
21	Set (VBS) treatments for organic aerosol (OA) formation in WRF/Chem is improved by
22	considering the secondary OA (SOA) formation from semi-volatile primary organic aerosol
23	(POA), a semi-empirical formulation for the enthalpy of vaporization of SOA, as well as
24	functionalization and fragmentation reactions for multiple generations of products from the
25	oxidation of VOCs. Two-month long simulations (May to June 2010) are conducted over
26	continental U.S. and results are evaluated against surface and aircraft observations during the
27	Nexus of Air Quality and Climate Change (CalNex) campaign. Among all the configurations
28	considered, the best performance is found for the simulation with the 2005 Carbon Bond
29	mechanism (CB05) and the VBS SOA module with semivolatile POA treatment, 25%





30 fragmentation, and the emissions of semi-volatile and intermediate volatile organic compounds 31 being 3 times of the original POA emissions. Among the three gas-phase mechanisms (CB05, 32 CB6, and SAPRC07) used, CB05 gives the best performance for surface ozone and PM<sub>2.5</sub> 33 concentrations. Differences in SOA predictions are larger for the simulations with different VBS 34 treatments (e.g., non-volatile POA vs. semivolatile POA) as compared to the simulations with different gas-phase mechanisms. Compared to the simulation with CB05 and the default SOA 35 36 module, the simulations with the VBS treatment improve cloud droplet number concentration 37 (CDNC) predictions (NMBs from -40.8% to a range of -34.6% to -27.7%), with large differences 38 between CB05/CB6 and SAPRC07 due to large differences in their OH and HO<sub>2</sub> predictions. An 39 advanced aerosol activation parameterization based on the FN05 series reduces the large negative 40 CDNC bias associated with the default ARG00 parameterization from -35.4% to a range of -0.8% 41 to 7.1%, it, however, increases the errors due to overpredictions of CDNC, mainly over 42 northeastern U.S. This work indicates a need to improve other aerosol-cloud-radiation processes 43 in the model such as the spatial distribution of aerosol optical depth and cloud condensation nuclei 44 in order to further improve CDNC predictions.

#### 45 **1. Introduction**

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The Intergovernmental Panel on Climate Change (IPCC) report on the AR5 scenario attributes the aerosol radiative forcing (RF) to be the dominant source of uncertainty contributing to the overall uncertainty in the net Industrial Era Radiative Forcing (RF) calculations (Myhre et al., 2013). Despite the inclusion of more aerosol processes in the current generation of atmospheric models, differences between atmospheric models and observations continue to persist. Aerosols affect the climate through the direct effect by absorbing or scattering radiation, or the indirect effect by acting as cloud condensation nuclei (CCN). According to Hallquist et al. (2009), the





54 formation of inorganic particulates such as sulfate, nitrate, and ammonium are well understood, 55 however, there are large uncertainties in the formation of secondary organic aerosol (SOA). As a 56 result, current models do not have a comprehensive treatment of OA, which usually result in an 57 underprediction of OA concentrations (Hodzic et al., 2010; Jathar et al., 2011; Bergstrom et al., 58 2012), due to missing key precursors and processes in OA formation (Ahmadov et al., 2012). Some of the missing key precursors in most models include semi-volatile primary organic aerosol (POA), 59 60 long-chain n-alkanes, polycyclic aromatic hydrocarbons (PAHs), and large olefins that have lower 61 volatilities compared to traditional SOA precursors (Chan et al., 2009). The organic carbon (OC) 62 component of the radiative forcing in the IPCC AR5 report also does not include SOA with the 63 reason that the formation is dependent on a number of factors that are not currently sufficiently 64 quantified (Myhre et al., 2013). However, SOA can form a significant percentage of total OA (up 65 to 95% in rural areas) (Zhang et al., 2007). Another large source of uncertainty is the quantification 66 of clouds as well as aerosol-cloud interactions due to incomplete or inaccurate representations of 67 these processes in climate models (Boucher et al., 2013). A major process in cloud formation from 68 aerosol is aerosol activation, which involves the condensational growth of aerosols in a cooling air 69 parcel until maximum supersaturation, and some of the wet particles reach a critical radius where 70 they are then able to grow spontaneously into cloud droplets (Ghan et al., 2011). Various 71 approaches have been developed to reduce the uncertainties associated with OA and aerosol 72 activation treatments in climate models. Those treatments are reviewed in the following section.

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# 1.1. VBS Treatments and Sensitivity to Different Gas-Phase Chemical Mechanisms in Regional and Global Models

76 Unlike inorganic aerosols such as sulfate, the physical and chemical properties of OA
77 dynamically evolve with age (Jimenez et al., 2009). The traditional approach to modeling SOA is





78 to assume that each VOC precursor forms several surrogate compounds (Odum et al., 1996). 79 However, the traditional method has several shortcomings, for example, two products are needed 80 for each VOC precursor causing this method to be computationally-expensive if many VOC 81 precursors are treated in the model (Murphy and Pandis, 2009). The assumption that the products 82 are unreactive also does not reflect the dynamic nature of the first generation products from the oxidation of VOCs that can undergo successive oxidation steps to further produce lower volatility 83 84 products (Jimenez et al., 2009). The volatility basis set (VBS) is a framework developed by 85 Donahue et al. (2006), which is able to simulate gas-phase partitioning and multiple generations 86 of gas-phase oxidation of organic vapors. This approach addresses the shortcomings of the 87 traditional SOA modeling approach as it can cover the complete volatility range of OA compounds 88 (Murphy and Pandis, 2009).

Table 1 summarizes some of the VBS treatments from current regional and global models. The 89 90 VBS treatment has been implemented into a number of regional models such as the Weather, 91 Research and Forecasting model with Chemistry (WRF/Chem) (Shrivastava et al., 2011; 92 Ahmadov et al., 2012), the Particulate Matter Comprehensive Air Quality Model with extensions 93 (PMCAMx) (Lane et al., 2008; Donahue et al., 2009; Murphy et al., 2009), and CHIMERE (Hodzic 94 et al., 2010). It has also been implemented in global models such as GISS II' GCM (Farina et al., 95 2010; Jathar et al., 2011) and the Community Earth System Model (CESM) (Shrivastava et al., 96 2015). Different studies define the classifications of the organic species slightly differently. 97 Donahue et al. (2009) defined primary organic vapors with effective saturation concentrations (C\*) of  $10^{-2}$  -  $10^{-1}$ ,  $10^{0}$  -  $10^{2}$ , and  $10^{3}$  -  $10^{6} \mu g \text{ m}^{-3}$  at 298 K to be low volatility organic compounds 98 99 (LVOCs), semi-volatile organic compounds (SVOCs), and intermediate volatility organic compounds (IVOCs), respectively. Shrivastava et al., (2011) and Jathar et al. (2011) defined 100





- 101 primary organic vapors with C\* values of  $10^{-2}$   $10^3$  and  $10^4 10^6 \,\mu g \,m^{-3}$  to be SVOCs and IVOCs,
- 102 respectively. All those studies defined VOCs to be gas-phase organic species with C\* larger than
- 103  $10^6 \,\mu g \, m^{-3}$  at 298 K.
- 104 The traditional emission inventories used in the chemical transport models consist of VOCs 105 but not SVOCs or IVOCs as both SVOCs and IVOCs are difficult to measure. This is most likely 106 because SVOCs and IVOCs tend to evaporate at high temperatures from combustion sources 107 (Donahue et al., 2009). As the traditional SOA approach usually underpredicts the SOA 108 concentration, the addition of the SVOC and IVOC emissions on top of the existing VOC 109 emissions in most emission inventories can improve model performance. To account for the 110 missing key precursors in OA formation, SVOC and IVOC emissions are usually estimated as a 111 factor of existing POA emissions in current emission inventories. For example, Shrivastava et al. (2011) estimated the sum of all SVOC and IVOC precursors to be 7.5 times the mass of traditional 112 113 POA emissions inventory over Mexico City, but indicated that the scaling factor of 3 for SVOC 114 emissions based on the POA emissions is poorly constrained. Shrivastava et al. (2008) and Jathar et al. (2011) assumed that SVOC emissions are represented by the traditional emission inventory 115 116 while IVOC emissions are 1.5 times the traditional emission inventory. Pye and Seinfeld (2010) 117 assumed that SVOC emissions are a subset of traditional POA emission inventories, and their POA 118 emissions were scaled up by 27% on a global scale. IVOC emissions are assumed to be spatially-119 distributed similar to naphthalene and are predicted to be roughly a factor of half of global POA emissions. Tsimpidi et al. (2014) assumed that the IVOC emissions are 1.5 times the traditional 120 POA emission inventory and are assigned to the 4<sup>th</sup> volatility bin with  $C^* = 10^5 \ \mu g \ m^{-3}$ . For 121 122 comparison, some studies such as Ahmadov et al. (2012) and Bergstrom et al. (2012) used the 123 VBS approach for OA modeling but did not include additional SVOC emissions. There are also





differences in the volatility distribution used in literature. Shrivastava et al. (2008) and Jathar et al. (2011) found that moving half the mass of SVOC from all bins to the lowest bin from the traditional "diesel exhaust" volatility distribution of Robinson et al. (2007) produced the lowest errors in simulated OA on an annual average basis.

The number of bins used can also result in differences in simulated SOA concentrations. Shrivastava et al. (2011) showed that the 2-species VBS performed better than the 9-species VBS in modeling oxygenated organic aerosol (OOA) and gave the closest agreement to the OOA calculated by the Positive Matrix Factorization (PMF) method. This indicates that SOA may be less volatile as compared to the volatility distribution in the 9-species VBS which allows for evaporation of SOA with dilution (Shrivastava et al., 2011).

The amount of oxygen added for each oxidation step may contain uncertainties. This factor 134 can influence the O:C ratio used for the model evaluation. O:C predictions from models need to 135 136 be improved by including fragmentation reactions (which could lead to an increase in O:C ratios) 137 and improving emission estimates (Shrivastava et al., 2011). Different rate constants can also result 138 in different predictions of SOA concentrations. For example, Farina et al. (2010) showed that the use of k value of  $1 \times 10^{-12}$  cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> compared to the default k value of  $10 \times 10^{-12}$  cm<sup>3</sup> 139 molecule<sup>-1</sup> s<sup>-1</sup> resulted in a reduced aged SOA formation by 71%. Hodzic et al. (2010) also showed 140 141 a case study based on Grieshop et al. (2009) in which each oxidation step reduced the volatility of 142 the S/IVOC vapors by two orders of magnitude and each successive oxidation step produced a 143 40% increase in mass due to the addition of oxygen. This case is inconclusive in urban areas - a 144 larger bias along with a higher correlation coefficient compared to the more common case where 145 each oxidation step reduced the volatility by one order of magnitude with a 7.5% increase in mass.





146 However, the model performed worse (with larger bias and lower correlation coefficient) in

147 suburban areas.

The aging process improves model performance in general in the United States (U.S.) but deteriorates the performance in several parts of Europe. Accounting for the aging process of OA will increase the OA concentrations and improve model results in the U.S. where OA is usually underpredicted, but increase the model bias for OA in several parts of Europe where OA concentrations are overpredicted (Farina et al., 2010; Bergstrom et al., 2012).

Shrivastava et al. (2013) studied the effects of the fragmentation and functionalization in VBS. 153 154 Functionalization increases the mass of OA for each successive oxidation step, while 155 fragmentation reduces the mass for each oxidation step. One such a case includes simulating first-156 order effects of the fragmentation and functionalization processes in VBS by assuming 157 functionalization of 100% of organic vapors for the first two generations of oxidation and both fragmentation and functionalization for the third and higher generations of oxidation. The 158 159 fragmentation reduces the SOA concentrations drastically. For example, Shrivastava et al. (2013) 160 showed that peak SOA concentrations can be reduced by factors of 2 to 4 for a 1-hour example on 161 10 March 2006 at 21 UTC over Mexico City Plateau.

