FORest Canopy Atmosphere Transfer (FORCAsT) 1.0: a 1-D

2 model of biosphere-atmosphere chemical exchange

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20 Abstract

21 Biosphere-atmosphere interactions play a critical role in governing atmospheric composition, mediating the concentration of key species such as ozone and aerosol, thereby influencing air 22 23 quality and climate. The exchange of reactive trace gases and their oxidation products (both 24 gas and particle phase) is of particular importance in this process. The FORCAsT (FORest 25 Canopy AtmoSphere Transfer) one-dimensional model is developed to study the emission, 26 deposition, chemistry and transport of volatile organic compounds (VOCs) and their oxidation 27 products in the atmosphere within and above the forest canopy. We include an equilibrium partitioning scheme, making FORCAsT one of the few canopy models currently capable of 28 29 simulating the formation of secondary organic aerosols (SOA) from VOC oxidation in a

forest environment. We evaluate the capability of FORCAsT to reproduce observed 1 2 concentrations of key gas-phase species and report modeled SOA concentrations within and above a mixed forest at the University of Michigan Biological Station (UMBS) during the 3 4 Community Atmosphere-Biosphere Interactions Experiment (CABINEX) field campaign in 5 summer 2009. We examine the impact of two different gas-phase chemical mechanisms on modelled concentrations of short-lived primary emissions, such as isoprene and 6 7 monoterpenes, and their oxidation products. While the two chemistry schemes perform 8 similarly under high-NO_x conditions, they diverge at the low levels of NO_x at UMBS. We 9 identify peroxy radical and alkyl nitrate chemistry as the key causes of the differences, 10 highlighting the importance of this chemistry in understanding the fate of biogenic VOCs 11 (bVOCs) for both the modelling and measurement communities.

12 **1** Introduction

Exchanges of energy and mass between the biosphere and atmosphere play a crucial role in 13 14 the Earth system. These interactions control the physical and chemical properties of the atmosphere, which in turn influence the characteristics of the land surface and ecosystems. 15 The biogeophysical and biogeochemical feedbacks initiated by these interactions are known 16 17 to mediate climate on both the global and local scale through their role in the hydrological, 18 and coupled carbon and nitrogen cycles (e.g. Grace et al., 2006; Pongratz et al., 2010; 19 Friedlingston and Prentice, 2010). These large-scale effects are generally included in 20 atmospheric chemistry and transport models and Earth system models. However, exchanges between the terrestrial biosphere and the atmosphere also include fluxes of many chemical 21 22 species with relatively short atmospheric lifetimes (of the order of fractions of a second to a 23 few days) and atmospheric concentrations measured in parts per billion (ppb) or less. In spite 24 of their relatively low concentrations, these trace gases (ozone, volatile organic compounds (VOCs), nitrogen oxides (NOx), and hydrogen oxides (HOx)) and aerosols can govern 25 26 atmospheric composition on both short timescales (i.e. days to weeks), affecting air quality, and longer timescales, impacting climate (e.g. Mellouki et al., 2015; Laothawornkitkul et al., 27 2009). 28

While the dominant shorter-lived species are included in atmospheric and Earth system models, their exchange between the land surface and the lowest atmospheric model layer are often treated in a simplified manner. Specifically, the interface between the land surface and the atmosphere in these models is essentially 2-dimensional, with mass typically injected into the atmosphere at the mid-point height of the lowest model level. While the modelled injection rates are usually dependent on the land cover or ecosystem classification at any given location, which take into account a generic surface roughness and leaf area index, there is no explicit consideration of coupling mechanisms between the land and atmosphere.

In reality, biosphere-atmosphere interactions take place in a dynamic rapidly changing bi-5 6 directional equilibrium within the canopy structure of the vegetation, where physical and chemical conditions can be very different from those in the atmosphere above and can change 7 8 on very short timescales. The potential importance of the individual processes occurring in 9 this space on both the atmosphere and the land surface has prompted a recent focus on the development and application of small-scale or single point models that explicitly consider the 10 canopy space and its processes (e.g. CACHE, Forkel et al. (2006), Bryan et al. (2012); 11 SOSA(A), Boy et al. (2011), Zhou et al. (2014); CAFE, Wolfe and Thornton (2011); MLC-12 13 Chem, Ganzeveld et al. (2002); ACCESS, Saylor (2013)). These models range in complexity 14 in terms of both vertical resolution and the chemical and physical mechanisms that are 15 included.

16 Here, we describe the canopy model FORCAsT (FORest Canopy-Atmosphere Transfer) 17 which has been developed from the original Canopy Atmospheric CHemistry Emission (CACHE) model (Forkel et al., 2006; Bryan et al., 2012). Major updates from CACHE 18 19 include: 1) adding the CACM (Caltech Atmospheric Chemistry Mechanism) gas-phase 20 chemistry scheme (Griffin et al., 2002; 2005; Chen and Griffin, 2005); 2) restructuring the 21 code to facilitate switching between chemistry mechanisms using codes generated by the Kinetic PreProcessor (KPPv2.1; Sandu and Sander, 2006); and 3) incorporation of the 22 23 MPMPO (Model to Predict the Multiphase Partitioning of Organics) aerosol module (Griffin 24 et al., 2003; 2005; Chen and Griffin, 2005). We evaluate FORCAsT's performance against its predecessor, the CACHE model, and observations from the CABINEX intensive field 25 26 campaign, conducted at the University of Michigan Biological Station (UMBS) during the summer of 2009. 27

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29 2 Model Description

The canopy exchange model FORCAsT is a single column (1-D) model incorporating both atmospheric chemistry and dynamics and land surface modelling, based on the CACHE canopy exchange model (Forkel et al., 2006). Energy balances and radiative transfer within the canopy are calculated following the algorithms of the CUPID soil-plant-atmosphere
 model (Norman 1979; Norman and Campbell, 1983).

From the atmospheric perspective, FORCAsT includes parameterisations of all of the processes occurring within and above the canopy space: emissions, advection, deposition, turbulent (vertical) exchange, and chemical production and loss (Figure 1b). One of the novel aspects of FORCAsT is that it includes both the gas-phase chemistry and subsequent partitioning of condensable species to the particle-phase, while the majority of canopy models consider only the gas-phase.

9 Fluxes of energy and mass are simulated by solving the continuity equations for energy and10 mass:

11 Heat (energy):
$$\frac{\partial T}{\partial t} = \frac{\partial}{\partial z} \left(K \frac{\partial T}{\partial z} \right) + S_h \tag{1}$$

12 where *T* is air temperature (K), *K* is the turbulent exchange coefficient (m² s⁻¹) and S_h 13 represents sources and sinks of heat (K s⁻¹).

14 Mass (gas-phase): $\frac{\partial c}{\partial t} = \frac{\partial}{\partial z}$

$$\frac{\partial c}{\partial t} = \frac{\partial}{\partial z} \left(K \frac{\partial c}{\partial z} \right) + S_c + C \tag{2}$$

where *c* is the concentration or mixing ratio of a chemical species, S_c represents all sources and sinks (i.e. emissions, deposition, and advection) of water vapour or chemical compound (s⁻¹), and *C* is chemical production or loss (s⁻¹).

18 Mass transfer of aerosols is modelled as for the gas-phase with an additional term accounting19 for gravitational settling of the aerosols.

- 20 Mass (aerosols): $\frac{\partial c}{\partial t} = \frac{\partial}{\partial z} \left(K \frac{\partial c}{\partial z} + V_S C \right) + S_c + C$ (3)
- 21 where V_S is the sedimentation velocity of a particle (m s⁻¹).

The vertical resolution of FORCAsT can be configured with a minimum of 20 and maximum 22 23 of 60 vertical layers, extending from the land surface to a maximum height set by the user 24 (Figure 1a). The default total number of above-ground model levels is 40, around half of which are in the vegetation canopy space, with the remainder of the levels representing the 25 planetary boundary layer above. The thickness of the layers increases with height, permitting 26 greater resolution in the canopy levels, which are further sub-divided into a trunk space and 27 crown space. The height of the trunk space and the top of the crown space are set by the user 28 for the specific location of interest. The lower boundary of the column represents the land 29

(soil) surface. In addition to the above-ground layers, the model includes 15 soil layers for
computing soil heat and moisture storage and transfer to the atmosphere, as well as root
extraction (Forkel et al. 2007).

4 As the CACHE model has been described extensively elsewhere (Forkel et al., 2006; Bryan et al., 2012), we mostly confine our descriptions to the improvements and updates to the original 5 model although we give a brief summary of the main processes. We outline the general or 6 default settings of FORCAsT simulations within the main text. Many of the parameters and 7 8 boundary or initial conditions (e.g. canopy architecture, foliage properties, meteorological 9 conditions, concentrations) in the model can be adjusted by the user for a specific site or timeperiod. The values for the simulation period used to evaluate FORCAsT are given in the 10 11 accompanying Supplementary Material, along with further information on the initialization 12 and use of FORCAsT.

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14 **2.1** Canopy structure and radiative transfer

Following the parameterisations of the CUPID model (Norman, 1979; Norman and Campbell, 16 1983) FORCAsT simulates the transfer of radiation through the vegetation canopy, allowing 17 an energy budget to be computed for each model level within the canopy space. Thus, 18 prognostic leaf temperatures, and latent and sensible heat fluxes are determined for both sunlit 19 and shaded foliage at each canopy level.

20 Incoming radiation at the top of the canopy is prescribed, either via user-provided radiation observations or by a default scheme within the code that includes provision for cloud 21 22 coverage (based on an average fractional coverage specified by the user). Solar radiation is 23 split between visible (0.4-0.7 μ m) and near-infrared (0.7-4 μ m), and the thermal radiation 24 contribution (4-100 µm) is calculated on-line (Norman, 1979). The visible component of the 25 incoming solar radiation is assumed equivalent to photosynthetically active radiation (PAR) 26 and is used to drive the biological vegetation processes linked to photosynthesis and biogenic emissions. Within the canopy, reflection, transmission and absorption of all incoming 27 28 radiation wavelength bands and the total back-scattered or up-welling radiation is dependent on the canopy structure and the angle of the leaves relative to the incident radiation. The 29 30 incoming solar radiation is further divided into direct and diffuse radiation based on the 31 proportion of back-scattered radiation.

The canopy architecture is constructed during the initialisation routines within FORCAsT. A 1 2 leaf angle distribution (i.e. the area fraction of leaves within each canopy layer whose normal 3 lines fall within a specified range of angles from the solar zenith angle) is calculated based on 4 the total projected LAI of the canopy, and the fraction of the total LAI in each canopy layer 5 (which may be set by the user via an input file). By default, the calculation assumes a spherical canopy (i.e. perfectly symmetrical in all directions) in terms of its response to 6 7 incoming radiation, but this may also be altered via the input file. FORCAsT currently 8 considers 9 angle classes of sunlit leaves, and designates an additional (tenth) class to the 9 shaded leaves within each layer. An initial attenuation factor for radiation within the canopy 10 is then calculated based on this leaf angle distribution and a user-provided foliage clumping 11 factor, describing the distribution of leaves along the branches, and hence the ease with which 12 radiation can penetrate the canopy.

13 The effective area of leaf surface intercepting solar radiation is then calculated at each model 14 timestep assuming a β -distribution relative to the solar zenith angle (Goel and Strebel, 1984) and a default azimuthal angle distribution (Strebel et al., 1985). Either of these distributions 15 can be altered by the user to fit site-specific observations of canopy structure. This effective 16 interception area for each angle class in each canopy layer provides the basis for the 17 simulation of light attenuation within the canopy (based on Beer's Law), and of the 18 19 absorption of thermal radiation at each model timestep. Leaves in the nine sunlit angle classes 20 are assumed to receive components of both direct and diffuse radiation; shaded leaves receive 21 only diffuse.

Radiation penetrating each canopy layer decays due to shading from leaves in the layers above. An energy balance is calculated for each leaf angle class within each canopy layer to determine leaf temperature and heat fluxes. Biogenic emissions, driven by PAR and leaf temperature, thus vary between layers and between angle classes within a single layer.

26 **2.2 Emissions**

Biogenic emissions of VOCs (bVOCs) from canopy vegetation are calculated on-line using
the parameterised light and temperature dependencies developed by Guenther et al. (1995)
and modified by Steinbrecher et al. (1999) to account for emissions from storage pools. Pool
emissions are dependent on temperature alone and are characteristic of most terpenoids,

- although isoprene is only emitted via direct synthesis. Site-specific direct synthesis and pool
 emission factors are prescribed for different vegetation types and bVOCs.
- 3 Emissions flux (F; nmol m⁻² s⁻¹) are calculated for each leaf-angle class and summed over 4 each layer in the canopy crown space using prognostic leaf temperature and accounting for 5 sunlit and shaded leaves in each level at every model timestep.
- 6 Synthesis (direct) emissions: $F = LAI. \varepsilon. \gamma_{TS}. \gamma_{LS}$ (4)
- 7 where LAI is the leaf area index in each leaf-angle class and layer, ε is the emission factor or 8 base emission rate (i.e. the emission rate at standard conditions of 30 °C and incoming PAR 9 of 1000 µmol m⁻² s⁻¹) and γ_{TS} and γ_{LS} are scaling factors accounting for the actual leaf 10 temperature and incoming radiation respectively. The scaling factors are calculated as:

11 Temperature scaling factor:
$$\gamma_{\text{TS}} = \frac{exp\left(\frac{C_{T1}(T_L - T_S)}{R.T_L.T_S}\right)}{x + exp\left(\frac{C_{T2}(T_L - T_M)}{R.T_L.T_S}\right)}$$
(5)

where C_{T1} , C_{T2} and *x* are empirically determined coefficients (95000 J mol⁻¹, 230000 J mol⁻¹) and 0.926 respectively). T_L is the leaf temperature, T_S is a standard temperature (here taken as 303K), and T_M is an optimum temperature (here set to 314K). *R* is the ideal gas constant (8.314 J K⁻¹ mol⁻¹).

