

1 Consistent assimilation of multiple data streams in a 2 carbon cycle data assimilation system

3
4 **Natasha MacBean¹, Philippe Peylin¹, Frédéric Chevallier¹, Marko Scholze²,**
5 **Gregor Schürmann³**

6 [1]{Laboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, CEA-CNRS-
7 UVSQ, Université Paris-Saclay, F-91191 Gif-sur-Yvette, France}

8 [2]{Department of Physical Geography and Ecosystem Science, Lund University, Lund,
9 Sweden}

10 [3]{Max Planck Institute for Biogeochemistry, Jena, Germany}

11 Correspondence to: N. MacBean (nlmacbean@gmail.com)

13 **Abstract**

14 Data assimilation methods provide a rigorous statistical framework for constraining
15 the parametric uncertainty of land surface models (LSMs), with the aim of improving our
16 predictive capability as well as identifying areas in which the models need improvement. The
17 increase in the number of available datasets in recent years allows us to address different
18 aspects of the model at a variety of spatial and temporal scales. However, combining data
19 streams in a DA system is not a trivial task. In this study we highlight some of the challenges
20 surrounding multiple data stream assimilation for the carbon cycle component of LSMs. We
21 give particular consideration to the assumptions associated to the type of inversion algorithm
22 typically used for optimising the parameters of global LSMs – namely, Gaussian error
23 distributions and linearity in the model dynamics. We explore the effect of biases and
24 inconsistencies between the observations and the model (resulting in non Gaussian error
25 distributions), and we examine the difference between performing a simultaneous assimilation
26 (in which all data streams are included in one optimisation) and a step-wise approach (in
27 which each data stream is assimilated sequentially) in the presence of non-linear model
28 dynamics. In addition, we perform a preliminary investigation into the impact of correlated
29 errors between two data streams for two cases, both when the correlated observation errors are

1 included in the prior observation error covariance matrix, and when the correlated errors are
2 ignored. We demonstrate some of these issues by assimilating synthetic observations into two
3 simple models: the first a simplified version of the carbon cycle processes represented in
4 many LSMs, and the second a non-linear toy model. Finally we provide some perspectives
5 and advice to other land surface modellers wishing to use multiple data streams to constrain
6 their models.

7 Keywords: data assimilation, carbon cycle, biogeochemical cycles, land surface model.

8

9 **1 Introduction**

10 The carbon cycle is an important component of the Earth system, especially when
11 considering the climatic impact of rising greenhouse gas concentrations from fossil fuel
12 emissions and land use change. It is estimated that the oceans and land surface absorb
13 approximately half of the CO₂ emissions due to anthropogenic activity, but uncertainties
14 remain in the strength and location of sources and sinks, as well as in predictions of future
15 trends (Ciais et al., 2013). Observations allow us to understand the system up until the present
16 day and provide inference about how ecosystems may respond to future change. However,
17 their use in estimating model state variables and boundary conditions is limited beyond
18 diagnostic purposes, and they can be restricted in their spatial coverage. They also do not
19 contain all the information we may need to distinguish between the complex interactions that
20 may occur between many different processes. Incorporating our current knowledge of
21 physical mechanisms of biogeochemical cycles, including carbon, C, dynamics, into land
22 surface models (LSMs) represents a promising approach to analyse these interacting effects,
23 to upscale observations to larger regions, and to make future predictions. However, the
24 models can be limited by the lack of process representation, either due to gaps in our
25 knowledge, or in our technical and computing capability. As a result, model evaluations
26 reveal that not all variables are well-captured by the model under current conditions (Anav et
27 al., 2013), and the spread between model projections is still very large (Sitch et al., 2015).

28 Aside from model structural and forcing errors, one source of uncertainty is related to the
29 parameter (i.e. fixed) values of a model. Model-data fusion, or data assimilation (DA), allows
30 the calibration, or optimisation, of these values by minimising a cost function (that quantifies
31 the model-data misfit) while accounting for the uncertainties inherent in both the model and
32 data in a statistically rigorous framework. The C cycle component of most LSMs is complex

1 and contains a large number of parameters; luckily however, there are an increasing number
2 of in-situ and remote sensing-based data streams that can be used for parameter optimisation.
3 These data bring information on different spatial and temporal scales, such as:

- 4 • Atmospheric CO₂ concentration data measured at surface stations at continental to
5 global scales, which provide information from synoptic timescales to inter-annual
6 variability (IAV) and long-term trends.
- 7 • Eddy covariance net CO₂ (net ecosystem exchange – NEE) and latent (LE) and
8 sensible heat fluxes measured at half-hourly intervals at many sites across different
9 ecosystems/regions, providing information at seasonal to inter-annual timescales.
- 10 • Satellite-derived measures of vegetation dynamics, including “greenness” indices (i.e.
11 the Normalised Difference Vegetation Index – NDVI), fraction of absorbed
12 photosynthetically active radiation (FAPAR) and leaf area index (LAI) at global scales
13 and at daily time steps spanning more than a decade, thus capturing IAV and long-
14 term trends (though usually with a trade-off between spatial and temporal resolution).
- 15 • Satellite-derived measurements of soil moisture and land surface temperature at the
16 same temporal and spatial scales as the satellite-derived observations of vegetation
17 productivity.
- 18 • Aboveground biomass measurements are currently taken at only one or a few points in
19 time at plot scale up to regional scale from aircraft and satellite data, or are estimated
20 from allometric relationships at each site.
- 21 • Soil C stock estimates usually are only taken at one point in time at plot scale.
- 22 • Ancillary data on vegetation characteristics such as tree height or budburst – only
23 measured at certain well-instrumented sites.

24
25 Increasingly, researchers are attempting to bring these sources of data together to
26 constrain different parts of a model at different spatio-temporal scales within a multiple data
27 stream assimilation framework (e.g. Richardson et al., 2010; Keenan et al., 2012; Kaminski et
28 al., 2012; Forkel et al., 2014; Bacour et al., 2015). However, whilst the potential benefit of
29 adding in extra data streams to constrain the C cycle of LSMs is clear, multiple data stream
30 assimilation is not as simple as it may seem. This is particularly true when considering a

1 regional-to-global scale, multiple data stream, multiple site optimisation of a complex LSM
2 that contains many parameters, and which typically takes on the order of minutes to an hour
3 to run a one year simulation. When using more than one data stream there is the option to
4 include all data streams together in the same optimisation (simultaneous approach), or to take
5 a sequential (step-wise) approach. Mathematically, the optimal approach is the simultaneous,
6 but computational constraints related to the inversion of large matrices or the requirement of
7 numerous simulations, especially for global datasets (e.g. Peylin et al., 2016), and/or the
8 weight of different data streams in the optimisation (e.g. Wutzler and Carvalhais, 2014), may
9 complicate a simultaneous optimisation. On the other hand, in a step-wise assimilation the
10 parameter error covariance matrix has to be propagated at each step, which implies that it can
11 be computed. If the parameter error covariance matrix can be properly estimated and is
12 propagated between each step, the step-wise approach should be mathematically equal to
13 simultaneous. However, many inversion algorithms (e.g. derivative-based methods that use
14 the gradient of the cost function to find its minimum) require assumptions of model (quasi-)
15 linearity and Gaussian parameter and observation error distributions (Tarantola, 1987, p195).
16 If these assumptions are violated, or the error distributions are poorly defined, it is likely that
17 the step-wise will not be equal to the simultaneous, because information will be lost at each
18 step due to an incorrect calculation of the posterior error covariance matrix at the end of the
19 first step. An incorrect description of the observation (– model) error distribution could result
20 from i) the wrong assumption about the distribution of the residuals between the observation
21 and the model, ii) a poor characterisation of the error correlations, iii) an incompatibility
22 between the model and the data (possibly due to a model structural issue or differences in how
23 a variable is characterised), or iv) a bias in the observations that is not unaccounted for (i.e. is
24 treated as a random error). As mentioned, whilst a simultaneous optimisation is
25 mathematically more rigorous in the sense that the error correlations are treated within the
26 same inversion, if the prior distributions are not properly characterised any bias may be
27 aliased to the wrong parameters (Wutzler and Carvalhais, 2014), more so than in a step-wise
28 approach.

29 This tutorial-style paper highlights some of the challenges of multiple data stream
30 optimisation of carbon cycle models discussed above. Note that we do not aim to explore all
31 possible issues related to a DA system, for example the choice of the cost function,
32 minimization algorithm, or the characterization of the prior error distributions; indeed
33 previous studies have investigated such aspects at length (e.g. Fox et al., 2009; Trudinger et

1 al., 2007), therefore we refer the reader to these papers for more information. Section 2
2 reviews recent carbon cycle multiple data stream assimilation studies with reference to some
3 of the aforementioned challenges. Section 3 demonstrates the issues related to multiple data
4 stream assimilation with synthetic experiments designed around two simple models: one a
5 simplified version of the carbon dynamics included in many LSMs, and the other a “toy”
6 model designed to demonstrate the issues that arise with complex, non-linear models.. Finally
7 Section 4 provides some advice to land surface modellers wishing to carry out multiple data
8 stream assimilation.

9 10 11 **2 Review of existing multiple data stream carbon cycle data assimilation** 12 **studies**

13 **2.1 Extra constraint from multiple data streams**

14 Most site-based carbon cycle data assimilation studies have used eddy covariance
15 measurements of NEE and LE fluxes to constrain the relevant parameters of ecosystem
16 models. However, a few studies have also made use of chamber flux soil respiration data and
17 field measurements of vegetation characteristics (e.g. tree height, budburst, LAI) or estimates
18 of litterfall and carbon stocks as ancillary information (e.g. Fox et al., 2009; Keenan et al.,
19 2012; Thum et al., in review; Van Oijen et al., 2005; Richardson et al., 2010; Williams et al.,
20 2005). Two recent studies combined high-resolution satellite-derived FAPAR data and in-situ
21 eddy covariance measurements to optimize parameters related to carbon, water and energy
22 cycles of the ORCHIDEE and BETHY LSMs at a couple of sites (Bacour et al., 2015; Kato et
23 al., 2013, respectively).

24 At global scales the number of studies that use multiple data streams from satellites or
25 large-scale networks to optimise LSMs has been increasing in recent years, although this
26 remains a relatively new area of research. CCDAS-BETHY was the first global carbon cycle
27 data assimilation system (CCDAS) making use of the high-precision measurements of the
28 atmospheric CO₂ concentration flask sampling network (Rayner et al., 2005; Scholze, 2003)
29 to constrain process parameters of the prognostic terrestrial carbon cycle model BETHY
30 (Knorr, 2000). Since its first application in which only atmospheric CO₂ concentration data
31 were assimilated, CCDAS-BETHY has been further developed to consistently assimilate

1 multiple data streams both at local and global scales. In particular, Kaminski et al. (2012)
2 optimised 70 process parameters plus one initial condition by simultaneously assimilating a
3 satellite-derived FAPAR product derived from the Medium Resolution Imaging Spectrometer
4 (MERIS; Gobron et al., 2008) and flask samples of atmospheric CO₂ at two sites from the
5 GLOBALVIEW product (GLOBALVIEW-CO₂, 2008) at coarse resolution. More recently,
6 Scholze et al. (2016) demonstrated the added value of assimilating remotely sensed soil
7 moisture data in addition to observations of atmospheric CO₂ concentration from the flask-
8 sampling network. They used the same coarse resolution set-up of CCDAS as Kaminski et al.
9 (2012) and CO₂ observations from 10 sites of the GLOBALVIEW product (GLOBALVIEW-
10 CO₂, 2012) together with the SMOS L3 daily soil moisture product (version 246; CATDS-
11 L3, 2012).

12 Two other global CCDAS based on different LSMs have been developed in recent years
13 (Peylin et al., 2016; Schürmann et al., 2016). Schürmann et al. (2016) optimized model
14 parameters and initial conditions of the land component JSBACH (Raddatz et al. 2007) of the
15 MPI Earth System Model (ESM) (Giorgetta et al. 2013) using atmospheric CO₂ concentration
16 data and the TIP-FAPAR product (Pinty et al., 2007) as joint constraints over a 5 year period,
17 in addition to evaluating the mutual benefit of each data stream in a fully factorial design.
18 Peylin et al. (2016) used three different data streams as global constraints for the ORCHIDEE
19 LSM (Krinner et al., 2005), which forms the land surface component of the IPSL ESM
20 (Dufresne et al., 2013), in a multi-site step-wise assimilation approach. First, satellite-derived
21 vegetation index data (NDVI) from the MODIS instrument were used to constrain the
22 phenology parameters at 60 sites for the temperate and boreal deciduous PFTs, followed by
23 NEE and LE observations at 78 FLUXNET sites for 7 PFTs to optimise all the carbon-related
24 parameters, and finally atmospheric CO₂ concentration measurements from 53 sites in the
25 GLOBALVIEW network (GLOBALVIEW-CO₂, 2013), which predominantly provided a
26 constraint on the initial magnitude of the soil carbon reserves in the model. These three global
27 multiple data stream CCDAS have allowed an improvement in both the mean seasonal cycle
28 as well as the trend of net land surface CO₂ exchange, especially with the inclusion of the
29 atmospheric CO₂ data (Kaminski et al., 2012; Peylin et al., 2016; Schürmann et al., 2016).
30 Atmospheric CO₂ concentration observations are one of the most accurate, long-term data sets
31 in environmental science and they provide important information about the global CO₂ sink
32 capacity by land and ocean.

1 Many of the aforementioned studies reported that adding extra data streams helped to
2 constrain unresolved sub-spaces of the total parameter space. Scholze et al. (2016) found that
3 simultaneously assimilating SMOS soil moisture data with CO₂ observations reduced the
4 ambiguity in the solution space when assimilating CO₂ only, and the multiple data constraint
5 was able to resolve a much larger sub-space in parameter space (about 30 parameters out of
6 the 101 compared to 15 without SMOS data). Bacour et al. (2015) and Schürmann et al.
7 (2016) both reported that the addition of FAPAR data brought extra information on the
8 phenology-related processes in the model, and therefore retrieved different posterior C flux-
9 related parameter values than when assimilating NEE or atmospheric CO₂ data alone. An
10 interesting aspect of the Kaminski et al. (2012) study was that the inclusion of FAPAR in
11 addition to atmospheric CO₂ concentration samples resulted in a particular improvement for
12 the hydrological fluxes in the model, thus demonstrating the importance of assessing the
13 potential benefit for model variables that may not have been the main target of optimisation.