The VBS framework for OA modeling in the latest version of WRF/Chem, v3.7.1, is coupled with several gas-phase mechanisms including the 2005 Carbon Bond Mechanism (CB05) (Yarwood et al., 2005), the Model for Ozone and Related chemical Tracers version 4 (MOZART-4) (Emmons et al., 2010), the Regional Atmospheric Chemistry Model (RACM) (Stockwell et al., 166 1997), and the 1999 version of the Statewide Air Pollution Research Centre (SAPRC99) mechanism (Carter, 2000). Different gas-phase mechanisms have different lumpings/groupings 168 for VOCs, which will affect OA formation. For example, VOCs are lumped according to their





169 carbon bonds (e.g., single or double bond) in CB05 (Yarwood et al., 2005) while VOCs in 170 SAPRC99 (Carter, 2000) are lumped according to their OH reactivities. A number of studies 171 examined the differences in predicting O<sub>3</sub> concentrations due to different gas-phase mechanisms 172 (e.g., Luecken et al., 2008; Li et al., 2012; Shearer et al., 2012; Zhang et al., 2012), but fewer 173 studies reported the impact of different gas-phase mechanisms on modeling SOA and PM25 174 concentrations (Kim et al., 2011; Zhang et al., 2012). SAPRC99 has more detailed organic 175 chemistry compared to CB05. SAPRC99 has been updated to SAPRC07 (and recently, to SAPRC11) based on newly available information regarding the reactions and influence of 176 177 individual VOCs on O<sub>3</sub>, as well as evaluations against chamber experiments (Carter, 2010). In 178 addition, SAPRC07 has reformulated reactions of peroxy radicals so that the effects of changes in 179 nitrogen oxides (NOx) on organic product formation is more accurately represented. SAPRC07 180 has the most extensive set of VOC species and reactions, as compared to CB05 and the Carbon 181 Bond version 6 (CB6). Shearer et al. (2012) reported that a condensed version of SAPRC07 182 predicted lower O<sub>3</sub> and OH concentrations in central California compared to SAPRC99 due to a decreased reaction rate coefficient in the reaction of OH and NO<sub>2</sub> to form HNO<sub>3</sub>. Li et al. (2012) 183 184 also showed that predicted O<sub>3</sub> concentrations from SAPRC07 were lower than those of SAPRC99 185 by up to 20% over Texas. The same study also reported that SAPRC07 gave lower OH 186 concentrations due to differences in the reaction rate constants in the reactions of O<sup>1</sup>D and H<sub>2</sub>O 187 between SAPRC07 and SAPRC99. Luecken et al. (2008) reported that SAPRC99 gave higher O<sub>3</sub> 188 concentrations compared to CB05 on average; however, the differences vary with locations, 189 VOC/NO<sub>x</sub> ratios, and the concentrations of precursor pollutants. This is consistent with the results 190 from Zhang et al. (2012), which predicted that SAPRC99 using WRF/Chem with the Model of 191 Aerosol, Dynamics, Reaction, Ionization and Dissolution (WRF/Chem-MADRID) produced the





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192 highest O<sub>3</sub> mixing ratios in July at the Southeastern Aerosol Research and Characterization 193 (SEARCH) sites. The CB6 (Yarwood et al., 2010) is an updated version of CB05 with improved 194 kinetic and photolysis data, additional explicit species for long-lived and abundant organic 195 compounds including propane, acetone, benzene and acetylene, as well as revised isoprene and 196 aromatics chemistry from CB05. Yarwood et al. (2010) showed that CB6 produces higher daily 197 maximum 8-hr O<sub>3</sub> as compared to CB05 over Los Angeles for one episode day in August with the 198 highest observed O<sub>3</sub> mixing ratios. CB6 was also shown to produce substantially higher OH 199 concentrations (25% to 50% higher at mid-day over large areas) over eastern U.S. compared to 200 CB05 over a few days in June, 2006. A summary of the main characteristics of CB05, CB6, and 201 SAPRC07 gas-phase mechanisms are listed in Table 2.

#### 202 1.2. Description of Aerosol Activation Parameterizations

204 Ghan et al. (2011) provided a comprehensive review on various aerosol activation treatments 205 in current climate models. Two main types of parameterizations are commonly used: the Abdul-206 Razzak and Ghan (2000) (AR-G00) and the Fountoukis and Nenes (2005) (FN05) and associated 207 updates described in Barahona et al. (2010) and Morales Betancourt and Nenes (2014). AR-G00 208 uses multiple lognormal or sectional distributions to approximate the aerosol size distribution. It 209 uses the Kohler theory to relate the aerosol size distribution and composition to the number of 210 aerosols activated as a function of maximum supersaturation  $(S_{max})$ . The complex function 211 involving  $S_{max}$  is parameterized based on the standard deviation  $\sigma$  from a large number of 212 numerical solutions using a cloud parcel model. The number and mass activated are particles with 213 critical supersaturation less than  $S_{max}$ . It also accounts for particle growth before and after the 214 particles are activated. However, the ARG treatment does not explicitly represent kinetic 215 limitations which tend to affect smaller or larger particles (with diameters far from their critical





216 size). Very small particles tend to lose water when supersaturation declines as they never exceed 217 the critical supersaturation for that particle size, and very large particles may not have achieved 218 the critical size before  $S_{max}$  is reached (Ghan et al., 2011). Kinetic limitations refer to the (i) inertial 219 mechanism – where particles with large dry diameters grow to be as large as activated particles 220 but have not been activated themselves, these particles should be considered together with 221 activated particles; (ii) evaporation mechanism - where particles with high critical supersaturation 222 evaporate before reaching their critical diameters; and the (iii) deactivation mechanism - where 223 initially activated particles that are deactivated to interstitial aerosols when the parcel 224 supersaturation falls below the equilibrium supersaturation (Nenes et al., 2001). Neglecting kinetic 225 limitations performs well for all conditions except in highly-polluted areas (Ghan et al., 2011). In 226 urban and highly-polluted cases, many particles fail to be activated due to strong evaporation and 227 deactivation processes (Nenes et al., 2001). Explicitly accounting for kinetic limitations reduces 228 CDNC at low updraft velocity (Ghan et al., 2011).

229 The Fountoukis and Nenes (2005) (FN05) scheme improved the ARG00 scheme by solving  $S_{max}$  analytically (with the exception of kinetically-limited particles) using a so-called "population 230 231 splitting" method. In addition, FN05 took into account the kinetic limitations, as well as the 232 influence of gas kinetics on water vapor diffusivity (Ghan et al., 2011). The other improved 233 treatments built on top of the FN05 scheme include the entrainment of ambient air, which could 234 reduce the supersaturation of the updraft (Barahona and Nenes, 2007) (BN07) (therefore reducing 235 CDNC); the adsorption of water vapor onto insoluble particles by Kumar et al. (2009) (KU09) 236 based on a modified Frenkel-Halsey-Hill (FHH) adsorption theorem (which will increase CDNC); 237 the growth of giant cloud condensation nuclei (CCN) (Barahona et al., 2010) (BA10) by 238 introducing an additional condensation rate term to account for condensation of giant CCN (which





will reduce CDNC); as well as the modification of the original population splitting concept in
FN05 and BA10 by Morales Betancourt and Nenes (2014) (MN14) by better accounting for the
size of inertially limited CCN, and removing a discontinuity in the calculation of the surface area
of cloud droplets.

The parameterization of Abdul Razzak and Ghan (2000) (ARG00) is used as the default aerosol activation module in WRF/Chem. It is not linked to the microphysics module or cumulus parameterization in WRF or WRF/Chem. However, for WRF/Chem, the cloud droplet number concentration (CDNC) generated in ARG00 is passed to the microphysics scheme, i.e., the Morrison two-moment microphysics scheme selected in this work.

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#### 1.3 Motivations and Objectives

The online-coupled meteorology and chemistry model, WRF/Chem, has recently been 250 251 applied for air quality and climate modeling on a decadal scale (Yahya et al., 2016a, b). 252 WRF/Chem can also simulate aerosol direct and indirect feedbacks, which are important considerations for climate modeling. However, as mentioned previously, the representations of 253 254 OA and aerosol-cloud interactions in current regional and global climate models are subject to 255 large uncertainties. In particular, while the VBS framework in WRF/Chem significantly improves 256 SOA performance (Wang et al., 2015), it lacks the semi-volatile POA treatment, as well as 257 fragmentation processes (Shrivastava et al., 2013). The first objective of this study is to reduce 258 uncertainties associated with OA predictions by improving the existing VBS module in 259 WRF/Chem and identifying the best gas-phase chemical mechanism to drive the VBS module for 260 the most accurate OA predictions. The impact of the improved OA predictions on CDNC in 261 WRF/Chem will be quantified. The second objective is to incorporate an improved aerosol





262 activation parameterization based on the FN05 series into WRF/Chem to study its impacts on

263 CDNC predictions.

#### 264 2. Model Configuration, Evaluation Protocol, and Observational Datasets

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### 2.1.Model Setup and Inputs

268 The model used in this study is a modified version of WRF/Chem v3.7.1 as described by Wang 269 et al. (2015). The 2005 Carbon Bond gas-phase mechanism (CB05) of Yarwood et al. (2005) with 270 additional chlorine chemistry is coupled with the Modal for Aerosol Dynamics in Europe -271 Secondary Organic Aerosol Model (MADE/SORGAM) (Ackermann et al., 1998; Schell et al., 272 2001) and the Volatility Basis Set (MADE/VBS) (Ahmadov et al., 2012). The CB05-VBS option 273 has also been coupled to existing model treatments including the aerosol direct effect, the aerosol 274 semi-direct effect on photolysis rates of major gases, and the aerosol indirect effect on CDNC and 275 resulting impacts on shortwave radiation. The physics options used in WRF/Chem include the 276 rapid and accurate radiative transfer model for GCM (RRTMG) for both shortwave and longwave 277 radiation, the Yonsei University (YSU) planetary boundary layer (PBL) scheme (Hong et al., 2006; 278 Hong, 2010), the Morrison et al. (2009) double moment microphysics scheme, as well as the Multi-279 scale Kain-Fritsch (MSKF) cumulus parameterization scheme (Zheng et al., 2016). Aqueous-280 phase chemistry module (AQCHEM) for both resolved and convective clouds is based on a similar 281 AQCHEM module in CMAQv4.7 of Sarwar et al. (2011). The anthropogenic emissions used are 282 from the 2010 emissions based on the 2008 U.S. Environmental Protection Agency (U.S. EPA) 283 National Emissions Inventory (NEI) from the Air Quality Model Evaluation International Initiative 284 (AQMEII) project (Pouliot et al., 2015). Dust emissions are based on the Atmospheric and 285 Environmental Research Inc. and Air Force Weather Agency (AER/AFWA) scheme (Jones and 286 Creighton, 2011). Emissions from sea salt are generated based on the scheme of Gong et al. (1997).





287 Biogenic emissions are simulated online by the Model of Emissions of Gases and Aerosols from

288 Nature v2.1 (MEGAN2.1) (Guenther et al., 2006).

289 The chemical initial and boundary conditions (ICONs/BCONs) come from the modified 290 CESM/CAM version 5.3 with updates by Gantt et al. (2014), He and Zhang (2014), and Glotfelty 291 et al. (2016). The meteorological ICONs/BCONs are from the National Center for Environmental 292 Protection Final Reanalyses (NCEP FNL) dataset, which is available every 6 hours. The chemical 293 fields are also allowed to run continuously while the meteorology is reinitialized every 5 days. The 294 simulations are performed at a horizontal resolution of 36-km with 148 × 112 horizontal grid cells 295 over the CONUS domain and parts of Canada and Mexico, and a vertical resolution of 34 layers 296 from the surface to 100-hPa. 297 A number of sensitivity simulations are designed to identify the model configuration with

results that are in the closest agreement to observations as well as the realistic model treatments of OA that are the closest to atmospheric processes. The baseline and sensitivity simulations are conducted from May to June 2010, during which the Nexus of Air Quality and Climate Change (CalNex) campaign was held in Bakersfield and Pasadena, California. The first 10 days from May 1<sup>st</sup> to May 10<sup>th</sup> are considered to be the spin-up period.

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#### 2.2. Model Evaluation Protocol and Available Measurements

305 Statistical measures including the Mean Bias (MB), Correlation Coefficient (Corr), 306 Normalized Mean Bias (NMB) and Normalized Mean Error (NME) (Yu et al., 2006) are used to 307 evaluate the simulations against observational data. Observational data are available for organic 308 carbon (OC) and total carbon (TC) from the Speciated Trends Network (STN) and the Interagency 309 Monitoring for Protected Visual Environments (IMPROVE). While both OC and TC from 310 IMPROVE are used for model evaluation, only TC data from STN are used as STN uses the





311 thermo-optical transmittance protocol for OC that is different from the one used by IMPROVE 312 (Zhang et al., 2012). In addition, the measurements for STN OC are not blank corrected for carbon 313 on the background filter (Wang et al., 2012). The ratios OA/OC ratios vary across locations in the 314 continental U.S. (CONUS) depending on whether the OA is dominated by secondary formation 315 (Aitken et al., 2008) or it contains more aliphatic hydrocarbons (Turpin and Lim, 2001). In this 316 study, two ratios, 1.4 and 2.1, are used to convert simulated OA to OC based on a number of 317 literature (Turpin and Lim, 2001; Aitken et al., 2008; Xu et al., 2015). As the simulations are based 318 on CONUS with varying OA properties (less or more oxidized OA), the use of two OA/OC ratios 319 can represent the different types of OA present for all locations in the U.S. Spatial plots, time series 320 plots at specific sites, as well as overlay plots are used to evaluate model performance. The 321 IMPROVE sites chosen for the time series plots include the visibility-protected areas in Brigantine 322 National Wildlife Refuge (NWR), NJ, Death Valley National Park (NP), CA, Swanqwarter 323 National Wildlife Refuge (NWR), NC, and the Tallgrass Prairie National Preserve, KS. The 324 Brigantine NWR is a tidal wetland and has a shallow bay, the Death Valley NP is a desert, and the 325 Swangwarter NWR is a coastal brackish marsh. The time series plots are made at four STN sites 326 including two urban sites: in Washington, DC and Boise, ID, one industrial site in Tampa, FL, and 327 one rural/agricultural site in Liberty, KS. SOA, hydroxyl radical (OH) and hydroperoxy radical 328 (HO<sub>2</sub>) data are also available for May to June 2010 as part of the California Research at the Nexus 329 of Air Quality and Climate Change (CalNex) campaign (Kleindienst et al., 2012; Lewandowski et 330 al., 2013) in Bakersfield, CA and Pasadena, CA, which are both urban locations. The Bakersfield 331 sampling site is located between the city center and areas of agricultural activity, while the 332 Pasadena site is located at the California Institute of Technology campus within the Los Angeles 333 metropolitan area to the southwest and mountains in the north (Baker et al., 2015).