16 Light scaling factor:
$$\gamma_{\rm LS} = \frac{\alpha.C_L.PAR}{\sqrt{1+\alpha.PAR^2}}$$
 (6)

17 where $C_{\rm L}$ and α are empirically determined coefficients (1.1066 and 0.0027 respectively). 18 PAR (µmol m⁻² s⁻¹) is that reaching the leaf surface.

19 Pool emissions: $F = LAI. \varepsilon. \gamma_{\rm TP}$ (7)

20 where γ_{TP} is the temperature correction factor accounting for the actual conditions, calculated 21 as:

22 Temperature correction factor: $\gamma_{TP} = \exp(\beta(T_{\rm L} - T_{\rm S}))$ (8)

23 where constant β is determined from observations (typically 0.09 K⁻¹) and T_S is taken as 24 293K.

25 **2.3 Advection**

Traditionally, box and 1-D canopy models do not include advection as they are not designed or intended to be atmospheric transport models. However, without additional advective sources or sinks of heat or mass, many such models cannot reliably capture observed
 fluctuations in concentrations of primary emitted species and their immediate oxidation
 products, which may accumulate in the column.

4 Robust data of nearby (upwind and downwind) temperatures and concentrations at numerous model layers, taken either from monitoring stations or atmospheric chemistry and transport 5 6 models, at a spatial and temporal resolution appropriate for application to a single-point column model are generally not available for most remote forest sites. This precludes the 7 8 inclusion of a rigorous mass-balance advection scheme. Bryan et al. (2012) therefore 9 incorporated a simple parameterisation of advection, based on wind direction, air mass origin and wind speed to account for potential anthropogenic influences on remote forested regions. 10 Site-specific observations of temperature and concentration at and above the canopy height 11 are used to define advection rates for specific wind directions, based on the simplified mass-12 13 balance approach shown in Eqn 9 and 10.

14 Advection rate (heat):
$$\frac{\partial T}{\partial t} = U.k.T$$
 (9)

15 Advection rate (mass):
$$\frac{\partial c}{\partial t} = U.k.c$$
 (10)

16 where k is a species-dependent advection coefficient that aligns the concentrations in the 17 model to the observed concentrations under different wind speeds and directions.

18 **2.4 Deposition**

19 Sedimentation of aerosol particles occurs at all model levels and is explicitly included in the continuity equations (see Eqn 3), using volume-averaged sedimentation velocities. Dry 20 21 deposition of gases and particles occurs to vegetation surfaces within the crown space and to 22 the ground, and is calculated following the resistance scheme (Meyers and Baldocchi 1988; 23 Wesely, 1989; Gao et al., 1993). The total leaf resistance to deposition is dependent on the 24 individual resistances of the quasi-laminar boundary layer on the leaf surface (level 25 dependent), and the mesophyll and cuticular resistances (species dependent), and stomatal 26 resistance (level and species dependent). The soil or surface resistance is modelled after Gao 27 et al. (1993).

Deposition is assumed to occur at a rate dependent on species-specific Henry's Law coefficient, diffusivity relative to water vapour and a nominal "reactivity" relative to ozone.
The ozone-relative reactivity has been increased for oxygenated VOCs and bVOC oxidation

products following Karl et al. (2010) to account for enhanced uptake due to reactions
 occurring within plant cells.

As FORCAsT includes a full multi-level representation of vegetation structure, the processes governing deposition rates are explicitly incorporated. In particular, stomatal conductance for each leaf angle class in each canopy layer is calculated according to the canopy environment at each time step, accounting for changes in temperature, light levels above and within the canopy, and vapour pressure deficit. Soil resistances are likewise calculated at each timestep based on the temperature and soil moisture profile at that time.

9 Deposition velocities of gases and particles are calculated by FORCAsT before being passed 10 to the chemistry scheme, where they are included as a loss term in the computation of reaction 11 rates. The mass of a species lost through deposition is calculated from its deposition velocity 12 or potential and its atmospheric concentration within any particular vertical layer.

As the simulation of stomatal conductance within FORCAsT occurs on-line, this provides the potential to estimate the flux of any species into the vegetation, allowing simulation of damage to plant cells due to the uptake of powerful oxidising agents such as ozone. This capability will be utilised in future studies.

17 2.5 Turbulent exchange

18 In FORCAsT, vertical turbulent exchange of mass and energy follows traditional K-theory 19 (Blackadar, 1962). Mixing within and above the canopy is simulated using the parameterisations of Baldocchi (1988) and Gao et al. (1993), respectively. The resulting 20 vertical profiles are further modified to improve the simulated exchange of heat and trace 21 22 gases by constraining the friction velocity with sonic anemometer observations near the 23 canopy following Bryan et al. (2012). For the simulation period presented here (see Section 3 24 for details), sonic data are incorporated at two heights (20.6 m - roughly the top of the 25 canopy, and 36.94 m – the top of the PROPHET flux measurement tower at UMBS; see 26 Section 3). The vertical exchange coefficient (K in Eqns. 2.1 - 2.3) within the crown space is calculated by linear interpolation between the modelled value at the crown base and the value 27 28 estimated from sonic data at the top of the canopy, following the approach of Stroud et al. 29 (2005). The same procedure is then performed between the top of the canopy and top of the 30 tower, and the top of the tower value is linearly interpolated to the value at 1 km modelled 31 using Gao et al. (1993). Bryan et al. (2012) demonstrate that the limitations of traditional K- theory within and just above the canopy make this method necessary to capture the observed
 vertical exchange and distribution of heat and mass.

3 **2.6 Gas-phase chemistry**

In FORCAsT, gas-phase chemistry can be calculated using either the Regional Atmopsheric
Chemistry Mechanism (RACM; Stockwell et al., 1997; Geiger et al., 2003) or CACM
mechanisms, but aerosol partitioning is only available when CACM is used, via the MPMPO
equilibrium-partitioning model (Griffin et al., 2003; Griffin et al., 2005; Chen et al., 2006).
The subroutines and modules within the CACM model included here have been generated
using the Kinetic PreProcessor (KPP; Sandu and Sander, 2006), facilitating the use of other
chemistry schemes within FORCAsT.

11 2.6.1 RACM

The version of RACM included in FORCAsT incorporates the key reactions of the Mainz Isoprene Mechanism (Pöschl et al., 2000) as described by Geiger et al. (2003). The concentrations of 84 gas-phase species are calculated at 1-minute timesteps. The scheme includes 249 reactions. Changes to RACM since its original description by Stockwell et al. (1997) are listed in Tables B1-2 in the Supplementary Material.

17 **2.6.2 CACM**

In order to achieve an improved representation of condensable species and simulate SOA 18 19 formation within the canopy, we add the CACM (Griffin et al., 2002; Griffin et al., 2005; Chen and Griffin, 2005; Chen et al., 2006) gas-phase chemical mechanism because of its 20 21 explicit treatment of SOA-relevant chemical species. CACM uses a mechanistic approach to simulate VOC-NO_x-HO_x chemistry while tracking condensable species that contribute to 22 23 SOA. This represents an intermediate complexity approach between those of a highly lumped, simplified mechanism such as RACM and a fully explicit chemical mechanism such as the 24 25 University of Leeds Master Chemical Mechanism (MCM) (Jenkin et al., 2003; Saunders et al., 2003). In principle, the MCM approach is most rigorous, but such a mechanism is 26 27 computationally expensive, and many of the required reaction rates, products, and thermodynamic parameters are still not accurately known. CACM is a condensed version of 28 29 MCM that simulates ozone chemistry as well as formation of individual organic oxidation products that are capable of forming SOA. The version of CACM incorporated into 30

FORCAsT includes the original mechanism of Griffin et al. (2002) with updates of Griffin et al. (2005) and addition of explicit treatments for SOA formation from the monoterpenes apinene, b-pinene, and *d*-limonene of Chen and Griffin (2005). It includes 300 prognostic species and 620 chemical reactions, with a full description listed in Tables A1-2 of the Supplementary Material. To simulate SOA, gas-phase species in CACM are categorized into condensable and non-condensable groups according to experimental or estimated vapour pressures or solubility.

8 2.6.3 Update of the CACM mechanism for low-NO_x conditions

9 The original CACM mechanism (i.e. as described by Griffin et al., 2002; 2005; Chen et al., 10 2005, and referred to as CACM0.0 hereafter) was updated based on the performance of the 11 for a two-day simulation period driven by observed conditions at UMBS. Full details of this 12 simulation period and location are given in Section 3 below. Output concentrations of key 13 species were compared against measurements made at UMBS during this period and against 14 those simulated by the RACM scheme. Figure 2 shows these concentrations at the height of 15 the top of the PROPHET tower (~35 m) for this period of 4th-5th August 2009.

As shown in Figure 2a and 2b CACM0.0 reproduces the observed concentrations of the 16 17 primary emitted terpenoids (isoprene and total monoterpenes which are lumped as α -pinene, β-pinene and δ-limonene in CACM0.0, and as α-pinene and δ-limonene in RACM) 18 19 effectively, for the most part capturing both the magnitudes and the diurnal profiles of the observations. The two chemistry schemes are also in close agreement. Both show a tendency 20 21 to overestimate isoprene concentrations during the afternoon and overnight, while failing to capture the high concentrations of the monoterpenes during the early morning of August 4th. 22 23 Likely reasons for these discrepancies are discussed in further detail in Section 4. The similarity of the modelled concentrations suggest that differences in terpenoid oxidation 24 25 pathways and hence oxidant availability between the two chemistry schemes is of little 26 importance compared to the magnitude of emissions and efficiency of vertical turbulent 27 transport at this site.

Likewise, modelled concentrations of ozone (Figure 2g) show little difference between the two mechanisms. This is attributable in part to the well-documented buffering of ozone in atmospheric chemistry mechanisms (e.g. Young et al., 2013; Emerson and Evans, 2009; Wild, 2007). In addition, the atmospheric lifetime of ozone relative to the timescale of its chemical production suggest that the ozone budget at UMBS is dominated by long-range transport from source regions, with in-situ production making only a minor contribution. In this case, the simulated concentrations are in good agreement with the observed levels although the diel cycle is not well captured, particularly on the first day of the simulation.

5 The success of CACM0.0 in simulating mixing ratios of the primary terpenoids is likely due 6 to factors other than the atmospheric oxidation reactions, and this is clear from the remaining 7 panels in Figure 2. Figures 2c and 2d show the concentrations of key products of isoprene 8 oxidation. Given the skill of both chemistry schemes in capturing isoprene concentrations 9 within the canopy, it might be expected that these species would be similarly well modelled. However, as is evident from Figure 2c, which shows the mixing ratio of methyl vinyl ketone 10 plus methacrolein (lumped as a single species in the RACM mechanism although treated 11 separately by CACM0.0, and referred to hereafter as MVK+MCR), neither chemistry 12 13 mechanism reproduces either the diurnal profile or the absolute concentrations of 14 MVK+MCR in the canopy space. In both cases, the modelled concentrations are far higher 15 than those observed and there is a tendency for accumulation within the canopy over the course of the two days. MVK+MCR concentrations on the second day of the RACM 16 17 simulation are substantially lower than those modelled by CACM0.0, but are still a factor of 2-3 higher than observations. 18

19 CACM0.0 displays the same difficulties with formaldehyde (Figure 2d), over-estimating the 20 concentration at the top of canopy by a factor of 4-5. RACM performs much better in terms of 21 capturing the absolute concentrations but fails to reproduce the diurnal profile seen in the 22 measurements on the second day of the simulation period.

23 Many of the differences in modelled concentrations between the two chemistry schemes were found to be attributable to the availability of oxidants in the two simulations. Following an 24 initial sharp decline in NO₂ (Figure 2e), which is also evident in the observations and the 25 RACM simulation, NO_x concentrations in CACM0.0 fail to recover indicating that loss rates 26 27 far exceed the rates of production or recycling of NO_x in the scheme. NO mixing ratios 28 (Figure 2f) behave similarly, following the measured concentrations and those simulated by 29 RACM early on the first day, but failing to recover once exhausted. Low NO_x conditions at 30 UMBS occur under northerly (clean) air flow; at these times, soil NO emissions are the only 31 source of NO_x. This is insufficient to outweigh the loss of NO_x to PAN and other unreactive nitrate species via the reactions of peroxy radicals with NO₂ in CACM0.0. 32

The picture is more complex when considering the HO_x oxidants. CACM0.0 and RACM 1 2 produce very similar mixing ratios of the OH radical (Figure 2h), although both appear to fall 3 well below measured concentrations. The final panels of Figure 2 show concentrations of HO₂ (Figure 2i) and HO₂^{*} (the sum of HO₂ and the peroxy radicals derived from the isoprene+OH 4 reaction; Figure 2j). In both cases, the model concentrations are displayed against 5 measurements of HO₂ made at the site. It is thought that HO₂ sampling instruments detect 6 both HO₂ and these peroxy radicals on the same channel, and that modelled output of HO₂^{*} is 7 8 therefore more appropriate to use for comparison (Griffith et al., 2013; Fuchs et al., 2011). HO₂ concentrations for both chemistry models are well below those measured (as would be 9 expected if the observations include the peroxy radicals). CACM0.0 mixing ratios are lower 10 than those in RACM, from the point on 4th August when NO_x levels reach zero in the 11 CACM0.0 simulation, as the lack of NO limits oxidant recycling via radical reactions. 12 Interestingly, however, while RACM mixing ratios of HO_2^* agree well with the measurement 13 data, the combined concentrations in the CACM0.0 scheme exceed the measured values by a 14 factor of 20-30, suggesting a significant over-estimation of isoprene-derived peroxy radicals, 15 likely due to lack of NO, reactions with which are their primary sink. 16

