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dynamics of northern
vegetation in the
ORCHIDEE
ecosystem model

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Improving the dynamics of northern vegetation in the ORCHIDEE ecosystem model

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Abstract

Processes that describe the distribution of vegetation and ecosystem succession after disturbance are an important component of dynamic global vegetation models (DGVMs). The vegetation dynamics module (ORC-VD) within the process-based ecosystem model ORCHIDEE (Organizing Carbon and Hydrology in Dynamic Ecosystems) has not been updated and evaluated since many years and does not match the progress in modeling the rest of the physical and biogeochemical processes. Therefore, ORC-VD is known to produce unrealistic results. This study presents a new parameterization of ORC-VD for mid-to-high latitude regions in the Northern Hemisphere, including processes that influence the existence, mortality and competition between tree functional types. A new set of metrics is also proposed to quantify the performance of ORC-VD, using up to five different datasets of satellite land cover, forest biomass from remote sensing and inventories, a data-driven estimate of gross primary productivity (GPP) and two gridded datasets of soil organic carbon content. The scoring of ORC-VD derived from these metrics integrates uncertainties in the observational datasets. This multi-dataset evaluation framework is a generic method that could be applied to the evaluation of other DGVM models. The results of the original ORC-VD published in 2005 for mid-to-high latitudes and of the new parameterization are evaluated against the above-described datasets. Significant improvements were found in the modeling of the distribution of tree functional types north of 40° N. Three additional sensitivity runs were carried out to separate the impact of different processes or drivers on simulated vegetation distribution, including soil freezing which limits net primary production through soil moisture availability in the root zone, elevated CO₂ concentration since 1850, and the return frequency of cold climate extremes causing tree mortality during the spin-up phase of the model.

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1 Introduction

The terrestrial biosphere plays an important role in the carbon (Schimel, 1995; Ciais et al., 2013), water (Oki and Kanae, 2006) and energy balance of the Earth (Trenberth et al., 2009). Interactions between vegetation and the atmosphere involve complex biophysical and biogeochemical processes and feedbacks (Heimann and Reichstein, 2008; Foley et al., 2003). To simulate past and future changes on long time scales, Earth system models must represent how the distribution of terrestrial ecosystems adjusts in response to changes in climate, CO₂ and land use. This need provides the motivation for the development of dynamic global vegetation models (DGVM), which couple fast processes (canopy exchange, soil heat and moisture dynamics, photosynthesis), intermediate processes (vegetation phenology, carbon allocation and growth, soil carbon decomposition) and slow processes (vegetation dynamics, recovery from disturbances) to simulate the distribution of vegetation, its carbon stocks and the fluxes exchanged with the atmosphere (Sitch et al., 2003; Krinner et al., 2005). DGVMs have been used to study the response of ecosystems to recent climate change (e.g., Piao et al., 2006) and to project the evolution of the coupled carbon-climate system (e.g., Cox et al., 2000).

The representation of vegetation structural dynamics in DGVMs builds on principles previously applied in biogeography models and “gap models” (Sitch et al., 2003). Biogeography models define the patterns of vegetation physiognomy based on plant functional types (PFT) driven by temperature, precipitation, CO₂, climate-related disturbances, and soil properties (Prentice et al., 1992; Haxeltine and Prentice, 1996). Gap models on the other hand simulate forest dynamics at patch scale, including demographic processes (recruitment, growth, death), competition, and disturbance (Prentice and Leemans, 1990; Bugmann, 2001). The coupling of vegetation dynamics with a climate model allows for the inclusion of vegetation-atmosphere interactions related to ecosystem migration in global climate simulations (Quillet et al., 2010).

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Vegetation distribution largely depends on bioclimatic limits and competition between species, which are regrouped into PFTs in most DGVMs (Woodward, 1987; Sitch et al., 2003; Krinner et al., 2005). Bioclimatic limits consist of direct limiting factors (e.g., minimum temperature for survival) and indirect limitations that control primary productivity and in turn the competitive ability of a PFT (e.g., optimal temperature for photosynthesis, various temperature and moisture phenological controls of leaf-out and senescence). PFTs with a better tolerance to extreme climate conditions and higher growth efficiency during the growing season are more competitive than others, and their distribution will therefore expand at the end of each time step. The competence of any PFT is dependent on the underlying plant traits that define this PFT. In most DGVMs, the traits for a given PFT are fixed, but Verheijen et al. (2013) conducted a variable trait simulation with the JSBACH model for three leaf traits (Specific Leaf Area, and the constants defining the maximum rate of photosynthesis, v_{cmax} , j_{max}) based on observed trait–climate relationships, emphasizing the need for climate-dependent and regional trait-variation modeling. Higgins et al. (2014) pointed out the inherent limitations in Verheijen et al. (2013) using a statistical method to parameterize plant trait diversity, and proposed that the focus should not be on trait values, but rather on the trade-offs between traits (Scheiter et al., 2013). In this study, we will use a fixed trait approach to describe the characteristics of each PFT in ORCHIDEE (the PFTs are listed in Table 1).

ORCHIDEE is the terrestrial surface component of the Institut Pierre Simon Laplace (IPSL) Earth system model. Since the first model description by Krinner et al. (2005), the representation of existing processes has been improved and new processes have been implemented, such as a physically-based multi-layer soil hydrology scheme (de Rosnay et al., 2002), and a scheme describing soil freezing and its effects on root-zone soil moisture and soil thermodynamics (Gouttevin et al., 2012). These new parameterizations have been evaluated for static runs in which PFT maps were prescribed based on observed satellite land-cover information. Yet, their influence on the simulated PFT distribution when the vegetation dynamics module is activated has not been addressed. The original vegetation dynamics module in ORCHIDEE (hereafter “ORC-

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VD”) described by Krinner et al. (2005) was adapted from the LPJ model (Sitch et al., 2003) with minor modifications. Unlike the rest of the model, ORC-VD has not been updated since the Krinner et al. (2005) description, which produces unrealistic results in dynamic runs. For example, Woillez et al. (2011) have shown that the boreal forest area is largely modeled as broadleaf deciduous, whereas in reality it is mainly comprised of needleleaf trees.

The work described here improves ORC-VD, with a focus on Northern Hemisphere vegetation dynamics. Different sets of recent observations have been used to evaluate model performance using quantitative metrics, either related directly to the spatial distribution of vegetation (satellite-observed land-cover and tree fraction) or resulting from it (data-driven spatial distribution of gross primary production (GPP), biomass and soil carbon stocks). The evaluation methodology developed here could be used for other DGVMs as well, and is thus of general interest for the DGVM modeling community.