334 POA/OA ratios are also used to evaluate the performance of the model. A number of studies 335 have reported observed POA/OA ratios which range from 15% to 40% over CONUS. For 336 example, over southeastern U.S., hydrocarbon-like OA (HOA) and cooking OA are found to 337 contribute to 21 – 38% of total OA in urban sites (Xu et al., 2015). HOA and oxygenated OA 338 (OOA) are found to account for 34% and 66% of measured OA from Pittsburgh in September 2002 339 (Zhang et al., 2005). HOA and cooking OA are assumed to be synonymous to POA, and OOA is 340 assumed to be synonymous to SOA. Particulate matter sampled during August and September 341 2006 in Houston as part of the Texas Air Quality Study II Radical and Aerosol Measurement 342 Project showed that approximately 32% of OA comes from HOA (Cleveland et al., 2012). Results 343 from positive matrix factorization analysis from the Pasadena ground site during May and June 344 2010 showed that the primary components contribute 29% of the total OA mass (Hayes et al., 345 2013). Based on Zhang et al. (2007), the percentages of HOA mass at urban sites in Riverside, 346 CA, from mid-July to mid-August 2005, in Houston, TX, from mid-August to mid-September 347 2000, and in New York City in July 2001 are 15%, 38%, and 30%, respectively. In addition, Zhang 348 et al. (2011) compiled a large number of field campaigns across the globe where the average 349 POA/OA ratios for urban, downwind and rural/remote areas are found to be 0.42, 0.18 and 0.10 350 respectively.

For the aerosol activation sensitivity and production simulations, additional variables that will be analyzed in this study include maximum 1-hour and 8-hour O<sub>3</sub> against the Clean Air Status and Trends Network (CASTNET) and Air Quality System (AQS), aerosol optical depth (AOD), CDNC and cloud condensation nuclei (CCN) against MODIS.

- 355 **3. Model Development and Improvement**
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- 357 A number of modifications have been made to the standard version of WRF/Chem model
- 358 v3.7.1. Those modifications and treatments are described below.
- 359 3.1. OA Treatments

361 The CB05-VBS treatment in the default WRF/Chem v3.7.1 assumes that POA is nonreactive 362 and nonvolatile. In this study, POA is assumed to be semivolatile, and can undergo gas-particle 363 partitioning, similar to anthropogenic SOA (ASOA) and biogenic SOA (BSOA) in VBS. While the volatility of ASOA and BSOA is represented by 4 bins with C\* from  $10^0$  to  $10^3 \,\mu g \, m^{-3}$ . The 364 POA is distributed into 9 bins, with C\* from  $10^{-2}$  to  $10^{6}$  µg m<sup>-3</sup>, following the set-up of Shrivastava 365 366 et al. (2011). The POA is oxidized to form semi-volatile OA (SVOA), which can also undergo 367 gas-particle partitioning. For the POA, bin-resolved enthalpies of vaporizations are used, ranging from 64 kJ mol<sup>-1</sup> for the 9<sup>th</sup> bin to 112 kJ mol<sup>-1</sup> for the 1<sup>st</sup> bin according to Shrivastava et al. (2011). 368 369 The default enthalpy of vaporization ( $\Delta H_{vap}$ ) for SOA in WRF/Chem is 30 kJ mol<sup>-1</sup> according to 370 Lane et al. (2008). A more accurate alternative is to use the  $\Delta H_{vap}$  values calculated from the semi-371 empirical correlation from Epstein et al. (2010):

$$\Delta H_{vap} = -11 \, le$$

 $H_{vap} = -11 \log_{10} C^*_{300} + 129 \tag{1}$ 

The values of  $\Delta H_{vap}$  Epstein et al. (2010) are used in a number of sensitivity simulations and final production simulation.

375 Shrivastava et al. (2013, 2015) also implemented several cases of fragmentation and 376 functionalization (FF) processes into VBS. For this study, the FF set-up is similar to the method 377 employed by Shrivastava et al. (2013), with the exception that fragmentation percentages of 10%, 378 25%, and 50% are used in sensitivity simulations. Shrivastava et al. (2013) used fragmentation 379 percentages of 50% (intermediate fragmentation) and 85% (high fragmentation) in his simulations 380 over Mexico City. For example, for the 10% FF case, 10% of the mass in the VBS species is





functionalized and moved to the next lower volatility bin, 80% is fragmented and moved to the highest volatility bin, and the remaining 10% is fragmented and becomes more volatile than the highest volatility bin (i.e., it is lost from the current volatility bins). For the 50% FF case, 50% is functionalized and moved to the next lower volatility bin, 40% is fragmented and moved to the highest volatility bin, and 10% is lost.

2hao et al. (2014) measured IVOCs in Pasadena, CA during CalNex and found that the 2kao et al. (2014) measured IVOCs are similar to those of single-ring aromatics, and they produce 2kao about 30% of newly formed SOA in the afternoon. With the semivolatile POA and FF cases in this 2kg additional IVOC and SVOC emissions are added as three times of the traditional POA 2kg emissions from NEI, to account for missing IVOC and SVOC species in the traditional POA 2kg emission inventory. The fraction of IVOC/SVOC emissions assigned to each volatility bin is 2kg summarized in Table 3.

393 The mass fraction of organics in each volatility bin determined in laboratory studies also differs 394 significantly according to the sources of organics. For example, May et al. (2013a, b, c) has 395 different volatility distributions of mass fractions of organics for gasoline vehicle exhaust, diesel 396 exhaust, and biomass burning. To take into account the different sources of organic compounds 397 into a single volatility distribution for the purpose of this work, a new volatility distribution is 398 calculated based on the mass fractions reported by Shrivastava et al. (2011), May et al. (2013a, c) 399 and the percentages of VOC emissions from various sources from the 2008 NEI. According to the 400 2008 NEI report (Rao et al., 2013), total VOC emissions from stationary, mobile and fire 401 (prescribed and wildfire) sources are ~7.6, ~5.6, and ~49.6 million tons, respectively. The 402 corresponding percentages for VOC emissions are ~12%, ~9%, and ~79% for stationary, mobile, 403 and fire sources, respectively. Based on the U.S. EPA (2013), the percentages of diesel emissions





404 from mobile sources are low compared to gasoline sources (~7% of total diesel and gasoline

405 sources); they are thus not included in this study.

406 An example calculation for the mass fraction of the lowest volatility bin for POA and 407 IVOC/SVOC emissions are as follows:

408 
$$\text{Log } C_{-2}^* (\text{at } 298\text{K}) = 0.04 \times 12\% + 0.14 \times 9\% + 0.79 \times 79\% = 0.1754$$
 (2)

where  $C_{-2}^*$  refers to the lowest volatility bin with a value of  $10^{-2} \mu \text{g m}^{-3}$ , 12%, 9%, and 79% refer 409 to the percentages for VOC emissions from stationary, mobile, and fire sources, respectively from 410 NEI, 0.04 refers to the original mass fraction for stationary emissions based on anthropogenic 411 emissions from Shrivastava et al. (2011) for the lowest volatility bin with a value of  $10^{-2} \mu \text{g m}^{-3}$ , 412 413 0.14 refers to the original mass fraction for gasoline emissions from May et al. (2013a) for the lowest volatility bin with a value of  $10^{-2} \,\mu g \, m^{-3}$ , 0.2 refers to the original mass fraction for biomass 414 burning emissions from May et al. (2013c) for the lowest volatility bin with a value of  $10^{-2} \,\mu g \,m^{-1}$ 415 <sup>3</sup>, and 0.1754 refers to the newly-calculated mass fraction of POA and IVOC/SVOC emissions for 416 417 this study. The mass fractions used by Shrivastava et al. (2011), May et al. (2013a, c), and this 418 work can be found in Table 3.

419 420

#### 3.2. Gas-Phase Chemical Mechanisms

Three gas-phase mechanisms are used: CB05, CB6, and SAPRC07. The gas-phase mechanisms for CB6 and SAPRC07 are coupled to the MADE/VBS in WRF/Chem v3.7.1 in this work following the coupling of CB05 with MADE/VBS by Wang et al. (2014). The emissions for all cases are based on the CB05 chemical species from the 2010 emissions based on the 2008 NEI. For SAPRC07, slight modifications had to be made to account for the different VOC species or groups. The mapping of emission species from CB05 to SAPRC07 is based on the grouping of





427 emitdb.xls species from from Henderson et al. (2014)as well as from 428 http://www.cert.ucr.edu/~carter/emitdb/old-emitdb.htm. CB05 emissions are used for the CB6 429 case, with the exception of the VOCs including propane, benzene, ethyne, acetone, and ketone that 430 are mapped based on fractions of existing CB05 VOCs according to Yarwood et al. (2010). 431 In VBS, the SOA precursors for CB6 are similar to those for CB05. The SOA precursors for 432 CB05 (and therefore CB6) are mapped from the default SAPRC99 precursors by Wang et al. (2015). The SAPRC07 SOA precursors follow the existing mapping of SAPRC99-MOSAIC/VBS 433 434 in WRF/Chem. The chemical equations and rate parameters from ENVIRON (2013) and Carter 435 (2010) for CB6 and SAPRC07 gas-phase mechanisms, respectively, were included in the 436 chem/KPP/mechanisms directory in WRF/Chem. The SAPRC07 gas-phase mechanism 437 implemented in WRF/Chem in this case is the uncondensed and expanded version C, which includes reactions for peroxy radical operators (Carter, 2010). Species in both CB6-MADE/VBS 438

and SAPRC07-MADE/VBS undergo dry deposition, aqueous chemistry, photolysis, and wet
 scavenging that are similar to CB05-MADE/VBS.

#### 441 **3.3. Aerosol Activation**

442

443 The FN05 series aerosol activation parameterizations (with the exclusion of MN14) have been 444 incorporated into 3-D regional air quality models and global climate and Earth system models such 445 as the WRF-Community Atmosphere Model version 5 (WRF-CAM5) (Zhang et al., 2015), and in 446 the global-through-urban WRF/Chem (GU-WRF/Chem) (Zhang et al., 2012) and CESM (Gantt et 447 al., 2014). In this study, the FN series parameterizations are incorporated into WRF/Chem 448 following the methods of Gantt et al. (2014) and Zhang et al. (2015) as described in detail in Zhang et al. (2015). However, in WRF/Chem, the aerosol activation module is only linked to the 449 450 microphysics module through the variable CDNC, which is read by the microphysics module. It





451 is not coupled to the cumulus parameterization scheme unlike in WRF-CAM5 and CESM. The 452 FN05 series has been incorporated into module mixactivate. F in the physics directory in 453 WRF/Chem. As BN07 involves the entrainment effect for convective clouds and has very small 454 impacts on non-convective CDNC (Zhang et al., 2015), it is not included in this study. In addition, 455 unlike Gantt et al. (2014) and Zhang et al. (2015), the KU09 treatment is also not included in this study as the empirical constants AFHH and BFHH used in the formulation, which are compound-456 457 specific, have not been experimentally determined for black carbon, although those constants have 458 been determined for dust and confirmed by Laaksonen et al. (2016). The additional MN14 459 treatment incorporated in this study involves a small modification to the original FN05 series 460 parameterizations (without KU09), and helps to better account for the size of inertially limited 461 CCN, and to remove a discontinuity in the calculation of the surface area of cloud droplets (Morales Betancourt and Nenes, 2014). The updated treatments are about 20% more 462 463 computationally expensive to run as compared to ARG00 (Zhang et al., 2016), but capture the 464 sensitivity of CDNC to all aspects of the aerosol with comparable accuracy to numerical parcel 465 models, which was shown to be an underlying reason for biases from ARG (Morales Betancourt 466 et al., 2014).

467 **4. Res** 

468

470

## 4. Results and Discussions

#### 469 4.1.Sensitivity Simulations with VBS Treatments Coupled with CB05

As listed in Table 4, a number of sensitivity simulations are designed to identify the best model configuration for OA treatments with the closest agreement to observations over CONUS. Those sensitivity simulations consider (i) two SOA modules (MADE/SORGAM vs MADE/VBS), (ii) two types of VBS treatment for POA (nonvolatile POA vs. semivolatile POA), (iii) two  $\Delta H_{vap}$ treatments (default vs. the semi-empirical  $\Delta H_{vap}$  equation by Epstein et al. (2010)), (iv) three





476 different percentages of functionalization and fragmentation (FF) (10%, 25%, and 50%), (v) three sets of POA emissions (default vs. 1.5 or 3 times the original NEI POA emissions), (vi) three 477 478 different gas-phase mechanisms (CB05, CB6, and SAPRC07), and (vii) two different aerosol 479 activation schemes (ARG00 vs. combinations of different aerosol activation schemes of the FN05 480 series: FN05, FN05/BA10, and MN14) All simulations except for CB05-SORG-DH contain the 481 VBS treatments for OA. CB05-SORG-DH and CB05-VBS-DH treat POA emissions as 482 nonvolatile. In addition, the impact of two different cumulus parameterization schemes: Grell -Freitas (Grell and Freitas, 2014) and the Multi-scale Kain Fritsch (MSKF) (Zheng et al., 2016) 483 484 scheme were also tested.