17 The time of divergence of modelled concentrations coincides with meteorological changes at the site. As outlined in Sections 3 and 4, the prevailing conditions at UMBS changed with the 18 passage of a cold front on the morning of the 4th August, bringing cooler cleaner air from the 19 north. Around mid-morning of the 4th August therefore marks a transition from what could be 20 considered a high-NO_x to a low-NO_x regime at the site, suggesting that CACM0.0 fails to 21 22 represent low-NO_x VOC oxidation chemistry effectively. Previous studies using and 23 evaluating the CACM scheme (see for example Griffin et al., 2002; Chen et al., 2007; Chen et al., 2010) were all conducted in regions and time periods when NO_x levels were high relative 24 25 to bVOC concentrations. Under such conditions, CACM0.0 has been shown to perform well. 26 In addition, the mechanism was developed and tested in a region in which the VOC budget is 27 dominated by anthropogenic sources, with the bVOC contribution predominantly from monoterpenes rather than isoprene. Applying the model for this two-day period at UMBS, 28 which can be characterised as a combination of low NO_x and high isoprene concentrations 29 therefore represented a profound change from previous simulations. 30

Figure 2 suggests that the key difference between the mechanisms is the production and lossof peroxy radicals formed from the initial oxidation reactions of VOCs. The main chemical

sinks for peroxy radicals are through reactions with NO, HO₂ and with other peroxy radicals
 (see e.g. Atkinson and Arey, 2003; Jenkin et al., 1997; Perring et al., 2013):

| 3 | $\text{RO}_2^* + \text{NO} \rightarrow \text{carbonyl} \text{ (via alkoxy radical)} + \text{NO}_2$ | (R1) |
|---|----------------------------------------------------------------------------------------------------|------|
| 4 | $\mathrm{RO_2}^* + \mathrm{NO} \rightarrow \mathrm{alkyl} \ \mathrm{nitrate}$ | (R2) |
| 5 | $\mathrm{RO_2}^* + \mathrm{HO_2} \rightarrow \mathrm{peroxide}$ | (R3) |
| 6 | $\mathrm{RO_2}^* + \mathrm{RO_2}^* \rightarrow \mathrm{carbonyl}$ | (R4) |

Sensitivity studies were conducted for high-NO_x conditions, in which the performance of 7 8 CACM0.0 was found to be closely comparable to that of the RACM scheme, indicating that the discrepancies shown in Figure 2 were due to the low levels of NO_x at UMBS (results not 9 10 shown). These studies strongly suggested that the source of the discrepancy was the relative 11 rates of reactions R1-R4. This hypothesis is consistent with current understanding of the difference in radical termination reactions at high and low-NO_x levels. When NO_x 12 concentrations are high relative to those of VOCs, the RO_2^* peroxy radicals formed from the 13 14 initial oxidation of VOCs are oxidised to stable species through their reactions with NO (R1 and R2). At relatively lower levels of NO_x, termination reactions of the peroxy radicals with 15 HO_2 (R3) and other RO_2^* (R4) dominate. Evidence from recent field campaigns and 16 laboratory experiments indicates that the self- and cross-reactions between RO_2^* radical 17 species (R4) are of particular importance in locations where the VOC:NO_x ratio is very high 18 19 (such as forested ecosystem in remote environments; see e.g. Whalley et al., 2014; Wolfe et 20 al., 2014; Perring et al., 2013; Peeters et al., 2009).

The original isoprene chemistry mechanism in CACM0.0 (Griffin et al., 2002) was based on knowledge that is now almost two decades old and, unlike the monoterpene chemistry mechanism (Chen and Griffin, 2005), has not been updated. We update CACM0.0 to include some recent advances in modelling low-NO_x atmospheric VOC oxidation.

CACM0.0 relies on the NO reactions to continue the degradation of VOCs after the initial oxidation by OH, O₃ or NO₃, with few peroxy radicals channelled through the HO₂/RO₂ pathways, even at very low concentrations of NO. Once NO_x levels fall, the rates of RO₂^{*} + NO reactions slow and peroxy radicals accumulate in the system, resulting in further depletion of NO and feeding back to further accumulation of peroxy radicals.

The rates of equivalent or similar reactions involving peroxy radicals (RO2*) in the 1 2 CACM0.0 mechanism were compared against those in RACM and MCM v3.2 (http://mcm.leeds.ac.uk/MCM/). The rates of RO₂^{*}+NO reactions were similar across all 3 4 mechanisms, as might be expected given that such reactions are well-studied and well-5 constrained, and that CACM0.0 performed similarly to RACM under high-NO_x conditions. The HO₂ reaction rates were generally lower (by a factor of around 3) in the CACM0.0 6 7 scheme, accounting in part for slightly higher HO₂ concentrations in the CACM0.0 simulation 8 under high-NO_x conditions (not shown).

The most substantial discrepancies between the mechanisms were the rates of the $RO_2^* + RO_2^*$ 9 reactions. Direct comparison with the RACM scheme was difficult as CACM0.0 employs the 10 11 technique of using a generic peroxy radical species (referred to as RO₂T - see Table A2) that is effectively the sum of all peroxy radicals to represent all possible permutations of R4. 12 13 There are fewer distinct peroxy radical species in RACM, and other than the methyl and 14 isoprene-derived peroxy radicals, there are no self- or cross-reactions included. Comparison 15 with the MCM showed that other than for the reactions involving radicals produced from monoterpene oxidation (which were updated more recently by Chen et al., 2005) the reaction 16 rates used in CACM0.0 were several orders of magnitude too low. The reaction rates of the 17 peroxy radical reactions with HO₂ (R3) and RO_2^* (R4) were therefore increased to better 18 match those in the MCM (see Table A2 of the Supplementary Material). 19

20 Recent data from field campaigns also suggest that the formation and loss of organic nitrates produced from alkyl peroxy radicals play an important role in governing nitrogen cycling and 21 22 availability over relatively short timescales particularly in low-NO_x environments (Beaver et al., 2012, Brown et al., 2010; Browne et al., 2010; Perring et al., 2013). While the CACM0.0 23 24 mechanism included formation of alkyl nitrates from the reactions of many of the alkyl 25 peroxy radicals with NO, not all of the isoprene peroxy radicals produced nitrates. Given the 26 relative abundance of isoprene at this site, the clear over-production (or reduced loss) of isoprene peroxy radicals, and the low NO_x conditions, the products of these reactions were 27 altered to include the formation of isoprene nitrates at a yield of ~3-5% (see Table A2 of the 28 Supplementary Material). 29

30 The subsequent reactions of alkyl nitrates with OH, which release NO₂ at timescales likely to

31 be relevant to in-canopy chemistry, included in the original CACM scheme (Griffin et al.,

32 2002) but later removed (Griffin et al., 2005), were re-introduced. Equivalent reactions for the

new isoprene nitrates were also added, as nitrates formed from bVOCs are known to have
 particularly short lifetimes with respect to the OH radical (Müller et al., 2014, Perring et al.,
 2013, Paulot et al., 2009), suggesting that these reactions occur on timescales likely to be of
 relevance to in-canopy chemistry.

A new theoretical study based on pervious laboratory experiments has also demonstrated that photolysis of isoprene-derived nitrates may occur at a timescale competitive with their reactions with the OH radical (Müller et al., 2014). The breakdown of the isoprene nitrates via photolysis has therefore also been included, with photolysis rates following the suggestions of Müller et al. (2014) (see Table A2 of the Supplementary Material).

10 As shown in Figure 2d, CACM0.0 also produces too much formaldehyde compared to both the observations and the RACM scheme. While the initial problem may stem from excessive 11 reaction rates or formaldehyde yields from RO₂^{*}+NO reactions, it was found that bias 12 13 increases were larger under low-NO_x than high-NO_x conditions, suggesting this is associated with $RO_2^* + RO_2^*$ and $RO_2^* + HO_2$ reactions. When NO_x is abundant relative to VOCs, the 14 reaction with NO dominates with minor contributions from HO₂ and RO₂^{*} pathways (R1 and 15 R2). In low-NO_x environments, the competing HO₂ and RO₂^{*} reactions form the main sink of 16 peroxy radicals (R3 and R4). As reactions between peroxy radicals and HO₂ do not produce a 17 significant yield of carbonyl compounds as first-generation products (see however Liu et al., 18 19 2013), overall yields of aldehydes and ketones are reduced when NO_x levels are low and a 20 greater proportion of oxidation occurs via reaction with HO₂ (see, e.g., Sumner et al., 2001). 21 Experiments of isoprene peroxy radical reactions conducted under high-NO_x and NO_x-free 22 conditions, for example, suggest that the overall yield of formaldehyde is around 0.57 when 23 NO_x is abundant; dropping to around 0.34 when no NO_x is present (Miyoshi et al., 1994).

Although the peroxy radical+HO₂ reactions initially form organic peroxides, subsequent 24 25 photolysis releases carbonyls and HO_x. The reaction scheme in CACM0.0 combines these into a single step with peroxy radicals reacting with HO₂ to form aldehydes (mostly 26 27 formaldehyde) immediately, in addition to a proxy species that then photolyses to recycle 28 HO_x. By contrast, RACM forms an organic peroxide that can then photolyse to form an aldehyde and HO_x, with formaldehyde only being produced from the peroxide produced from 29 30 methane oxidation. While the approach in CACM0.0 should in theory permit better aldehyde 31 speciation without the introduction of numerous separate photolysis reactions, the overall effect is to increase the production of formaldehyde and to alter the time at which it is 32

produced. As photolysis only occurs during daylight hours, the inclusion of this as a separate
 reaction could be expected to introduce a diurnal profile that is currently absent from
 CACM0.0 formaldehyde concentrations.

The peroxy radical reactions in CACM0.0 were modified as outlined above to bring them closer in line with those included in RACM. The formaldehyde yield from the peroxy radical+HO₂ reactions was set to zero, and a yield of unity added to the photolysis reactions of the proxy species formed from the peroxy radical+HO₂ reactions. The photolysis rate of this reaction was also increased to match that in the RACM mechanism.

9 These updates to CACM0.0, hereafter referred to as CACM, are included in FORCAsT 1.0.

10 2.7 Aerosol partitioning

11 One of the most significant capabilities of FORCAsT 1.0 is the inclusion of the partitioning of 12 condensable species to the particle phase. Of the 300 prognostic species in CACM, 99 are treated as condensable in MPMPO (highlighted in Supplementary Material Table A3). For 13 14 biogenic SOA precursors, CACM includes explicit gas-phase chemistry for a-pinene, bpinene, and *d*-limonene (Chen and Griffin, 2005); other monoterpenes are lumped into a low 15 SOA yield or a high SOA yield group, represented by a-terpineol and g-terpinene, 16 respectively (Griffin et al., 2002). Explicit formation of SOA from isoprene is not considered 17 18 in this version of MPMPO, driven by the CACM gas-phase mechanism (hereafter referred to as CACM-MPMPO). However, oxidation reactions of methyl vinyl ketone and methacrolein, 19 20 two major oxidation products of isoprene, form keto-propanoic acid and oxalic acid, 21 respectively, which are considered condensable and form SOA in CACM-MPMPO. 22 Anthropogenic SOA and primary organic aerosols (POA) are also included in MPMPO (Griffin et al. 2003, 2005). For the simulations of UMBS during the CABINEX campaign 23 presented here, POA concentration is assumed to be a constant value of 0.5 mg m⁻³ and 24 25 anthropogenic VOC concentrations are set to zero.

Condensable species formed from VOC oxidation in CACM create a "reservoir" of potential SOA. In MPMPO, the 99 condensable species are lumped into 12 surrogate species according to their structures, sources (biogenic or anthropogenic), volatilities, and dissociative capabilities. These surrogate species are the original S1 to S9 groups described in Griffin et al., (2003), an updated S10 group described in Griffin et al. (2005), a new S11 group for bpinene oxidation product 2,10-dinitrato-pinane (Chen et al., 2006), and a new S12 nonvolatile group representing dimers formed from multifunctional acid species generated from
 oxidation of monoterpenes (Chen et al., 2006). Characteristics, surrogate species, and list of
 species for each surrogate group are summarized in Table A3 of the Supplementary Material.

4 The MPMPO aerosol module calculates the partitioning of the CACM gas-phase condensable 5 oxidation products. Absorption into the organic phase is governed by the an absorption 6 coefficient, $K_{\text{om},i}$ (m³ mg⁻¹) (Pankow, 1994):

7
$$K_{\text{om},i} = \frac{O_i}{G_i M_0} = \frac{RT}{MW_{\text{om}} 10^6 \gamma_i p_{L,i}^0}$$
(11)

where O_i (mg m⁻³ air) and G_i (mg m⁻³ air) are the organic aerosol- and gas-phase 8 concentrations of surrogate species *i*, respectively, M_0 (mg m⁻³ air) is the total organic aerosol-9 phase concentration, R is the ideal gas constant (8.206 \times 10⁻⁵ m³ atm mol⁻¹ K⁻¹), T is 10 temperature (K), MW_{om} is the average molecular weight of the organic phase (g mol⁻¹), g_i is 11 the activity coefficient of surrogate *i*, and $p_{L,i}^{0}$ is the pure component vapor pressure (atm) of 12 surrogate *i* at temperature *T*. The method of Myrdal and Yalkowsky (1997) calculates p_{Li}^{o} , and 13 the UNIFAC method is employed to calculate activity coefficients γ_i (Fredenslund et al., 14 1977; Smith and Van Ness, 1987; Saxena and Hildemann, 1996; Pankow et al., 2001; 15 Seinfeld et al., 2001). 16

17 The partitioning between the gas and the aqueous phase is determined by

18
$$A_i = \frac{G_i(\text{LWC})H_i}{\gamma_{\text{aq},i}}$$
(12)

where $A_i \text{ (mg m}^{-3}\text{)}$ is the aqueous-phase concentration of surrogate species *i*, LWC (mg H₂O 19 m^{-3} air) is the liquid water content in the aqueous phase, H_i (mg mg⁻¹ H₂O) is the Henry's Law 20 coefficient of surrogate species i, and $g_{aq,i}$ is the activity coefficient (normalized by that at 21 22 infinite dilution) of surrogate species *i* in the aqueous phase. The group contribution method 23 of Suzuki et al. (1992) is used to estimate the Henry's Law coefficients H_i . The UNIFAC method is employed to calculate activity coefficients $g_{aq,i}$. The liquid water content due to the 24 25 presence of aqueous phase organics is determined using the Zdanovskii-Stokes-Robinson 26 (ZSR) method (Meng et al., 1998; Pun et al., 2002). Total aerosol liquid water content (ALWC) associated with inorganic and organic phases is an input to the MPMPO module and 27 is needed to determine organic aerosol aqueous-phase concentrations. For the simulations 28 29 presented here, we used hourly ALWC calculated using the hygroscopity parameter, κ , which is based on observed CCN concentrations at 0.3% supersaturation and observed particle size
 distributions during CABINEX (VanReken et al., 2015).