14 Richardson et al. (2010) and Keenan et al. (2012) concluded that using ancillary
15 information (e.g. woody biomass increment, field-based LAI and chamber measurements of
16 soil respiration) as in addition to NEE data provided a valuable extra constraint on many
17 model parameters, which improved both the bias in model predictions and reduced the
18 associated uncertainties. The results of the REFLEX model-data fusion inter-comparison
19 project also indicated that observations of the different carbon pools would help to constrain
20 parameters such as root allocation and woody turnover that were not well resolved using NEE
21 and LAI data alone (Fox et al., 2009). Thum et al. (in review) found that the addition of
22 aboveground biomass stocks brought a longer-term constraint on allocation parameters and
23 mortality/turnover processes. However, they noted an incompatibility when assimilating both
24 annual increment and total biomass data, as the total stocks take into account losses related to
25 disturbance and management (e.g. canopy thinning) – processes that were not included in that
26 version of the model. On the other hand, Williams et al. (2005) observed that one-off, or
27 rarely taken, measurements of carbon stocks were unable to constrain components of the
28 carbon cycle to which they were not directly related. This raises the issue of the relative
29 influence of different data streams in a joint assimilation, especially if the number of
30 observations for each is vastly different, which will be the case when assimilating both half-
31 hourly C flux data in addition to soil C stock observations that are typically available at an
32 annual time scale. The spatial distribution of each data stream is also important, especially for
33 heterogeneous landscapes (Barrett et al., 2005; Alton, 2013).

1 Although a number of multiple data stream assimilation studies exist at various scales,
2 very few studies have specifically investigated the added benefit of different combinations of
3 data streams in a factorial study, with a few notable exceptions (Barrett et al., 2005;
4 Richardson et al., 2010; Kato et al., 2013; Keenan et al., 2013; Bacour et al., 2015;
5 Schürmann et al., 2016). Kato et al. (2013) and Bacour et al. (2015) both evaluated the
6 complementarity of eddy covariance and FAPAR data streams at site level, i.e. the impact of
7 assimilating one individual data stream on the other model state variable, as well as when
8 both data streams were included in the optimization (see discussion in Section 3.2). The study
9 of Keenan et al. (2013) was particularly notable in its aim to quantify which data streams
10 provide the most information (in terms of model-data mismatch) and how many data streams
11 are actually needed to constrain the problem. They reported that of the 17 field-based data
12 streams available, projections of future carbon dynamics were well-constrained with only 5 of
13 the data sources, and crucially, not with eddy covariance NEE measurements alone. These
14 results may be specific to this site or type of ecosystem, but their study highlights the need for
15 further research in this area, and in particular, for synthetic data experiments that allow us to
16 understand which data will be the most useful for a given scientific question. This will also
17 enable researchers to plan more efficient measurement campaigns with experimentalists, as
18 also pointed out by Keenan et al. (2012).

19

20 **2.2 Issue of bias and inconsistencies between the observations and the** 21 **model**

22 Despite the theoretical benefit of adding data streams into an assimilation system as
23 additional constraints, several of the aforementioned studies at both site and global scale have
24 reported a bias or inconsistency either between the different observation data streams, or
25 between the observations and the model. This is easily detected when the optimisation of one
26 data stream results in a worse fit than the prior in one or more of the other data streams, as
27 seen in Section 2.2.2. Kato et al. (2013) assimilated SeaWiFS FAPAR (Gobron et al., 2006)
28 and eddy covariance LE measurements at the FLUXNET site in Maun, Botswana. They
29 showed that the individual assimilation of each the two data streams resulted in a perfect (i.e.
30 within the observational uncertainty) fit to the assimilated data set, but a considerable
31 degradation of the fit to the non-assimilated data set compared to the prior. A comparison
32 against eddy covariance measurements of gross carbon uptake (gross primary production –

1 GPP) pointed to a bias in the FAPAR data because the fit to the independent GPP data was
2 degraded after assimilating FAPAR data only, while the fit improved after assimilating the LE
3 data only. Nevertheless, the simultaneous assimilation of both data streams achieved a
4 compromise between the two suboptimal results achieved after assimilating only one data
5 stream. The calibration further limited the number of parameters with correlated errors, and
6 yielded a higher theoretical reduction in parameter uncertainty and a decrease in the RMS
7 difference by 16% for the GPP data compared to the prior.

8 Bacour et al. (2015) also noted that when assimilating both in-situ and satellite-derived
9 FAPAR data (from the SPOT and MERIS instruments) and in-situ NEE and LE flux data
10 from two French FLUXNET sites into the ORCHIDEE LSM both separately and together, the
11 posterior parameter values changed significantly for the photosynthesis and phenology-related
12 parameters, depending on the bias between the model and the observations and the correlation
13 between the parameter errors. If NEE data were assimilated alone there was an even stronger
14 positive bias (model–observations) in the start of leaf onset in the FAPAR data than in the
15 prior simulations, and no improvement in the maximum value. This was likely due to the fact
16 that there were enough degrees of freedom to fit the NEE without changing the phenology-
17 related parameters. Similarly, the fit to the NEE was degraded when the model was only
18 optimized with FAPAR data. The model was able to fit the maximum FAPAR but this
19 resulted in an adverse effect on the carbon assimilation capacity of the vegetation. The
20 authors argued this was related to incompatibilities between the FAPAR and both the model
21 and NEE measurements, possibly due to its larger spatial footprint of the satellite-derived
22 FAPAR data and/or inaccuracies in the retrieval algorithm. However, given that assimilating
23 in-situ FAPAR also degraded the fit to the NEE, another culprit may be an inconsistency
24 between the model and the data. The authors suggested this could be due to the different
25 characterisation of FAPAR or LAI in the model compared to what is described in the data.
26 For example, satellite-derived greenness measures (FAPAR/NDVI) also contain information
27 on the non-green elements of vegetation, but the model only simulates green LAI.
28 Furthermore parameters and processes in models have been developed at certain temporal and
29 spatial scales. Vegetation is often simply represented as a “big leaf” model in LSMs, taking
30 no account of vertical canopy structure or the spatial heterogeneity in a scene, which is an
31 additional source of inconsistency with what is measured. The joint (simultaneous)
32 assimilation of all three data streams in Bacour et al. (2015) reconciled the different sources
33 of information, with an improvement in the model-data fit for NEE, LE and FAPAR.

1 However, the compromise achieved in the joint assimilation was only possible when the
2 FAPAR data were normalised to their maximum and minimum values, which thus partially
3 accounted for any bias in the magnitude of the FAPAR or inconsistency with the model.

4 The story of biases and apparent inconsistencies in FAPAR data does not end there. A
5 bias correction was also necessary in the study by Kaminski et al. (2012) with CCDAS-
6 BETHY using the MERIS FAPAR product in addition to atmospheric CO₂ data (see above).
7 They found that optimisation procedure failed when using the original FAPAR product
8 because the FAPAR values were biased towards higher values. Only after applying a bias
9 correction on the FAPAR data before assimilation was the optimisation successful.
10 Schürmann et al. (2016) also reported the need to reduce a prior model bias in FAPAR. Even
11 though the assimilation successfully corrected for this FAPAR bias, the impact of the prior
12 bias was evident in the spatial patterns of the modelled heterotrophic respiration. Assimilating
13 FAPAR data alone therefore resulted in a slight degradation in the net C flux and
14 consequently led to incorrect simulations of the atmospheric CO₂ growth rate. The addition of
15 CO₂ as a constraint prevented this degradation and resulted in a compromise in which FAPAR
16 helped to disentangle these processes and find different parameter values compared to the
17 CO₂-only case, thus improving the fit to both data streams. Forkel et al. (2014) discovered an
18 apparent inconsistency between satellite-derived FAPAR and GPP data in tundra regions
19 when using these data (plus satellite-derived albedo) to optimise the LPJmL LSM. They too
20 speculated that the data might be positively biased, in this case due to issues with satellite
21 measurements taken at high sun zenith angles. However, they gave alternative suggestions,
22 one being that an inadequate model structure may be at fault – for example, LPJmL does not
23 include vegetation classes corresponding to shrub, moss and lichen species that are dominant
24 in these ecosystems. They also noted that the GPP product they used, which is based on a
25 model tree ensemble up-scaling of FLUXNET data (Jung et al., 2011), might contain
26 representation-related biases, given that there are very few FLUXNET stations in tundra
27 regions. The issue of representation errors of sites has been touched upon before (e.g.
28 Raupach et al., 2005). Alton (2013), who performed a global multi-site optimisation of the
29 JULES LSM with a diverse range of data including satellite-derived LAI, FLUXNET, soil
30 respiration and global river discharge, raised the point that FLUXNET sites are known to be
31 large carbon sinks, which could potentially result in biased global NEE estimates. Resolving
32 these apparent inconsistencies was beyond the scope of most of these studies, aside from
33 applying a bias correction where one was evident. Aside from simple corrections, Quaife et

1 al. (2008) and Zobitz et al. (2014) suggested that LSMs should be coupled to radiative
2 transfer models to provide a more realistic and mechanistic observation operator between the
3 quantities simulated by the model and the raw radiance measured by satellite instruments.
4 This proposition followed the experience gained in the case of atmospheric models for several
5 decades (Morcrette, 1991).

6

7 **2.3 Step-wise versus simultaneous assimilation**

8 The paper by Alton (2013) documents the only previous study to have used a step-wise
9 assimilation approach with more than two data streams, stating that the final parameter values
10 were independent of the order of data streams assimilated. No studies in the LSM community
11 to date have explicitly examined a step-wise versus simultaneous assimilation framework
12 with the same optimisation system and model. The step-wise assimilation with the
13 ORCHIDEE-CCDAS detailed in Peylin et al. (2016) has been compared to a simultaneous
14 optimisation using the same three data streams as part of an on-going study. At each step, the
15 resulting simulations (using the posterior parameters) were compared to the data stream from
16 the previous steps. The fit to the MODIS NDVI (used in a similar manner to FAPAR as a
17 proxy for vegetation greenness) was unchanged after further optimization of the phenology-
18 related parameters in the second and third steps using in-situ flux and atmospheric CO₂
19 concentration data. In the simultaneous optimisation, the addition of NEE or atmospheric CO₂
20 concentration measurements resulted in a lower improvement to the fit to MODIS NDVI
21 (Bacour et al. submitted?). As the NDVI data were normalised this was not a result of a
22 simple bias in the magnitude of the data. Rather, it was likely due to inconsistencies between
23 the model and data as discussed by Bacour et al. (2015). It is important to reiterate that there
24 should be no difference between the step-wise and the simultaneous given an adequate
25 description of the error covariance matrices and compliance with the assumptions associated
26 with the inversion algorithm used. However, in practice it is very difficult to define a
27 probability distribution that properly characterises the model structural uncertainty and
28 observation errors accounting for biases and non-Gaussian distributions. This leads to issues
29 within a simultaneous assimilation, and a greater risk of differences between a step-wise and
30 simultaneous assimilation. However, as discussed in the introduction a step-wise assimilation
31 may be useful on a provisional basis for dealing with possible inconsistencies. For example in
32 the step-wise approach of Peylin et al. (2016) as the error covariance of the phenology-related

1 parameters was strongly constrained by the satellite data in the first step (and was propagated
2 to the second step), the later optimisations with NEE and atmospheric CO₂ data in steps 2 and
3 3 found alternative solutions for the C flux-related parameters that provided a reasonable fit to
4 all data streams. Wherever possible however, a simultaneous optimisation is favourable
5 because the strong parameter linkages between different processes are maintained, and
6 therefore biases and inconsistencies between the model and observations should be addressed
7 prior to optimisation.

8

9

10 **3 Demonstration with two simple models and synthetic data**

11 The three sub-sections in Section 2 highlight examples within a carbon cycle modeling
12 context of the three main challenges faced when performing a multiple data stream
13 assimilation, namely, i) the possible negative influence of including additional data streams
14 into an optimization on other model variables; ii) the impact of bias in the observations,
15 missing model processes or incompatibility between the observations and with the model (as
16 discussed in Section 2.2), and iii) the difference between a step-wise and simultaneous
17 optimization (and the order of data stream assimilation) if the assumptions of the inversion
18 algorithm are violated, which is more likely to be the case with non-linear models when using
19 derivative-based algorithms and least-squares formulation of the cost function (as discussed in
20 Section 2.3). The latter point is important because derivative methods (compared to global
21 search) are the only viable option for large-scale, complex LSMs given the time taken to run a
22 simulation. In addition to the above three challenges we have performed a preliminary
23 investigation into the impact of correlated errors between the two data streams, which is a
24 topic that has not yet been studied in the context of carbon cycle models.

25 This section aims to demonstrate these challenges using simple toy models and synthetic
26 experiments where the true values of the parameters are known. Thus the following sections
27 include a description of the toy models together with the derivation of synthetic observations,
28 the inversion algorithm used to optimise the model parameters and the experiments
29 performed, followed by the results for each test case.

1 3.1 Methods

2 3.1.1 Simple carbon model

3 To demonstrate the challenges of multiple data stream assimilation in a carbon cycle
4 context, we have chosen a test model that represents a simplified version of the carbon cycle
5 dynamics typically implemented in most LSMs. The model has been well-documented in
6 Raupach (2007) and has been used previously in the OptIC DA inter-comparison project
7 (Trudinger et al., 2007). It is based on two equations that describe the temporal evolution (on
8 a daily time step) of two living biomass (carbon) stores, s_1 and s_2 , and the biomass fluxes
9 between these two stores:

$$10 \quad \frac{ds_1}{dt} = F(t) \left(\frac{s_1}{p_1 + s_1} \right) \left(\frac{s_2}{p_2 + s_2} \right) - k_1 s_1 + s_0 \quad (1)$$

$$11 \quad \frac{ds_2}{dt} = k_1 s_1 - k_2 s_2 \quad (2)$$

12 In this model formulation, s_1 and s_2 are approximately equivalent to above- and belowground
13 biomass stocks. The unknown parameters p_1 , p_2 , k_1 and k_2 will be optimised in the inversions.
14 The first term on the right-hand side of Eq. (1) corresponds to the Net Primary Production
15 (NPP) i.e. the carbon input to the system as a function of time, represented by $F(t)$, weighted
16 by factors (the two fractions in parentheses) that account for the size of both pools, in order to
17 introduce a limitation on NPP. The $F(t)$ forcing term is a random function of time (“log-
18 Markovian” random process) representing the effect of fluctuating light and water availability
19 due to climate on the NPP (Raupach, 2007 – Section 5.3). The litterfall is an output of s_1
20 (aboveground biomass store) and an input to s_2 (belowground biomass store) and is calculated
21 as a constant fraction (k_1) of s_1 (defined by $k_1 s_1$). Heterotrophic respiration (Rh) is a constant
22 fraction (k_2) of the belowground carbon reserve s_2 and is represented $k_2 s_2$. The constant s_0 is a
23 “seed production” term set to 0.01 (i.e. not optimised) to ensure the model does not verge
24 towards zero. A more detailed description of the properties of the model is given in Trudinger
25 et al. (2007 – Section 2.1) and an in-depth analysis of the dynamical behaviour of the model is
26 provided in Raupach (2007). Synthetic observations of both s_1 and s_2 variables were used to
27 optimise all the unknown parameters in the model (see Section 2.1.5).