485 Table 5 summarizes the main statistics for all sensitivity simulations in terms of mean obs, 486 mean sim, Corr, NMB, and NME for hourly OC and TC concentrations against IMPROVE and 487 hourly TC concentrations against STN, respectively, over the whole CONUS domain. Figure 1 488 compares the domain mean hourly averaged observed OC or TC concentrations based on 489 IMPROVE and STN with simulated concentrations calculated based on the ratios of OA/OC 1.4 490 and 2.1 for each sensitivity simulation. The domain mean hourly averaged obs OC concentration 491 is 0.88  $\mu$ g m<sup>-3</sup> for IMPROVE, and the domain mean hourly averaged obs TC concentration is 1.03  $\mu g$  m<sup>-3</sup> for IMPROVE and 2.71  $\mu g$  m<sup>-3</sup> for STN. As shown in Figure 1, the simulation 492 CB05 SORG DH with the default SOA module SORG largely underpredicts OC and TC with the 493 494 largest NMBs and NMEs and the lowest Corr as compared to all other simulations with a SOA 495 module based on the VBS method. The remaining VBS simulations significantly reduce the biases 496 and errors in OC and TC from CB05\_SORG\_DH and also improve the correlation. Compared to 497 CB05 SORG DH, CB05 VBS DH with nonvolatile POA seems to perform relatively well in 498 terms of NMBs and Corr against IMPROVE OC, IMPROVE TC, and STN TC.





499 Adding the semivolatile POA treatment with 1.5 times the NEI POA emissions (CB05 POA DH) reduces simulated OC and TC concentrations as compared to CB05 VBS DH, 500 501 due to the loss of mass from the semivolatile POA. As the POA mass is reduced, less surface area 502 is available for SOA precursors to condense onto, resulting in decreased OA (thus decreased OC 503 and TC) for CB05 POA DH. Using the semi-empirical correlation of Epstein et al. (2010) for 504 △H<sub>vap</sub> increases the OC and TC concentrations (CB05\_POA vs. CB05\_POA\_DH). Compared to 505 the default  $\Delta H_{vap}$  of 30 kJ mol<sup>-1</sup> used in CB05\_POA\_DH, the semi-empirical correlation of Epstein 506 et al. (2010) gives much higher  $\Delta H_{vap}$  values, resulting in more of the organic vapors in the 507 particulate phase than in the gas phase. Compared to CB05-POA, the simulations with various FF 508 treatments decrease the OA concentrations, as part of the OA mass is fragmented to higher 509 volatility bins. The 10%FF case (CB05\_10%FF) does not differ significantly from the no FF case 510 (CB05\_POA). However, increasing the percentage of FF (from 10% to 25%, then to 50%) 511 decreases the OA concentrations. The FF treatments, however, even if they are more representative 512 of actual SOA atmospheric formation processes, reduce the Corr slightly (compared to the cases 513 CB05\_POA and CB05-10%FF). By doubling the POA emissions (from 1.5 to 3.0 times the 514 original POA NEI emissions) for the 25% FF case (CB05\_FF25%\_EM3), the predicted OC and 515 TC concentrations are closer to the observations. When evaluated against IMPROVE OC, 516 IMPROVE TC, and STN TC, among for simulations using CB05, the simulations 517 CB05\_VBS\_DH, CB05\_POA, CB05\_FF10%, and CB05\_FF25%\_EM3 perform better than other 518 cases. The differences in the OC and TC predictions from the simulations with different gas-phase 519 mechanisms will be discussed later in Section 2.

520 Figure 2 shows the spatial distributions of simulated OC and TC concentrations overlaid with 521 observed OC from IMPROVE and TC from STN for the case CB05\_25%FF\_EM3 for the two





522 OA/OC ratios. The model performs much better for IMPROVE OC with an OA/OC ratio of 2.1 523 as compared to 1.4, especially over eastern U.S. where the use of an OA/OC ratio of 1.4 results in 524 large overpredictions. However over the central U.S. and parts of the western U.S., the use of an 525 OA/OC ratio of 1.4 shows slightly better predictions of IMPROVE OC compared to the use of 526 OA/OC ratio of 2.1 that gives underpredictions of OC. On the other hand, the model performs 527 better for STN OC with an OA/OC ratio of 1.4 as compared to 2.1. The use of an OA/OC ratio of 528 1.4 gives better agreement with STN TC over eastern U.S. where the use of an OA/OC ratio of 2.1 529 results in large underpredictions of TC. Evaluation of OC and TC against IMPROVE and STN, 530 respectively, therefore depends heavily on the OA/OC ratio, which is site-specific. Therefore in 531 more rural sites (IMPROVE), the OA/OC ratio is more likely to be high (~2.1) with more 532 oxygenated OA, while in more urban sites (STN), the OA/OC ratio is more likely to be lower 533 (~1.4) due to fresher emissions and less oxidized species.

534 Figure 3 shows the POA/OA ratios for six sensitivity simulations. As mentioned earlier, the 535 observed ratio of POA/total OA is approximately 15% to 40% during the summer period over various locations in the CONUS. As SOA concentrations from field campaigns are sparse at 536 537 different locations and at different time periods, the POA/OA ratio is used to evaluate the model's 538 capability to reproduce POA and SOA concentrations. The simulation CB05\_SORG\_DH with 539 default SORGAM SOA module largely overpredicts the POA/OA ratio, due to significant 540 underpredictions of SOA. The simulations CB05 VBS DH, CB05 50%FF, and 541 CB05 25%FF EM3 with various VBS treatments all have POA/OA ratios that fall within the 542 range of 0.15 to 0.4, with lower POA/OA ratios over more rural areas and higher POA/OA ratios 543 over urban areas. CB05\_VBS\_DH, however, might give too high POA concentrations over the western portion of the domain as it does not consider POA to be semivolatile. Considering 544





545 semivolatile POA, however, without considering the fragmentation and functionalization 546 processes in the simulation CB05 POA results in too low POA/OA ratio (< 0.1 over most areas). 547 Similarly, the CB05 FF25% case also results in a large portion of CONUS with POA/OA ratios 548 of < 0.1, due to the loss of POA mass. CB05\_FF50%, however, predicts reasonable POA/OA 549 ratios, even with fragmentation/functionalization due to balanced loss of both POA and SOA mass 550 through fragmentation to higher volatility bins. The simulation CB05 FF25% EM3 also improves 551 from CB05\_FF25% by increasing the POA mass contributing to higher POA/OA ratios. 552 Figure 4 shows the observed and simulated temporal variations of SOA concentrations at the 553 two CalNex sites: Bakersfield and Pasadena in CA from May to June 2010 for the simulations 554 CB05 SORM DH, CB05 VBS DH, CB05 25%FF EM3, CB6 25%FF EM3. and 555 SAPRC07\_25%FF\_EM3. There are large underpredictions of SOA by all runs on some days (e.g. 556 May 15 - 16, June 2 - 6, June 13 - 14) likely due to missing SOA precursor emissions. Table 6 557 shows the statistics of the simulations presented in Figure 4. The results using CB6 and SAPRC07

558 gas-phase mechanisms will be discussed in section 4.2. The observed SOA was derived based on

the tracer method of Kleindienst et al. (2012) which contains some uncertainties. For example, it

assumes mass fraction of the tracers in secondary organic carbon is the same in the field as that in

the laboratory, and the tracers are assumed to be inert and are unlikely to undergo oxidation in the

562 atmosphere, which might not be the case. In addition, the SOA data from the CalNex campaign

563 only consider contributions from a small number of precursors including biogenic precursors (i.e.,

564 isoprene,  $\alpha$ -pinene, and  $\beta$ -caryophyllene), and the anthropogenic precursors (i.e., toluene,

565 polycyclic aromatic hydrocarbons (PAHs) and methyl butenol (MBO)).

As shown in Figure 4 and Table 6, the simulation CB05\_SORG\_DH with the default SORGAM SOA module significantly underpredicts observed SOA concentrations at both sites.





568 The model configuration of CB05\_VBS\_DH has been used in a number of WRF/Chem 569 simulations published in literature (e.g., Yahya et al., 2015a; Campbell et al., 2015; Wang et al., 570 2015a, b). At Bakersfield, the simulation CB05 VBS DH overpredicts the SOA concentrations 571 for almost all the days. The simulation CB05\_25%FF\_EM3, however, underpredicts the SOA 572 concentrations at Bakersfield, especially in June. The CB05 25%FF EM3 case also shows low 573 SOA concentrations throughout May and June, without much variability in SOA concentrations, 574 likely due to underestimations of original POA emissions at Bakersfield. As the S/IVOC emissions 575 for CB05\_25%FF\_EM3 are a factor of 3 of the original POA emissions from NEI, if the original 576 POA emissions from NEI are underestimated, the S/IVOC emissions will be low, resulting in low 577 SOA concentrations due to low concentrations of condensable material. At Pasadena, both CB05 VBS DH and CB05 25%FF EM3 overpredict the obs SOA from May 15<sup>th</sup> to May 30<sup>th</sup>, 578 but are unable to capture the high SOA concentrations from 2<sup>nd</sup> to 6<sup>th</sup> June. The CB05 VBS DH 579 580 case seems to perform better than the CB05 25%FF EM3 case when observed SOA 581 concentrations are high. The results from this study are consistent with those from Baker et al. (2015), which showed that measured  $PM_{2.5}$  OC at Bakersfield is largely underpredicted compared 582 583 to Pasadena. Baker et al. (2015), however, attributed to the underpredictions of OC at Bakersfield 584 and Pasadena mainly to primary OC predicted by the baseline model, as compared to the Aerosol 585 Mass Spectrometer measurements, suggesting that OC is mostly secondary in nature in Pasadena. 586 In addition, as mentioned earlier, the simulated SOA from WRF/Chem does not consider 587 contributions from all the SOA precursors identified by their trace compounds (e.g., the biogenic 588 precursor, b-caryophyllene, and the anthropogenic precursor MBO, are not included in 589 WRF/Chem), which can help to account for the discrepancies between the simulated and observed 590 SOA concentrations.





591 4.2. Sensitivity of OA predictions to different gas-phase mechanisms 592 593 Figure 1 shows that CB05 FF25% EM3 produces the highest OC and TC concentrations at the IMPROVE sites, followed by CB6\_FF25%\_EM3 and SAPRC07\_FF25%\_EM3, while 594 595 CB6 FF25% EM3 produces the highest TC concentrations at the STN sites. However, the 596 differences in domain-mean simulated OC and TC between the simulations with the three different 597 gas-phase mechanisms are small, compared to the differences in simulated OC and TC due to 598 differences in VBS treatments (e.g., nonvolatile vs. semivolatile POA). Figure 4 also shows that 599 there are not much differences between simulated SOA concentrations with different gas-phase 600 mechanisms at Bakersfield, but larger differences are found at Pasadena. . For example, 601 SAPRC07 25%FF EM3 produces much higher SOA concentrations compared to 602 CB05\_25%FF\_EM3 and CB6\_25%FF\_EM3 at Pasadena on several days (e.g., June 6-8). Figure 603 5 shows the time series of hydroxyl radical (OH) mixing ratios as well as diurnal plots of OH and 604 hydroperoxyl radical (HO<sub>2</sub>) at Pasadena from the CalNex field campaign. The time series of HO<sub>2</sub> 605 is not shown due to irregularity of the observational data. The model is able to reproduce the diurnal variation of OH radicals but significantly overpredicts the daytime and peak OH mixing 606 607 ratios, especially for CB05 and CB6. All gas-phase mechanisms underpredict OH mixing ratios at 608 night. Among all simulations, SAPRC07 produces the closest simulated OH mixing ratios 609 compared to CB05 and CB6 gives the largest overpredictions. Similarly, the HO<sub>2</sub> mixing ratios 610 are generally overpredicted by all gas-phase mechanisms with SAPRC07 performing the best. The 611 overpredictions in OH and HO<sub>2</sub> mixing ratios do not help explain the underpredictions of SOA for 612 several days at Pasadena where underpredictions of VOCs may be the main cause, which is 613 consistent with the findings of Baker et al. (2015).