For the aqueous phase, equilibrium is also constrained by dissociation of the dissolved organic species. The concentrations of the singly charged ion from surrogate species *i*, A_{1i} (µg m⁻³ air), and the concentration of the doubly charged ion from surrogate species *i*, A_{2i} (µg m⁻³ air), can be represented as

$$A_{1i} = \frac{K_{1i}A_i(MW_i - MW_{H^+})}{[H^+]MW_i}$$
(13)

$$A_{2i} = \frac{K_{2i}A_{1i}(MW_i - 2MW_{H^+})}{[H^+]MW_i}$$
(14)

9 where $[H^+]$ (mol kg⁻¹ H₂O) is the proton concentration in the aqueous phase, and K_{1i} and K_{2i} 10 (mol kg⁻¹ H₂O) are the dissociation constants, which are estimated using structure-activity 11 relationships (Harris and Hayes, 1982; Schwarzenbach et al., 1993).

12 **3 Evaluation**

The performance of FORCAsT was evaluated with output from the CACHE canopy and 13 14 chemistry model described in Bryan et al. (2012) and observations from UMBS during the 15 CABINEX campaign in 2009 (ACP special issue: Eds. Williams et al., 2011). This intensive 16 field campaign was primarily focused at the PROPHET flux tower with further measurements 17 made at the nearby Ameriflux tower. Full details of this field measurement site and the PROPHET tower can be found in Carroll et al. (2001), with the 2009 campaign described in 18 19 Kanawade et al. (2011), Kim et al. (2011), Steiner et al. (2011), Zhang et al. (2012), Bryan et al. (2012), Griffith et al. (2013), and VanReken et al. (2015). 20

The results presented here are based on a two-day model simulation for Aug 4th-5th 2009, coinciding with the simulation period in Bryan et al. (2012). The driving meteorology, land surface and vegetation characteristics are derived from UMBS measurements and are identical to those used in the previous study. Initial and boundary conditions were also set following Bryan et al. (2012) with the addition of aerosol measurements (VanReken et al., 2015) to allow full and robust comparison of the models' skill in reproducing conditions at the site. Further details of the model settings are given in the Supplementary Material.

UMBS is located near Pellston, Michigan and consists of around 4000 hectares of natural
habitat containing a range of ecosystems. The 2009 CABINEX field campaign was conducted
at the PROPHET flux tower (34 m high, located at 84.7145°W, 45.5588°N), in an area of

transition forest containing a mix of northern hardwood, aspen and conifer. The short-term
measurements made at PROPHET, including micro-meteorology and concentrations and
fluxes of gases and aerosols, were supplemented by additional meteorological data taken from
the Ameriflux Tower (46 m high, located at 84.7138°W, 45.5598°N).

The summer of 2009 was uncharacteristically cool and wet, with daytime maximum 5 6 temperatures during the CABINEX campaign that were around 4°C below the long-term 7 average for the site and precipitation or fog recorded on over 60% of the measurement days 8 (Bryan et al., 2012). In total, only 6 days were categorised as sunny or partly sunny. The 9 cloudy conditions were brought about in part by strong synoptic influences with predominantly southerly winds, a wind direction that is also associated with the long-range 10 11 transport of air pollutants to the site from the cities of Chicago, Milwaukee and Detroit, resulting in elevated background concentrations of NO_x, ozone, and anthropogenic VOCs in 12 13 particular.

The two-day period of Aug 04-05, 2009 was selected for the evaluation of FORCAsT. Full details of the prevailing conditions at the UMBS measurement site during this time are provided in Bryan et al. (2012). We briefly summarise the salient points here.

Skies were clear throughout the 48 hours of the simulation, although the prevailing wind direction changed with the passage of a frontal system at around 07:00 (EST) on Aug 04. Prior to that time, winds from the southwest brought relatively warmer temperatures to the site. For the remainder of the simulation period, winds were northerly bringing cooler air from Minnesota and southern Canada. High temperatures were less than 21°C on both days, slightly below the average high temperature of 22°C for the CABINEX period (Bryan et al., 2012).

The change in wind direction also resulted in different chemical conditions at the site, with southerly air mass bringing anthropogenic pollutants from Detroit and Chicago. Air masses from the North are associated with clean (low-NO_x) conditions.

27 4 Results

28 4.1 Air temperature

29 The air temperature (Figure 3) modeled by the energy balance routines in FORCAsT shows a 30 typical diurnal cycle and is generally in reasonable agreement with the observed temperatures

in the canopy. However, during the first 6 hours of the simulation period, modeled 1 2 temperatures are well below those experienced at UMBS. The passage of the frontal system 3 from the north discussed above (Sec. 3) brought cooler temperatures to the site. Conditions 4 prior to this had been relatively stagnant with temperatures remaining elevated overnight due 5 to a warm air mass over the site. As FORCAsT is a 1-D model, without prescriptive meteorology, it cannot be expected to capture this. However, the canopy energy balance also 6 7 appears to over-predict canopy air temperatures at all heights during the middle of the day and 8 also fails to reproduce accurately the variation of temperature with height overnight within the 9 canopy. Specifically, simulated overnight temperatures are 2-3°C above those observed at 20.4 m and 34 m. 10

The discrepancy between the modeled and observed air temperatures during the first 8 hours of the simulation period is sufficiently great to affect simulated emission and reaction rates (see Section 4.2). The time until 8am on 4th August is therefore treated as a spin-up period and not included in our evaluation of model performance.

15 **4.2 Gas-phase chemistry**

16 The gas-phase chemistry scheme was modified to improve performance under low-NOx 17 conditions. Section 4.2.1 compares output from FORCAsT (i.e. the updated CACM scheme) against UMBS observations and RACM output data. FORCAsT is also evaluated under high-18 19 NO_x conditions to ensure that the modifications to CACM0.0 do not adversely affect its performance in these situations. The same two-day period was modelled as a high-NO_x 20 environment by artificially advecting NO₂ throughout the simulation period. The rate of NO₂ 21 advection to the site was consistent with an assumption of continual southerly winds bringing 22 23 pollution from Detroit, as outlined by Bryan et al. (2012). The results of these simulations are 24 presented in Section 4.2.2.

25 4.2.1 Low-NO_x

Figure 4 shows concentrations of key species involved in VOC oxidation at the top of the tower (~35m) for 4th-5th August 2009 as observed and modeled with RACM and the updated CACM-MPMPO chemistry mechanisms. The grey shaded region in all Figures denotes the spin-up period as explained above in Section 3.

1 FORCAsT reproduces both the magnitude and the diurnal profile of the observed isoprene 2 concentrations reasonably well (Figure 4a). However, the modelled mixing ratio of isoprene is 3 higher than that observed during the middle of the day and this may be due to an over-4 estimation of temperature (Figure 3). While the daytime discrepancies between modelled and 5 measured concentrations can be ascribed to incorrect emissions, the biggest difference occurs during the night. As isoprene emissions are light dependent, night-time emissions are zero and 6 7 observed concentrations approach zero. In both chemical mechanisms, concentrations are still 8 about 1 ppb at night, suggesting inadequate oxidation in both of the chemistry scheme(s), 9 either at night (possibly due to insufficient NO₃ radical concentrations) or during the late 10 afternoon (resulting in an accumulation of isoprene that persists overnight). Both mechanisms 11 show virtually identical diel cycles, demonstrating the relative insignificance of chemistry 12 compared with other canopy processes over the time and spatial scales involved.

Monoterpene concentrations (Figure 4b) are similarly relatively well reproduced by FORCAsT, as might be expected given that they are also a primary emission in the canopy. Both chemistry mechanisms simulate the same diurnal profile, with maximum concentrations at night when chemistry is slow and vertical mixing out of the canopy is negligible.

17 Concentrations of MVK+MCR simulated by FORCAsT (Figure 4c) with all chemical 18 mechanism options remain well above those observed (by a factor of ~3-5) and show a 19 tendency to accumulate over the course of the day. The updates to the CACM scheme have 20 brought the diurnal profile of MVK+MCR more into line with that of RACM, although 21 neither scheme captures the observed pattern well. Although production is initially more rapid 22 in CACM, mixing ratios are not significantly above those in CACM0.0.

The elevated concentrations of MVK+MCR are most likely the result of the over-production 23 24 of peroxy radicals, with many of the peroxy radical reactions in CACM producing further peroxy radicals. While both methyl vinyl ketone and methacrolein are direct reaction products 25 of the initial oxidation of isoprene by O₃, their primary sources are the reactions of isoprene-26 27 derived peroxy radicals. Figures 5b and 5d show the concentrations of the peroxy radicals 28 produced in the initial oxidation of isoprene by the OH radical, and of the summed peroxy 29 radicals (RO2T) in CACM with those simulated by RACM for comparison. The mixing ratios 30 of peroxy radicals in the CACM simulations are a factor of 2-3 above those estimated by the 31 RACM mechanism (Figure 5a). While the improvements made to CACM0.0 bring both the magnitude and diurnal profile of the peroxy radical concentrations in closer agreement with 32

1 the RACM scheme. CACM still shows a tendency to over-produce and/or under represent 2 their losses. The diurnal profiles of mixing ratios of the isoprene+OH-derived peroxy radicals are in close agreement and strongly reflect the diel cycle of isoprene emissions (Figure 5c). 3 4 CACM concentrations, although well below those simulated by CACM0.0, still exceed those 5 generated in RACM by 100-200%. Although the model output cannot be directly evaluated due to the lack of observations, the relative overestimation of MVK+MCR concentrations in 6 7 CACM compared to both measured levels and those simulated by the RACM scheme, suggest 8 that these radicals are over-produced by the CACM mechanism.

9 Formaldehyde concentrations (Figure 4d), on the other hand, are close to observed mixing ratios and to those simulated by the RACM mechanism, supporting the hypothesis that it is 10 11 over-production of isoprene peroxy radicals that is the cause of the elevated MVK+MCR concentrations in CACM. The elevated formaldehyde concentrations in CACM are the result 12 of the lumping of all RO_2^* + HO₂ peroxides as a single proxy species that photolyses to 13 produce formaldehyde, when in reality many of these would produce higher aldehydes. The 14 15 diurnal profile of formaldehyde concentrations is still not a good match to measured concentrations with a marked over-production at night. This is likely due to the over-estimates 16 17 of peroxy radical concentrations discussed above leading to excessive peroxy radical-peroxy 18 radical reactions.

19 The changes implemented in the CACM gas-phase chemistry scheme, particularly the increase in the rate of $RO_2^* + RO_2^*$ reactions, had a substantial effect on the HO_x-NO_x species. 20 Concentrations of NO₂ (Figure 4e) and NO (Figure 4f) now show typical diurnal profiles for 21 22 each, with NO₂ depletion during the day and production from NO conversion overnight. Daytime NO₂ concentrations are in good agreement with those observed, but overnight 23 recovery is too low with night-time concentrations around a factor of 2-3 below measured 24 mixing ratios. In spite of the increased competition between $RO_2^* + RO_2^*$ and $RO_2^* + NO_2^*$ 25 26 reactions, NO concentrations are still a factor of ~2-5 too low throughout the simulation, 27 showing that there is still too much dependency on the NO reaction channel in the updated reaction scheme. 28

After the passage of the frontal system, ozone concentrations (Figure 4g) are in close agreement with both measurements, and RACM and CACM0.0 simulated values, pointing again to the powerful buffering inherent in most atmospheric chemistry schemes, and dominance of transportation of ozone over local production. OH concentrations (Figure 4h) are little affected by the alterations made to CACM0.0 indicating that it is the initial oxidation reactions and production via ozone that dominate the OH budget in current atmospheric chemistry schemes. However, mixing ratios are well below those observed, consistent with many field campaigns in low-NO_x environments (e.g. Ganzeveld et al., 2008; Stone et al., 2011; Wolfe et al., 2011; Lu et al., 2012). This shows the urgent need to fully update the CACM chemistry scheme (in particular the isoprene oxidation reactions) to reflect more recent understanding of reaction paths under such conditions.

8 CACM HO₂ concentrations (Figure 4i) are substantially lower and HO₂^{*} slightly higher 9 (Figure 4j) in comparison with observed levels of HO₂. In both cases, however, the changes 10 implemented in CACM0.0 have brought CACM mixing ratios into much closer agreement 11 with those simulated by RACM. The two schemes now display virtually identical diurnal 12 profiles. The elevated HO₂^{*} concentrations are most likely the result of the excessive peroxy 13 radical production in CACM discussed above.