We present a new parameterization of vegetation dynamics in the ORCHIDEE High Latitude version (ORC-HL) described by Gouttevin et al. (2012), with modifications to the equations and parameters describing tree mortality, thermal constraints and a calibration of $v_{\text{cmax}}/j_{\text{max}}$ (Sect. 2.2). The results of the original module (ORC-HL-OVD) and of the new parameterization (ORC-HL-NVD) are evaluated against different satellite land-cover products, forest inventory data for forest area and biomass, and data-driven GPP and soil carbon products (Sects. 4 and 5). Because the biogeochemical and physical processes that characterize high latitudes interact in a complex way with the processes that control vegetation structure, in Sect. 6 we performed and analyzed factorial model simulations changing one process or driver at a time, to isolate their impacts on vegetation distribution. In addition, because the initial distribution of the vegetation in 1850 is sensitive to pre-industrial climate conditions, we also tested the effect of the return frequency of cold extremes relating to tree mortality during the spin-up phase of the model and discussed its implications.

2 Model description

2.1 ORCHIDEE High Latitude

ORCHIDEE consists of two main modules: SECHIBA (the surface-vegetation-atmosphere transfer scheme) which simulates energy and water exchanges between the atmosphere and land surface at a half-hourly time-step, as well as photosynthesis based on enzyme kinetics (Ducoudré et al., 1993; de Rosnay and Polcher, 1998), and STOMATE (Saclay Toulouse Orsay Model for the Analysis of Terrestrial Ecosystems, Viovy, 1997) which simulates carbon dynamics at a daily time-step, including carbon allocation, biomass accumulation, litter and soil carbon decomposition, and phenology. STOMATE includes a dynamic vegetation module with equations adapted from the LPJ model (Sitch et al., 2003) as described by Krinner et al. (2005).

ORCHIDEE High Latitude version (ORC-HL) is an evolution of ORCHIDEE including additional high latitude processes, described by Gouttevin et al. (2012). In particular, the simple 2-layer soil hydrology (Ducoudré et al., 1993) was replaced by an 11-layer diffusion scheme (de Rosnay et al., 2002), which describes water infiltration and diffusion through soil in a physically-based way. A soil-freezing scheme is implemented in the 11-layer model to calculate liquid and ice water fractions in each soil layer. This scheme has been shown to improve the representation of pan-Arctic river discharge and soil thermal regimes in permafrost regions (Gouttevin et al., 2012). The version number of the ORC-HL used in this study is ORCHIDEE rev1322. Its basic structure is shown in Fig. S1 in the Supplement, in which different processes from Krinner et al. (2005) are marked red.

2.2 Modifications to ORCHIDEE vegetation dynamics

Figure 1 is a schematic of ORC-VD, which simulates the dynamic area covered by each PFT as functions of bioclimatic limitation, competition, mortality and establishment. The modifications made in this study are described in the following, shown red in Fig. 1.

2.2.1 Tree mortality

Mortality is defined as the percentage reduction in population density at the end of each day. The overall tree mortality rate (maximum 1) is the summation of each component including background mortality (M_{BG}), extreme coldness (M_{EC}) and spring frost (M_{SF}) related mortalities, fire-induced mortality, and light competition-induced mortality.

Background mortality

In ORC-HL-OVD, the default calculation of mortality rate for tree PFTs was the inverse of a PFT-specific longevity parameter (30 years for tropical trees, 40 years for temperate trees, 80 years for boreal trees). An alternative calculation in ORC-HL-OVD was a dynamic mortality related to growth efficiency, inherited from LPJ (Sitch et al., 2003):

$$M_{BG} = \left(\frac{k_{BG}}{1 + 0.035V} \right) / 365 \quad (1)$$

where M_{BG} is the dynamic background mortality for tree PFTs (d^{-1}); k_{BG} is maximum background mortality rate (yr^{-1}), set to 0.1 for all tree PFTs in ORC-HL-OVD; and V is vigor or growth efficiency, defined as the ratio of the net annual biomass increment to maximum LAI of the preceding year.

The default calculation defines a constant mortality for each PFT in all grid cells, without considering the variations in mortality of that PFT caused by adaptation to different climate conditions. The dynamic mortality formulation M_{BG} takes into account the influence of growth efficiency on tree mortality, and thus can simulate the competition between tree PFTs under various climates, but it does not consider longevity differences between PFTs. In the new version, ORC-HL-NVD, the dynamic M_{BG} formulation, Eq. (1), is again adopted, but k_{BG} is set to different values for tropical (0.14), temperate (0.1) and boreal (0.05) tree PFTs, proportional to the inverse of their respective longevity in the original ORC-HL-OVD model code.

Tree mortality during extremely cold days

In ORC-HL-OVD, when instantaneous minimum temperature on each day (T_{\min}) drops below a PFT-dependent threshold ($T_{\min,\text{crit}}$, Table 1), the corresponding tree PFT was eliminated instantly. This assumption makes the vegetation distribution highly sensitive to the minimum temperature during a few extremely cold days, which varies from year to year. In reality, trees within a grid cell are unlikely to all die during a single extremely cold event, and moreover, at the resolution at which global models usually run (0.5° or coarser), a single minimum temperature cannot depict the heterogeneity within each grid cell. Therefore, we replaced the original threshold-based LPJ equation by a linearly increasing mortality rate as a function of daily minimum temperature, such that when $T_{\min} < T_{\min,\text{crit}}$

$$M_{\text{EC}} = k_{\text{EC}}(T_{\min,\text{crit}} - T_{\min}) \quad (2)$$

where M_{EC} is mortality caused by extreme coldness in winter (d^{-1}); $k_{\text{EC}} = 0.04$, estimated by trial and error based on the return frequency of below-threshold T_{\min} both within and between years according to the CRU-NCEP climate forcing.

The PFT-specific $T_{\min,\text{crit}}$ (Table 1) confines the distribution of each tree PFT to their adaptable temperature zones. Boreal needleleaf deciduous trees (PFT9) have no $T_{\min,\text{crit}}$ value, meaning that they are insensitive to extreme coldness, and thus can prevail over other boreal tree PFTs in the model in regions with extreme winters such as eastern Siberia.

Broadleaf tree mortality caused by spring frost

Broadleaf species have the specific property of being vulnerable to freezing events that occur after the spring leaf-out. Spring frost can cause damage to leaf buds, developing shoots and flowers, leading to reproductive failure and reduced peak growing-season leaf area index. These effects may result in a natural selection of species with a higher frost resistance, and affect species distribution in the long term (Augsburger, 2009).

Kollas et al. (2013) found that minimum temperature during bud-break was a better predictor of the climate space of seven broadleaf tree species in Europe than winter temperature or mean growing-season temperature.

The change of temperature variability projected by climate models (Cohen et al., 2012; Screen, 2014) may increase or alleviate the risk of spring frost damage. Warmer winters and springs and earlier leaf presence may lead to a greater exposure of mid-latitude broadleaf species to spring frost events (Bokhorst et al., 2009; Gu et al., 2007), while the severity of individual cold spells may also decrease because of a faster warming of the Arctic compared to mid-latitudes (Screen, 2014). DGVMs must therefore represent spring frost induced mortality if they are to account for the response of broadleaf trees to altered climate variability.