614 Figure 6 shows spatial distributions of average concentrations of oxidants including ozone 615 (O<sub>3</sub>), OH, HO<sub>2</sub>, as well as the OA species including anthropogenic SOA (ASOA), biogenic SOA 616 (BSOA), TSOA, and POA. SAPRC07-25%FF-EM3 produces the highest domain average O<sub>3</sub> 617 mixing ratios but the lowest domain average OH+HO<sub>2</sub> mixing ratios while CB6-25%FF-EM3 618 produces the highest domain average and maximum OH+HO<sub>2</sub> mixing ratios but the lowest domain 619 average  $O_3$  mixing ratios. These findings are mostly consistent from literature. For example, 620 maximum O<sub>3</sub> and OH mixing ratios over the Los Angeles area are higher for CB6 compared to 621 CB05, which are consistent with the results from Yarwood et al. (2010). SAPRC07 also generally 622 produces higher O<sub>3</sub> mixing ratios compared to CB05. However, average O<sub>3</sub> mixing ratios from 623 CB6 are expected to be higher than CB05 (rather than lower as shown in Figure 6), according to 624 the study from Nopmongcol et al. (2012) which showed higher  $O_3$  mixing ratios over Europe for 625 January and July using the Comprehensive Air Quality Model with Extensions (CAMx). CB6 is a 626 relatively new gas-phase mechanism, there are very few studies that evaluated its performance 627 over a longer period of time, e.g., for the whole summer, and over CONUS. In addition, there are 628 other uncertainties in this study. For example, the emissions for CB05 are used for CB6, the 629 additional explicit VOC species in CB6 such as benzene and acetylene are not considered, which 630 can also contribute to O<sub>3</sub> formation. In addition, most locations in the U.S. in 2010 are considered 631 to by NOx-limited with localized VOC-limited regimes from May to September (Campbell et al., 632 2015), which means that  $O_3$  formation is more likely to depend on NO<sub>x</sub> rather than VOC 633 concentrations.

Table 7 shows the statistics for maximum 1-hr and 8-hr O<sub>3</sub> mixing ratios evaluated against CASTNET and AQS. CASTNET sites are mainly rural sites, while AQS consists of urban, suburban, and rural sites. As expected, SAPRC07 consistently produces the highest maximum 1-





637 hr and maximum 8-hr  $O_3$  mixing ratios and overpredicts at AQS sites with an NMB of ~16%. 638 However, SAPRC07 performs the best at CASTNET sites, as both CB05 and CB6 significantly 639 underpredict maximum 1-hr and maximum 8-hr O<sub>3</sub> mixing ratios. At CASTNET sites, CB6 640 performs the poorest with the largest underpredictions for both maximum 1-hr and 8-hr O<sub>3</sub> mixing 641 ratios. However, CB6 predicts higher maximum 1-hr and 8-hr O<sub>3</sub> mixing ratios at AQS sites, while 642 CB05 predicts the lowest maximum 1-hr and 8-hr O<sub>3</sub> mixing ratios at AQS sites. It is likely that 643 CB6 predicts higher O<sub>3</sub> mixing ratios at more VOC-limited sites in urban areas, while CB05 644 predicts higher O<sub>3</sub> mixing ratios at more NO<sub>x</sub>-limited areas, due to the improvement in VOC 645 speciation in CB6 compared to CB05. Overall, however, CB05 has the highest Corr and the lowest NMEs for CASTNET maximum 1-hr and AQS maximum 1-hr and 8-hr O3 mixing ratios. For 646 PM2.5 concentrations, CB6 produces the best performance against IMPROVE (highest Corr, 647 lowest NMB and NME) while CB05 produces the best performance against STN (highest Corr 648 649 and lowest NME). All 3 cases perform poorly for PM10 against AQS, with large underpredictions 650 due to the non-consideration of the coarse mode inorganic species in MADE-VBS treatments.

651 Anthropogenic SOA (ASOA) concentrations are lower for CB6 and SAPRC07 compared to 652 CB05. This is likely partially due to the emissions which are mapped from CB05 to CB6 and 653 SAPRC07. The CB05 emissions are not likely to account for all anthropogenic VOC emissions in 654 CB6 and SAPRC07, resulting in lower ASOA concentrations for CB6 and SAPRC07 compared 655 to CB05. Biogenic SOA (BSOA) concentrations, however, are the largest for CB6, followed by 656 SAPRC07 and CB05. BSOA concentrations are likely the highest for CB6 due to the highest 657 OH+HO<sub>2</sub> mixing ratios for CB6. The more extensive VOC representation and high O<sub>3</sub> mixing 658 ratios for SAPRC07 also likely contribute to the high BSOA concentrations for SAPRC07





659 compared to CB05. However, overall, the total SOA (TSOA) and POA concentrations for all three

660 gas-phase mechanisms do not vary much, resulting in similar OA concentrations.

Figures 7 and 8 show the time series of simulated vs, observed OC from IMPROVE and 661 662 simulated vs, observed TC from STN at several representative sites over CONUS for the different 663 gas-phase mechanisms. In general, at IMPROVE sites, CB05 gives the highest OC concentrations 664 compared to CB6 and SAPRC07 most of the time, resulting in overpredictions of OC 665 concentrations, while CB6 and SAPRC07 perform better against IMPROVE OC. The overpredictions of CB05 are likely due to overpredictions in ASOA (as CB05 produces the highest 666 667 ASOA concentrations compared to CB6 and SAPRC07 as shown in Figure 6). As these sites are 668 located in rural locations, the dominant SOA is likely to be BSOA, or downwind ASOA from more 669 urban areas. With the exception of Death Valley NP, CA, the model performs relatively well in 670 predicting IMPROVE OC concentrations. Simulations with all three gas phase mechanisms 671 overpredict OC concentrations over several days in May in Brigantine NWR, Death Valley and 672 Swanqwarter, but is able to predict several of the peaks in June. All three gas-phase mechanisms, however, largely underpredict OC concentrations over Death Valley from May 21<sup>st</sup> to June 30<sup>th</sup>. 673 674 As the Death Valley NP is a desert, the OC at Death Valley NP is most likely due to downwind 675 OC transported from upwind locations, for which the model is not able to capture due to 676 meteorological biases such as biases in wind fields. The differences between simulation results 677 from the gas-phase mechanisms are smaller for STN TC compared to IMPROVE OC, probably 678 due to similar elemental carbon (EC) concentrations for all gas-phase mechanisms, which can form 679 a significant percentage of TC. In general, all simulations with the three gas-phase mechanisms 680 also show similar trends (peaks and troughs) for simulated TC, likely due to influences from 681 meteorological parameters such as wind and precipitation. Overall, all three simulations are also





able to predict the magnitude and trends of STN TC concentrations relatively well. Similarly,
CB05 tends to produce the highest TC concentrations, however, CB6 also does produce the highest
TC concentrations for several days, for example, for some days in May in Washington, DC and
Tampa, FL, as well as in June in Liberty, KS, likely due to influences of BSOA where CB6
produces the highest concentrations as shown in Figure 6.

687 688

#### 4.3. Impact of Different VBS treatments on CDNC

Table 7 shows the statistics for model evaluation for simulated CDNC against MODIS-derived 689 690 CDNC from Bennartz (2007) for May to June 2010. All simulations underpredict CDNC due likely 691 to underpredictions in PM and CCN concentrations and uncertainties and/or assumptions in the 692 derived CDNC based on MODIS retrievals of cloud properties by Bennartz (2007) (Zhang et al., 693 2015). For example, Bennartz (2007) derived the CDNC from cloud optical depths and cloud 694 effective radius assuming adiabatically-stratified clouds. Among all simulations with CB05, CB05\_SORG\_DH produces the lowest CDNC due to underestimated OA concentrations. 695 696 Increasing the OA concentrations helps to reduce the negative biases for CDNC. There are small 697 differences, however, among simulated CDNC with different VBS treatments for CB05 in CDNC 698 predictions, with similar Corr ~ 0.29, NMBs of ~-29% to -27% and NMEs of ~ 47%. Figure 9 699 shows the spatial differences in predictions in warm clouds between the several simulations and 700 the simulation CB05\_VBS\_DH. CB05\_SORGAM\_DH gives the lower CDNC than 701 CB05 VBS DH, indicating that the VBS treatment in CB05 VBS DH helps to increase CDNC 702 significantly. While other simulations with semivolatile POA treatments further increase domain 703 average CDNC comparing to CB05\_VBS\_DH, the differences between CDNC predictions from 704 those simulations and CB05\_VBS\_DH are quite similar. In general, CDNC with the semivolatile





721

705 POA cases are higher over western U.S. but lower over eastern U.S. due to decreases in column 706 OA concentrations for the semivolatile POA cases compared to CB05\_VBS\_DH over eastern U.S. 707 The large differences in CDNC predictions, however, are found between simulations with the 708 different gas-phase mechanisms. SAPRC07\_25%FF\_EM3 has the largest negative bias (NMB of 709 -52%) compared to all other simulations with CB05 and the simulation with CB6. Figure 10 710 compares the spatial plots for CDNC predictions for simulations with different gas-phase 711 mechanisms, as well as the surface spatial plots for total OA and inorganic PM<sub>2.5</sub> concentrations. 712 The simulation with SAPRC07 shows significantly lower CDNC over northeastern U.S. 713 comparing to CDNC predictions from the other two simulations. While all three simulations show 714 similar total OA concentrations, large differences are found for their total inorganic PM2.5 715 concentrations, with SAPRC07 producing the lowest domain mean and maximum total inorganic 716 PM<sub>2.5</sub> concentrations. Compared to CB05 and CB6, the lower inorganic PM<sub>2.5</sub> concentrations 717 simulated with SAPRC07 are likely due to the low OH+HO<sub>2</sub> mixing ratios for SAPRC07 as shown 718 in Figure 6, resulting in a lower PM number concentration and lower cloud condensation nuclei 719 (CCN), thus lower CDNC.

#### 720 4.4.Sensitivity Simulations for Aerosol Activation Parameterizations

Among all OA sensitivity simulations, the simulation CB05-25%FF-EM3 gives an overall best performance in terms of OC, TC, O<sub>3</sub>, PM<sub>2.5</sub>, and CDNC evaluation, it is thus selected to test various aerosol activation parameterizations. As listed in Table 4, four sensitivity simulations are designed to test the FN05 series aerosol activation parameterizations with improved treatments comparing to the default ARG00 aerosol activation parameterization. These sensitivity simulations include the default ARG00, the FN05, the combination of FN05 and BA10, and the MN14. These simulations use the MSKF scheme instead of the Morrison microphysics schemes in the previous





729 SOA runs as the MSKF scheme has a better correlation with MODIS CDNC as compared to the 730 Morrison microphysics scheme. Table 8 summarizes the model evaluation results against MODIS-731 derived CDNC from Bennartz (2007). The simulation ARG00 underpredicts CDNC with an NMB 732 of -35%. The FN05 series helps to reduce the underpredictions of CDNC significantly, because 733 they in general give higher activation fractions compared to the ARG00 parameterization under 734 most atmospheric conditions (Ghan et al., 2011). The addition of BA10 to the FN05 takes into 735 account the effects of condensation on giant CCN, which reduces the CDNC predictions and leads to a negligible underprediction of CDNC (with an NMB of -0.8%) compared to a slight 736 737 overprediction by the FN05 with an NMB of 7.1%. MN14, which revises the original population 738 splitting method in FN05 and BA10, slightly increases the CDNC to an NMB of 4.2% comparing 739 to the FN05/BA10 simulation. The trends in the predictions of FN05, BA10, and MN14 are 740 consistent with the reported bias of ~+8%, -10% and -3%, respectively, by Morales Betancourt 741 and Nenes (2014) against the CDNC concentrations simulated from the cloud parcel model. 742 However, the Corr and NME are worse with the FN05 series and MN14. The NMEs are almost doubled for the FN05 series and MN14, compared to that from the default ARG00. Figure 11 743 744 compares the spatial distributions of the simulated CDNC in warm clouds from ARG00, FN05 745 series, and MN14 and the MODIS-derived CDNC from Bennartz (2007). As shown in Figure 11, 746 the lower Corr and higher NMEs for the FN05 series as compared to ARG00 shown in Table 9, as 747 compared to ARG00, are due to the large overpredictions over northeastern U.S. but 748 underpredictions over other parts of the domain. The simulated CDNC from the default ARG00 749 case is similar to that from Bennartz (2007) over eastern U.S., the underpredictions are mainly 750 over western U.S. and over the ocean because of the known bias when large CCN are not present 751 (Morales Betancourt et al., 2014). The simulations with the FN05 series increase CDNC where





752 CCN is high, i.e., over the northeastern U.S., resulting in overpredictions in CDNC over
753 northeastern U.S., and does not help to improve CDNC predictions over other parts of the U.S. as
754 well as over the ocean.