14 **4.2.2 High-NO_x**

Model output from CACM is compared to the RACM and CACM0.0 mechanisms for high-15 NO_x conditions in Figure 6. For most species considered here, the alterations to the scheme 16 make little difference to modelled mixing ratios. The biggest changes occur at night, with the 17 increased $RO_2^* + RO_2^*$ stimulating night-time chemistry. This results in greater overnight 18 losses of the primary terpenoids (Figures 6a,b) and increased MVK+MCR production (Figure 19 20 6c) in particular. Although MVK+MCR concentrations remain well above those simulated in 21 RACM, formaldehyde concentrations (Figure 6d) are in much closer agreement. Concentrations of the HO_x-NO_x oxidant species are also brought more in line with the RACM 22 23 output, with a marked increase in concentrations of both NO₂ (Figure 6e) and NO (Figure 6f) as the $\mathrm{RO_2}^*$ and $\mathrm{HO_2}$ reaction channels become competitive at relatively higher levels of $\mathrm{NO_x}$. 24

HO₂ concentrations (Figure 6i) are reduced to levels in line with those in RACM, but the most notable change is in the simulation of HO_2^* (Figure 6j). Not only are the absolute levels in excellent agreement with RACM, the diurnal profile is now also a good match, with the tendency to over-accumulate isoprene peroxy radicals at night seen in CACM0.0 removed due to the increased night-time peroxy radical loss via the $RO_2^* + RO_2^*$ reactions.

1 4.3 Secondary Organic Aerosols

2 We applied the updated CACM gas-phase chemistry with the MPMPO aerosol module to simulate the production of SOA for the same two-day period under the observed low-NO_x 3 conditions. For the simulations, primary organic aerosol (POA) concentration was set at a 4 constant value of 0.5 μ g m⁻³, consistent with simulated background concentrations during July 5 for the region (Barsanti et al., 2013). Observed hourly submicron particle size distribution 6 7 data for the simulation period, interpolated to 30-minute intervals, were used to calculate 8 volume-weight sedimentation velocities; aerosol aqueous phase pH was set at 4, consistent 9 with the high sulfate to ammonium ratio measured at PROPHET during CABINEX (VanReken et al., 2015). Calculated hourly aerosol liquid water content (ALWC) data, also 10 interpolated to 30-minute intervals, based on hourly observed particle size distributions, CCN 11 12 concentrations, and ambient relative humidity (see Supplementary Material), were used as input to MPMPO. The lowest model layer was initialized with 2 μ g m⁻³ of condensable gases 13 split equally among the 99 condensable species; above the first layer, initial concentrations 14 15 decreased exponentially with height (see Table S4).

16 Figure 7 shows the vertical and temporal profiles of predicted total (gas and aerosol phases) and aerosol-phase concentrations of condensable bVOC oxidation products. Time series of all 17 18 condensable species and selected categories at 835 m (model layer 24) are shown in Figure 8. In Figure 7, the sum of gas- and aerosol-phase concentrations represents the total semi- and 19 non-volatile material simulated by CACM. The oxidation of biogenic emissions produces up 20 to $\sim 3 \,\mu g \, m^{-3}$ (or $\sim 300 \, ppt$) of condensable material from within the canopy to the top of the 21 daytime boundary layer at ~ 1 km above the ground. The two-day CACM-MPMPO 22 simulations indicate that below ~400 m condensable material tends to accumulate during the 23 24 night and early morning and decrease slightly around noon. This diurnal pattern is consistent 25 with the accumulation of oxidation products, especially from monoterpenes which are emitted throughout the day, and the decomposition of PAN and non-PAN alkyl nitrates during noon 26 27 and early afternoon. Between ~ 400 m and ~1 km, however, modeled concentrations increase 28 continuously in time. This accumulation may be an artifact of initial concentrations being too 29 low but cannot be verified due to lack of observational data.

Generally, between 5 and 25% of the condensable material partitions into the aerosol phase,
with the highest SOA concentrations occurring around ~ 900 m (Figure 8), which is near the
mixed layer height and coincides with the buildup of keto-propanoic acid from oxidation of

MVK. The model exhibits the tendency to accumulate SOA as in the case of gas-phase 1 2 oxidation products. Among the bVOC precursors considered in CACM, oxidation products of 3 *d*-limonene, which are predominantly in surrogate group S5 (biogenic, non-dissociative) with 4 some in group S4 (biogenic, dissociative), contribute the largest amount of condensable 5 material (maximum of about 150 ppt or 50%). However, only a small portion of S5 partitions into the aerosol phase as it is non-dissociative and has relatively high vapor pressure. 6 7 Surrogate group S12 contributes to 20-50% of the SOA. This group represents non-volatile 8 dimers of multifunctional acids from monoterpene oxidation and it starts accumulating after 9 sunrise on the first day of simulation. Surrogate group S1, which consists predominantly of MVK oxidation product keto-propanoic acid, contributes to 20-50% of the SOA. The highest 10 11 contribution from S1 and highest total concentrations of SOA occur during the second half of August 5 as MVK concentrations build up from isoprene oxidation and the aerosol water 12 13 content is high enough to draw oxalic acid, the surrogate species for group S1, into the aerosol 14 aqueous phase.

15 The only reported data set of aerosol composition at UMBS as measured by an Aerosol Mass 16 Spectrometer is the data taken during the PROPHET 2001 field campaign from July to early August of 2001. Organic aerosol concentrations within and near the canopy top varied from 17 below 1 µg m⁻³ during clean periods to 3.5 µg m⁻³ at peak of polluted events (Delia, 2004). 18 19 There are no data available for total organic aerosol concentrations at UMBS during 20 CABINEX. Submicron aerosol size distributions, CCN concentrations and water-soluble aerosol components, including water-soluble organic carbon (WSOC), sampled from the 21 22 understory (6 m) of PROHET during CABINEX are reported in VanReken et al. (2015). During CABINEX WSOC concentrations averaged 2.5 ± 2.9 µg C m⁻³ (approximately 5.2) 23 $\pm 6.1 \ \mu g \ m^{-3}$ assuming carbon mass to total organic mass ratio of 2.1), much higher than Delia 24 25 (2004) observed in 2001; however, concentrations were often below detection limits during CABINEX. The large standard deviation relative to the mean is due to the high temporal 26 variability. For the two-day simulation period, anthropogenic influences were small and 27 observed WSOC concentrations ranged from 0.4 to 6.4 C μ g m⁻³ (9 two-hour samples). The 28 CACM-MPMPO predictions of less than $1 \mu g m^{-3}$ in the canopy are therefore an 29 30 underestimation. One reason for the underestimation is that the model currently does not 31 include explicit treatment of SOA from isoprene, despite the buildup of the S1 surrogate from 32 MVK oxidation. Alternatively the over-prediction of temperatures at both the mid-day peak 33 and at night could result in a higher portion of condensable species remaining in the gas

phase. Uncertainties in aqueous-phase pH and POA concentrations (associated with advection) may also contribute to the underestimation. Incorporation of an explicit treatment of SOA formation from isoprene and sesquiterpene oxidations and detailed evaluation with more comprehensive sets of gas, aerosol, and meteorological measurements, such as those from BEACHON (Ortega et al., 2014) and SOAS (e.g. Xu et al., 2015; Nguyen et al., 2014) are needed to elucidate the mechanism for SOA formation and to better understand measuredmodeled discrepancies.

8 **5** Conclusions

9 The 1-D CACHE canopy model (Forkel et al., 2006; Bryan et al., 2012) has been updated to include a modified version of the CACM gas-phase chemistry scheme (Griffin et al., 2002; 10 Chen et al., 2005) and MPMPO aerosol partitioning mechanism (Griffin et al., 2003; Chen et 11 al., 2005). This new model, FORCAsT 1.0, is one of the few canopy exchange models that 12 incorporate both the gas-phase oxidation of VOCs and the production of condensable 13 14 products that can lead to SOA formation. Thus FORCAsT represents a substantial step forward in canopy-atmosphere exchange modeling, with the potential to significantly enhance 15 16 our understanding of the processes involved, their relative importance under different regimes, and the ability to validate our knowledge against site-specific measurement data. 17 18 Recent laboratory experiments and field measurement campaigns have shown that we still 19 lack understanding of many of the fundamental processes involved in the exchange of gases 20 and particles between the forest canopy and atmospheric boundary layer: from primary emissions (e.g. Jardine et al., 2013), to VOC oxidation chemistry (e.g. Rohrer et al., 2014; 21 22 Perring et al., 2013, Surratt et al., 2014; Mellouki et al, 2015), to deposition of reactive species (e.g. Nguyen et al., 2015) and the mechanisms of turbulent vertical exchange (e.g. 23 24 Steiner et al., 2011). It is only through the application of 1-D canopy models such as FORCAsT, in which all of the processes are prognostically included, that we can fully 25 investigate the relative importance of each of these processes and assess the validity of 26 proposed mechanisms. Insights gained from the application of FORCAsT can be used to 27 improve 3-D models of regional and global atmospheric chemistry and climate. 28

Previous evaluation of model performance at the UMBS field station for the CABINEX field campaign (Bryan et al., 2012) demonstrated that its predecessor CACHE was able to reproduce environmental conditions at the site. We show here that FORCAsT 1.0 also effectively reproduces mixing ratios of many key species associated with the oxidation of

bVOCs. However, the initial performance of the CACM0.0 chemistry scheme was poor under 1 2 the low-NOx, high isoprene conditions found at UMBS and substantial modifications were 3 made, in particular to the handling of peroxy radical oxidation and organic nitrate formation 4 in order to improve the performance of CACM for low NO_x environments. Given the 5 substantial NO_x emissions decreases due to implementation of emissions control strategies in many mid-latitude areas it will become increasingly important in future applications to 6 7 address lower NO_x scenarios in many rural and even urban areas previously considered to be 8 high-NO_x regions.

9 The sensitivity studies and chemistry mechanism updates included here have provided valuable insight into the importance of peroxy radicals and organic nitrates in VOC oxidation 10 11 under low-NOx conditions, and further suggest that nighttime chemistry plays a vital role in controlling the oxidative capacity of the atmosphere within and above forest ecosystems. We 12 13 find that peroxy radical self and cross reactions dominate VOC degradation under low-NOx 14 conditions, but due to complexity are necessarily crudely modelled either by considering a 15 small subsection of the possible permutations or by representing many peroxy radicals as a single species. This study points to the urgent need to constrain concentrations of key short-16 17 lived radical species such as organic peroxy radicals in and above forest ecosystems, and to elucidate the mechanisms and processes governing their production and loss. 18

19 Discrepancies between observed and simulated concentrations of the primary HOx-NOx 20 oxidants and a tendency to accumulate the products of VOC oxidation, in particular methyl vinyl ketone and methacrolein, and formaldehyde (see e.g. Ganzeveld et al., 2008) suggest 21 22 that further improvement is required in the representation of gas-phase reaction pathways 23 under low-NO_x conditions to better capture the degradation of VOCs and formation of SOA 24 in such environments. Future development work for FORCAsT includes additional 25 improvements in its simulation of gas-phase chemistry and SOA formation under low- NO_x 26 conditions, viz.:

- Updating the isoprene oxidation scheme to include the production of isoprene epoxide and
subsequent formation of SOA (see e.g. Paulot et al., 2009, Surratt et al., 2006); regeneration
of HOx via HPALD (see e.g. Peeters et al., 2009; Crounse et al., 2011); formation of SOA
from methacrolein (see e.g. Carlton et al., 2009);

Including primary emissions and atmospheric oxidation of MBO (2-methyl-3-buten-2-ol),
 known to influence atmospheric oxidative capacity and ozone production similarly to

- 1 isoprene (see e.g. Steiner et al., 2007) and recently shown to produce SOA via MBO epoxides
- 2 and 2,3-dihydroxyisopentanol (see e.g. Mael, et al., 2015, Zhang et al., 2014);
- Including primary emissions and reactions of key sesquiterpenes (β-caryophyllene, and αfarnesene), a highly reactive group of compounds with high SOA yields (see e.g. Lee et al.,
 2006a,b).

New knowledge of the mechanisms of production and loss of VOCs and their oxidation 6 7 products gained from theoretical and experimental studies will also be incorporated. 8 FORCAsT will be extensively tested against gas-phase and aerosol measurements from field 9 and long-term campaigns from many more sites under a spectrum of NO_x concentrations. It is through fully integrated field measurement-modelling campaigns, the establishment of long-10 term comprehensive measurement networks and datasets, and the application of 1-D canopy 11 12 exchange models such as FORCAsT 1.0 that the biosphere-atmosphere community will gain insight into the fundamental processes involved. 13

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17 Code Availability

18 FORCAsT 1.0 is available by request to the corresponding author. Users of the code will be

- 19 asked to cite this work, and include appropriate references to CACHE, CUPID and CACM-
- 20 MPMPO, in publications based on its application.