We added a frost damage limitation to the distribution of the two broadleaf deciduous tree PFTs (PFT6 and PFT8). After leaf-out in the model, if daily minimum temperature drops below a threshold of -3°C (Kollas et al., 2013), tree mortality is assumed to increase with decreasing temperature. This frost-induced mortality is multiplied by the period elapsed since leaf-out, because the more time that has elapsed, the larger the mass of vulnerable foliage. Thus, during the consecutive 40 days after leaf-out when,

$$T_{\min} < T_{\text{SF,crit}} \quad \text{and} \quad t - t_{\text{leaf-out}} < 40 \text{ days}$$

$M_{\text{SF}}(t)$, the spring frost induced mortality for broadleaf deciduous trees in PFT6 and PFT8 (d^{-1}), is given by:

$$M_{\text{SF}}(t) = 0.01(T_{\text{SF,crit}} - T_{\min}) \left\{ \frac{t - t_{\text{leaf-out}}}{40} \right\} \quad (3)$$

where $T_{\text{SF,crit}} = -3^{\circ}\text{C}$; and $t_{\text{leaf-out}}$ is the day of the year when leaf-out was simulated in the model.

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2.2.2 Growing-season temperature limits to tree extension

In the version of ORCHIDEE described by Krinner et al. (2005), a warm season air temperature (T_{ws}) limit was set to exclude trees from cold Arctic regions, with T_{ws} being required to exceed 7°C for trees to become established or be able to stay at a grid point. T_{ws} was calculated using a linear relaxation method (a substitute for the running mean method to reduce computer memory requirement) given by:

$$T_{ws,t} = \frac{(\tau - \Delta t)T_{ws,(t-\Delta t)} + \Delta t T_{\text{daily}}}{\tau} \quad (4)$$

where Δt = time-step, 1 day; τ = relaxation time of 60 days; and T_{daily} = daily mean air temperature.

In ORC-HL-OVD, used as a starting point for this study, this T_{ws} criterion had been removed. In ORC-HL-NVD, we re-introduced a growing-season temperature criterion to constrain tree extension to Arctic regions, but modified the original formulation using recent results. In their global study of temperature controls on high altitude treelines, Körner et al. (2004) found a growing-season mean soil temperature of $6.7 \pm 0.8^{\circ}\text{C}$ to be the most consistent criterion to predict treelines across different climate zones. Other predictors tested (growing-season length, thermal sums and thermal extremes) were shown to have too large amplitudes and therefore be less suitable indicators of the altitudinal treeline position (Körner et al., 2004). We assumed that the cold limits of trees at both high altitude and high latitude are similar, which is supported by the recent study of Randin et al. (2013), and thus used the Körner et al. (2004) empirical results to re-define the thermal constraint on the existence of trees (treeline) in ORCHIDEE.

Combining the same definition of growing season as Körner et al. (2004), i.e., the period during which 10 cm depth soil temperature exceeds 3.2°C , with their linear relationship between soil temperature in the root zone and canopy air temperature, we prescribe the large-scale thermal limitation of trees in ORC-HL-NVD as follows: mean weekly air temperature during the growing season (T_{GS}) must exceed 7°C , corresponding to $T_{GS,root}$ larger than 6.7°C ; the growing season is calculated as the period when

weekly air temperature is greater than 0°C, which corresponds closely to Troot above 3.2°C. The new T_{GS} criterion shows more consistency with the current treeline positions than the earlier T_{ws} criterion described by Krinner et al. (2005) (Fig. S2).

2.2.3 Modifying v_{cmax} and j_{max}

5 The values of the maximum rate of Rubisco carboxylase ($v_{cmax,opt}$) and maximum rate of photosynthetic electron transport ($j_{max,opt}$) for each PFT were revised using the results of the ORCHIDEE parameter optimization against flux tower measurements from Kuppel et al. (2012). Corresponding values are given in Table 1. In ORC-HL-OVD, v_{cmax} (or j_{max}) is the product of $v_{cmax,opt}$ (or $j_{max,opt}$) and a leaf efficiency factor (e_{rel}), itself
10 determined by relative leaf age (a_{rel}). a_{rel} is defined as the ratio of the calculated leaf age since leaf-out considering four leaf cohorts to a PFT-dependent leaf longevity (a_{crit} in Table 1) (Krinner et al., 2005). As the value of a_{rel} increases with time since $t_{leaf-out}$, e_{rel} increases from 0 to 1 quickly at the beginning of the growing season, and then gradually decreases if $a_{rel} > 0.5$ when leaves become senescent near the end of the
15 growing season. This rule was originally implemented to simulate the influence of seasonal variation in leaf age on photosynthetic activity for all tree PFTs. However, unlike deciduous trees, temperate and boreal evergreen needleleaf trees can keep their needles for 4–6 consecutive years, or even longer for some species (Richardson et al., 2000), resulting in a rather constant leaf age. Thus, we removed the dependence of
20 v_{cmax} and j_{max} on leaf age for temperate and boreal evergreen needleleaf trees (PFTs 4 and 7) in ORC-HL-NVD.

2.3 Code availability

The ORCHIDEE model used as a starting point in this study is ORCHIDEE rev1322. The source code can be obtained at <http://forge.ipsl.jussieu.fr/orchidee/browser/branches/ORCHIDEE-MICT/ORCHIDEE?rev=1322>. A detailed documentation and
25 the forcing data needed to drive ORCHIDEE can be found at <http://forge.ipsl.jussieu>.

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fr/orchidee/wiki/Documentation and <http://forge.ipsl.jussieu.fr/orchidee/wiki/Forcings>. ORC-HL-NVD is derived from rev1322 with the modifications presented in the Sect. 2.2, the source code of which can be obtained upon request.

3 Datasets and methods

3.1 Simulation protocol

Six different runs with ORC-HL (Table 2) were performed to test the impact of the new dynamic vegetation parameterizations and parameter calibrations. Since the modifications in vegetation dynamics module were mainly for temperate and boreal PFTs, the simulation domain is Northern Hemisphere from 20 to 90° N. All runs were conducted at 2° resolution. The climate forcing files were from the 6 hourly CRU-NCEP dataset (http://dods.extra.cea.fr/store/p529viov/cruncep/V4_1901_2012/readme.htm), resampled from their original 0.5° data. CRU-NCEP is widely used as standard climate forcing in current offline terrestrial models, such as MsTMIP (Multi-scale synthesis and Terrestrial Model Intercomparison Project, Huntzinger et al., 2013) and TRENDY (trends in net land-atmosphere carbon exchange over the period 1980–2010). Tests with different resolutions were carried out, showing quite similar results in the simulated vegetation distribution and carbon fluxes and pools (results not shown), indicating that the results presented below do not depend significantly on the spatial resolution of input climate and soil property data within the tested resolution range [0.5, 2°].