28

1 3.1.2 Non-linear toy model

2 Although the simple carbon model contains a non-linear term it is essentially still a
3 quasi-linear model. In order to illustrate the challenges associated with multiple data stream
4 data assimilation for more complex non-linear models, especially when using derivative
5 methods, we defined a simple non-linear toy model based on two equations with two
6 unknown parameters:

$$7 \quad s_1 = a \exp^b + at^2 \quad (3)$$

$$8 \quad s_2 = \sin(10a + 10b) + 10t^2 \quad (4)$$

9 where s_1 and s_2 also correspond to two model state variables (as for the simple C model), a
10 and b are the unknown parameters included in the optimisation, and t is the independent
11 variable, which could represent time in a real-world scenario. Note that this model is not
12 based on any particular physical process associated with land surface biogeochemical cycles,
13 but it does contain typical mathematical functions that are observed in reality and
14 implemented in LSMs. For example, the sinusoidal function (Eq. (4)) could represent diurnal
15 variations of various processes such as photosynthesis and respiration. Exponential response
16 functions (such as in Eq. (3)) are also observed for certain processes, including the
17 temperature sensitivity of soil microbial decomposition. As for the simple carbon model,
18 synthetic observations corresponding to the s_1 and s_2 variables were used to optimise both
19 parameters (see Section 2.1.5).

20

21 3.1.3 Bayesian inversion algorithm

22 Most data assimilation approaches follow a Bayesian formalism which, simply put,
23 allows prior knowledge of a system (in this case the model parameters) to be updated, or
24 optimised, based on new information (from the observations). In order to achieve this we
25 define a “cost function” that describes the misfit between the data and the model, taking into
26 account their respective uncertainties, as well as the uncertainty on the prior information. If
27 we follow a Bayesian formalism and least-squares minimisation approach, and assume
28 Gaussian probability distributions for the model parameter and observation error
29 variance/covariance, we derive the following cost-function (Tarantola, 1987):

$$1 \quad J(\mathbf{x}) = \frac{1}{2}[(H(\mathbf{x}) - \mathbf{y})^T \cdot \mathbf{R}^{-1} \cdot (H(\mathbf{x}) - \mathbf{y}) + (\mathbf{x} - \mathbf{x}^b)^T \cdot \mathbf{B}^{-1} \cdot (\mathbf{x} - \mathbf{x}^b)] \quad (5)$$

2 where \mathbf{y} is the observation vector, $H(\mathbf{x})$ the model outputs given parameter vector \mathbf{x} , \mathbf{R} the
3 observation error covariance matrix (including measurement and model errors), \mathbf{x}^b the a priori
4 parameter values, and \mathbf{B} the prior parameter error covariance matrix. This framework leads to
5 a Gaussian posterior parameter probability distribution function and requires that the model
6 and its observation operator are linear.

7 The aim of the inversion algorithm is to find the minimum of this cost function,
8 thereby achieving the best possible fit between the model simulations and the measurements,
9 conditioned on their respective uncertainties and prior information. For cases where there is a
10 strong linear dependence of the model to the parameters (at least for variations in \mathbf{x} of the size
11 of those expected in the data assimilation system), and where the dimensions of the problem
12 are not too large, the solution can be derived analytically. If not, as is usually the case with
13 LSMs, there are different numerical methods to find the most optimal parameter values.
14 These include global search methods that randomly search the parameter space and test the
15 likelihood of a particular parameter set at each iteration, and derivative methods, which
16 calculate the gradient of the cost function at each iteration to find its minimum. In this study
17 we use the latter class of methods. More specifically we use a quasi-Newton algorithm that
18 uses both the gradient of the cost function and its derivative (Hessian) to evaluate if the
19 minimum has been reached (i.e. where the gradient is zero). Thus we obtain the following
20 algorithm for iteratively finding the minimum (Tarantola, 1987, p195):

$$21 \quad \mathbf{x}_{i+1} = \mathbf{x}_i - \varepsilon_i (\mathbf{H}^T \mathbf{R}^{-1} \mathbf{H} + \mathbf{B}^{-1})^{-1} (\mathbf{H}^T \mathbf{R}^{-1} (\mathbf{y} - H(\mathbf{x})) + \mathbf{B}^{-1} (\mathbf{x}_i - \mathbf{x}^b)) \quad (6)$$

22 where i is the iteration number and \mathbf{H} is the Jacobian, or first-order derivatives, of H , which in
23 this study is determined using a finite difference method. Note that as we are potentially
24 dealing with non-linear models, the quasi-Newton method has been slightly adapted to
25 include the constant scaling factor ε_i (with a value < 1.0) to ensure that the algorithm will
26 converge.

27 Of course no inversion algorithm is perfect, and therefore it is possible that the true
28 “global” minimum of the cost function has not been found. Derivative methods in particular
29 can get stuck in so-called “local minima”, preventing the algorithm from finding the true
30 minimum. To address this issue we carry out a number of assimilations with different random

1 first guess points in the parameter space. If they all result in the same reduction in cost
2 function value, we can have more confidence that the true minimum has been found.

3 Once the minimum of the cost function has been found, the posterior parameter error
4 covariance can be approximated (using the linearity assumption) from the inverse Hessian of
5 the cost function around its minimum, which is calculated using the Jacobian of the model at
6 the minimum of $J(\mathbf{x})$ (for the set of optimized parameters), \mathbf{H}_∞ , following Tarantola (1987):

$$7 \quad \mathbf{A} = [\mathbf{H}_\infty^T \mathbf{R}^{-1} \mathbf{H}_\infty + \mathbf{B}^{-1}]^{-1} \quad (7)$$

8 Note that the posterior error covariance matrix can be propagated into the model space to
9 determine the posterior uncertainty on the simulated state variables as a result of the
10 parametric uncertainty (as shown in the coloured error bands in the time series plots – Figures
11 1 and 5) using the following matrix product and the hypothesis of local linearity (Tarantola,
12 1987):

$$13 \quad \mathbf{R}_{post} = \mathbf{H}_\infty \cdot \mathbf{A} \cdot \mathbf{H}_\infty^T \quad (8)$$

14

15 3.1.4 Step-wise versus simultaneous assimilation

16 *Step-wise approach*

17 In the step-wise approach each data stream (in our cases s_1 and s_2 , see above) is
18 assimilated sequentially, and the posterior error covariance matrix of Eq. (7) is propagated to
19 the next step as the prior in Eq. (6). Note that the error covariance matrix can only be
20 propagated if it is calculated within the inversion algorithm, which is the case here but may
21 not be possible in other studies. The following details an example for two data streams.

22 Step 1: Assimilation of the first data stream, s_1 . The prior parameters, including their values
23 and error covariance (\mathbf{x}^b and \mathbf{B}), are optimised to produce a first set of posterior
24 optimised parameters \mathbf{x}_1 with error covariance \mathbf{A}_1 .

25 Step 2: Assimilation the second data stream, s_2 . The parameters, \mathbf{x}_1 , and their error
26 covariance, \mathbf{A}_1 , are used as a prior to the optimisation system and further optimised
27 to produce the second (and final) set of posterior optimised parameters, \mathbf{x}_{post} , and the
28 associated error covariance \mathbf{A} .

1 *Simultaneous approach*

2 Both data streams s_1 and s_2 are included in the optimisation and all parameters are optimised
3 at the same time. The prior parameters, including their values and error covariance (\mathbf{x}^b and \mathbf{B})
4 are optimised to produce the posterior parameter vector (\mathbf{x}_{post}) and associated uncertainties \mathbf{A} .

5

6 3.1.5 Optimisation set-up: parameter values and uncertainty, and generation 7 of synthetic observations

8 In this study we used synthetic observations that were generated by running the model
9 with known (or ‘true’) parameter values and adding random Gaussian noise corresponding to
10 the defined observation error for both s_1 and s_2 (see Table 1). We optimised a ten-year time
11 window for the simple carbon model, in order to capture the dynamics of the s_1 and s_2 pools
12 over a time period compatible with typically available observations. For the non-linear toy
13 model, which did not correspond to physical processes in the terrestrial biosphere, we ran a
14 simulation over a window of 100 integrations (steps) of the equations. The observation
15 frequency was daily, corresponding to the time-step of the simple carbon model (a value of 1
16 for the non-linear toy model), and the observation error was set to 10% of the mean value for
17 each set of pseudo-observations derived from multiple first guesses of the model.

18 The true values of all parameters for both models are given in Table 1, together with
19 their upper and lower bounds (following Trudinger et al., 2007). We have not performed a
20 prior sensitivity analysis to decide to which parameters are important to include in the
21 optimisation, as the model variables are sensitive to all of the (small set of) parameters.
22 However, in the case of a more complex, large-scale LSM it is advisable to carry out such an
23 analysis, particularly given the computational burden of optimising many parameters. In this
24 study the parameter uncertainty (1 sigma) was set to 40% of the parameter range following
25 recent studies (e.g. Bacour et al., 2015). Prior values were chosen from a uniform random
26 distribution bounded by the parameter bounds.

27

28 3.1.6 Experiments

29 The specific objective of the following experiments is to test the impact of a bias in the
30 observations that is not accounted for in the R matrix, and the impact of using derivative

1 methods with non-linear models (as may be necessary with large-scale LSMs), particularly
2 with reference the differences that may arise between step-wise and simultaneous
3 optimisations.

4 Table 2 details the experiments that were carried out based on all possible combinations for
5 assimilating the two data streams. Three approaches were compared: i) separate – where only
6 one data stream was included in the optimisation; ii) step-wise – where each data stream was
7 assimilated sequentially; and iii) simultaneous – where both data streams were included in the
8 optimisation. All parameters for both models were optimised in all experiments, therefore in
9 the step-wise cases the parameters were optimised twice. Tests for the step-wise were also
10 carried out with and without the propagation of the full posterior parameter error covariance
11 matrix, \mathbf{A}_1 , in between steps 1 and 2 (test cases 2b and d – see Table 2) – i.e. for these tests
12 only the posterior variance was propagated. An additional test was included for the
13 simultaneous assimilation in order to test the impact of having a substantial difference in the
14 number of observations for the data stream included in the optimisation, as may be the case
15 for belowground (e.g. soil) biomass observations in reality. Therefore in test case 3b, only one
16 observation was included for data stream s_2 .

17 The differences in the parameter values and the theoretical reduction in their uncertainty
18 ($1 - (\sigma_{\text{post}} / \sigma_{\text{prior}})$) were examined for all eight test cases, as well as the fit (RMSE) to both
19 data streams after the optimisation. For the step-wise approach we investigated if the fit to the
20 first data stream is degraded in the second step by comparing the RMSE after each step. Note
21 that the reduction in uncertainty is a theoretical, or approximate, estimate of the real
22 uncertainty reduction because of the assumptions made in the inversion scheme.

23 In a second stage the impact of an unknown, un-accounted for bias in the model was
24 examined. This bias could be a systematic bias in the observations due to the algorithm used
25 for their derivation, the result of missing or incomplete processes in the model, or an
26 incompatibility between the observations and the model, for example due to differences in
27 spatial resolution or an inconsistent characterisation of a variable between the model and the
28 observations. To test the impact of such an occurrence, we introduced a constant scalar bias
29 into the modelled s_2 variable with a value of 10 (i.e. twice the magnitude of the defined
30 observation uncertainty). All eight experiments were repeated, but a bias was introduced into
31 the model calculation of s_2 that was not accounted for in the cost function (i.e. the error
32 distributions retained a mean of zero). This was treated as an unknown bias, and therefore not

1 corrected or accounted for in the inversion scheme and the defined observation uncertainty
2 (Table 1) was not changed for this set of experiments.

3 In all experiments for both models twenty assimilations were performed starting from
4 different random “first guess” points in the parameter space. As discussed in Section 2.1.3
5 this was done to test the ability of the algorithm to converge to the true global minimum of the
6 cost function. Note that the global minimum and possible reduction in $J(x)$ will be different
7 for each experiment, as each is based on a different cost function.

8 For all the above tests we assumed independence (i.e. uncorrelated errors) for both the
9 parameters and observation covariance matrices, thus the \mathbf{R} and \mathbf{B} matrices were diagonal. In
10 a final test we performed a simultaneous optimisation to examine the impact of having
11 correlated errors between the s_1 and s_2 observations. Thus the random Gaussian noise added
12 to s_1 for each time step was correlated to the noise added to s_2 . The correlated observation
13 errors were generated following the method used by Trudinger et al. (2007 – paragraph 22).
14 The added noise was time invariant, i.e. there was no correlation between one time step and
15 the next as we were specifically looking at correlations between the observations. We tested
16 both accounting for the correlated errors by populating the corresponding off-diagonal
17 elements of the \mathbf{R} (observation error covariance) matrix, and ignoring the correlated errors by
18 keeping \mathbf{R} diagonal. The reason for performing both tests was to demonstrate the possible real
19 world scenario where correlated observation errors exist, but this information is not included
20 in the optimisation due to a lack of knowledge as to how to characterise the errors. For both
21 tests we performed optimisations using a combination of different of observation error and
22 correlation magnitudes (observations errors between 0.05 and 20 in 9 uneven intervals, and
23 observation correlations between -0.9 and 0.9 with an interval of 0.4). As in the above
24 experiments, twenty random first guesses in the parameter space were used and 15 iterations
25 of the inversion algorithm were performed.

27 **3.2 Results**

28 The twenty random first guess assimilations were examined for each set of experiments
29 for both models (before the results for each test were examined in more detail), in order to
30 check that the algorithm converged to a global minimum. As shown in the supplementary
31 information (Fig. S1), a high proportion of the twenty first guess assimilations across all test
32 cases for both models resulted in a similar reduction in $J(x)$, even though the overall

1 magnitude of the reduction was sometimes different between tests. This indicates that the
2 algorithm does not easily get stuck in any local minima (if they exist). The examples shown in
3 the results below were taken from one first guess parameter set for each model that belonged
4 to the cluster that had the highest cost function reduction. Any differences seen in the
5 parameter values, their posterior uncertainty or the resultant RMSE reduction described below
6 therefore are due to the specific details of each test and not the inability of the algorithm to
7 find the minimum.