1 References

- Atkinson, R., Arey, J.: Gas-phase tropospheric chemistry of biogenic volatile organic
 compounds: a review, *Atmos. Environ.*, 37 (Supplement No. 2), S197-219, 2003. doi:
 10.1016.S1352-2310(03)00391-1
- 5 Baldocchi, D.: A Multi-layer model for estimating sulfur dioxide deposition to a deciduous
 6 oak forest canopy, *Atmos. Environ.*, 22 (5), 869-884, 1988.
- Barsanti, K. C., Carlton, A. G. and Chung, S. H.: Analyzing experimental data and model
 parameters: implications for predictions of SOA using chemical transport models, *Atmos. Chem. Phys.*, 13(23), 12073–12088, **2013**.
- 10 Beaver, M. R., St Clair, J. M., Paulot, F., Spencer, K. M., Crounse, J. D., LaFranchi, B. W.,
- 11 Min, K. E., Pusede, S. E., Wooldridge, P. J., Schade, G. W., Park, C., Cohen, R. C.,
- 12 Wennberg, P. O.: Importance of biogenic precursors to the budget of organic nitrates:
- 13 observations of multifunctional organic nitrates by CIMS and TD-LIF during BEARPEX
- 14 2009, Atmos. Chem. Phys., 12 (13), 5773-5785, 2012.
- Blackadar, A.K.: Vertical distribution of wind and turbulent exchange in a neutral
 atmosphere, J. Geophys. Res., 67 (8), 3095-3102, 1962. doi: 10.1029/JZ067i008p03095
- 17 Boy, M., Sogachev, A., Lauros, J., Zhou, L., Guenther, A., Smolander, S., SOSA a new
- 18 model to simulate the concentrations of organic vapours and sulphuric acid inside the ABL -
- 19 Part 1: Model description and initial evaluation, *Atmos. Chem. Phys.*, 11, 43-51, **2011**.
- 20 Brown, S. S., deGouw, J. A., Warneke, C., Ryerson, T. B., Dube, W. P., Atlas, E., Weber, R.
- 21 J., Peltier, R. E., Neuman, J. A., Roberts, J. M., Swanson, A., Flocke, F., McKeen, S. A.,
- 22 Brioude, J., Sommariva, R., Trainer, M., Fehsenfeld, F. C., Ravishankara, A. R.: Nocturnal
- 23 isoprene oxidation over the Northeast United States in summer and its impact on reactive
- nitrogen partitioning and secondary organic aerosol, *Atmos. Chem. Phys.*, 9 (9), 3027-3042,
 2009.
- Browne, E. C., Cohen, R. C.: Effects of biogenic nitrate chemistry on the NO_x lifetime in
 remote continental regions, *Atmos. Chem. Phys.*, 12 (24), 11917-11932, 2012.
- 28 Bryan, A. M., Bertman, S. B., Carroll, M. A., Dusanter, S., Edwards, G. D., Forkel, R.,
- 29 Griffith, S., Guenther, A. B., Hansen, R. F., Helmig, D., Jobson, B. T., Keutsch, F. N., Lefer,
- 30 B. L., Pressley, S. N., Shepson, P. B., Stevens, P. S. and Steiner, A. L.: In-canopy gas-phase

- 1 chemistry during CABINEX 2009: sensitivity of a 1-D canopy model to vertical mixing and
- 2 isoprene chemistry, *Atmos. Chem. Phys.*, 12 (18), 8829-8849, **2012**.
- Carroll, M.A., Bertman, S.B., Shepson, P.B., Overview of the Program for Research on
 Oxidants: PHotochemistry, Emissions, and Transport (PROPHET) summer 1998
 measurements intensive, *J. Geophys. Res.*, 106(D20), 24275-24288, 2001.
- 6 Carlton, A.G., Wiedinmyer, C., Kroll, J.H.: A review of secondary organic aerosol (SOA)
 7 formation from isoprene, *Atmos. Chem. Phys.*, 9, 4987-5005, 2009. doi: 10.5194/acp-9-49878 2009
- 9 Chen, J. and Griffin, R. J.: Modeling secondary organic aerosol formation from oxidation of 10 α -pinene, β -pinene, and *d*-limonene, *Atmos. Environ.*, 39(40), 7731–7744,
- 11 doi:10.1016/j.atmosenv.2005.05.049, **2005**.
- 12 Chen, J., Mao, H., Talbot, R. W. and Griffin, R. J.: Application of the CACM and MPMPO
- modules using the CMAQ model for the eastern United States, *J. Geophys. Res.*, 111, 12 PP.,
 doi:200610.1029/2006JD007603, **2006**.
- Chen, J., Griffin, R.J., Grini, A., Tulet, P.: Modeling secondary organic aerosol formation
 through cloud processing of organic compounds, *Atmos. Chem. Phys.*, 7, 5343-5355, 2007.
- 17 Chen, J.J., Ying, Q., Kleeman, M.J.,: Source apportionment of wintertime secondary organic
- aerosol during the California regional PM10/PM2.5 air quality study, Atmos. Environ., 44
- 19 (10), 1331-1340, **2010**. doi: 10.1016/j.atmosenv.2009.07.010
- Crounse, J.D., Paulot, F., Kjaergaard, H.G., Wennberg, P.O.: Peroxy radical isomerization in
 the oxidation of isoprene, *Phys. Chem. Chem. Phys.*, 13, 13607–13613, 2011. doi:
 10.1039/c1cp21330j
- 23 Delia A.: Real-Time Measurements of Non-Refractory Particle Composition and Interactions
- 24 at Rural and Semi-Rural Sites, PhD Thesis, University of Colorado-Boulder, 2004.
- Emmerson, K. M. and Evans, M. J.: Comparison of tropospheric gas-phase chemistry schemes for use within global models, *Atmos. Chem. Phys.*, 9, 1831-1845, **2009**.
- 27 Forkel, R., Klemm, O., Graus, M., Rappenglück, B., Stockwell, W. R., Grabmer, W., Held, A.,
- Hansel, A., and Steinbrecher, R.: Trace gas exchange and gas phase chemistry in a Norway
- 29 spruce forest: A study with a coupled 1-dimensional canopy atmospheric chemistry emission
- 30 model, Atmos. Environ., 40, 28-42, 2006. doi:10.1016/j.atmosenv.2005.11.070

- 1 Fredenslund, A., J. Gmehling, and P. Rasmussen, Vapor-Liquid Equilibrium Using UNIFAC,
- 2 Elsevier Sci., New York, 1977.
- 3 Friedlingston, P. and Prentice, I.C.: Carbon-climate feedbacks: a review of model and 4 observation based estimates, *Curr. Opin. Environ. Sustainability*, 2 (4), 251-257, **2010**.
- 5 Fuchs, H., Bohn, B., Hofzumahaus, A., Holland, F., Lu, K. D., Nehr, S., Rohrer, F., and
- 6 Wahner, A.: Detection of HO₂ by laser-induced fluorescence: calibration and interferences
- 7 from RO₂ radicals, *Atmos. Meas. Tech.*, 4, 1209–1225, **2011**. doi:10.5194/amt-4-1209-2011
- 8 Ganzeveld, L.N., Lelieveld, J., Dentener, F.J., Krol, M.C., Roelofs, G.J.: Atmosphere-
- 9 biosphere trace gas exchanges simulated with a single-column model, J. Geophys. Res., 107
- 10 (D16), 4297, **2002**. doi: 10.1029/2001JD000684
- 11 Ganzeveld, L., Eerdekens, G., Feig, G., Fischer, H., Harder, H., Konigstedt, R., Kubistin, D.,
- 12 Martinez, M., Meixner, F. X, Scheeren, H. A., Sinha, V., Taraborrelli, D., Williams, J., de
- 13 Arellano, J. V.-G., Lelieveld, J.: Surface and boundary layer exchanges of volatile organic
- 14 compounds, nitrogen oxides and ozone during the GABRIEL campaign, Atmos. Chem. Phys.,
- 15 8 (20), 6223-6243, **2008**.
- Gao, W., Wesely, M.L., Doskey, P.V.: Numerical modeling of the turbulent diffusion and
 chemistry of NO_x, O₃, isoprene, and other reactive trace gases in and above a forest canopy, *J. Geophys. Res.*, 98(D10), 18339–18353, 1993. doi:10.1029/93JD01862
- Geiger, H., Barnes, I., Bejan, I., Benter, T., Spittler, M., The tropospheric degradation of
 isoprene: an updated module for the regional atmospheric chemistry mechanism, *Atmos. Environ.*, 37 (11), 1503-1519, 2003.
- Goel, N.S., Strebel, D.E. Simple beta distribution representation of leaf orientation in
 vegetation canopies. *Agron. J.*, 76 (5), 800-802, **1984**.
- Grace, J., Cox, P. and Meir, P: The influence of terrestrial ecosystems on climate, *Trends Ecol. & Evol.*, 21 (5), 254-260, 2006.
- 26 Griffin, R.J., Dabdub, D., Seinfeld, J.H.: Secondary organic aerosol 1. Atmospheric chemical
- 27 mechanism for production of molecular constituents, J. Geophys. Res., 107(D17), 4332, 2002.
- 28 doi:10.1029/2001JD000541.

- 1 Griffin, R.J., Nguyen, K., Dabdub, D., Seinfeld, J.H.: A Coupled Hydrophobic-Hydrophilic
- 2 Model for Predicting Secondary Organic Aerosol Formation, J. Atmos. Chem., 44, 171-190,
- 3 **2003**.
- 4 Griffin, R. J., Dabdub, D. and Seinfeld, J. H.: Development and initial evaluation of a 5 dynamic species-resolved model for gas phase chemistry and size-resolved gas/particle 6 partitioning associated with secondary organic aerosol formation, *J. Geophys. Res.*, 110, 16
- 7 PP., doi:200510.1029/2004JD005219, **2005**.
- 8 Griffith, S. M., Hansen R. F., Dusanter, S., Stevens, P. S., Alaghmand, M., Bertman, S. B.,
- 9 Carroll, M. A., Erickson, M., Galloway, M., Grossberg, N., Hottle, J., Hou, J., Jobson, B. T.,
- 10 Kammrath, A., Keutsch, F. N., Lefer, B. L., Mielke, L. H., O'Brien, A., Shepson, P. B.,
- 11 Thurlow, M., Wallace, W., Zhang, N., and Zhou, X. L. OH and HO₂ radical chemistry during
- 12 PROPHET 2008 and CABINEX 2009 Part 1: Measurements and model comparison, Atmos.
- 13 *Chem. Phys.*, 13, 5403–5423, **2013**.
- 14 Guenther, A., Hewitt, C. N., Erickson, D., Fall, R., Geron, C., Graedel, T., Harley, P.,
- 15 Klinger, L., Lerdau, M., Mckay, W. A., Pierce, T., Scholes, B., Steinbrecher, R., Tallamraju,
- 16 R., Taylor, J., and Zimmerman, P.: A global model of natural volatile organic compound
- 17 emissions, J. Geophys. Res., 100, 8873-8892, **1995**. doi: 10.1029/94JD02950
- 18 Harris, J. C., and M. J. Hayes, Acid dissociation constant, in Handbook of Chemical Property
- 19 Estimation Methods: Environmental Behavior of Or- ganic Compounds, edited by W. J.
- 20 Lyman, W. F. Reehl, and D. H. Rosenblatt, McGraw-Hill, New York, 1982.
- 21 Jenkin, M.E., Saunders, S. M., Pilling, M. J.: The tropospheric degradation of volatile organic
- compounds: a protocol for mechanism development, Atmos. Environ., 31 (1), 81-104, 1997.
- 23 doi:10.1016/S1352-2310(96)00105-7
- 24 Karl, T., Harley, P., Emmons, L., Thornton, B., Guenther, A., Basu, C., Turnipseed, A.,
- 25 Jardine, K.: Efficient atmospheric cleansing of oxidized organic trace gases by vegetation,
- 26 Science, 330, 6005, 816-819, **2010**. 10.1126/science.1192534
- 27 Kanawade, V. P., Jobson, B. T., Guenther, A. B., Erupe, M. E., Pressley, S. N., Tripathi, S.
- 28 N., and Lee, S.-H.: Isoprene suppression of new particle formation in a mixed deciduous
- 29 forest, Atmos. Chem. Phys., 11, 6013-6027, 2011

- 1 Kim, S., Guenther, A., Karl, T. and Greenberg, J.: Contributions of primary and secondary
- 2 biogenic VOC total OH reactivity during the CABINEX (Community Atmosphere-Biosphere
- 3 INteractions Experiments)-09 field campaign, Atmos. Chem. Phys., 11, 8613-8623, 2011.
- Laothawornkitkul, J., Taylor, J.E., Paul, N.D., Hewitt, C. N.: Biogenic volatile organic
 compounds in the Earth system, *New Phytol.*, 183 (1), 27-51, 2009. doi: 10.1111/j.14698137.2009.02859.x
- Lee, A., Goldstein, A.H., Keywood, M.D., Gao, S., Varutbangkul, V., Bahreini, R., Ng, N.L.,
 Flagan, R.C., Seinfeld, J.H.: Gas-phase products and secondary aerosol yields from the
 ozonolysis of ten different terpenes, *J. Geophys. Res.*, 111, D07302, 2006a.
 doi:10.1029/2005JD006437
- 11 Lee, A., Goldstein, A.H., Kroll, J.H., Ng, N.L., Varutbangkul, V., Flagan, R.C., Seinfeld, J.H:
- 12 Gas-phase products and secondary aerosol yields from the photooxidation of 16 different
- 13 terpenes, J. Geophys. Res., 111, D17305, **2006b**. doi:10.1029/2006JD007050
- 14 Lelieveld, J., Butler, T. M., Crowley, J. N., Dillon, T. J., Fischer, H., Ganzeveld, L., Harder,
- 15 H., Lawrence, M. G., Martinez, M., Taraborrelli, D. and Williams, J.: Atmospheric oxidation
- 16 capacity sustained by a tropical forest, *Nature*, 452 (7188), 737-740, **2008**. doi:
- 17 10.1038/nature06870
- Liu, Y. J., Herdlinger-Blatt, I., McKinney, K. A., and Martin, S. T.: Production of methyl
 vinyl ketone and methacrolein via the hydroperoxyl pathway of isoprene oxidation, *Atmos. Chem. Phys.*, 13, 5715-5730, 2013.
- Lu, K. D., Rohrer, F., Holland, F., Fuchs, H., Bohn, B., Brauers, T., Chang, C. C., Haeseler,
 R., Hu, M., Kita, K., Kondo, Y., Li, X., Lou, S. R., Nehr, S., Shao, M., Zeng, L. M., Wahner,
 A., Zhang, Y. H., Hofzumahaus, A.: Observation and modelling of OH and HO2
 concentrations in the Pearl River Delta 2006: a missing OH source in a VOC rich atmosphere, *Atmos. Chem. Phys.*, 12 (3), 1541-1569, 2012.
- 26 Mael, L.E., Jacobs, M.I., Elrod, M.J.: Organosulfate and nitrate formation and reactivity from
- 27 epoxides derived from 2-methyl-3-buten-2-ol, J. Phys. Chem. A, 119, 4464-4472, 2015.
- 28 doi.org/10.1021/jp510033s