Each simulation was preceded by a spin-up from bare ground. For the standard run with the new vegetation dynamics parameterizations (NEW), in spin-up, ORC-HL-NVD was forced by cycling CRU-NCEP 1901–1920 climate data and constant pre-industrial CO₂ concentration (285 ppm) for 250 years. Then the soil carbon sub-model was driven by the previous outputs for 1000 years for the soil carbon pools to reach equilibrium; this was followed by another 50 years of ORC-HL-NVD to complete the spin-up. Each transient simulation from 1850 to 2010 was started from the last year of the spin-up,

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forced by historical CRU-NCEP climate and rising CO₂ concentration. No climate data were available before 1901, so for that period, randomly selected years between 1901 and 1920 were used. The OLD run used the original vegetation dynamics equations from Krinner et al. (2005) in the ORC-HL version so that comparing NEW and OLD allows us to evaluate the improvements listed. The other four runs (EXP1–3, STAT) were similar to NEW except for one different setting for each run (Table 2). In EXP1, we deactivated soil freezing to test its impact on vegetation distribution. In EXP2, we used fixed CO₂ concentration at 285 ppm to test the sensitivity of vegetation distribution to rising CO₂. In EXP3, the model spin-up was forced by the CRU-NCEP 1901–1920 averaged climatology instead of the 20 year cycle, in order to examine the impact of interannual climate variability on the initial PFT distribution after spin-up. In STAT runs, dynamic vegetation was deactivated and a fixed land-cover map was prescribed, in order to separate the effect of simulated vs. observed PFT fractions on GPP, biomass and soil carbon. In STAT1 and STAT2, the PFT map was prescribed from ESA CCI land cover v1.1 (ESA, Bontemps et al., 2013, <http://maps.elie.ucl.ac.be/CCI/viewer/index.php>) and a synergetic land-cover product (SYNMAP, Jung et al., 2006), respectively.

Fires play an important role in determining vegetation patterns by preventing trees from achieving their climate potentials of height, biomass and fractional cover (Bond et al., 2005). Fire occurrence in ORC-HL is formulated using the fire model of Thonicke et al. (2001), based on litter quantity and moisture (Krinner et al., 2005). In this study, the fire module was activated in all the runs. But in a separate test, ORC-HL-NVD was run without the fire module. Compared to NEW, this simulation showed a small increase (5%) in the total temperate and boreal forest area in Northern Hemisphere (20–90° N) without fire. In this study, we used the relatively simple Thonicke et al. (2001) fire module, but compared the results with those obtained with SPITFIRE (Thonicke et al., 2010), a more sophisticated fire model, which explicitly simulates natural and human ignition, fire propagation and fuel combustion (Yue et al., 2014). The average annual burned area during 1981–2010 simulated by the Thonicke et al. (2001) fire module (as implemented in ORC-HL) is 2.7Mkm² in Northern Hemisphere forests,

similar to that simulated by SPITFIRE (2.1 Mkm², implemented in ORCHIDEE standard version).

In this study, agriculture is excluded from all the dynamic runs in order to simulate the potential vegetation distribution without croplands and pasture. The results were post-processed for comparison with observed vegetation cover or carbon stocks. For vegetation cover, this is done by subtracting the observed cropland fraction from the simulated natural PFT fraction in each grid:

$$V_{k,c} = V_{k,c,orig} \times (1 - V_{crop,c}) \quad (5)$$

where $V_{k,c,orig}$ is the model simulated fractional vegetation cover for PFT k (except C3 and C4 crops) and for grid cell c ; $V_{k,c}$ is the fraction of PFT k for grid cell c , after post-processing; and $V_{crop,c}$ is observed fraction of cropland for grid cell c , in this study we use croplands estimated from the ESA land-cover map.

For total GPP and soil carbon stocks, since ORCHIDEE outputs the values per unit PFT, which are multiplied by PFT fractions and summed up to derive the total amount, the results from dynamic runs were post-processed using the following equation (taking GPP as an example), to compare with observational data:

$$GPP_c = \sum_k^n (GPP_{k,c} \times V_{k,c}) + GPP_{crop,c} \times V_{crop,c} \quad (6)$$

where $GPP_{k,c}$ is GPP for natural PFT k and for grid cell c ($gCm^{-2}yr^{-1}$ PFT-1), simulated by dynamic runs; $GPP_{crop,c}$ is GPP for crops (including C3 and C4) for grid cell c ($gCm^{-2}yr^{-1}$ PFT-1), simulated by STAT1 (prescribed from the ESA map); GPP_c is total GPP for grid cell c ($gCm^{-2}yr^{-1}$), after post-processing; and $n = 11$, the number of natural PFTs.

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3.2 Evaluation datasets

We use satellite observations of land cover translated into the PFTs of ORCHIDEE to evaluate the simulated vegetation distribution. In order to account for uncertainties of observations, we used three different land-cover maps: the ESA CCI land cover v1.1 for year 2010, GLC2000 (JRC, 2003) and ISLSCP II vegetation continuous field for 1992–1993 (Defries and Hansen, 2009). The first two land-cover products (hereafter “ESA” and “GLC”) were converted from their original classifications (22 categories based on LCCS system) into PFT maps, using the cross-walking method of Poulter et al. (2011). The third product (hereafter “VCF”) provides the fractional cover of bare ground, herbaceous vegetation and forest (further split into evergreen or deciduous, and broadleaf or needleleaf), and was merged with climate zones of the Köppen–Geiger classification system to resolve to PFT classes, based on Poulter et al. (2011). For Siberia, two additional regional land-cover maps were used, the PFT map of Siberia at 1 km scale from Otlé et al. (2013) based on the GlobCover2005 product (Bicheron et al., 2006), hereafter “OSIB”, and the Russian land-cover dataset produced by International Institute for Applied Systems Analysis (Schepaschenko et al., 2011), hereafter “IIASA”, which was converted into PFT map using the cross-walking method of Poulter et al. (2011). Along with ESA, GLC and VCF, the five land-cover products were used to evaluate the model skill at simulating the vegetation distribution across Siberia. The PFT maps were aggregated at $2^\circ \times 2^\circ$, matching the resolution run by ORCHIDEE in this study. Figure 2 displays an RGB composite-color map of the vegetation fractional cover partitioned between broadleaf (including evergreen and deciduous, red), needleleaf evergreen (green), and needleleaf deciduous (blue) trees, from the five PFT maps.