755 Figure 12 compares the simulated CCN and AOD from the CB05\_25%FF\_EM3 + MN14 case 756 with those derived from the MODIS. The model largely underpredicts CCN, especially over the western part of the domain, which explains the large underprediction of CDNC also over the 757 758 western part of the domain. Condensation of the available water vapor occurs over CCN which are 759 concentrated over northeastern U.S., resulting in overpredictions of CDNC over northeastern U.S. 760 The lack of CCN over the ocean and the western part of the domain is related to the 761 underpredictions of AOD over the same areas. This indicates biases in number (and probably mass) 762 concentrations of column PM concentrations, especially over the ocean and western U.S. PM<sub>2.5</sub> 763 and PM<sub>10</sub> observational data are available over the surface and are both underpredicted, however, 764 there are no observational data for column concentrations of  $PM_{2.5}$  and  $PM_{10}$  for evaluation. 765 Improving the spatial distribution and magnitude of emissions for PM species and precursors for 766 the model layers at the surface and above the surface can help improve AOD and CCN predictions, 767 therefore CDNC predictions.

768 769

### 5. Summary and Conclusions

Current regional air quality models including WRF/Chem have large uncertainties in modeling OA and aerosol-cloud feedback mechanisms such as the aerosol activation process. Comparing to the traditional OA method, the VBS treatment helps to improve OA predictions by reducing the underpredictions of OA. By including a semivolatile POA treatment, using a semi-empirical formation of Epstein et al. (2010) for  $\Delta H_{vap}$ , including 25% fragmentation and functionalization as well as including additional S/IVOC emissions, the VBS treatment in WRF/Chem simulates the





776 atmospheric OA formation processes more realistically and can perform relatively well in predictions of OC and TC against IMPROVE and STN. POA/OA ratios for the 777 778 CB05 25%FF EM3 and CB05 FF50% treatments are within the range of POA/OA ratios of 779 ~0.15 to 0.40 from literature. Compared to the simulation with default SORGAM SOA module, 780 the simulations with various new VBS treatments also give better agreement with observed SOA 781 at Bakersfield and Pasadena during the CalNex field campaign from May to June 2010. However, 782 biases exist in those simulations with the VBS treatments due to several possible reasons, including 783 underestimated POA emissions, underpredicted VOC concentrations, as well as differences in the 784 SOA precursors used in the model and those contributing to the observed SOA concentrations. 785 The simulations with different gas-phase mechanisms (i.e., CB05, CB6, and SAPRC07) produce 786 in general different ASOA and BSOA concentrations. SAPRC07 produces the highest O<sub>3</sub> mixing 787 ratios, while CB6 produces the lowest  $OH + HO_2$  mixing ratios. CB6 also performs the best when 788 evaluated against IMPROVE PM2.5 while CB05 performs the best when evaluated against STN 789 PM2.5 concentrations. All 3 cases perform poorly against AQS PM10 evaluation. Due to the 790 significant differences between  $O_3$ , OH, and  $HO_2$  mixing ratios for the three gas-phase 791 mechanisms, inorganic PM concentrations vary widely, especially between the carbon bond 792 mechanisms (CB05 and CB6) and SAPRC07, resulting in significantly different predictions of 793 CDNC. The CDNC predictions do not vary much among simulations with CB05 and different 794 VBS treatments, for example, for simulations with nonvolatile vs. semivolatile POA, and with and 795 without fragmentation and functionalization treatments. The simulation with SAPRC07 produces 796 the lowest CDNC compared to those with CB05 and CB6, due to the lowest inorganic PM number 797 and mass concentrations resulted from the lowest OH and HO2 mixing ratios among all





- 798 simulations. CB05 gives the best performances when evaluated against CASTNET and AQS ozone
- 799 mixing ratios, STN PM2.5 concentrations and against MODIS CDNC.
- 800 With the default ARG00 treatment in the model, in general, all simulations with VBS 801 treatments underpredict the MODIS-derived CDNC by Bennartz (2007). By including the FN05 802 series (i.e., FN05, FN05/BA10, and MN14), the underpredictions for CDNC are greatly reduced. 803 However, the correlation coefficient and errors are worse with the FN05 series, with large 804 overpredictions over the northeastern U.S., where CCN is high. The model performs poorly for 805 AOD and CCN, likely due to inaccuracies in spatial distribution and magnitudes of PM and PM 806 precursor emissions in the model layers at the surface and above the surface. The CDNC 807 predictions can be improved by improving AOD and CCN undepredictions over western U.S. and 808 over the ocean.

#### 809 **Code and Data Availability**

810 The WRF/Chem v3.7.1 code used in this paper will be available upon request. The inputs 811 including the meteorological files, meteorological initial and boundary conditions, chemical initial 812 and boundary conditions, model setup and configuration, and the namelist setup and instructions 813 on how to run the simulations for a 1-day test case, as well as a sample output for a 1-day test, can 814 be provided upon request.

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Tables

Source	Ahmadov et al., 2012	Shrivastava et al., 2011	Farina et al., 2010	Jathar et al., 2011	Hodzic et al., 2010
Model	WRF-Chem	WRF-Chem	GISS II' GCM	GISS II' GCM	CHIMERE
Domain	CONUS	Mexico City	Global	Global	Mexico City
Spatial Resolution	20 km and 60 km	Nested 3 km within 12 km	$4^{\circ} \times 5^{\circ}$	$4^{\circ} \times 5^{\circ}$	$5 \text{ km} \times 5 \text{ km}$ and $35 \text{ km} \times 35 \text{ km}$
Emissions of SVOCs, IVOCs, and VOCs	Only VOCs; no emissions of SVOCs and IVOCs	SVOC emissions 3 times POA emissions for both anthropogenic and biomass burning emissions. IVOC emissions 1.5 times POA emissions	POA is treated as nonvolatile and nonreactive, but acts as absorbing phase for SOA condensation, forming 1 OA phase	SVOC emissions are represented by the traditional emission inventory. IVOC emissions are 1.5 times traditional emissions	SVOC emissions 3 times POA emissions. IVOC emissions 1.5 times POA emissions
No. of VBS bins	4	2 and 9	4	9	9
Aging	Yes and No Simulations with aging: each oxidation step produces 7.5% additional mass	Yes and No Simulations with aging: each oxidation step produces 15% additional mass	Yes	Yes and No. Each oxidation step does not produce any additional mass	Yes. 2 cases below: (i)Each oxidation step produces 7.5% additional mass (ii)Each oxidation step reduces the volatility by 2 orders of magnitude and 40% of additional mass produced
Observational data	SEARCH, STN, IMPROVE	MILAGRO 2006 field campaign	IMPROVE, EMEP	IMPROVE, FAME, MILAGRO, SOAR	MILAGRO 2006 field campaign
Variables evaluated	OC and EC	OA, HOA, OOA O:C ratio	OM:OC of 1.8	OA (surface), HOA, OOA	HOA, OOA, BBOA, O:C ratio
Summary of results with VBS framework with/without aging compared to the traditional SOA approach	-Improved diurnal variability -Results without the aging process underestimate OC throughout the day	-HOA and OOA: Lower negative bias with addition of S/IVOC emissions -OOA: 2 bin VBS better results than 9 bin VBS -Underprediction of O:C ratio in both cases	-IMPROVE: improved with aging -EMEP: aging further biases already high OA predictions	-Adding IVOC emissions improves performance, however underprediction remains in winter months	-HOA overpredicted during nightime -Case(i):Modeled O:C 3 times lower than observed -Case(ii):Better agreement for O:C but SOA generally overestimated

Table 1. Summary of several literature studies of VBS treatments in various regional and global models.

Note: HOA: Hydrocarbon-like OA – Reduced specie of OA, generally understood as a surrogate for urban combustion-related POA; OOA: Oxygenated OA – Characterized by its high oxygen content and generally understood as a surrogate for SOA; BBOA: Biomass burning OA





Table 1. (cont). SSummary of several literature studies of VBS treatments in various regional and global models.

Source	Bergstrom et al., 2012	Lane et al., 2008	Donahue et al., 2009	Murphy et al., 2009	This work
Model	EMEP	PMCAMx	PMCAMx	PMCAMx	WRF/Chem
Domain	Europe, a large part of the North Atlantic and Arctic areas	Eastern U.S.	Eastern U.S.	Eastern U.S.	CONUS with parts of Canada and Mexico
Spatial Resolution	$50 \text{ km} \times 50 \text{ km}$	36 km × 36 km	NA	$36 \text{ km} \times 36 \text{ km}$	36 km × 36 km
Emissions of SVOCs,	VOCs are present.	Only VOCs; SVOCs	Additional IVOCs	IVOC emissions are 0.2	S/IVOCs are 1.5 to 3
IVOCs and VOCs	S/IVOCs are 2.5 times the POA emissions	and IVOCs not added	added but details are not given	to 0.8 times the nonvolatile POA emission rates	times the nonvolatile POA NEI emissions
No. of VBS bins	4 for SOA components and 9 for POA	4	9	10	4 for SOA components and 9 for POA
Aging	Yes and No. Each oxidation step produces 7.5% additional mass	Yes. No additional mass produced for each oxidation step.	Yes. No additional mass produced for each oxidation step.	Yes. No additional mass produced for each oxidation step.	Yes. Each oxidation step produces 7.5% additional mass
Observational data	CARBOSOL, SORGA, Gote-2005	STN, IMPROVE	NA	STN, IMPROVE	STN, IMPROVE, field data
Variables evaluated	TC, OC	OA	POA, OPOA, SOA	OA	TC, OC, POA/OA
Summary of results with VBS framework with/without aging compared to the traditional SOA approach	-Addition of aging reactions improve summertime results but has little or negative consequences in wintertime -Deteriorations of model results with	-Addition of aging reactions overpredicts the OA concentrations in rural IMPROVE stations but improves the model performance in urban areas	-Aging results in better model predictions	-Slight overprediction with IMPROVE -Underprediction with STN	-Large improvements in predictions
	increased aging at urban influenced sites in southern Europe				

Note: TC: total carbon; OC: Organic carbon; OPOA: oxidized POA





	CB05-Cl <sup>1</sup>	CB6	SAPRC07 <sup>2</sup>
No. of species	70	114	118
No. of reactions	156	218	599
Lumping method	Lumped structure based on carbon bonds	Lumped structure based on carbon bonds	Lumped species based on their reactivity towards hydroxyl (OH)
Kinetic Data for rate constants	Mostly from IUPAC (Atkinson et al., 2005). NASA/JPL (Sander et al., 2003) values were used in some cases where IUPAC data was not available.	New information from IUPAC (Atkinson et al., 2010) and NASA (Sander et al., 2006)	Mainly from IUPAC (2006) and NASA (Sander et al., 2006).
Photolysis data	Mainly from SAPRC99 chemical mechanism. IUPAC (Atkinson et al., 2005) was used if it differs significantly from SAPRC99.	New information from IUPAC (Atkinson et al., 2010) and NASA (Sander et al., 2006)	Mainly from IUPAC (2006) and NASA (Sander et al., 2006).
Ozone chemistry	Slightly underpredict O <sub>3</sub> mixing ratios with isoprene and in synthetic urban mixtures in chamber experiments.	Reduced underprediction in $O_3$ mixing ratios from benzene, toluene, and xylene, but forms $O_3$ from isoprene too slowly compared to CB05.	Slightly underpredict O <sub>3</sub> mixing ratios at low NO <sub>x</sub> levels in chamber experiments.
Organic nitrate	2 reactions involving organic nitrate (NTR).	Additional NO <sub>x</sub> recycling from organic nitrate to represent fate of NO <sub>x</sub> over multiple days.	Added peroxy+NO reactions to form organic nitrate.
Chlorine chemistry	20 additional reactions for Cl chemistry involving species Cl <sub>2</sub> , HOCl, Cl, ClO, and FMCl.	CB05 chlorine chemistry included in this work.	22 base chlorine reactions involving CL <sub>2</sub> , CLNO, CLONO, CLNO <sub>2</sub> , CLONO <sub>2</sub> , HOCl, and 26 additional reactions involving organic products
Organic chemistry	<ul> <li>Explicit organic aerosol precursors, e.g., isoprene, toluene, xylene, α-pinene, β- pinene.</li> </ul>	<ul> <li>Explicit long-lived and abundant organic compounds including propane, acetone, benzene and acetylene added</li> <li>Extensive revision of isoprene and aromatics chemistry</li> <li>Formation of alpha-dicarbonyl compounds (glyoxal, glycoaldehyde, methylglyoxal)</li> <li>Updates to peroxy radical chemistry that will improve formation of H<sub>2</sub>O<sub>2</sub></li> </ul>	<ul> <li>Reformulated reactions of peroxy radicals so that effects of changes in NO<sub>x</sub> conditions on organic product formation is more accurately represented</li> <li>Most comprehensive representation of VOCs compared to other gas-phase mechanisms</li> </ul>
3-D host models	Implemented into WRF/Chem v3.6.1 by Wang et al. (2014). Also available in WRF/Chem v3.7.1	Implemented in CAMx by ENVIRON (2013)	Implemented in CMAQ (Carter, 2010)
Reference	Yarwood et al. (2005)	Yarwood et al. (2010) ENVIRON (2013)	Carter (2010)

Table 2. Summary of main characteristics of CB05, CB6, and SAPRC07 gas-phase mechanisms.

<sup>1</sup>CB05 gas-phase mechanism with reactive chlorine chemistry (Yarwood et al., 2005)

<sup>2</sup> SAPRC07 uncondensed and expanded version C, which includes reactions for peroxy radical operators (Carter, 2010).