- 1 Mellouki, A., Wallington, T. J., Chen, J.: Atmospheric chemistry of oxygenated Volatile
- Organic Compounds: Impacts on air quality and climate, *Chem. Rev.*, 115, 3984–4014, 2015.
 doi:10.1021/cr500549n
- Meng, Z.Y., Dabdub, D., Seinfeld, J.H.: Size-resolved and chemically resolved model of
 atmospheric aerosol dynamics, *J. Geophys. Res.*, 103 (D3), 3419-3435, 1998.
 doi:10.1029/97JD02796
- 7 Meyers, T. P. and Baldocchi, D. D.: A comparison of models for deriving dry deposition
- 8 fluxes of O_3 and SO_2 to a forest canopy, *Tellus B*, 40B, 270–284, **1988**. doi:10.1111/j.1600-
- 9 0889.1988.tb00297.x
- 10 Miyoshi, A., Hatakeyama, S. and Washida, N.: OH radical-initiated photooxidation of
- 11 isoprene: An estimate of global CO production, J. Geophys. Res., 99, 18,779-18,787, 1994.
- 12 doi:10.1029/94JD01334
- Müller, J.-F., Peeters, J. and Stavrakou, T.: Fast photolysis of carbonyl nitrates from isoprene, *Atmos. Chem. Phys.*, 14 (5), 2497-2508, 2014. *Ind. Eng. Chem. Res.* doi: 10.1021/ie9502421
- Nguyen, T. K. V., Petters, M. D., Suda, S. R., Guo, H., Weber, R. J., Carlton, A. G.: Trends in
 particle-phase liquid water during the Southern Oxidant and Aerosol Study, *Atmos. Chem. Phys.*, 14 (20), 10911-10930, **2014**.
- 18 Nguyen, T.B., Crounse, J.D., Teng, A.P., Clair, J.M.S., Paulot, F., Wolfe, G.M., Wennberg,
- 19 P.O.: Rapid deposition of oxidized biogenic compounds to a temperate forest, Proc. Natl.
- 20 Acad. Sci. U.S.A., 112 (5), E392-E40, 2015. doi: 10.1073/pnas.1418702112
- Norman, J.M. 1979. Modeling the complete crop canopy. p. 249- 277. In B.J. Barfield and
 J.F. Gerber (ed.) Modification of the aerial environment of plants. *ASAE Monogr. Am. Soc.*
- 23 Agric. Eng., St. Joseph, MI.
- Norman, J.M., and G.S. Campbell. 1983. Application of a plant-environment model to
 problems in irrigation. p. 155-188. In D.I. Hillel (ed.) *Advances in irrigation, Vol. II.*Academic Press, New York.
- 27 Ortega, J., Turnipseed, A., Guenther, A. B., Karl, T. G., Day, D. A., Gochis, D., Huffman, J.
- A., Prenni, A. J., Levin, E. J. T., Kreidenweis, S. M., DeMott, P. J., Tobo, Y., Patton, E. G.,
- 29 Hodzic, A., Cui, Y. Y., Harley, P. C., Hornbrook, R. S., Apel, E. C., Monson, R. K., Eller, A.
- 30 S. D., Greenberg, J. P., Barth, M. C., Campuzano-Jost, P., Palm, B. B., Jimenez, J. L., Aiken,

- A. C., Dubey, M. K., Geron, C., Offenberg, J., Ryan, M. G., Fornwalt, P. J., Pryor, S. C.,
 Keutsch, F. N., DiGangi, J. P., Chan, A. W. H., Goldstein, A. H., Wolfe, G. M., Kim, S.,
 Kaser, L., Schnitzhofer, R., Hansel, A., Cantrell, C. A., Mauldin, R. L., Smith, J. N.:
 Overview of the Manitou Experimental Forest Observatory: site description and selected
 science results from 2008 to 2013, *Atmos. Chem. Phys.*, 14 (12), 6345-6367, 2014.
- Pankow, J.F., Seinfeld, J.H., Asher, W.E., Erdakos, G.B.: Modeling the formation of
 secondary organic aerosol. 1. Application of theoretical principles to measurements obtained
 in the alpha-pinene-, beta-pinene-, sabinene-, Delta(3)-carene-, and cyclohexene-ozone
- 9 systems, *Environ. Sci. Technol.*, 35 (6), 1164-1172, **2001**. doi: 10.1021/es001321d
- 10 Pankow, J. F.: An absorptive model of gas-particle partitioning of organic compounds in the
- 11 atmosphere, *Atmos. Environ.*, 28(2), 185–188, doi:10.1016/1352-2310(94)90093-0, **1994**.
- 12 Paulot, F., Crounse, J. D., Kjaergaard, H. G., Kroll, J. H., Seinfeld, J. H., Wennberg, P. O.:
- 13 Isoprene photooxidation: new insights into the production of acids and organic nitrates,
- 14 Atmos. Chem. Phys., 9 (4), 1479-1501, 2009.
- 15 Peeters, J., Nguyen, T. L., Vereecken, L.: HOx radical regeneration in the oxidation of 16 isoprene, *Phys. Chem. Chem. Phys.*, 11 (28), 5935-5939, **2009**. doi: 10.1039/b908511d
- Perring, A. E., Pusede, S. E. and Cohen, R. C.: An Observational Perspective on the
 Atmospheric Impacts of Alkyl and Multifunctional Nitrates on Ozone and Secondary Organic
- 19 Aerosol, *Chem. Rev.*, 113 (8), 5848-5870, **2013**. doi: 10.1021/cr300520x
- 20 Pongratz, J., Reick, C. H., Raddatz, T., Claussen, M.: Biogeophysical versus biogeochemical
- climate response to historical anthropogenic land cover change, *Geophys. Res. Lett.*, 37,
 L08702, 2010. doi: 10.1029/2010GL043010
- Pöschl, Ü., von Kuhlmann, R., Poisson, N., Crutzen, P.J.: Development and intercomparison
 of condensed isoprene oxidation mechanisms for global atmospheric modeling, *J. Atmos.*
- 25 Chem., 37 (1), 29-52, 2000. doi: 10.1023/A:1006391009798
- 26 Pun, B. K., Griffin, R. J., Seigneur, C. and Seinfeld, J. H.: Secondary Organic Aerosol 2.
- Thermodynamic model for gas/particle partitioning of molecular constituents, *J. Geophys. Res.*, 107, 15 PP., doi:200210.1029/2001JD000542, 2002.
- 29 Rohrer, F., Lu, K.D., Hofzumahaus, A., Bohn, B., Brauers, T., Chang, C.C., Fuchs, H.,
- 30 Haseler, R., Holland, F., Hu, M., Kita, K. Kondo, Y., Li, X., Lou, S.R., Oebel, A., Shao, M.,

- Zeng, L.M., Zhu, T., Zhang, Y.H., Wahner, A.: Maximum efficiency in the hydroxyl-radical based self-cleansing of the troposphere, *Nature Geosci.*, 7 (8), 559-563, 2014. doi:
 10.1038/NGEO2199
- Sandu, A. and Sander, R. Technical note: Simulating chemical systems in Fortran90 and
 Matlab with the Kinetic PreProcessor KPP-2.1, *Atmos. Chem. Phys.*, 6, 187-195, 2006.
- 6 Saunders, S.M., Jenkin, M.E., Derwent, R.G., Pilling, M.J.: Protocol for the development of
- 7 the Master Chemical Mechanism, MCM v3 (Part A): tropospheric degradation of non-
- 8 aromatic volatile organic compounds, *Atmos. Chem. Phys.*, 3, 161-180, 2003.
- 9 Saxena, P. and Hildemann, L.M.: Water absorption by organics: Survey of laboratory
- 10 evidence and evaluation of UNIFAC for estimating water activity, *Environ. Sci. Technol.*, 31
 - 11 (11), 3318-3324, **1997**. doi: 10.1021/es9703638
 - 12 Saylor, R. D.: The Atmospheric Chemistry and Canopy Exchange Simulation System
 - 13 (ACCESS): model description and application to a temperate deciduous forest canopy, *Atmos*.
 - 14 *Chem. Phys.*, 13, 693-715, **2013**.
 - Schwarzenbach, R. P., P. M. Gschwend, and D. M. Imboden, Environmental OrganicChemistry, John Wiley, New York, 1993.
 - Seinfeld, J.H., Erdakos, G.B., Asher, W.E., Pankow, J.F.: Modeling the formation of
 secondary organic aerosol (SOA). 2. The predicted effects of relative humidity on aerosol
 formation in the alpha-pinene-, beta-pinene-, sabinene-, Delta(3)-Carene-, and cyclohexeneozone systems, *Environ. Sci. Technol.*, 35 (9), 1806-1817, 2001. doi: 10.1021/es001765+
 - Smith, J. M. and Van Ness, H. C., 1987: Introduction to Chemical Engineering
 Thermodynamics, McGraw-Hill, Inc., New York.
 - Steinbrecher, R., Hauff, K., Hakola, H., and Rössler, J.: A Revised Parameterisation for
 Emission Modelling of Isoprenoids for Boreal Plants, in: Biogenic VOC emissions and
 photochemistry in the boreal regions of Europe –Biphorep, edited by Laurila, T. and Lindfors,
 V., no. 70 in Air pollution research report, pp. 29-43, Commission of the European
 Communities, EUR 18910 EN. EC, Brussels, 1999.
 - 28 Steiner, A. L., Pressley, S. N., Botros, A., Jones, E., Chung, S. H. and Edburg, S. L.: Analysis
 - 29 of coherent structures and atmosphere-canopy coupling strength during the CABINEX field
 - 30 campaign, Atmos. Chem. Phys., 11, 11921-11936, 2011

- 1 Stockwell, W.R., F. Kirchner, F., Kuhn, M., Seefeld, S.: A new mechanism for regional
- 2 atmospheric chemistry modeling, J. Geophys. Res., 102, 25847-25879, 1997. doi:
- 3 10.1029/97JD00849
- 4 Stone, D., Evans, M. J., Edwards, P. M., Commane, R., Ingham, T., Rickard, A. R., Brookes,
- 5 D. M., Hopkins, J., Leigh, R. J., Lewis, A. C., Monks, P. S., Oram, D., Reeves, C. E., Stewart,
- 6 D., Heard, D. E.: Isoprene oxidation mechanisms: measurements and modelling of OH and
- 7 HO2 over a South-East Asian tropical rainforest during the OP3 field campaign, Atmos.
- 8 Chem. Phys., 11 (13), 6749-6771, **2011**.
- 9 Strebel, D. E., Goel, N. S., Ronson, K. J. Two-dimensional leaf orientation distributions,
- 10 *IEEE T. Geosci. Remote*, GE-23 (5), 640-647, **1985**.
- 11 Stroud, C., Makar, P., Karl, T., Guenther, A., Geron, C., Turnipseed, A., Nemitz, E., Baker,
- 12 B., Potosnak, M., Fuentes, J.D.: Role of canopy-scale photochemistry in modifying biogenic-
- 13 atmosphere exchange of reactive terpene species: Results from the CELTIC field study, J.
- 14 Geophys. Res., 110 (D17), D17303, 2005. doi: 10.1029/2005JD005775
- 15 Sumner, AL., Shepson, P.B., Couch, T.L., Thornberry, T., Carroll, M.A., Sillman, S., Pippin,
- 16 M., Bertman, S., Tan, D., Faloona, I., Brune, W., Young, V., Cooper, O., Moody, J.,
- 17 Stockwell, W.: A study of formaldehyde chemistry above a forest canopy, J. Geophys. Res.,

18 106 (D20), 24387-24405, **2001**. doi: 10.1029/2000JD900761

- 19 Surratt, J.D., Murphy, S.M., Kroll, J.H., Ng, N.L., Hildebrandt, L., Sorooshian, A.,
- 20 Szmigielski, R., Vermeylen, R., Maenhaut, W., Claeys, M., Flagan, R.C., Seinfeld, J.H.:
- 21 Chemical composition of secondary organic aerosol formed from the photooxidation of
- 22 isoprene, J. Phys. Chem. A, 110 (31), 9665-9690, 2006. doi: 10.1021/jp061734m
- Suzuki, T., K. Ohtaguchi, and K. Koide, Application of principal components analysis to
 calculate Henry's constant from molecular structure, *Comput. Chem.*, 16, 41–52, 1992.
- 25 VanReken, T. M., Mwaniki, G. R., Wallace, H. W., Pressley, S. N., Erickson, M. H., Jobson,
- 26 B. T., Lamb, B. K.: Influence of air mass origin on aerosol properties at a remote Michigan
- 27 forest site, *Atmos. Environ.*, 107, 35-43, **2015**. doi:10.1016/j.atmosenv.2015.02.027
- 28 Wesely, M.L.: Parameterization of surface resistances to gaseous dry deposition in regional-
- 29 scale numerical-models, Atmos. Environ., 23 (6), 1293-1304, 1989. doi 10.1016/0004-
- 30 6981(89)90153-4