Simulated GPP was evaluated using the data-derived field obtained from FLUXNET data, satellite fAPAR and gridded climate and land-cover data using a model tree ensemble (Jung et al., 2011), hereafter “MTE”. A recent forest carbon density map (Thurner et al., 2013) for Northern Hemisphere boreal and temperate forests (30–80° N), derived from radar remote sensing of growing-stock volume (GSV), was used

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to evaluate modeled forest biomass. For soil carbon stocks, the simulated soil carbon density was compared with the Harmonized World Soil Database (HWSD, 0–1 m depth, FAO/IIASA/ISRIC/ISSCAS/JRC, 2012) and the Northern Circumpolar Soil Carbon Database (NCSCD, Hugelius et al., 2013). Since the model results for soil carbon are not fully comparable to NCSCD due to lack of peatland carbon accumulation and cryoturbation processes in ORC-HL, metrics were not applied to soil carbon for establishing a model score. All gridded observation-derived data were aggregated at $2^\circ \times 2^\circ$.

Apart from gridded data products based on satellite observations, independent forest inventory data at country/region level as compiled by Pan et al. (2011), including forest area and biomass, were also compared with model results.

3.3 Metrics for model evaluation

Different metrics can be used to quantify the agreement between model results and observations, including Pearson correlation, model-to-data deviation, mean error, root mean square error (see Kelly et al., 2013; Cadule et al., 2010). However, most of these metrics do not consider observational uncertainty. When there are multiple observations available and no particular dataset can be proved to be more accurate than others, which is the case for land cover, the choice of an observational dataset for model evaluation may have a large influence on the model performance score. In order to quantify the agreement between simulated and observed fields, as well as to integrate the uncertainty of observations, a metric normalized by observational uncertainty (Skill, S) was defined to evaluate model performances in terms of PFT fractional cover, GPP and forest biomass. For the following equations, M refers to the model results and O to observational data.

3.3.1 Metrics for PFT fractional abundance evaluation

For PFT fractions, a beta diversity metric (β) was used to calculate the disagreement between two different PFT maps, defined as the Euclidian distance of PFT classes

(Poulter et al., 2011; Otlé et al., 2013). For every grid cell c , beta diversity between model and observational dataset i (β_{c,M_O_i}) was calculated as:

$$\beta_{c,M_O_i} = \sqrt{2 \sum_k^n (V_{k,c,M} - V_{k,c,O_i})^2} \quad (7)$$

where $V_{k,c,M}$ is fractional abundance for PFT k and for grid cell c , simulated by model; V_{k,c,O_i} is fractional abundance for PFT k and for grid cell c , from observational dataset i ; and $n = 11$, the number of natural PFTs.

Similarly, the disagreement between two observations was quantified using $\beta_{c,O_i_O_j}$, as:

$$\beta_{c,O_i_O_j} = \sqrt{2 \sum_k^n (V_{k,c,O_i} - V_{k,c,O_j})^2} \quad (8)$$

where V_{k,c,O_i} and V_{k,c,O_j} are fractional abundances from different observations i and j separately.

β is bound to the interval $[0, \sqrt{2}]$, with higher values representing larger discrepancies between two PFT maps. To take into consideration uncertainties of the different satellite land-cover products (Sect. 3.2), we use the mean β_{c,M_O} of the model vs. all datasets normalized by the mean β_{c,O_O} of all combinations between different datasets. In order to derive a bounded score, with higher values representing better model performance, the metric for the model skill at simulating vegetation distribution in every grid cell ($S_{V,c}$) was defined as:

$$S_{V,c} = \left(\frac{1}{P} \sum_{i \neq j} \beta_{c,O_i_O_j} \right) / \left(\frac{1}{Q} \sum_{i=1} \beta_{c,M_O_i} \right) \quad (9)$$

where P is the number of all combinations between different datasets; Q is the number of datasets.

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Figure 3 shows the spatial pattern of β between the three observational datasets (ESA, GLC and VCF), and mean D among them for different PFT groups. The β between different datasets show a higher agreement for ESA vs. GLC (an average β of 0.25) and lower agreement for VCF vs. ESA or GLC (average β of 0.37 and 0.35 respectively). ESA and GLC legends are based on the FAO Land Cover Classification System (LCCS); while in VCF, the original 1 km continuous field data (DeFries et al., 2000), in which the forest fractional area is given for each grid cell instead of a discrete classification scheme, was aggregated to 0.5° resolution for ISLSCP II under the guidance of IGBP (International Geosphere-Biosphere Programme), by DeFries and Hansen (2009). LCCS uses a low threshold (15 %) of tree cover for forest definition, whereas IGBP uses a threshold of 60 % (Poulter et al., 2011), resulting in relatively lower tree cover in VCF than in either ESA or GLC land-cover maps.

For the PFT groups, higher D values were found for grassland, indicating significant discrepancies in observed grassland fractions. The difference may come from uncertainties in the remotely sensed land-cover products, as well as from uncertainty in the reclassification of land-cover classes into PFT categories. The overlap of broadly defined arid-land classifications (i.e., grassland, shrubland, barren) of land-cover products can introduce errors in partitioning between trees, grass and bare land, in deserts and tundra regions (Poulter et al., 2011).

3.3.2 Metrics for GPP and forest biomass evaluation

GPP and forest biomass were evaluated using gridded observational data containing uncertainty estimates. The metric for model performances was defined as:

$$S_{G,c} \text{ or } S_{B,c} = \frac{\sigma_O}{|X_{c,M} - X_{c,O}|} \quad (11)$$

where $S_{G,c}/S_{B,c}$ is model skill at simulating GPP or forest biomass for grid cell c ; $X_{c,M}$ is GPP or forest biomass for grid cell c , simulated by model; $X_{c,O}$ is GPP or forest biomass for grid cell c , from observation; and σ_O is the SD of the observation.

In grid cells where $X_{c,M} - X_{c,O} < \sigma O$, indicating a model-data difference within the uncertainty of the observational data, $S_{G,c}$ or $S_{B,c}$ is set to 1. The $S_{G,c}$ or $S_{B,c}$ of each grid cell were averaged over the Northern Hemisphere to get an overall score (S_G or S_B).

4 Modeled and observed vegetation distribution

4.1 Northern Hemisphere vegetation distribution

The present-day vegetation distributions simulated by OLD and NEW are shown in Fig. 4 as RGB composite-color maps the same as Fig. 2. Compared with OLD, NEW introduces two major improvements to the results. First, the tree distribution in cold subarctic regions has a northern boundary consistent with observations, mostly due to the introduction of a growing season temperature constraint (Sect. 2.2.2). Second, the observed dominance of needleleaf evergreen trees over broadleaf deciduous trees in northern Europe and North America is reproduced by NEW and not by OLD, an improvement mainly due to the introduction of the spring frost limitation for broadleaf deciduous trees (Eq. 3) and the removal of the v_{cmax} (and j_{max}) leaf-age dependency for evergreen needleleaf trees (Sect. 2.2.3).