8

9 3.2.1 Typical performance with a quasi-linear model and no bias

10 Figures 1a and b show the simple carbon model simulations for test case 3a (in which
11 both data streams are assimilated simultaneously) for the s_1 and s_2 variables. A large reduction
12 in RMSE is achieved after optimisation (blue curve) with respect to the observations (black
13 curve). Overall, there is a good reduction in RMSE for all test cases (including the individual
14 assimilations 1a and 1b) with a reduction of $\sim 80\%$ for s_1 and s_2 . In addition, the optimisation
15 of the s_1 and s_2 variables resulted in a good or moderate reduction in RMSE for variables not
16 included in any assimilation: $\sim 60\%$ for the litterfall (Eqn. 1) and $\sim 16\%$ for the heterotrophic
17 respiration (R_h – Eqn. 2) across all test cases (not shown), although there was already a good
18 prior fit to the data. As would be expected from these results, the parameter values and the
19 theoretical reduction in parameter uncertainty do not vary between the tests (Figures 2a and b
20 blue symbols), except for a slight difference in the value of the k_2 parameter in test cases 1a
21 and 3b, for which there is also a lower reduction in uncertainty ($\sim 82\%$ compared to $>95\%$).
22 Note that Fig. 2a shows the normalised parameter values to account for differences in the
23 magnitude of the different parameters and their range (the zero line represents the “true”
24 parameter value – see caption). In this situation therefore, where we have a relatively simple
25 linear model and two data streams to which the model parameters are highly sensitive, we see
26 that the differences between the step-wise and simultaneous approaches are minimal. This is
27 even the case when the error covariance is not propagated between the two steps (test cases 2b
28 and d), suggesting that under this assimilation set-up with this model both s_1 and s_2
29 individually contain enough spatio-temporal information to retrieve the true values of all
30 parameters, as we can see from the separate test cases 1a and b. However, we cannot
31 definitively say whether this is due to the simplicity or relative linearity of the model – it is

1 possible that observations of variables in more complex linear model would not be able to
2 retrieve the true values of all parameters.

3

4 3.2.2 Impact of unknown bias in one data stream – example with a simple 5 carbon model

6 In Section 2.2.1 we saw that there is little difference between a step-wise and
7 simultaneous optimisation if there is no bias in the model or observations, and if the model is
8 quasi-linear and therefore the critical assumptions behind the inversion approach were not
9 violated. However, it is not uncommon to have a bias between your observations and model
10 that is not obvious and therefore not accounted for in the optimisation, as the cost function
11 used in most inversion algorithms (and in this study) assume Gaussian error distributions with
12 zero mean. Note that this is also the case when defining a likelihood function for accepting or
13 rejecting parameter values in a global search method. To test the impact of a bias, we added a
14 constant value to the simulated s_2 variable in a second test (see Section 3.1.6) that was treated
15 as an unknown bias, and therefore not corrected or accounted for in the inversion scheme. The
16 impact of this bias on s_1 and s_2 is shown in Figures 1c-d, and the reduction in RMSE between
17 the model and observations is seen in Fig. 3 for all variables (including Rh and litterfall). The
18 red symbols in Fig. 2 show the resultant parameter values and theoretical reduction in
19 uncertainty as a result of the bias. The inversion cannot accurately find the correct values for
20 all parameters in any test case and there are now considerable differences between the
21 simultaneous and step-wise approach. Furthermore the order in which the data streams are
22 assimilated in the step-wise cases also results in different posterior parameter values (test
23 cases 2a and b versus 2c and d in Fig. 2a and Fig. 3). Nevertheless the optimisation results in
24 a similar reduction in uncertainty on the parameters, except in test case 1b where only s_2 data
25 are assimilated (Fig. 2b).

26 The main impact of the bias in the modelled s_2 variable is on the value of k_2 parameter
27 (Fig. 2a), which is consistently offset from the true value (dashed line in Fig. 2a) in all test
28 cases. This was expected given that it is the parameter most directly related to the calculation
29 of s_2 . However, in test cases 2a and 3a, the values of p_1 and p_2 are also incorrect (and p_1 for
30 test case 2b). Note that these parameters only indirectly influence the s_2 pool in the model,
31 and therefore we might have expected that they would be less affected by the bias. This nicely

1 demonstrates one issue that could arise in all DA studies, where the bias in a particular
2 variable (in the observations or the model) is aliased onto another process in the model
3 (Wutzler and Carvalhais, 2014). Such an aliasing of bias onto indirectly related parameters is
4 even more evident when only s_2 is included in the assimilation and s_1 does not provide any
5 constraint (test case 1b) – in this case all parameters are incorrect but the p_2 parameter in
6 particular shows a strong deviation from the true value (Fig. 2a). As a result we see a
7 deterioration in the RMSE for the s_1 , litterfall and Rh variables in test case 1b and in the step-
8 wise cases where s_2 is assimilated in the second step (Figures 3a, c and d – test case 1b, 2a
9 and 2b). However, the RMSE reduction remains high for the s_2 variable for these test cases
10 (Fig. 3b), as the inversion has found a solution that accounts for the bias even though all
11 inferred parameter values are incorrect. The assimilation of s_1 in the second step lowers the
12 reduction in RMSE for s_2 gained in the first step to $\sim 70\%$, but it is not a considerable
13 degradation.

14 Even though the posterior parameter values are incorrect, and despite the fact that the
15 first step results in a degradation, the final reductions in RMSE are largely the same as the
16 situation with no bias for all variables when s_1 is included in a simultaneous assimilation or
17 optimised in the second step (test cases 2c, d and 3a in Fig. 3). This shows that the inclusion
18 of s_1 observations can find a solution to counter the bias in s_2 and prevents a degradation in
19 the fit to the data. If s_2 is assimilated in the second step there is a negative impact on all other
20 variables as discussed above, demonstrating again that the order of data stream assimilation
21 can matter when there are biases or inconsistencies between the data and the model.

22 The analysis of the impact of the bias presented here is specific to this model and the
23 type and magnitude of the bias that was added, but the broader findings can be generalised to
24 any situation in which there is a bias or inconsistency between a model and data that is not
25 accounted for in the assigned error distributions. Exactly what might constitute a bias or
26 inconsistency is discussed more in Section 3.2. Also note that it is important to examine the
27 impact on the other variables. For the separate test case 1b in which only s_2 data are used to
28 optimise the model, the negative impact on the other variables (Fig. 3) would have been
29 concealed if we had only examined the posterior reduction in RMSE for the s_2 variable. Again
30 this is a concern that is inherent to all DA experiments, whether single- or multi-data stream,
31 but we can see from these results (i.e. by comparing the separate test cases 1b with 2a and b)

1 that adding another data stream in a multi-constraint approach does not always reduce the
2 problem.

3

4 3.2.3 Difference between the step-wise and simultaneous approaches in the 5 presence of a non-linear model

6 As discussed in Section 2.2.1, there is little difference between the step-wise and the
7 simultaneous assimilation approaches for simple, relatively linear models, unless the
8 observation error (including measurement and model errors) distribution deviates strongly
9 from the Gaussian assumption. However in reality, large-scale, complex LSMs may contain
10 highly non-linear responses to certain model parameters. To demonstrate the impact of non-
11 linearity in a multiple data stream assimilation context, we used a non-physically based toy
12 model chosen for its non-linear characteristics (see Section 2.1.2).

13 Fig. 4a shows the posterior parameter values for both the a and b parameters of the
14 non-linear toy model for all test cases. The values were not normalised as both parameters
15 have the same range. The horizontal dashed line shows the “true” known values of the
16 parameters (both equal to 1.0) that were used to generate the synthetic observations. Note that
17 no bias has been introduced into the model in the results described here. The prior and
18 posterior model s_1 and s_2 simulations for the non-linear toy model are compared to the
19 synthetic observations in Fig. 5 for both step-wise cases in which the posterior error
20 covariance matrix from step 1 (\mathbf{A}_1 – see section 2.1.4) was propagated to step 2 (experiments
21 2a and c – Fig. 5a-d) and both simultaneous cases 3a and b (Fig. 5 e-h). Finally Fig. 6
22 summarises the reduction in RMSE between the simulated and observed s_1 and s_2 variables
23 for the non-linear toy model for all test cases and, in the step-wise cases, the reduction in
24 RMSE after both the first and second steps (light versus dark green bars).

25 Assimilating each data stream individually (test cases 1a and b) does not result in an
26 accurate retrieval of the posterior parameters (Fig. 4a), nor in a strong constraint on either
27 parameter, as shown by the lack of theoretical reduction in the parameter uncertainty after the
28 optimisation (Fig. 4b). Despite this, there is a 91-92% reduction in RMSE for the data stream
29 that was included in the optimisation (i.e. for s_1 in test case 1a – Fig. 6a, and s_2 in test case 1b
30 – Fig. 6b). However, the improvement on the other data stream is much less (28% reduction
31 in RMSE for s_1 when s_2 is assimilated) or even results in a degradation compared to the prior

1 fit (e.g. in the case of s_2 when s_1 is assimilated – Fig. 6b). Lack of improvement, or even
2 degradation, in the RMSE of other variables in the model is a common issue for data
3 assimilation in general, one that is not often evaluated in model-data fusion studies. It is also
4 is not necessarily the result of a bias or incompatibility between the observations and the
5 model.

6 Only the simultaneous case, in which all s_1 observations have been included in the cost
7 function (test case 3a), manages to retrieve the correct parameter values after the optimisation.
8 All other posterior parameter values are incorrect, and are considerably different between
9 each case, unlike for the simple carbon model (without a model bias). Most step-wise test
10 cases (particularly 2b-d) do not result in the same parameter values as the simultaneous test
11 case 3a in which all the observations are included (Fig. 4a). This highlights that strong non-
12 linearity in the model sensitivity to parameters, together with the use of an algorithm that is
13 only adapted to weakly non-linear problems, can result in differences between a step-wise and
14 simultaneous approach in multiple – data stream assimilation (see Section 1).

15 In the simultaneous optimisation in which all observations are included (test case 3a)
16 the posterior fit to the data dramatically improves for both the s_1 and s_2 data streams after the
17 assimilation (blue dashed line in Fig. 5e and f). This was expected given that the correct
18 values of the parameters were found. For the step-wise cases (test case 2a in Figures 5a and b,
19 and test case 2c in Fig. 5c and d), the black dashed line shows the prior, and the posterior after
20 step 1 is shown by green dashed line. In the step-wise assimilation we see two different
21 scenarios depending on which data stream was assimilated first. In the first step the results are
22 the same as the case where each individual data stream is assimilated separately. In both cases
23 the first step results in a good fit to the data that was included in the optimisation in that step.
24 When the s_1 data was assimilated in the first step (Fig. 5 first row), the fit to s_2 deteriorated
25 after the optimisation (Fig. 5b green dashed line and Fig. 6b – test case 2a_s1), but when the
26 s_2 data were assimilated first (Fig. 5 second row) the optimisation step did manage to achieve
27 an improvement in the s_1 data stream (Fig. 5c green dashed line and Fig. 6a – test case 2c_s1).

28 In the second step the optimisation of s_2 in test cases 2a and b does not degrade the fit
29 to s_1 when the full parameter error covariance matrix (\mathbf{A}_1) is propagated between step 1 and 2
30 (Figures 5a blue curve and 6a 2a_s2). Furthermore optimising s_2 in the second step reverses
31 the deterioration in s_2 caused by assimilating s_1 in the first step (Figures 5b blue curve and 6b
32 2a and b dark green bars). However, when s_1 data were assimilated in the second step (test

1 cases 2c and d), we found that the good fit achieved with s_2 observations in the first step was
2 effectively reversed (Fig. 5d blue curve). Therefore assimilating s_1 in the second step
3 degraded the fit to the s_2 observations, even compared to the prior case (Fig. 6b, dark green
4 bars for test cases 2c and d). This nicely highlights one of the main possible issues with a
5 step-wise assimilation framework.

6 The fact that the final reduction in RMSE values after both steps was $\sim 90\%$ for most
7 cases, even though the values were not correct for all but case 3a (Fig. 4), indicates that the
8 error correlation between the two parameters (~ -1.0 – calculated from the posterior error
9 covariance matrix but not shown) led to alternative sets of values that resulted in a similar
10 improvement to the data – a phenomenon known as model equifinality.

11

12 3.2.4 Order of assimilation of data streams and propagation of parameter 13 error covariance matrices in a step-wise approach

14 Comparing the step-wise cases 2a and b with 2c and d for the non-linear toy model
15 reveals that neither order in the assimilation, s_1 then s_2 , or s_2 then s_1 , results in the correct
16 posterior parameter values that match the simultaneous test case (Fig. 4a). This is not a result
17 that can be generalised to all step-wise assimilations as it will depend on the data stream
18 involved and whether they contain enough spatio-temporal information to accurately
19 constrain all the parameters included in the optimisation, as well as any biases in the model or
20 observations (as discussed in Section 3.2.2) or model non-linearity (section 3.2.3). In the case
21 of the non-linear toy model, neither s_1 nor s_2 find the right parameter values when assimilated
22 individually, therefore it is not surprising that neither order manages to achieve the right
23 posterior parameter values. Nevertheless, the theoretical uncertainty of both parameters is
24 reduced by $>95\%$ for the step-wise cases in which \mathbf{A}_1 from step 1 is propagated between step
25 1 and 2 (test cases 2a and c – Fig. 4b), even though the posterior values for the step-wise
26 cases are incorrect. This demonstrates that a good theoretical reduction in uncertainty is not
27 always indicative that the right parameters have been found by the optimisation. The lower
28 theoretical reduction in parametric uncertainty for cases 2b and d (Fig. 4b) demonstrates that
29 information is lost between the steps if the posterior error covariance terms of \mathbf{A}_1 after step 1
30 are not propagated to step 2, and therefore cannot be used to further constrain the
31 optimisation.

1 From a mathematical standpoint the most rigorous approach is to propagate the full
2 parameter error covariance matrices between each step. Without that constraint not only is
3 information lost in the second step, but the information contained in the second data stream
4 may have a stronger influence compared to a simultaneous or step-wise case with a
5 propagated error covariance matrix. The inversion may therefore be more vulnerable to any
6 strong biases or incompatibilities between the model and the observations of the second data
7 stream, or indeed the particular sensitivity of its corresponding model state variable to the
8 parameters. This is one possible explanation for the degradation seen in s_1 in the non-linear
9 toy model when s_2 is optimised in the second step and \mathbf{A}_1 is not propagated between the steps
10 (Fig. 6a test case 2b_s2). The same was also true for the simple carbon model for test case 2b
11 when a bias was introduced into the s_2 simulation (see Section 3.2.2 and Fig. 3a).