- 1 Whalley, L., Stone, D., Heard, D.: New Insights into the Tropospheric Oxidation of Isoprene:
- 2 Combining Field Measurements, Laboratory Studies, Chemical Modelling and Quantum
- 3 Theory, in: Atmospheric and Aerosol Chemistry, Topics in Current Chemistry Series, 339,
- 4 55-95, Eds McNeill, V.F., Ariya, P.A., Springer-Verlag Berlin, Berlin, Germany, 2014. doi:
- 5 10.1007/128_2012_359
- 6 Williams, J., Fuentes, J., Hofzumahaus, Abbatt, J. (eds.) Community Atmosphere-Biosphere
 7 Interactions Experiment 2009 (CABINEX) [Special Issue]. *Atmos. Chem. Phys.*, 2011.
- 8 Wolfe, G. M., Cantrell, C., Kim, S., Mauldin, R. L., III, Karl, T., Harley, P., Turnipseed, A.,
- 9 Zheng, W., Flocke, F., Apel, E. C., Hornbrook, R. S., Hall, S. R., Ullmann, K., Henry, S. B.,
- 10 DiGangi, J. P., Boyle, E. S., Kaser, L., Schnitzhofer, R., Hansel, A., Graus, M., Nakashima,
- 11 Y., Kajii, Y., Guenther, A., Keutsch, F. N.: Missing peroxy radical sources within a
- summertime ponderosa pine forest, Atmos. Chem. Phys., 14 (9), 4715-4732, 2014.
- 13 Wild, O.: Modelling the global tropospheric ozone budget: exploring the variability in current
- 14 models, Atmos. Chem. Phys., 7 (10), 2643-2660, 2007.
- 15 Wolfe, G. M. and Thornton, J. A.: The Chemistry of Atmosphere-Forest Exchange (CAFE)
- Model Part 1: Model description and characterization, *Atmos. Chem. Phys.*, 11, 77–101,
 2011.
- 18 Wolfe, G. M., Thornton, J. A., Bouvier-Brown, N. C., Goldstein, A. H., Park, J.-H., Mckay,
- 19 M., Matross, D. M., Mao, J., Brune, W. H., LaFranchi, B. W., Browne, E. C., Min, K.-E.,
- 20 Wooldridge, P. J., Cohen, R. C., Crounse, J. D., Faloona, I. C., Gilman, J. B., Kuster, W. C.,
- 21 de Gouw, J. A., Huisman, A., Keutsch, F. N.: The Chemistry of Atmosphere-Forest Exchange
- 22 (CAFE) Model Part 2: Application to BEARPEX-2007 observations, Atmos. Chem. Phys.,
- 23 11 (3), 1269-1294, **2011**.
- 24 Xu, L., Guo, H., Boyd, C. M., Klein, M., Bougiatioti A., Cerully, K. M., Hite, J. R.,
- 25 Isaacman-VanWertz, G., Kreisberg, N. M., Knote, C., Olson, K., Koss, A., Goldstein, A. H.,
- 26 Hering S. V., de Gouw, J., Baumann, K., Lee, S.-H., Nenes A., Weber, R. J., and Ng N. L.:
- 27 Effects of anthropogenic emissions on aerosol formation from isoprene and monoterpenes in
- the southeastern United States, Proc. Natl. Acad. Sci. U.S.A., 112 (1), 37-42, 2015.
- 29 Young, P. J., Archibald, A. T., Bowman, K. W., Lamarque, J.-F., Naik, V., Stevenson, D. S.,
- 30 Tilmes, S., Voulgarakis, A., Wild, O., Bergmann, D., Cameron-Smith, P., Cionni, I., Collins,
- 31 W. J., Dalsøren, S. B., Doherty, R. M., Eyring, V., Faluvegi, G., Horowitz, L. W., Josse, B.,

- Lee, Y. H., MacKenzie, I. A., Nagashima, T., Plummer, D. A., Righi, M., Rumbold, S. T.,
 Skeie, R. B., Shindell, D. T., Strode, S. A., Sudo, K., Szopa, S., and Zeng, G.: Pre-industrial
 to end 21st century projections of tropospheric ozone from the Atmospheric Chemistry and
 Climate Model Intercomparison Project (ACCMIP), *Atmos. Chem. Phys.*, 13, 2063-2090,
 2013.
- Zhang, H., Zhang, Z., Cui, T., Lin, Y.-H., Bhathela, N.A., Ortega, J., Worton, D.R.,
 Goldstein, A.H., Guenther, A.B., Jimenez, J.L., Gold, A., Surratt, J.D.: Secondary Organic
 Aerosol formation via 2-methyl-3-buten-2-ol photooxidation: Evidence of acid-catalyzed
 reactive uptake of epoxides, *Environ. Sci. Technol. Lett.*, 1, 242–247, 2014.
 doi.org/10.1021/ez500055f
- 11 Zhang, N., Zhou, X., Bertman, S., Tang, D., Alaghmand, M., Shepson, P. B. and Carroll, M.
- A.: Measurements of ambient HONO concentrations and vertical HONO flux above a
 northern Michigan forest canopy, *Atmos. Chem. Phys.*, 12, 8285-8296, 2012.
- Zhou, L., Nieminen, T., Mogensen, D., Smolander, S., Rusanen, A., Kulmala, M., and Boy,
 M.: SOSAA a new model to simulate the concentrations of organic vapours, sulphuric acid
 and aerosols inside the ABL Part 2: aerosol dynamics and one case study at a boreal forest
 site, *Boreal Environment Research*, 19, 237-256, **2014**.
- 18

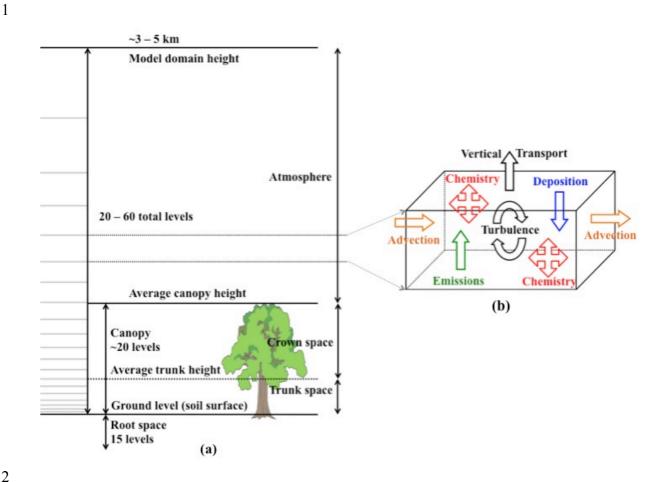
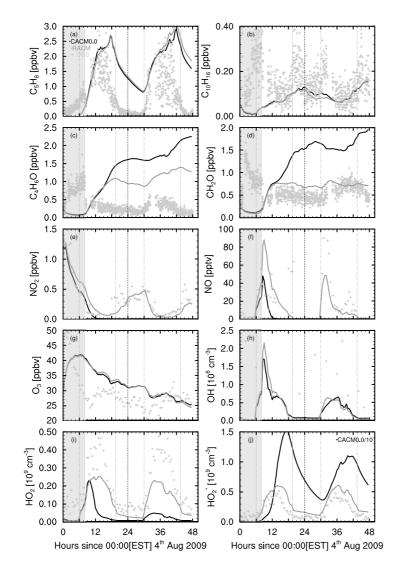


Figure 1. (a) A schematic of the FORCAsT column model. Each level within the column is a box model (b) incorporating the processes involved in canopy-atmosphere exchange of energy and mass appropriate for that level.



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Figure 2. Concentrations of (a) isoprene, (b) summed monoterpenes, (c) MVK+MCR, (d) formaldehyde, (e) NO₂, (f) NO, (g) ozone, (h) OH, (i) HO₂, and (j) HO₂^{*} for Aug 4th-5th, 2009 at the top of the flux tower (corresponding to 36.94 m for model output data and 34 m for measurements). Model data from CACM0.0 is shown in black, and RACM in grey; measurement data are shown by crosses. Note the scale for CACM0.0 in panel (j). The grey shaded region denotes the spin-up period, which is shown here for completeness but is not discussed in the text. Dashed vertical lines mark dawn and dusk.

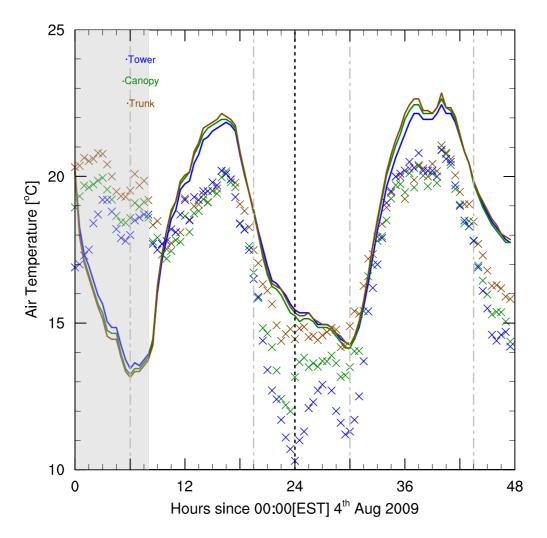


Figure 3. Air temperatures at the trunk height (6 m), canopy top (20.4 m for observations and
19.47 m for FORCAsT output) and tower top (34 m for observations and 36.94 m for
FORCAsT output).

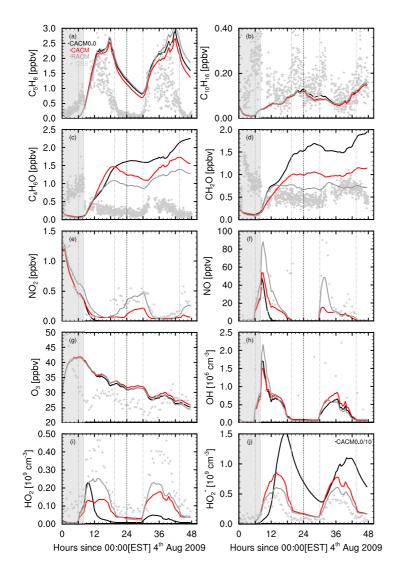
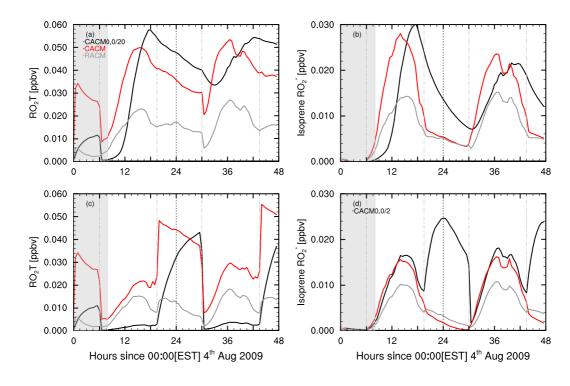


Figure 4. Concentrations of (a) isoprene, (b) summed monoterpenes, (c) MVK+MCR, (d)
formaldehyde, (e) NO₂, (f) NO, (g) ozone, (h) OH, (i) HO₂, and (j) HO₂^{*} for Aug 4th-5th, 2009
at the top of the flux tower (corresponding to 36.94 m for model output data and 34 m for
measurements). Model data from CACM are shown in red; measurement data by crosses.
Data for CACM0.0 (black) and RACM (grey) are shown for comparison.



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Figure 5. Concentrations of peroxy radicals at the top of the canopy (19.47 m for model output data, 20.4 m for observations for low-NO_x conditions (top) and high-NO_x conditions (bottom). The left panels show the total peroxy radical (RO2T) and the right panels peroxy radicals formed from the reactions of isoprene + OH. The data shown are output from the simulation using the optimised CACM chemistry scheme (red) in addition to the original CACM scheme (black) and RACM scheme (grey). Note that the concentrations in the CACM base simulations are scaled by a factor of 20 in panels a-c and by 2 in panel d.

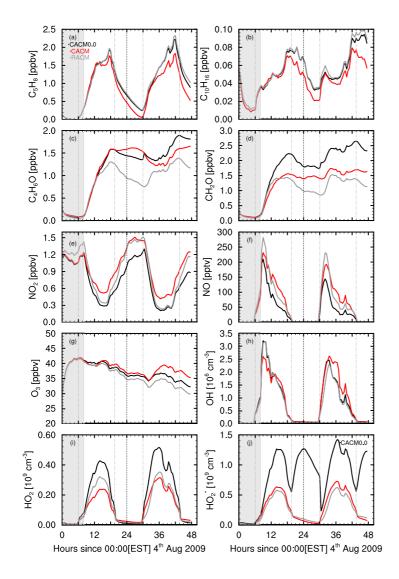
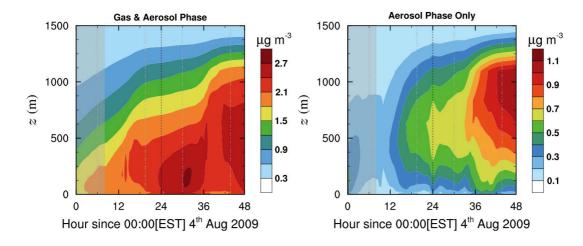
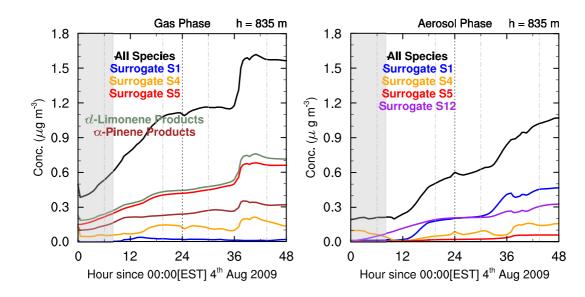


Figure 6. Concentrations of (a) isoprene, (b) summed monoterpenes, (c) MVK+MCR, (d)
formaldehyde, (e) NO₂, (f) NO, (g) ozone, (h) OH, (i) HO₂, and (j) HO₂^{*} for Aug 4th-5th, 2009
at the top of the flux tower (corresponding to 36.94 m for model output data and 34 m for
measurements). Data from CACM are shown in red, RACM grey, and CACM0.0 black.





2 Figure 7. Modelled total condensable (left) and aerosol-phase (right) concentrations.



4 Figure 8. Modelled concentrations at 835 m in the gas (left) and aerosol (aerosol) phases.