Figure 5 displays the spatial pattern of β index for OLD, NEW and different satellite land-cover products. Compared with OLD, the NEW results significantly reduce β in the boreal forests of Canada, western Siberia and northern Europe, consistent with results shown in Fig. 4. The disagreement is also reduced in pan-arctic tundra regions, after correction of the unrealistically high fraction of trees in these regions originally present in OLD. The average β over the Northern Hemisphere land surface (20–90° N, excluding bare ground and agricultural grid cells) for NEW vs. ESA, GLC and VCF are 0.56, 0.48 and 0.47 respectively, equivalent to a 3.5, 13 and 28% reduction (i.e., improvement) compared with OLD. The large variation of β for different observations shows the importance of accounting for observed land-cover uncertainty in DGVM evalua-

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tions, because the arbitrary choice of a specific land-cover product may result in quite different scores.

Accounting for uncertainty in observed PFT maps, the model skill at simulating the vegetation distribution (S_V) is shown in Fig. 6 for OLD and NEW. The average S_V for the major Northern Hemisphere forested countries or regions are listed in Table 3, showing improvement in all countries/regions. Larger improvements of NEW over OLD are found in European Russia (42%), Asian Russia (29%) and Canada (33%). The overall S_V for the Northern Hemisphere is 0.72 in NEW compared to 0.61 in OLD, equivalent to 18% improvement. In OLD, 13% of the land grid cells have a β_{c,M_O} value less than the uncertainty between different satellite products (β_{c,O_O}); in NEW, this fraction increases to 27%.

The forest areas simulated by the dynamic simulations and estimated from the land-cover products were aggregated to country level and compared with independent forest area from national forest inventories (Pan et al., 2011) (Table 4). In OLD, forest areas are systematically overestimated, especially for Asian Russia and Canada. The bias is decreased in NEW, for which most of the differences are less than 30% except for an overestimation in Canada (50%). This overestimation is, however, within the differences between the three land-cover products and the forest inventory data at country scale (Table 4). Forest areas estimated by VCF are systematically lower than inventory data, due to the difference in forest definition mentioned previously. The largest underestimation of VCF occurs in Asian Russia, where the vast taiga-tundra transition zones with relatively sparse trees make the definition-related biases more prominent.

4.2 Distribution of specific groups of Plant Functional Types

For the different PFT groups described in Sect. 3.3.1, the Northern Hemisphere average dissimilarity index (D) is plotted in Fig. 7 for OLD and NEW vs. observations, as well as between different observational datasets. For the broadleaf evergreen group, D is small for both OLD and NEW, and similar to the uncertainty in the data, because the broadleaf evergreen fraction is smaller than other tree PFT groups in temperate

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and cold zones. For the broadleaf deciduous, needleleaf evergreen and needleleaf deciduous groups, the average D for NEW vs. the three datasets is reduced (i.e., improved) by 53, 13 and 67 % respectively, compared with OLD. The OLD overestimation of broadleaf deciduous area in Canada, Scandinavia and European Russia is corrected in NEW (Fig. 8b). The large underestimation of needleleaf evergreen in OLD is partly corrected in NEW, but a significant underestimation of the needleleaf evergreen coverage still exists in southern Siberia and western Canada (Fig. 8c). For needleleaf deciduous, the unrealistically high fractions in subarctic regions in OLD are corrected in NEW, but needleleaf deciduous fractions in southern Siberia and Canada are still higher than observations, at the cost of needleleaf evergreen (Fig. 8d).

A strong disagreement between simulated and observed grassland fractions persists in NEW (average D of 0.35), but the data-data comparison also shows significant discrepancy (average D of 0.19) (Figs. 7 and 3). Since there are no specific shrubland and tundra PFTs in ORCHIDEE, the NEW simulation has high fractions of C3 grass (PFT10) in both arid and cold areas, including subarctic regions, the western USA and the middle of Eurasia (Fig. 8e). The average D for the grass fraction between OLD and observed land-cover maps is 0.27, lower than NEW, because the overestimations of tree cover in OLD decrease the distribution ranges of grassland, leading to a relatively higher agreement with observations for grassland cover than NEW.

4.3 Case study for Siberia, using regional land-cover datasets

For Siberia, the same metrics were calculated based on five observational datasets (ESA, GLC, VCF, OSIB and IIASA). As shown in Fig. 9a, the average β for NEW vs. all datasets is significantly reduced compared to OLD along all longitudes, with a larger reduction (improvement) in central Siberia and the most eastern part of Russia. The average values of β over Siberia for NEW vs. ESA, GLC, VCF, OSIB and IIASA are 0.59, 0.46, 0.38, 0.35 and 0.41 respectively, equivalent to 0, 10, 51, 45 and 26 % reduction compared with OLD, respectively. The average β between different datasets is 0.37, with larger β between ESA and VCF (0.50) and between GLC and VCF (0.47),

ern Siberia (Fig. 11a), indicating that the similarity in total Northern Hemisphere GPP between NEW and MTE masks compensating regional biases. The STAT1 and STAT2 runs produce very similar patterns of GPP to those from NEW (not shown), suggesting that the regional bias of GPP in ORCHIDEE is not related to the modeled PFT distribution, but to other non-modeled factors such as nitrogen interactions.

The model skill at simulating annual GPP (SG) averaged over different countries is given in Table 3. The average SG for the Northern Hemisphere in OLD and NEW are similar (~ 0.6). The improvement in vegetation distribution in NEW does not lead to a significant improvement of GPP, probably because simulated GPP in the same grid cells for high latitudes has only a weak dependence on the modeled PFT. For example, in Canada and northern Europe needleleaf evergreen trees (PFT7) are dominant in NEW, but broadleaf deciduous trees (PFT8) are dominant in OLD, the GPP differences between these two PFTs are less than $1.5 \text{ g C m}^{-2} \text{ yr}^{-1}$ per PFT (or 25 %), explaining why different modeled PFT fractions in this region do not result into large differences in GPP. This result means that GPP is not a discriminant variable for evaluating the performance of a vegetation dynamics module at high latitudes.

5.2 Forest biomass

The country-level forest biomass (above- and belowground) simulated by OLD, NEW and the two static runs with prescribed PFT maps were compared with forest inventory data from Pan et al. (2011) (Table 5). The satellite-based spatially explicit forest biomass estimates from Thurner et al. (2013) over temperate and boreal forests in $30\text{--}80^\circ \text{ N}$ were also aggregated to country level, showing generally good agreement with the data from Pan et al. (2011). The results in NEW are lower than the inventory for all countries, with the largest underestimation by 61 % in Asian Russia. OLD gives a higher total forest biomass in Asian Russia, but the biomass density of OLD and NEW are similar ($\sim 2.4 \text{ kg C m}^{-2}$ forest) and both lower than Pan et al. (2011) (4.1 kg C m^{-2} forest). The large overestimation of biomass in Canada by OLD is reduced in NEW, due to both reductions in forest area (Table 4, from 6.0 to 3.4 M km^2) and in biomass density

(from 5.6 to 3.8 kg C m⁻² forest). Considering the 50 % overestimation of forest area in Canada by NEW compared to the inventory data from Pan et al. (Table 4), the small underestimation (6 %) in total biomass results from a negative bias in biomass density simulation in the model. It is notable, however, that the biomass density in Canada estimated by Thurner et al. (2013) (3.7 kg C m⁻² forest) is also significantly lower than that given by Pan et al. (2011) (6.1 kg C m⁻² forest).