Table 3. Factors to calculate S/IVOC emissions from POA emissions from Shrivastava et al. (2011), May et al. (2013a, c) and newly calculated factors for this study.

	et al. (2013a, c) and newly calculated factors for this study.								
Log Ci* at	Normalized	Fraction for	Fraction for	New calculated					
298K	fraction for	gasoline	biomass	fraction for all					
	stationary	emissions from	burning	sources based					
	emissions based	May et al.	emissions from	on Shrivastava					
	on	(2013a)	May et al.	et al. (2011),					
	anthropogenic		(2013c)	May et al.					
	emissions from			(2013a, c), and					
	Shrivastava et			% distribution					
	al. (2011)			of NEI					
				emissions					
-2	0.04	0.14	0.2	0.1754					
-1	0.02	0.13	0.0	0.0141					
0	0.03	0.15	0.1	0.0961					
1	0.05	0.26	0.1	0.1084					
2	0.07	0.15	0.2	0.1799					
3	0.11	0.03	0.1	0.0949					
4	0.16	0.02	0.3	0.258					
5	0.20	0.01	0.0	0.0249					
6	0.32	0.11	0.0	0.0483					





Name	Gas-	∆H <sub>vap</sub>	VBS	FF	POA	Aerosol	Cumulus
	Phase	up			emissions	activation	Scheme
CB05-	CB05	30 kJ mol <sup>-1</sup>	-	-	Original	ARG00	Grell-Freitas
SORG-DH					NEI		
CB05-VBS-	CB05	30 kJ mol <sup>-1</sup>	SOA	-	Original	ARG00	Grell-Freitas
DH					NEI		
CB05-POA-	CB05	30 kJ mol <sup>-1</sup>	SOA/	-	1.5×	ARG00	Grell-Freitas
DH			POA				
CB05-POA	CB05	Epstein et	SOA/	-	1.5×	ARG00	Grell-Freitas
		al. (2010)	POA				
CB05-	CB05	Epstein et	SOA/	50%	1.5×	ARG00	Grell-Freitas
50%FF		al. (2010)	POA				
CB05-	CB05	Epstein et	SOA/	10%	1.5×	ARG00	Grell-Freitas
10%FF		al. (2010)	POA				
CB05-	CB05	Epstein et	SOA/	25%	1.5×	ARG00	Grell-Freitas
25%FF		al. (2010)	POA				
CB05-	CB05	Epstein et	SOA/	25%	3.0×	ARG00	Grell-Freitas
25%FF-EM3		al. (2010)	POA				
CB6-25%FF-	CB6	Epstein et	SOA/	25%	3.0×	ARG00	Grell-Freitas
EM3		al. (2010)	POA				
SAPRC07-	SAPRC07	Epstein et	SOA/	25%	3.0×	ARG00	Grell-Freitas
25%FF-EM3		al. (2010)	POA				
CB05-	CB05	Epstein et	SOA/	25%	3.0×	FN05	MSKF
25%FF-EM3		al. (2010)	POA				
(FN05)							
CB05-	CB05	Epstein et	SOA/	25%	3.0×	FN05/	MSKF
25%FF-EM3		al. (2010)	POA			BA10	
(FN05/							
BA10)							
CB05-	CB05	Epstein et	SOA/	25%	3.0×	MN14	MSKF
25%FF-EM3		al. (2010)	POA				
(MN14)							

Table 4. Configurations for OA and aerosol activation sensitivity simulations. All simulations except for CB05-SORG-DH contain the VBS treatments for OA.

The suffix "DH" in the case names refer to cases with the default  $\Delta H_{vap}$  of 30 kJ mol<sup>-1</sup>, otherwise with the semi-empirical correlation by Epstein et al. (2010). The simulations without the suffix "POA" indicate cases with nonvolatile default POA emissions. The suffix "POA" in the case names refer to cases with semivolatile POA. The suffix "FF" in the case names refer to cases with semivolatile POA and with fragmentation and functionalization treatments, and the suffix "EM3" in the case names refer to cases with 3 times the original NEI POA emissions to take into account for missing S/IVOC species. "-" indicates not applicable.





Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)
	0	C against IMPI	ROVE		
CB05-SORG-DH	0.88	0.28/ 0.19	0.26	-68.1/ -78.7	73.9/ 80.9
CB05-VBS-DH	0.88	1.19/ 0.79	0.51	34.9/ -10.1	75.5/ 52.3
CB05-POA-DH	0.88	0.89/ 0.59	0.51	1.1/-32.6	52.4/ 59.0
CB05-POA	0.88	1.05/ 0.70	0.51	18.9/ -20.7	63.2/49.2
CB05-10%FF	0.88	1.05/ 0.70	0.51	19.4/ -20.4	63.0/49.1
CB05-25%FF	0.88	0.86/ 0.57	0.49	-2.9/ -35.2	54.6/51.4
CB05-50%FF	0.88	0.56/ 0.37	0.45	-36.4/ -57.6	54.4/ 62.6
CB05-25%FF-EM3	0.88	1.09/ 0.73	0.47	23.8/ -17.5	65.9/ 50.2
CB6-25%FF-EM3	0.88	1.06/ 0.71	0.48	20.5/ -19.6	49.4/ 63.7
SAPRC07-25%FF-EM3	0.88	1.00/ 0.67	0.46	13.3/ -24.4	60.1/ 50.4
	T	C against IMPI	ROVE	•	
CB05-SORG-DH	1.03	0.44/ 0.34	0.30	-57.6/ -66.7	67.9/72.3
CB05-VBS-DH	1.03	1.34/ 0.94	0.52	30.6/ -8.0	70.3/ 51.1
CB05-POA-DH	1.03	1.13/ 0.83	0.52	10.2/ -18.7	58.5/ 48.7
CB05-POA	1.03	1.29/ 0.94	0.53	25.6/ -8.5	63.8/48.3
CB05-10%FF	1.03	1.29/ 0.94	0.53	25.9/ -8.2	63.8/48.2
CB05-25%FF	1.03	1.09/ 0.83	0.51	6.8/ -21.6	55.2/48.2
CB05-50%FF	1.03	0.80/ 0.61	0.47	-22.0/ -40.2	50.8/ 53.4
CB05-25%FF-EM3	1.03	1.32/ 0.97	0.49	29.7/ -5.7	50.7/ 66.9
CB6-25%FF-EM3	1.03	1.30/ 0.95	0.50	27.2/ -7.3	65.2/ 50.0
SAPRC07-25%FF-EM3	1.03	1.23/ 0.90	0.48	20.6/-11.9	61.4/49.4
	•	TC against S	ΓN	-	
CB05-SORG-DH	2.71	1.34/ 1.10	0.29	-50.6/ -59.4	60.1/64.9
CB05-VBS-DH	2.71	3.35/ 2.44	0.47	23.7/ -5.8	53.1/42.0
CB05-POA-DH	2.71	2.88/ 2.19	0.47	6.2/ -19.0	45.5/ 41.6
CB05-POA	2.71	3.03/ 2.30	0.46	11.7/ -15.3	44.6/39.9
CB05-10%FF	2.71	3.03/ 2.30	0.46	11.8/ -15.3	44.5/ 39.8
CB05-25%-FF	2.71	2.66/ 2.05	0.44	-1.8/ -24.3	41.5/ 42.0
CB05-50%-FF	2.71	2.07/ 1.65	0.39	-23.8/ -39.1	43.9/49.4
CB05-25%FF-EM3	2.71	3.27/ 2.45	0.41	20.5/ -9.5	49.7/41.7
CB6-25%FF-EM3	2.71	3.39/ 2.45	0.34	24.9/-6.4	54.8/45.5
SAPRC07-25%FF-EM3	2.71	3.00/ 2.28	0.41	10.7/ -16.1	45.2/ 42.0

Table 5. Range of statistics for OA/OC ratios of 1.4 and 2.1 (1.4/2.1) for May to June 2010. All simulations use the ARG00 aerosol activation scheme and the Grell-Freitas cumulus parameterization.





Table 6. Statistics for evaluation at Bakersfield and Pasadena sites. A bar chart of daily average obs vs. sim values can be found in Figure 4.

Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
Bakersfield								
CB05-SORG-DH	0.51	5.9e-04	-0.15	-100	100%			
CB05-VBS-DH	0.51	0.67	0.41	32.5	62.0			
CB05-25%FF-EM3	0.51	0.24	-0.01	-52.0	61.0			
CB6-25%FF-EM3	0.51	0.28	-0.04	-45.8	59.0			
SAPRC07-25%FF-EM3	0.51	0.24	-0.16	-53.1	63.0			
		Pasadena						
CB05-SORG-DH	0.63	0.04	-0.07	-94.0	94.0			
CB05-VBS-DH	0.63	0.54	0.09	-14.5	64.3			
CB05-25%FF-EM3	0.63	0.54	-0.2	-14.4	66.2			
CB6-25%FF-EM3	0.63	0.62	-0.2	-2.1	70.0			
SAPRC07-25%FF-EM3	0.63	0.62	0.03	-1.4	70.5			





	CAS	TNET Max 1-	h O <sub>3</sub>					
Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
CB05-25%FF-EM3	51.8	43.3	0.54	-16.3	21.9			
CB6-25%FF-EM3	51.8	41.9	0.52	-19.1	24.1			
SAPRC07-25%FF-EM3	51.8	48.3	0.51	-6.7	21.1			
	CAS	TNET Max 8-	h O <sub>3</sub>					
Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
CB05-25%FF-EM3	47.4	43.0	0.54	-9.3	18.9			
CB6-25%FF-EM3	47.4	41.8	0.53	-11.8	20.6			
SAPRC07-25%FF-EM3	47.4	47.9	0.50	1.0	19.8			
	Α	QS Max 1-h C	)3					
Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
CB05-25%FF-EM3	51.0	49.9	0.55	-2.1	18.2			
CB6-25%FF-EM3	51.0	51.5	0.43	1.0	20.8			
SAPRC07-25%FF-EM3	51.0	59.3	0.44	16.4	26.1			
	Α	QS Max 8-h C	)3		•			
Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
CB05-25%FF-EM3	46.2	46.0	0.54	-0.4	18.6			
CB6-25%FF-EM3	46.2	47.4	0.47	2.6	20.3			
SAPRC07-25%FF-EM3	46.2	53.7	0.46	16.3	25.4			
	IN	<b>IPROVE PM</b>	2.5					
Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
CB05-25%FF-EM3	4.9	3.8	0.64	-22.0	40.6			
CB6-25%FF-EM3	4.9	4.1	0.65	-16.5	39.6			
SAPRC07-25%FF-EM3	4.9	3.5	0.60	-28.5	42.9			
		STN PM <sub>2.5</sub>						
Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)			
CB05-25%FF-EM3	11.1	8.8	0.48	-20.6	40.7			
CB6-25%FF-EM3	11.1	10.0	0.37	-9.3	44.3			
SAPRC07-25%FF-EM3	11.1	7.7	0.40	-30.5	45.2			
	AQS PM <sub>10</sub>							
Case	Mean Obs	Mean Sim	Corr	<b>NMB</b> (%)	NME (%)			
CB05-25%FF-EM3	24.6	7.3	0.08	-70.2	73.5			
CB6-25%FF-EM3	24.6	8.0	0.09	-67.7	71.8			
SAPRC07-25%FF-EM3	24.6	6.9	0.09	-71.9	74.8			

Table 7. Statistics for max 1-h and max 8-h  $O_3$  for simulations with different gas-phases against CASTNET and AQS for May to June 2010.





Table 8. Statistics for model evaluation for simulated CDNC against MODIS-derived CDNC from Bennartz (2007).

Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)
CB05-SORG-DH	162.1	96.0	0.28	-40.8	50.4
CB05-VBS-DH	162.1	106.0	0.28	-34.6	50.6
CB05-POA-DH	162.1	115.0	0.29	-29.1	47.4
CB05-POA	162.1	117.3	0.29	-27.7	47.3
CB05-10%FF	162.1	117.1	0.29	-27.8	47.2
CB05-25%-FF	162.1	116.4	0.29	-28.2	47.3
CB05-50%-FF	162.1	114.7	0.29	-29.2	47.4
CB05-25%FF-EM3	162.1	116.2	0.29	-28.3	47.3
CB6-25%FF-EM3	162.1	110.4	0.30	-31.9	47.3
SAPRC07-25%FF-	162.1	77.3	0.26	-52.3	55.8
EM3					

Table 9. Statistics for simulated CDNC for CB05-25%FF-EM3 against MODIS-derived CDNC from Bennartz (2007) for May to June 2010.