12 However, the reverse is also true – if the first data stream contains strong biases then
13 the associated error correlations will be also propagated with \mathbf{A}_1 . If autocorrelation in the
14 observation errors, or indeed correlation between the errors of the data streams, is not
15 accounted for, it is likely that the posterior simulations are over-tuned, i.e. we will
16 overestimate the reduction in parameter uncertainty. If this is the case and the first step results
17 in incorrect parameter values, the propagation of \mathbf{A}_1 could restrict the parameter values to the
18 wrong location in the parameter space and thus inhibit the ability of the inversion to find the
19 correct global minimum. These issues are likely to be more considerable for non-linear
20 models, as seen by the lack of difference between test cases 2a-d in the simple carbon model
21 example (Fig. 2).

22

23 3.2.5 Impact of accounting for correlated observation errors in the prior 24 observation error covariance matrix

25 In a final test we introduced time invariant correlated noise between the two data streams
26 (see Section 3.1.6). We investigated the impact of ignoring cross-correlation between two
27 data streams by comparing the results of i) an optimisation in which the correlated errors were
28 included in the off-diagonal elements of the prior observation error covariance matrix, \mathbf{R} , to
29 ii) an optimisation in which the correlated observation errors were excluded (i.e. \mathbf{R} was kept
30 diagonal). Note that this experiment is only relevant to simultaneous multiple data stream
31 assimilation, as it is not possible to account for cross-correlation between data streams when

1 one is assimilated after the other in a step-wise approach. The presence of error correlations
2 increases redundancy in the inversion, which would therefore reduce the expected theoretical
3 error reduction compared to uncorrelated observations (experiments not shown). We would
4 expect a further limitation on the expected error reduction with a sub-optimal system, as
5 represented by optimisation ii) in which there was cross-correlation between the data streams,
6 but the correlated observation errors were ignored (as seen in Chevallier, 2007).

7 Figure 7 shows the difference between the two optimisation, (i.e. including off-diagonal
8 minus diagonal elements in the \mathbf{R} matrix), for the reduction in the cost function value (Figures
9 7a and d) and posterior s_1 and s_2 observation errors (1 sigma – Figures 7b, c, e and f), for both
10 the simple C model (top row) and the non-linear toy model (bottom row). The plot shows the
11 median difference across all twenty random first guess parameters, and the reduction is
12 calculated as $1 - (\text{posterior}/\text{prior})$.

13 At low observation error there is no discernible difference between accounting for the
14 correlated observation errors in the \mathbf{R} matrix or not. This is likely because there is enough
15 information in the observations to find the global minimum of the cost function. Trudinger et
16 al. (2007) also found that similar posterior values were obtained when comparing
17 observations with correlated and uncorrelated Gaussian errors. However, at a certain point
18 with increasing observation error increases along the x-axis (i.e. decreasing information
19 content) there is a difference in the cost function and parameter error reduction between the
20 two optimisations for both models (Figure 7). The sub-optimal set-up in optimisation ii) is
21 clearly be seen as the optimisation with off-diagonal correlated errors in \mathbf{R} results in a higher
22 reduction in the cost function and posterior observation error (blue values). This pattern fits
23 our expectation, as detailed above. Furthermore we see a pattern emerging suggesting that the
24 difference between the two optimisations increases with higher observation correlation for the
25 same error magnitude. However, for some combinations of observation error and correlation,
26 the pattern is opposite to what we expect (red values), particularly for the s_1 data stream in the
27 simple C model (Figure 7b). This is likely because the accuracy of the solution becomes
28 limited by observation uncertainty at higher observation errors, and also due to presence of
29 model non-linearity, which prevents a fully accurate characterisation of the posterior error
30 covariance matrix.

31 The key finding of this preliminary investigation into correlated observation errors is that
32 it becomes increasingly important to properly characterise and account for correlations

1 between data streams if the observations do not contain enough information (i.e. too few
2 observations or high observation uncertainty). However, this is a wide topic that has received
3 little-to-no attention in the carbon cycle data assimilation literature to date, aside from the 2
4 out of 21 experiments in the wider-ranging study of Trudinger et al. (2007). We therefore
5 suggest that an investigation such as this should be extended in order to fully understand the
6 impact of cross-correlation between data streams; however, this is beyond the scope of this
7 paper.

8

9

10 **4 Perspectives and advice for Land Surface Modellers**

11 Although it is clear that in many cases, increasing the number of different observations in
12 a model optimisation provides additional constraints, challenges remain that need to be
13 addressed. Many of the issues that we have investigated are relevant to any data assimilation
14 study, including those only using one data stream. However, most are more pertinent when
15 considering more than one source of data. Based on the simple toy model results presented
16 here, in addition to lessons learned from existing studies, we recommend the following points
17 when carrying out multiple data stream carbon cycle data assimilation experiments:

18 • If technical constraints require that a step-wise approach be used, it is preferable
19 (from a mathematical standpoint) to propagate the full parameter error covariance
20 matrix between each step. Furthermore, it is important to check that the order of
21 assimilation of observations does not affect the final posterior parameter values,
22 and that the fit to the observations included in the previous steps is not degraded
23 after the final step (e.g. Peylin et al., 2016).

24 • Devote time to carefully characterising the parameter and observation error
25 covariance matrices, including their correlations (Raupach et al., 2005), although
26 we appreciate this is not an easy task (but see Kuppel et al., 2013 for practical
27 solutions). In the context of multiple data stream assimilation, this should include
28 the correlation between different data streams, particularly with higher
29 observational uncertainty, though note that this is not possible in a step-wise
30 assimilation.

- 1 • The presence of a bias in a data stream, or an incompatibility between the
2 observations and the model, will hinder the use of multiple observation types in an
3 assimilation framework. Therefore it is imperative to analyse and correct for biases
4 in the observations and to determine if there is an incompatibility between the
5 model and data. Alternatively, it may be possible account for any possible
6 bias/inconsistency in the observation error covariance matrix, \mathbf{R} , using the off-
7 diagonal terms or inflated errors (Chevallier, 2007), or by using the prior model-
8 data RMSE to define the observation uncertainty.
- 9 • Most optimisation studies with a large-scale LSM require the use of derivative-
10 based algorithms based on a least-squares formulation of the cost function, and
11 therefore rely on assumptions of Gaussian error distributions and quasi model
12 linearity. However, if these assumptions are not met it may not be possible to
13 find the true global minimum of the cost function and the characterisation of the
14 posterior probability distribution will be incorrect. This is a particular problem if
15 the posterior parameter error covariance matrix is then propagated in a step-wise
16 approach, although these issues are relevant to both step-wise and simultaneous
17 assimilation. Therefore it is important to assess the non-linearity of your model,
18 and if the model is strongly non-linear, use a global search algorithm for the
19 optimisation – although at the resolution of typical LSM simulations ($\geq 0.5 \times 0.5^\circ$)
20 this will likely only be computationally feasible at site or multi-site scale.

21

22 In addition to the above points, we have investigated the impact of a difference in the
23 number of observations in each data stream in this study. Test case 3b, in which only one
24 observation was included for the s_2 data stream instead of the complete time-series, shows
25 that a substantial difference in number of observations between the data streams can influence
26 the resulting parameter values and posterior uncertainty (compare test cases 3a and b in Fig. 2
27 for the simple C model and Fig. 4 for the non-linear toy model) as each data stream will have
28 a different overall “weight” in the cost function. Xu et al. (2006), among others, have
29 mentioned the possible need to weight the cost function for different data sets. Different
30 arguments abound on this issue. Some contend that the cost function should not be weighted
31 by the number of observations because the error covariance matrices (\mathbf{B} and \mathbf{R}) already define
32 this weight in an objective way (e.g. Keenan et al., 2013), and we would agree with this

1 assertion. It should not be necessary to weight by the number of observations in the cost
2 function if there is sufficient information to properly build the prior error covariance matrices
3 (**B** and **R**).

4 It is always useful to investigate the issues such as those discussed here by performing
5 synthetic experiments with pseudo observations, as in this study, to understand the possible
6 constraint brought by different data streams, and the impact of a possible bias and observation
7 or observation–model inconsistency. Note also that performing a number of tests starting
8 from different random “first guess” points in parameter space can help to diagnose if the
9 global minimum has been reached, as outlined in Section 3.1.6 and discussed at the beginning
10 of the results (Section 3.2). Furthermore, several diagnostic tests exist to help infer the
11 relative level of constraint brought about by different data streams, including the observation
12 influence and degrees of freedom of signal metrics (Cardinali et al., 2004). Performing these
13 tests was beyond the scope of this study, particularly given that the simple toy models
14 contained so few parameters, but such tests may be instructive when optimising many
15 hundreds of parameters in a large-scale LSM with a number of different data streams.

16 Aside from multiple data stream assimilation, other promising directions could also be
17 considered to help constrain the problem of lack of information in resolving the parameter
18 space within a data assimilation framework, including the use of other ecological and
19 dynamical “rules” that limit the optimisation (see for example Bloom and Williams, 2015), or
20 the addition of different timescales of information extracted from the data such as annual
21 sums (e.g. Keenan et al., 2012). Finally we should also seek to develop collaborations with
22 researchers in other fields who may have advanced further in a particular direction. Members
23 of the atmospheric and hydrological modelling communities, for example, have implemented
24 techniques for inferring the properties of the prior error covariance matrices, including the
25 mean and variance, but also potential biases, autocorrelation and heteroscedasticity, by
26 including these terms as “hyper-parameters” within the inversion (e.g. Michalak et al. 2005;
27 Evin et al., 2014; Renard et al., 2010; Wu et al. 2013). Of course this extends the parameter
28 space – making the problem harder to solve unless sufficient prior information is available
29 (Renard et al., 2010), but such avenues are worth exploring.

30

1 **5 Conclusions**

2 In this study we have attempted to highlight and discuss some of the challenges
3 associated with using multiple data streams to constrain the parameters of LSMs, with a
4 particular focus on the carbon cycle. We demonstrated some of the issues using two simple
5 models constrained with synthetic observations for which the ‘true’ parameters are known.
6 We performed a variety of tests in Section 2 to demonstrate the differences between
7 assimilating each data stream separately, sequentially (in a step-wise approach) and together
8 in the same assimilation (simultaneous approach). In particular we focused on difficulties that
9 may arise in the presence of biases or inconsistencies between the data and the model, as well
10 as non-linearity in the model equations. In Section 3 we discussed the experimental results
11 with reference to similar difficulties that have been documented in recent C cycle assimilation
12 studies.

13 Many of the issues faced are inherent to all optimisation experiments, including those in
14 which only one data stream is used. It is of utmost importance to determine if the
15 observations contain biases, and/or if inconsistencies or incompatibilities exist between the
16 model and the observations, and to correct for this or properly account for this in the error
17 covariance matrices. We further note that the consequence of not accounting for cross-
18 correlation between data streams in the prior error covariance matrix becomes more critical
19 with higher observation uncertainty. Secondly it is crucial to understand the assumptions and
20 limitations related to the inversion algorithm used. Without these two points being met, there
21 is a greater risk of obtaining incorrect parameter values, which may not be obvious by
22 examining the posterior uncertainty and model-data RMSE reduction. Furthermore it is more
23 likely that the implementation of a step-wise versus simultaneous approach will lead to
24 different results.

25 This study was not able to examine an exhaustive list of all possible challenges that may
26 be faced when assimilating multiple data streams, but we hope that this tutorial style paper
27 will serve as a guide for those wishing to optimise the parameters of LSMs using the variety
28 of C cycle related observations that are available today. Furthermore we hope that by
29 increasing awareness about the possible difficulties of model-data integration, we can further
30 bring the modelling and experimental communities together to work more closely on these
31 issues.

32

1 **Code availability**

2 The model and inversion code will be made available via the ORCHIDAS website (upon
3 registration): https://orchidas.lsce.ipsl.fr/multi_data_stream.php.

4

5 **Acknowledgements**

6 We acknowledge the support from the International Space Science Institute (ISSI). This
7 publication is an outcome of the ISSI's Working Group on "Carbon Cycle Data Assimilation:
8 How to Consistently Assimilate Multiple Data Streams". N. MacBean was also funded by the
9 GEOCARBON Project (ENV.2011.4.1.1-1-283080) within the European Union's 7th
10 Framework Programme for Research and Development. The authors wish to thank colleagues
11 and collaborators in the atmospheric inversion and C cycle DA communities with whom they
12 have had numerous past conversations that have led to an improvement in their understanding
13 of the issues presented here.