In order to separate the bias of simulated biomass density from the bias of modeled tree cover, the spatial distributions of forest biomass per unit forest area (kg C m⁻² forest) simulated by OLD and NEW are shown in Fig. 11b and compared with the satellite-based estimates by Thurner et al. (2013). The original overestimation in eastern Canada, northern Europe and European Russia by OLD is improved in NEW, although underestimation in western Canada and Siberia still exists in NEW. Biomass at equilibrium is defined by the product of woody-NPP multiplying the turnover time of carbon in biomass pools. Natural disturbances and forest management can thus lower biomass by reducing the turnover time (Jandl et al., 2007; Litton et al., 2004). Since older forests store more biomass carbon than younger forests (Wei et al., 2013; Luysaert et al., 2008), managed and frequently burned forests may not be able to reach their climate-dependent maximum biomass.

In order to diagnose the possible causes of the biomass deviation from data, the ratio of forest biomass from NEW to that from Thurner et al., as well as the ratio of forest NPP (average during 2001–2010) from NEW to MODIS-NPP (NTSG), is plotted in Fig. S4. In eastern Canada, forest biomass is overestimated by NEW, while NPP is close to MODIS NPP, indicating an overestimation of biomass carbon turnover time in ORCHIDEE compared to reality. This bias may be caused by non-modeled forest management in this region. In western Canada and southern Siberia, the underestimation of biomass is attributable to underestimation of NPP.

The model skill at simulating forest biomass (S_B) averaged over different countries is given in Table 3. S_B is improved in NEW for all countries compared to OLD, with the

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largest improvement found in Canada (66%). The overall S_B for 30–80° N is 0.59 in NEW, compared to 0.46 in OLD, equivalent to 28% improvement.

5.3 Soil carbon

The spatial patterns of soil carbon density simulated by OLD and NEW (0–2 m depth) are shown in Fig. 11c, compared with that from HWSO (0–1 m depth, FAO/IIASA/ISRIC/ISSCAS/JRC, 2012) and NCSCD (0–1 m depth, Hugelius et al., 2013). Over the grid cells present in NCSCD, the total soil carbon is 285 Pg in HWSO, markedly lower than that in NCSCD (460 PgC for the upper meter of soil), indicating large uncertainties in the empirical soil carbon data. Since the ORC-HL in this study does not include processes of peatland and wetland carbon accumulation, whereas in NCSCD the peat deposits contain about 30% of the total soil organic carbon mass in the upper meter (Tarnocai et al., 2009), and wetland carbon stock is estimated to account for 20% of the total 1 m-deep soil organic carbon pool in Russia (Schepaschenko et al., 2013), the model results are not fully comparable to NCSCD. The spatial patterns of soil carbon from OLD and NEW are similar (Fig. 11c). Over the grid cells present in NCSCD, the total soil carbon simulated by OLD and NEW is 263 and 283 PgC, respectively.

A comparison of soil carbon simulations from several land surface models coupled with climate models in CMIP5 (Todd-Brown et al., 2013) suggested that most models cannot reproduce grid-scale variation in soil carbon; and that the substantial disagreement between the HWSO and NCSCD datasets and their lack of quantitative uncertainty estimates limit their ability for benchmarking land carbon models. Given the large carbon storage in northern high latitude soils, the ability to accurately simulate high latitude processes such as permafrost, wetland and peatland carbon accumulation, is a prerequisite for realistic projections of future climate-carbon feedbacks.

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6 Critical model processes influencing vegetation distribution

6.1 Soil freezing

The area of seasonally frozen ground covers 50 % of the Northern Hemisphere land, or 48 Mkm² (Zhang et al., 2003). Soil freezing limits plant access to soil moisture, and thus impacts the simulated PFT distribution through a set of complex interactions between productivity, tree-grass competition, and soil water limitations. In permafrost regions, the limitation of growing-season water availability due to soil freeze–thaw processes was shown to substantially contribute to the low vegetation carbon densities (Beer et al., 2007). In ORCHIDEE, a soil heat diffusion equation with latent heat (Gouttevin et al., 2012) is solved for each soil layer that impacts soil temperature and liquid water content. In this study, we tested the effects of soil freezing on the vegetation distribution by comparing NEW and EXP1 in which soil freezing processes were not activated (all other parameters being the same). In EXP1, soil temperature can drop below 0 °C, but liquid water continues to be available in the root zone irrespective of soil temperature conditions. Figure 12 shows the difference in tree fraction and in water availability (WA) during the growing season (May–September) between NEW and EXP1. In the model, soil moisture available to plants is defined by WA, the relative soil moisture in the root zone, weighted by PFT-specific root profiles. A value of WA = 0 defines the wilting point, and WA = 1 the field capacity. A stress factor is applied to stomatal conductance and canopy photosynthesis if WA drops below a critical value of 0.4, and this stress factor increases linearly for $0 < WA \leq 0.4$ (Krinner et al., 2005).

When soil freezes in autumn and winter, the amount of liquid water in the root zone is reduced as water is immobilized as ice in soil pores. In the growing season, WA in NEW is also lower than that in EXP1 (Fig. 12b). This is consistent with previous results of model validation at site scale (Gouttevin et al., 2012), in which the upper layer (0–20 cm) soil moisture in summer was found to be more depleted if the soil freezing module was activated. In regions underlain by permafrost, there is a spring peak in runoff originating from meltwater which does not infiltrate into frozen soils (Gouttevin

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et al., 2012). If soil freezing is not modeled as in EXP1, meltwater will infiltrate into soil, leading to overestimated soil water content in the growing season. The reduction of tree fraction in the presence of freezing occurs where there is significant reduction of WA (Fig. 12a). In areas with a small reduction (less than 0.1) in WA, however, there is a slight increase in tree fraction. The tree fraction in the model is related to population density and individual crown area. On the one hand, as WA decreases, GPP, LAI and crown area are smaller; yet on the other hand, reduced LAI leads to increased available space for establishment, resulting in a subsequent increase in population density, compensating for the loss of crown area.