Case	Mean Obs	Mean Sim	Corr	NMB (%)	NME (%)
ARG00	162.1	104.8	0.31	-35.4	49.9
FN05	162.1	173.8	0.26	7.1	93.0
FN05/BA10	162.1	160.8	0.27	-0.8	87.9
MN14	162.1	168.9	0.27	4.2	89.6





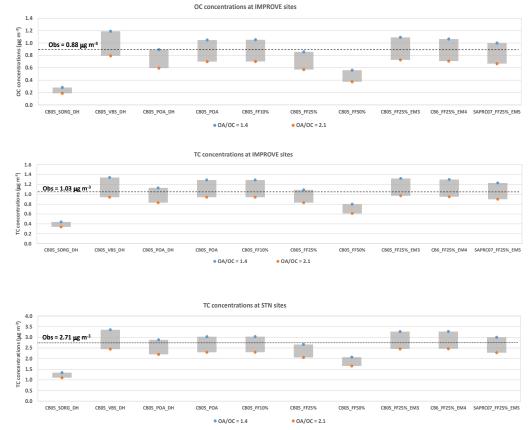


Figure 1. Sim OC and TC concentrations against observations from IMPROVE and STN under two A/OC ratios: 1.4 and 2.1, , resulting in a range of possible OC or TC values denoted by the gray bars. The obs OC or TC is denoted by the horizontal dotted line.





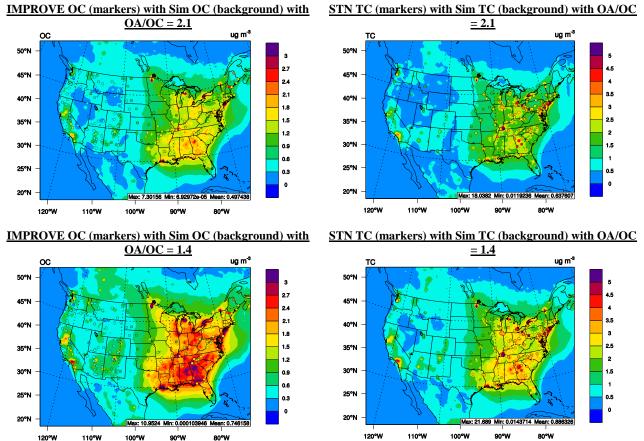
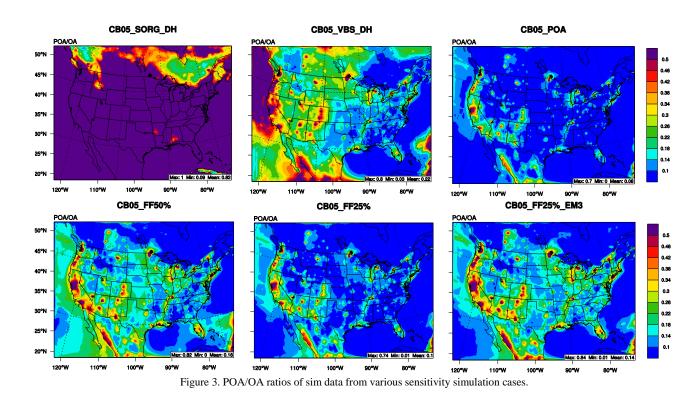


Figure 2. Overlay of obs data (markers) vs. sim data (background) for IMPROVE OC and STN TC and for OA/OC ratios of 1.4 and 2.1 for the case CB05\_25%FF\_EM3.











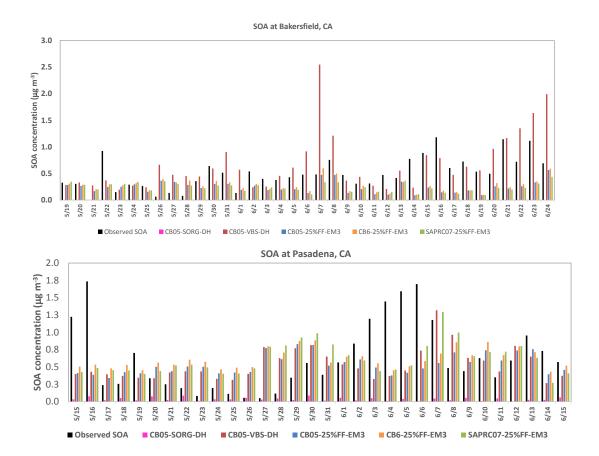


Figure 4. Comparison of obs SOA vs. sim SOA at CalNex sites in Bakersfield and Pasadena in California.





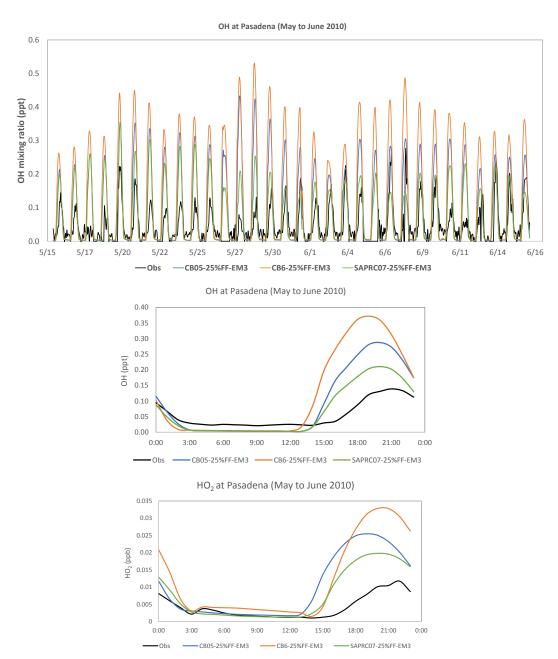


Figure 5. Time series of OH and diurnal plots of OH and HO<sub>2</sub> at Pasadena, CA during CALNEX, 2010.





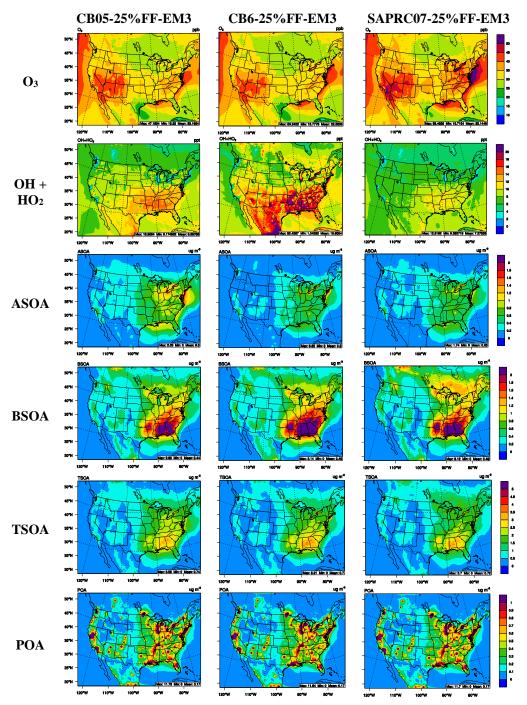


Figure 6. Spatial plots of several gas and aerosol species for the three cases with different gas-phase mechanisms.





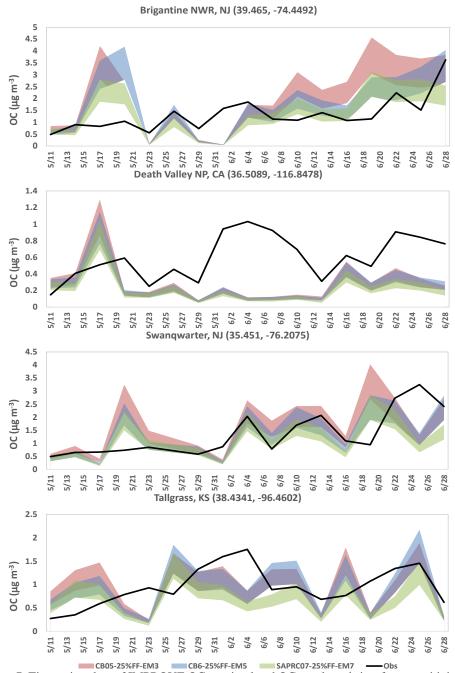


Figure 7. Timeseries plots of IMPROVE OC vs. simulated OC at selected sites from sensitivity simulations of different gas-phase mechanisms. The colored bands represent the range of OC values for ratios 1.4 to 2.1.





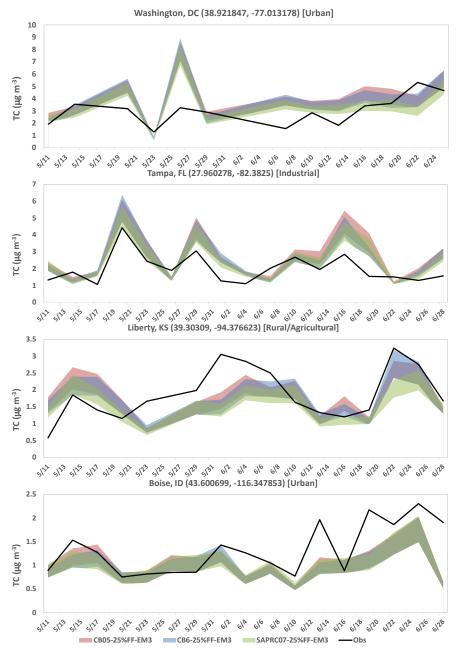


Figure 8. Timeseries plots of STN TC vs. simulated TC at selected sites from sensitivity simulations of different gas-phase mechanisms. The colored bands represent the range of OC values for ratios 1.4 to 2.1.





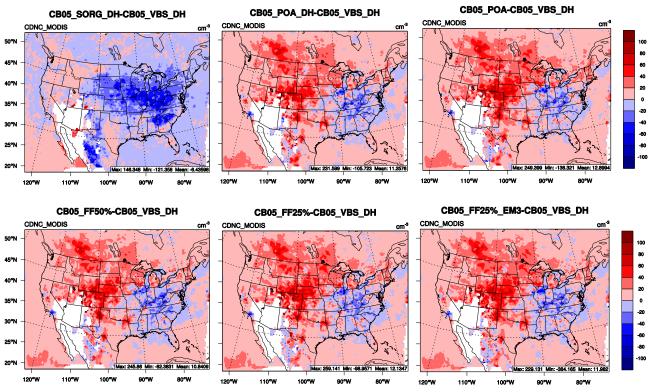


Figure 9. Impact of different VBS case on CDNC in warm clouds. The plots show the differences between the different sensitivity simulations and CB05\_VBS-DH.





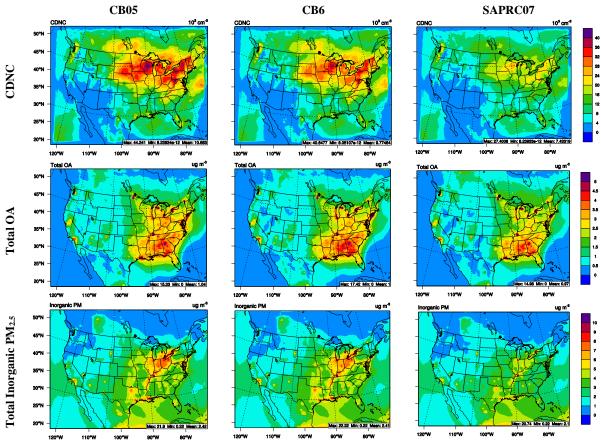


Figure 10. Spatial plots of CDNC, total surface OA and total inorganic PM<sub>2.5</sub> concentrations from different gas-phase mechanisms.







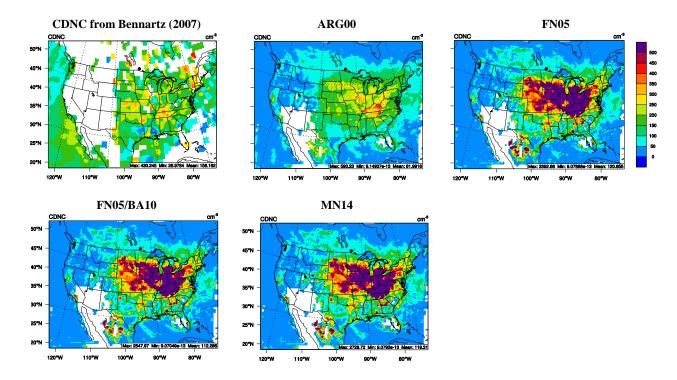


Figure 11. Spatial plots for MODIS-derived CDNC from Bennartz (2007) and simulated CDNC from CB05\_25%FF\_EM3 ARG00, FN series, and MN14 from May to June 2010.





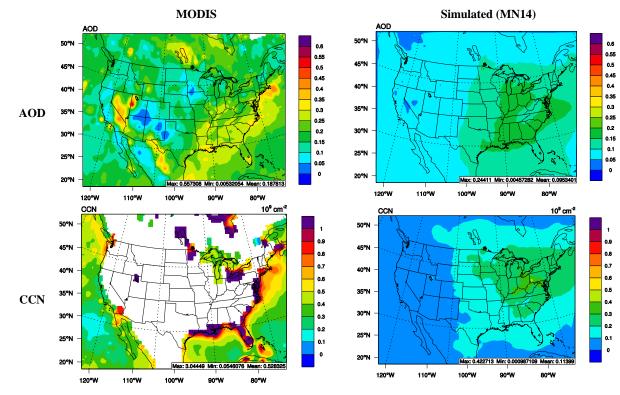


Figure 12. Plots of MODIS AOD and CCN against simulated AOD and CCN from MN14 with CB05\_25%FF\_EM3.