14

1 **References**

- 2 Alton, P. B.: From site-level to global simulation: Reconciling carbon, water and energy
3 fluxes over different spatial scales using a process-based ecophysiological land-surface
4 model, *Agric. For. Meteorol.*, 176, 111–124, doi:10.1016/j.agrformet.2013.03.010, 2013.
- 5 Anav, A., Friedlingstein, P., Kidston, M., Bopp, L., Ciais, P., Cox, P., Jones, C., Jung, M.,
6 Myneni, R. and Zhu, Z.: Evaluating the land and ocean components of the global carbon cycle
7 in the CMIP5 earth system models, *J. Clim.*, 26(18), 6801–6843, doi:10.1175/JCLI-D-12-
8 00417.1, 2013.
- 9 Bacour, C., Peylin, P., MacBean, N., Rayner, P. J., Delage, F., Chevallier, F., Weiss, M.,
10 Demarty, J., Santaren, D., Baret, F., Berveiller, D., Dufrêne, E. and Prunet, P.: Joint
11 assimilation of eddy covariance flux measurements and FAPAR products over temperate
12 forests within a process-oriented biosphere model, *J. Geophys. Res. Biogeosciences*, 120,
13 1839–1857, doi:10.1002/2015JG002966. Received, 2015.
- 14 Barrett, D. J., Michael J Hill, I., Hutley, L. B., Beringer, J., Xu, J. H., Cook, G. D., Carter, J.
15 O. and Williams, R. J.: Prospects for improving savanna biophysical models by using
16 multiple-constraints model-data assimilation methods, *Aust. J. Bot.*, 53(7), 689–714,
17 doi:10.1071/BT04139, 2005.
- 18 Bloom, A. A. and Williams, M.: Constraining ecosystem carbon dynamics in a data-limited
19 world: integrating ecological ‘common sense’ in a model–data fusion framework,
20 *Biogeosciences*, 12(5), 1299–1315, doi:10.5194/bg-12-1299-2015, 2015.
- 21 Cardinali, C., S. Pezzulli, E. Andersson (2004), Influence-matrix diagnostic of a data
22 assimilation system, *Q. J. R. Meteorol. Soc.*, 130: 2767–2786, doi: 10.1256/qj.03.205
- 23 Chevallier, F., 2007: Impact of correlated observation errors on inverted CO₂ surface fluxes
24 from OCO measurements, *Geophys. Res. Lett.*, 34, L24804, doi:10.1029/2007GL030463.
- 25 Dufresne, J. L., Foujols, M. a., Denvil, S., Caubel, a., Marti, O., Aumont, O., Balkanski, Y.,
26 Bekki, S., Bellenger, H., Benschila, R., Bony, S., Bopp, L., Braconnot, P., Brockmann, P.,
27 Cadule, P., Cheruy, F., Codron, F., Cozic, a., Cugnet, D., de Noblet, N., Duvel, J. P., Ethé,
28 C., Fairhead, L., Fichet, T., Flavoni, S., Friedlingstein, P., Grandpeix, J. Y., Guez, L.,
29 Guilyardi, E., Hauglustaine, D., Hourdin, F., Idelkadi, a., Ghattas, J., Jousaume, S.,
30 Kageyama, M., Krinner, G., Labetoulle, S., Lahellec, a., Lefebvre, M. P., Lefevre, F., Levy,

1 C., Li, Z. X., Lloyd, J., Lott, F., Madec, G., Mancip, M., Marchand, M., Masson, S.,
2 Meurdesoif, Y., Mignot, J., Musat, I., Parouty, S., Polcher, J., Rio, C., Schulz, M.,
3 Swingedouw, D., Szopa, S., Talandier, C., Terray, P., Viovy, N. and Vuichard, N.: Climate
4 change projections using the IPSL-CM5 Earth System Model: From CMIP3 to CMIP5., 2013.

5 Evin, G., Thyer, M., Kavetski, D., McInerney, D. and Kuczera, G.: Comparison of joint
6 versus postprocessor approaches for hydrological uncertainty estimation accounting for error
7 autocorrelation and heteroscedasticity, *Water Resour. Res.*, 50(3), 2350–2375,
8 doi:10.1002/2013WR014185, 2014.

9 Forkel, M., Carvalhais, N., Schaphoff, S., v. Bloh, W., Migliavacca, M., Thurner, M. and
10 Thonicke, K.: Identifying environmental controls on vegetation greenness phenology through
11 model-data integration, *Biogeosciences*, 11, 7025–7050, doi:10.5194/bg-11-7025-2014, 2014.

12 Gobron, N., Pinty, B., Ausedat, O., Chen, J. M., Cohen, W. B., Fensholt, R., Gond, V.,
13 Huemmrich, K. F., Lavergne, T., Mélin, F., Privette, J. L., Sandholt, I., Taberner, M., Turner,
14 D. P., Verstraete, M. M. and Widlowski, J. L.: Evaluation of fraction of absorbed
15 photosynthetically active radiation products for different canopy radiation transfer regimes:
16 Methodology and results using Joint Research Center products derived from SeaWiFS against
17 ground-based estimations, *J. Geophys. Res.*, 111, D13110, doi:10.1029/2005JD006511, 2006.

18 Gobron, N., Pinty, B., Ausedat, O., Taberner, M., Faber, O., Mélin, F., Lavergne, T.,
19 Robustelli, M. and Snoeij, P.: Uncertainty estimates for the FAPAR operational products
20 derived from MERIS - Impact of top-of-atmosphere radiance uncertainties and validation
21 with field data, *Remote Sens. Environ.*, 112(4), 1871–1883, doi:10.1016/j.rse.2007.09.011,
22 2008.

23 Kaminski, T., Knorr, W., Scholze, M., Gobron, N., Pinty, B., Giering, R. and Mathieu, P. P.:
24 Consistent assimilation of MERIS FAPAR and atmospheric CO₂ into a terrestrial vegetation
25 model and interactive mission benefit analysis, *Biogeosciences*, 9(8), 3173–3184,
26 doi:10.5194/bg-9-3173-2012, 2012.

27 Kato, T., Knorr, W., Scholze, M., Veenendaal, E., Kaminski, T., Kattge, J. and Gobron, N.:
28 Simultaneous assimilation of satellite and eddy covariance data for improving terrestrial water
29 and carbon simulations at a semi-arid woodland site in Botswana, *Biogeosciences*, 10(2),
30 789–802, doi:10.5194/bg-10-789-2013, 2013.

31 Keenan, T. F., Davidson, E., Moffat, A. M., Munger, W. and Richardson, A. D.: Using

1 model-data fusion to interpret past trends, and quantify uncertainties in future projections, of
2 terrestrial ecosystem carbon cycling, *Glob. Chang. Biol.*, 18(8), 2555–2569,
3 doi:10.1111/j.1365-2486.2012.02684.x, 2012.

4 Keenan, T. F., Davidson, E. a., Munger, J. W. and Richardson, A. D.: Rate my data:
5 Quantifying the value of ecological data for the development of models of the terrestrial
6 carbon cycle, *Ecol. Appl.*, 23(1), 273–286, doi:10.1890/12-0747.1, 2013.

7 Knorr, W.: Annual and interannual CO₂ exchanges of the terrestrial biosphere: process-based
8 simulations and uncertainties, *Glob. Ecol. Biogeogr.*, 9(3), 225–252, doi:10.1046/j.1365-
9 2699.2000.00159.x, 2000.

10 Krinner, G., Viovy, N., de Noblet-Ducoudré, N., Ogée, J., Polcher, J., Friedlingstein, P.,
11 Ciais, P., Sitch, S. and Prentice, I. C.: A dynamic global vegetation model for studies of the
12 coupled atmosphere-biosphere system, *Global Biogeochem. Cycles*, 19(1), 1–33,
13 doi:10.1029/2003GB002199, 2005.

14 Kuppel, S., F. Chevallier and P. Peylin,: Quantifying the model structural error in Carbon
15 Cycle Data Assimilation Systems. *Geosci. Model Dev.*, 6, 45-55, doi:10.5194/gmd-6-45-
16 2013, 2013.

17 Michalak, A. M., Hirsch, A., Bruhwiler, L., Gurney, K. R., Peters, W. and co-authors:
18 Maximum likelihood estimation of covariance parameters for Bayesian atmospheric trace gas
19 surface flux inversions. *J. Geophys. Res.* 110, D24107. DOI: 10.1029/2005JD005970, 2005.

20 Morcrette, J.-J.: Evaluation of Model-generated Cloudiness: Satellite-observed and Model-
21 generated Diurnal Variability of Brightness Temperature. *Mon. Wea. Rev.*, **119**, 1205–1224,
22 1991.

23 van Oijen, M., Rougier, J. and Smith, R.: Bayesian calibration of process-based forest models:
24 bridging the gap between models and data, *Tree Physiol.*, 25(7), 915–927,
25 doi:10.1093/treephys/25.7.915, 2005.

26 Peylin, P., Bacour, C., MacBean, N., Leonard, S., Rayner, P. J., Kuppel, S., Koffi, E. N.,
27 Kane, A., Maignan, F., Chevallier, F., Ciais, P., and Prunet, P.: A new step-wise Carbon
28 Cycle Data Assimilation System using multiple data streams to constrain the simulated land
29 surface carbon cycle, *Geosci. Model Dev. Discuss.*, doi:10.5194/gmd-2016-13, in review,
30 2016.

1 Quaife, T., Lewis, P., De Kauwe, M., Williams, M., Law, B. E., Disney, M. and Bowyer, P.:
2 Assimilating canopy reflectance data into an ecosystem model with an Ensemble Kalman
3 Filter, *Remote Sens. Environ.*, 112(4), 1347–1364, doi:10.1016/j.rse.2007.05.020, 2008.

4 Raupach, M. R.: Dynamics of resource production and utilisation in two-component
5 biosphere-human and terrestrial carbon systems, *Hydrol. Earth Syst. Sci.*, 11, 875–889,
6 doi:10.5194/hess-11-875-2007, 2007.

7 Raupach, M. R., Rayner, P. J., Barrett, D. J., Defries, R. S., Heimann, M., Ojima, D. S.,
8 Quegan, S. and Schimmlus, C. C.: Model-data synthesis in terrestrial carbon observation:
9 Methods, data requirements and data uncertainty specifications, *Glob. Chang. Biol.*, 11(3),
10 378–397, doi:10.1111/j.1365-2486.2005.00917.x, 2005.

11 Rayner, P. J., Scholze, M., Knorr, W., Kaminski, T., Giering, R. and Widmann, H.: Two
12 decades of terrestrial carbon fluxes from a carbon cycle data assimilation system (CCDAS), ,
13 19, doi:10.1029/2004GB002254, 2005.

14 Renard, B., Kavetski, D., Kuczera, G., Thyer, M. and Franks, S. W.: Understanding predictive
15 uncertainty in hydrologic modeling: The challenge of identifying input and structural errors,
16 *Water Resour. Res.*, 46(5), 1–22, doi:10.1029/2009WR008328, 2010.

17 Richardson, A. D., Williams, M., Hollinger, D. Y., Moore, D. J. P., Dail, D. B., Davidson, E.
18 a., Scott, N. a., Evans, R. S., Hughes, H., Lee, J. T., Rodrigues, C. and Savage, K.: Estimating
19 parameters of a forest ecosystem C model with measurements of stocks and fluxes as joint
20 constraints, *Oecologia*, 164(1), 25–40, doi:10.1007/s00442-010-1628-y, 2010.

21 Schürmann, G. J., Kaminski, T., Köstler, C., Carvalhais, N., Voßbeck, M., Kattge, J., Giering,
22 R., Rödenbeck, C., Heimann, M., and Zaehle, S.: Constraining a land surface model with
23 multiple observations by application of the MPI-Carbon Cycle Data Assimilation System,
24 *Geosci. Model Dev. Discuss.*, doi:10.5194/gmd-2015-263, in review, 2016.

25 Scholze, M., T. Kaminski, W. Knorr, S. Blessing, M. Vossbeck, J.p. Grant, and K. Scipal.
26 "Simultaneous Assimilation of SMOS Soil Moisture and Atmospheric CO2 In-situ
27 Observations to Constrain the Global Terrestrial Carbon Cycle." *Remote Sensing of*
28 *Environment* 180, 334-45, 2016.

29 Sitch, S., Friedlingstein, P., Gruber, N., Jones, S. D., Murray-Tortarolo, G., Ahlström, A.,
30 Doney, S. C., Graven, H., Heinze, C., Huntingford, C., Levis, S., Levy, P. E., Lomas, M.,
31 Poulter, B., Viovy, N., Zaehle, S., Zeng, N., Arneeth, A., Bonan, G., Bopp, L., Canadell, J. G.,

- 1 Chevallier, F., Ciais, P., Ellis, R., Gloor, M., Peylin, P., Piao, S., Le Quéré, C., Smith, B.,
2 Zhu, Z. and Myneni, R.: Recent trends and drivers of regional sources and sinks of carbon
3 dioxide, *Biogeosciences*, 12, 653–679, doi:10.5194/bgd-12-653-2015, 2015.
- 4 Thum, T., N. MacBean, P. Peylin, C. Bacour, D. Santaren, B. Longdoz, D. Loustau and P.
5 Ciais, The potential benefit of using forest biomass data in addition to carbon and water flux
6 measurements to constrain ecosystem model parameters: case studies at two temperate forest
7 sites. In revision for *Agric. For. Meteorol.*
- 8 Trudinger, C. M., Raupach, M. R., Rayner, P. J., Kattge, J., Liu, Q., Park, B., Reichstein, M.,
9 Renzullo, L., Richardson, A. D., Roxburgh, S. H., Styles, J., Wang, Y. P., Briggs, P., Barrett,
10 D. and Nikolova, S.: OptIC project: An intercomparison of optimization techniques for
11 parameter estimation in terrestrial biogeochemical models, *J. Geophys. Res. Biogeosciences*,
12 112(2), doi:10.1029/2006JG000367, 2007.
- 13 Williams, M., Schwarz, P. a, Law, B. E., Irvine, J. and Kurpius, M. R.: An improved analysis
14 of forest carbon dynamics using data assimilation, *Glob. Chang. Biol.*, 11(1), 89–105,
15 doi:10.1111/j.1365-2486.2004.00891.x, 2005.
- 16 Wu, L., M. Bocquet, F. Chevallier, T. Lauvaux, and K. Davis, 2013: Hyperparameter
17 estimation for uncertainty quantification in mesoscale carbon dioxide inversions. *Tellus B*, 65,
18 doi:10.3402/tellusb.v65i0.20894.
- 19 Xu, T., White, L., Hui, D. and Luo, Y.: Probabilistic inversion of a terrestrial ecosystem
20 model: Analysis of uncertainty in parameter estimation and model prediction, *Global*
21 *Biogeochem. Cycles*, 20(2), 1–15, doi:10.1029/2005GB002468, 2006.
- 22 Zobitz, J. M., Moore, D. J. P., Quaipe, T., Braswell, B. H., Bergeson, A., Anthony, J. a. and
23 Monson, R. K.: Joint data assimilation of satellite reflectance and net ecosystem exchange
24 data constrains ecosystem carbon fluxes at a high-elevation subalpine forest, *Agric. For.*
25 *Meteorol.*, 195-196, 73–88, doi:10.1016/j.agrformet.2014.04.011, 2014.

26
27
28
29
30

1 Table 1: The optimisation set-up for both models, including the true parameter values, their
 2 range and the observation uncertainty (1 sigma), which was set to 10% of the mean value for
 3 each set of pseudo-observations derived from multiple first guesses of the model. The
 4 parameter uncertainty (1 sigma) was set to 40% of the range for each parameter.

5

Model	Parameter value (range)				Observation uncertainty	
Simple carbon model	p_1 1 (0.5,5)	p_2 1 (0.5,5)	k_1 0.2 (0.03,0.9)	k_2 0.1 (0.01,0.12)	s_1 0.5	s_2 5
Non-linear toy model	a 1 (0,2)		b 1 (0,2)		s_1 0.5	s_2 0.5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

1 Table 2: List of experiments performed for both models with synthetic data. All parameters
 2 are optimised in all cases (therefore in both steps for the step-wise approach).

3

Test case	Step 1	Step 2	Parameter error covariance terms propagated in step 2?
<i>Separate</i>			
1a	s_1	-	-
1b	s_2	-	-
<i>Step-wise</i>			
2a	s_1	s_2	yes
2b	s_1	s_2	no
2c	s_2	s_1	yes
2d	s_2	s_1	no
<i>Simultaneous</i>			
3a	s_1 and s_2	-	-
3b	s_1 and only 1 obs for s_2	-	-

4

5

6

7

8

9

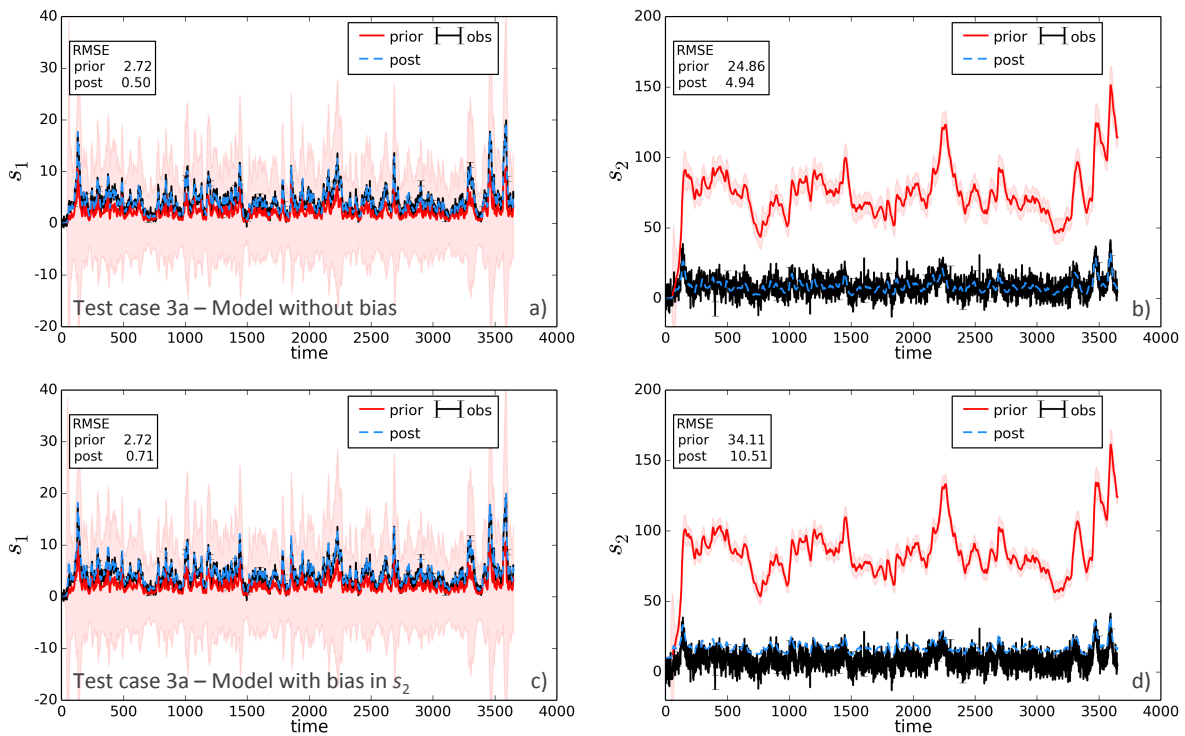
10

11

12

13

14



1

2

Figure 1: Prior and posterior model simulations compared to the synthetic observations for the simple carbon model for test case 3a for a) s_1 and b) s_2 simulations without any model bias, and c and d) with bias in the simulated s_2 variable. The coloured error band on the prior and posterior represents the propagated parameter uncertainty (1 sigma) on the model state variables (in the equivalent colour as the mean curve). This is mostly visible for the prior model simulation (pink band) as there is a high reduction in model uncertainty reduction as a result of the assimilation.

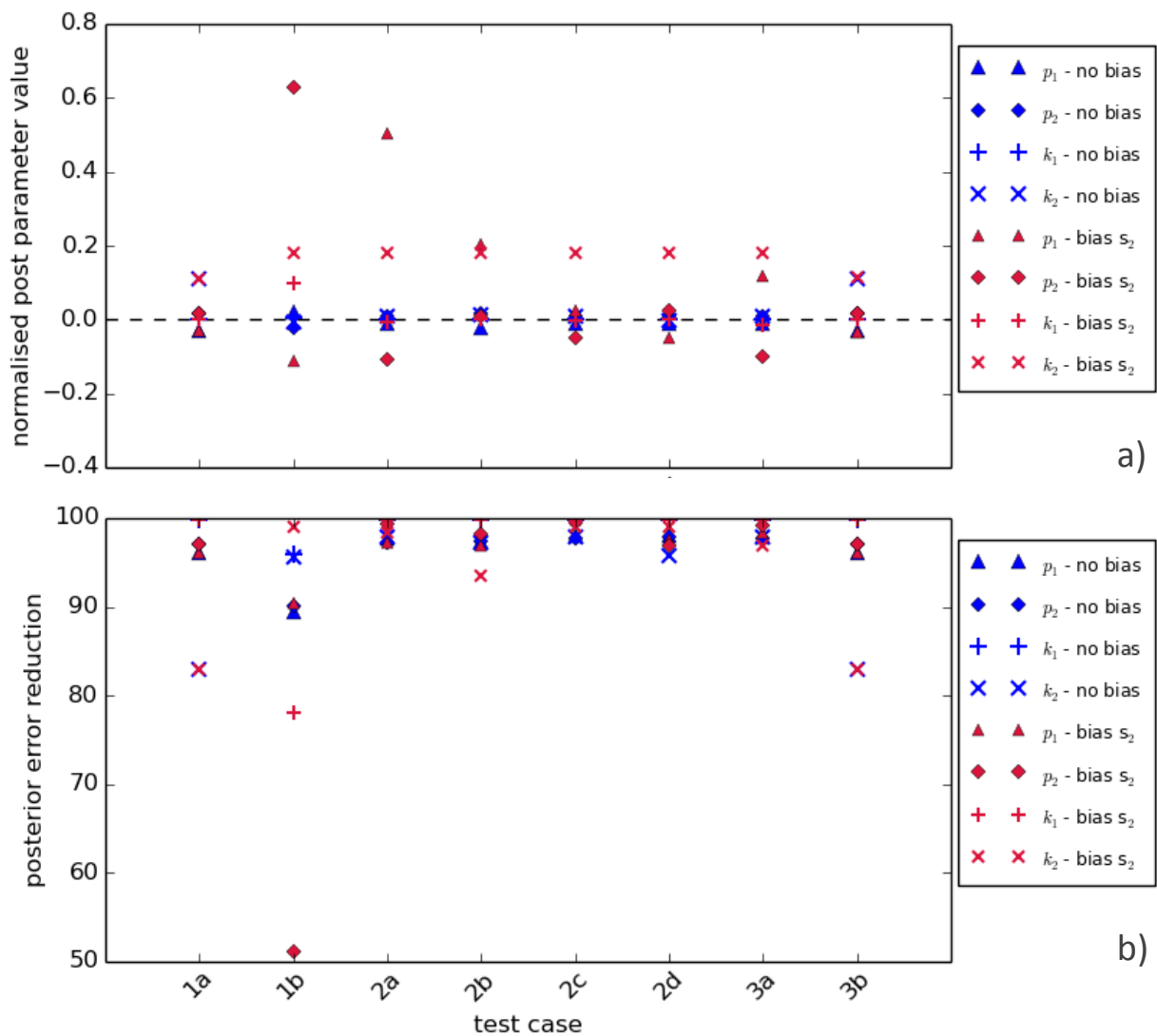
9

10

11

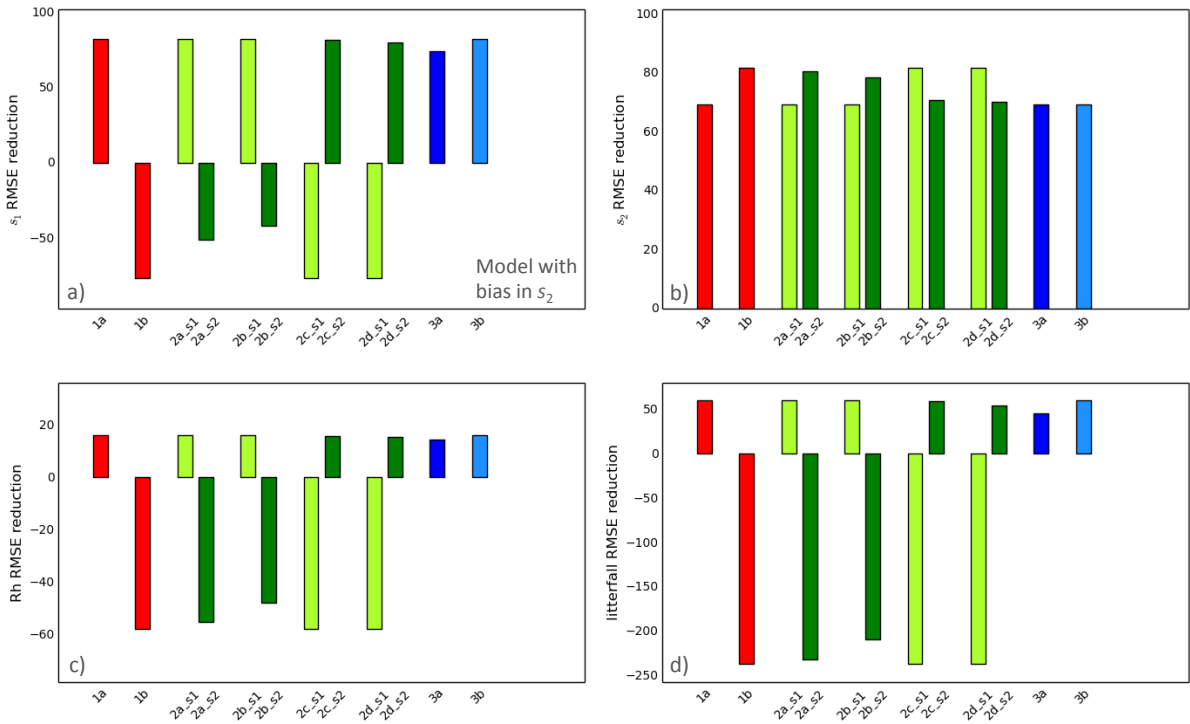
12

13



1
2
3
4
5
6
7
8
9
10
11
12
13

Figure 2: a) Normalised posterior parameter values and b) posterior parameter error reduction for all parameters of the simple carbon model for each test case, and for both the simulations with no bias (blue) and simulations with a bias in the s_2 variable that was not accounted for in the inversion (red). In a) parameters values were normalised to account for differences in the magnitude of the different parameters and their range, thus it is a measure of the distance from the true value as a fraction of the range and is calculated as: (posterior value – true value / max parameter value – minimum parameter value). The closer the value to the zero dashed line represents a better match to the “true” parameter value. To give an indication of the optimisation performance, the following are the normalised first guess parameter values for this particular example test (compare with posterior values in Fig. 2a): p_1 0.09, p_2 0.29, k_1 0.1, k_2 0.15.



1

2 Figure 3: Reduction in RMSE for all test cases for simulations with a bias in the s_2 variable: a)
 3 s_1 , b) s_2 , c) litterfall and d) heterotrophic respiration (Rh). For the step-wise cases (2a, b, c and
 4 d) the reduction after both step 1 and step 2 are shown in light and dark green respectively,
 5 and are denoted in the x-axis labels with ‘_s1’ for step 1 and ‘_s2’ for step 2. The reduction
 6 (in %) is calculated as $1 - (\text{RMSE}_{\text{post}} / \text{RMSE}_{\text{prior}})$.

7

8

9

10

11

12

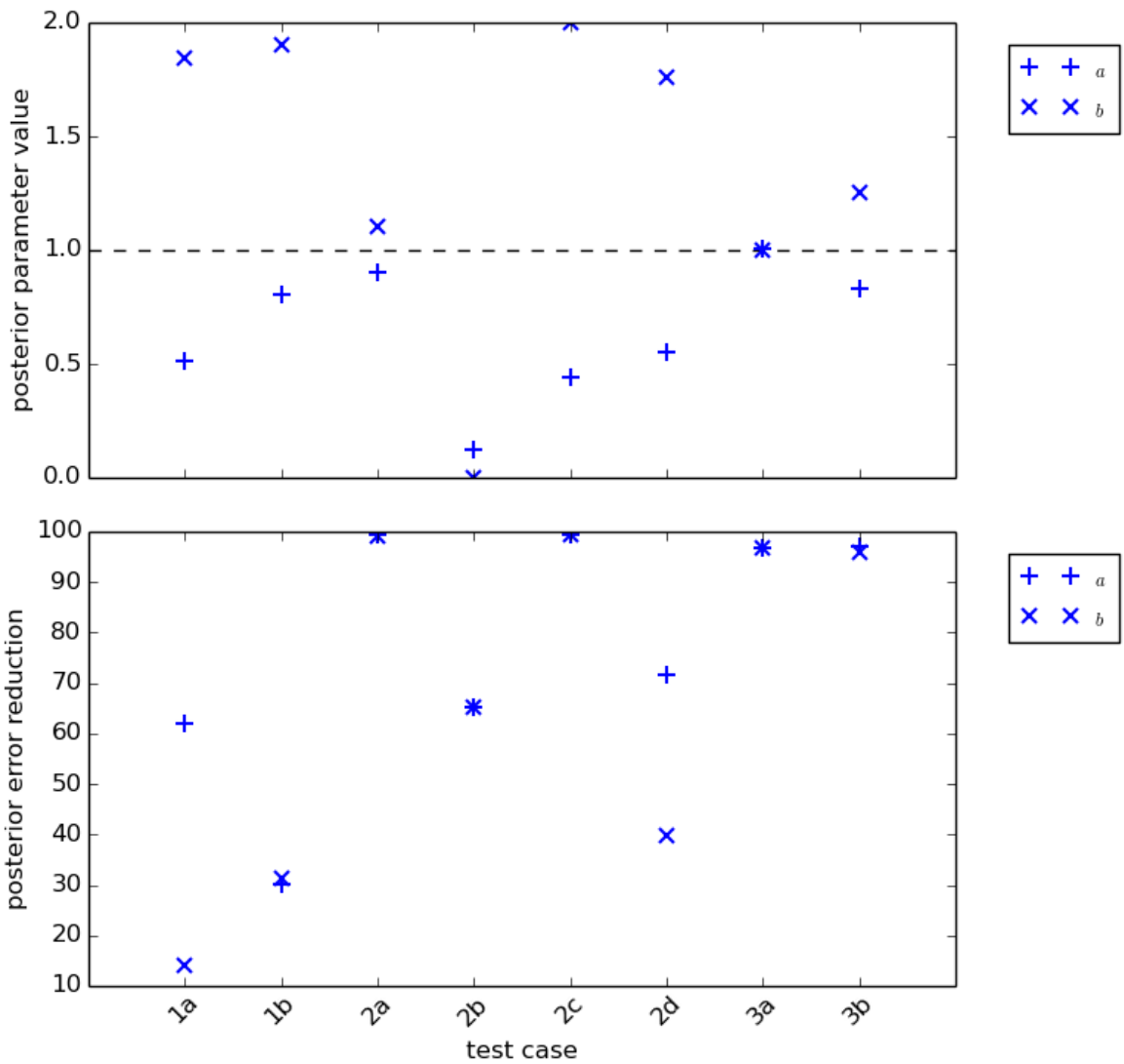
13

14

15

16

17



1

2 Figure 4: Posterior parameter values of both the non-linear toy model a and b parameters for
 3 each test case for the simulations with no model bias. The y-axis range corresponds to the
 4 parameter bounds and the dashed horizontal line represents the “true” known value of both
 5 parameters. To give an indication of the optimisation performance, the following are the first
 6 guess parameter values for this particular example test (compare with posterior values in Fig.
 7 4a): a 0.87, b 1.98. b) Posterior uncertainty reduction for both parameters for all test cases.

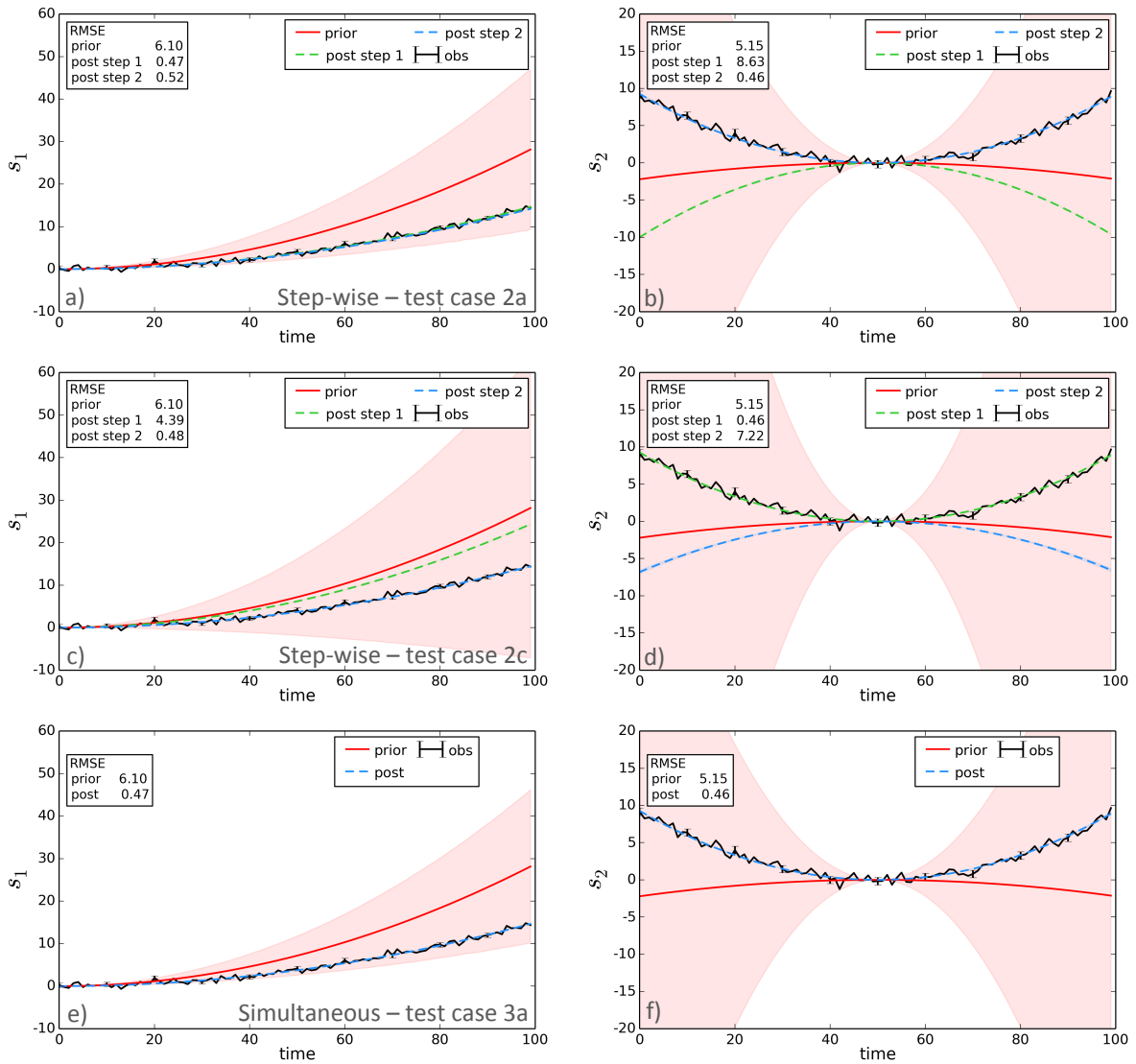
8

9

10

11

12



1

2

3

4

5

6

7

8

9

10

11

Figure 5: Prior and posterior model simulations compared to the synthetic observations for the non-linear toy model (with no bias) for both the s_1 (left column) and s_2 (right column) variables for a) and b) test case 2a (1st row) – step-wise approach with s_1 observations assimilated in the first step, followed by the s_2 observations in the second step; c) and d) test case 2c (2nd row) – step-wise approach with s_2 observations assimilated in the first step, followed by s_1 observations in the second step; and e) and f) test case 3a (3rd row) – the simultaneous case in which both data streams were included. For both step-wise examples A_1 was propagated between the 1st and 2nd steps. The coloured error band on the prior and posterior represents the propagated parameter uncertainty (1 sigma) on the model state variables (in the equivalent colour as the mean curve). This is mostly visible for the prior

1 model simulation (pink band) as there is a high reduction in model uncertainty reduction as a
2 result of the assimilation.

3

4

5

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

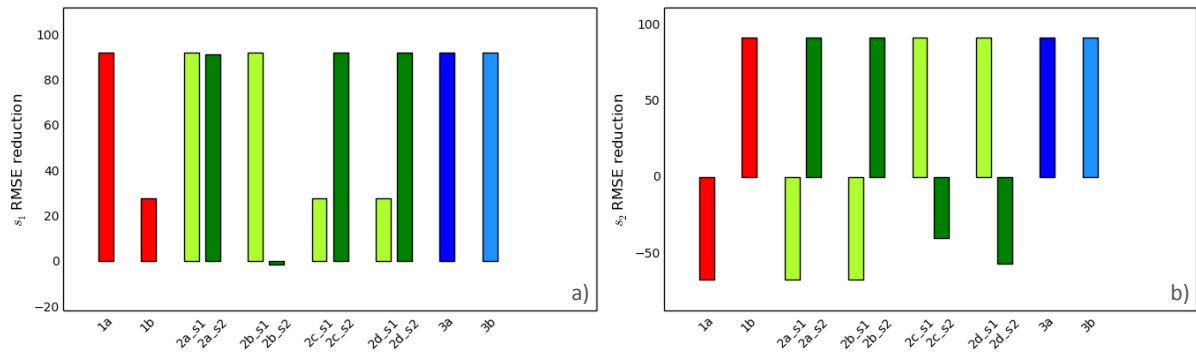
22

23

24

25

26



1

2 Figure 6: Reduction in RMSE for all test cases for both a) s_1 and b) s_2 variables for the non-
 3 linear toy model simulations with no model bias. For the step-wise cases (2a, b, c and d) the
 4 reduction after both step 1 and step 2 are shown in light and dark green respectively, and are
 5 denoted in the x-axis labels with ‘_s1’ for step 1 and ‘_s2’ for step 2. The reduction (in %) is
 6 calculated as $1 - (\text{RMSE}_{\text{prior}} / \text{RMSE}_{\text{post}})$.

7

8

9

10

11

12

13

14

15

16

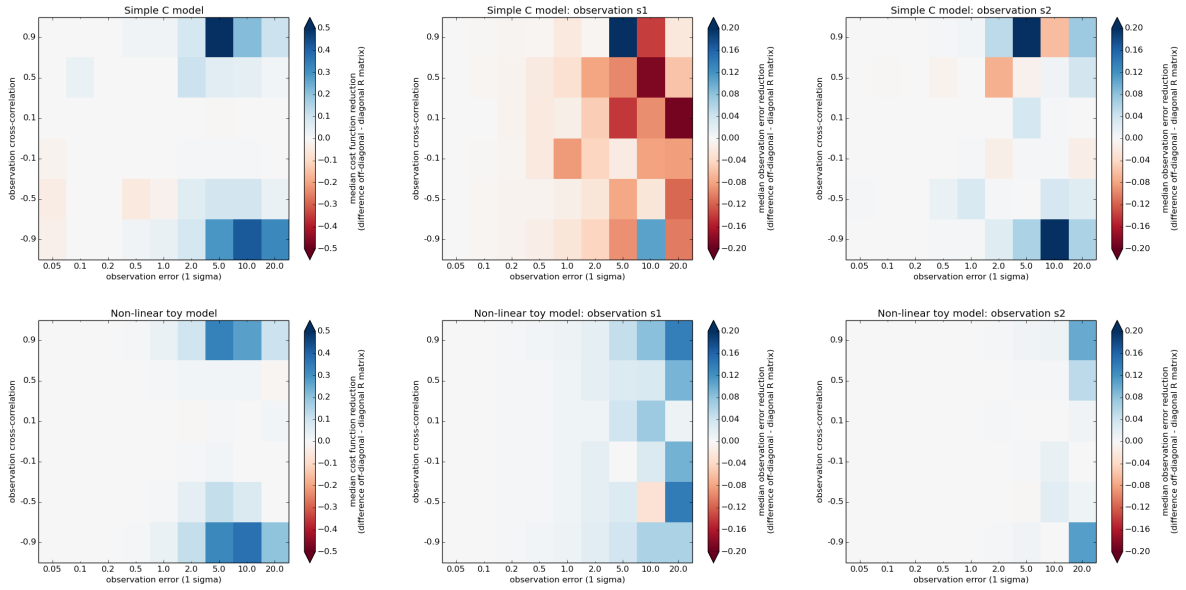
17

18

19

20

21



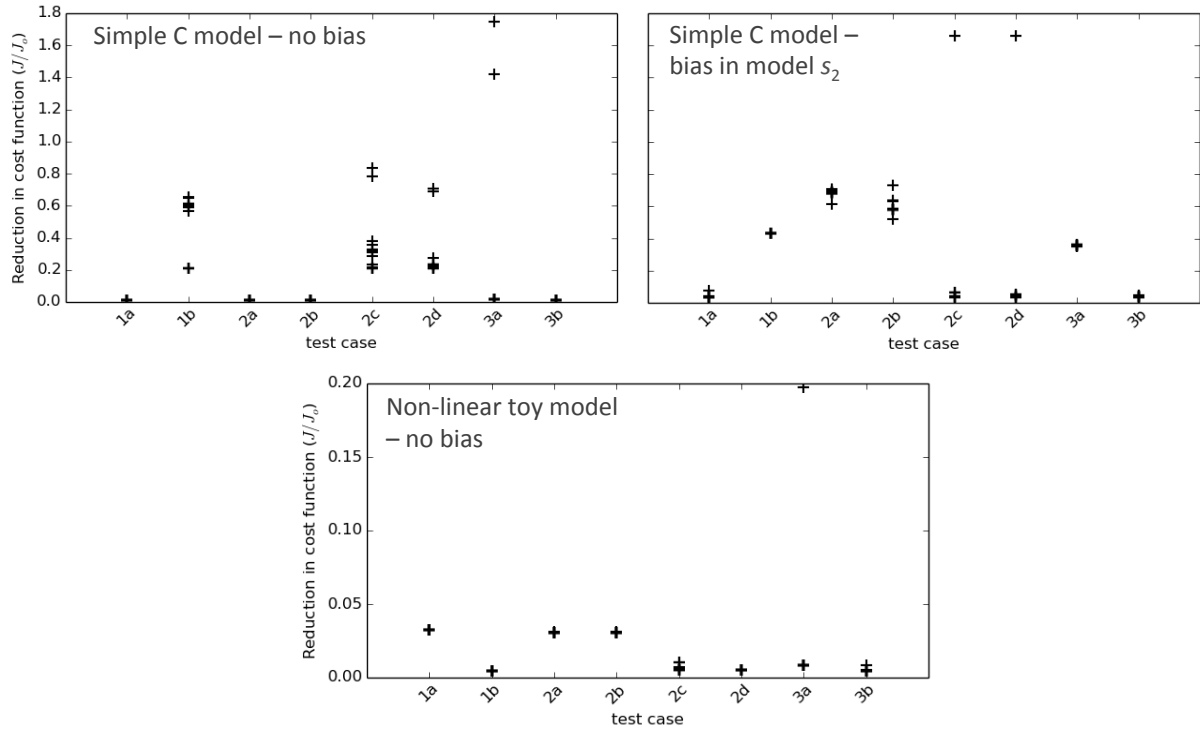
1

2 Figure 7: Median difference (across 20 first guess parameters) between including correlated
 3 observation errors in the R matrix (off-diagonal elements) minus ignoring the correlated
 4 observation errors (keeping R diagonal) for the reduction cost function (a and d: left column)
 5 and the reduction in s_1 and s_2 observation errors (b, c, e and f: middle and right columns), for
 6 both the simple C model (a, b and c: top row) and the non-linear toy model (d, e and f: bottom
 7 row). The reduction is calculated as $1 - (\text{posterior}/\text{prior})$.

8

9

1 Supplementary material



2

3 Figure S1: Reduction in the cost function (J/J_0) for each model and each test for all 20
4 assimilations with different random “first guess” points in the parameter space (i.e. each cross
5 represents the 20 random “first guess” tests). Top panel – simple C model without bias (left)
6 and with bias added to the simulated s_2 variable (right). Bottom panel – non-linear toy model
7 with no added bias. Note that the majority of the random “first guess” assimilations achieve
8 the same reduction in the cost function even though the final value is different for each test,
9 which is to be expected as each test (for each model) has a different cost function.