6.2 Changing CO₂ since 1850

Terrestrial plants respond to elevated atmospheric CO₂ concentration by increasing assimilation rate and reducing diffusive stomatal conductance (Lammertsma et al., 2012), both processes are included in ORCHIDEE (Krinner et al., 2005). Under elevated CO₂ concentration, the enhanced photosynthetic capacity and thus increased NPP of forest (Norby et al., 2005; Hickler et al., 2008) leads to higher growth efficiency of trees and thus higher tree fractional coverage. In the model, tree PFTs are superior to grass PFTs in terms of light competition, i.e., when trees expand, grass PFTs will give way to trees. Therefore, tree cover is expected to increase at the cost of grasslands under elevated CO₂. Here we conducted a sensitivity test (EXP2) with fixed pre-industrial CO₂ concentration (285 ppm). Compared with EXP2, the simulation NEW forced by historical CO₂ concentration produces higher tree fractions (Fig. 13a) by 2010, the spatial pattern of which mirrors the pattern of tree NPP increase (Fig. 13b) in the model. In NEW, total temperate and boreal forest area in the studied region (20–90° N) are modeled to increase by 2.6 Mkm² (11.5 %) from 1850 to 2010. In EXP2 the increase is only 1.1 Mkm² (4.8 %) indicating that about 58 % of the increase in forest area is attributable to the historical increase of CO₂, the rest being attributable to climate warming (longer growing seasons) and changes in rainfall.

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Since the processes of CO₂ uptake by photosynthesis and water loss by transpiration are tightly coupled, increasing CO₂ concentration results in increased water use efficiency (Lammertsma et al., 2012; O'ishi et al., 2009). Figure 13c displays the difference of WA for trees between NEW and EXP2. Compared to the fixed-CO₂ simulation, NEW produces higher WA by ~ 5 % in mesic regions such as Europe, western Siberia and the eastern part of North America, and similar WA in drier regions such as middle and eastern Siberia and the western part of North America.

6.3 Effects of the return frequency and severity of extreme cold events during the spin-up

As mentioned in Sect. 2.2.1, the distribution range of tree PFTs in ORCHIDEE is influenced by extremely cold days in winter that varies from year to year. When the PFT-dependent threshold $T_{\min, \text{crit}}$ (Table 1) is applied (Eq. 2), this mechanism results in a considerable difference in modeled tree fraction between the results of a spin-up forced by cycling multi-year climatic data vs. an average climatology. In EXP3, the model spin-up used the average climatology of the period 1901–1920 from CRU-NCEP, and was compared with NEW where interannually variable climate from years 1901–1920 was repeated in a loop. The minimum temperature in winter (T_{\min}) derived from the climatology is significantly higher than T_{\min} considering the 20 individual years (Fig. 14a). Since the intra-annual variations among different years are not synchronous, a low temperature of a day in one year is offset by a higher temperature of the same day during another year; this leads to a milder climate in the climatology.

The vegetation distributions after spin-up are very different between NEW and EXP3 as shown in Fig. 14. In EXP3, temperate tree PFTs can extend northward, taking up the boreal tree positions, while the distribution of boreal needleleaf evergreen (PFT7) and broadleaf deciduous (PFT8) trees is squeezed to the climatic range of needleleaf deciduous tree (PFT9). The large variance induced by interannual climate variability during the spin-up also holds if one single year is used instead of the multi-year average. Figure 14d shows the considerable difference in the fraction of PFT 7 and 8

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between two spin-ups forced by two different single years arbitrarily chosen (1914 and 1901). Similar results were obtained when three sets of forcings (one-year, climatological mean, and cycling of the whole period 1960–1999) were used in the spin-up process of CLM-DGVM (Li et al., 2011). Since climatology or recycled one-year climatic data are sometimes used in the spin-up of land surface models, it is notable that this may bias DGVMs to produce unrealistic or unstable results, considering the sensitivity of vegetation distribution to extreme temperatures. Thus, it is more appropriate to cycle multi-year climatic data to force DGVMs in a spin-up.

7 Conclusions

This study has presented an improved parameterization and a calibration of Northern Hemisphere vegetation dynamics in the ORCHIDEE process-based ecosystem model, based on a version that includes frozen soil moisture and its impacts on plant productivity. Keeping the original model's concept of plant functional types, we modified the processes that influence tree existence, mortality and competition. A new performance metric applicable for DGVM evaluation in terms of vegetation fractional cover was used to evaluate ORCHIDEE, which integrates uncertainties in different land-cover maps. The new version of the ORCHIDEE vegetation dynamics module shows marked improvement in the simulated PFT distribution compared to the previous version. A more realistic simulation of the northern tree limit is obtained, as well as of the distribution of evergreen and deciduous conifers in the boreal zone. The model still overestimates grass fraction in dry regions of central Asia and western North America, possibly because of the lack of a specific shrubland PFT. Grass fraction was also overestimated in the Arctic tundra. Considering the large coverage of shrubland and tundra in northern middle and high latitudes, a proper representation of shrub and tundra plant functional types in DGVMs, as well as their biophysical and biogeochemical processes, should be a priority for future development. The better PFT distribution results in improvements in simulated forest biomass, while significant regional biases still remain for GPP, forest

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biomass and soil carbon distributions, indicating other structural biases in the carbon cycle parameterizations in the model. Incorporating PFT trait variation into DGVMs, which allows the functional properties to vary within PFTs based on trait–climate relationships, might be a promising method to simulate vegetation acclimation that impacts both vegetation competition and the carbon cycle, and be an interesting future development.

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Table 1. PFT-specific parameters in ORC-HL-NVD.

PFT	$T_{\min,\text{crit}}$	k_{BG}	$v_{\text{cmax,opt}}$	$j_{\text{max,opt}}$	a_{crit}
2: tropical broadleaf evergreen trees	0	0.14	65	130	730
3: tropical broadleaf dry-season deciduous trees	0	0.14	65	130	180
4: temperate needleleaf evergreen trees	−30	0.1	35	70	910
5: temperate broadleaf evergreen trees	−14	0.1	45	90	730
6: temperate broadleaf summergreen trees	−30	0.1	55	110	180
7: boreal needleleaf evergreen trees	−45	0.05	33	66	910
8: boreal broadleaf summergreen trees	−45	0.05	30	60	180
9: boreal needleleaf summergreen trees	−	0.05	35	70	180
10: natural C3 grass	−	−	70	140	120
11: natural C4 grass	−	−	70	140	120
12: agricultural C3 grass	−	−	100	200	90
13: agricultural C4 grass	−	−	100	200	90

$T_{\min,\text{crit}}$: minimum temperature limitation ($^{\circ}\text{C}$), below which the mortality rate will increase as Eq. (2). k_{BG} : maximum background mortality rate (yr^{-1}) for tree PFTs. $v_{\text{cmax,opt}}$: optimal maximum rubisco-limited potential photosynthetic capacity ($\mu\text{mol m}^{-2} \text{s}^{-1}$). $j_{\text{max,opt}}$: maximum rate of photosynthetic electron transport ($\mu\text{mol m}^{-2} \text{s}^{-1}$). a_{crit} : critical leaf age for leaf senescence (days).

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Table 2. Characteristics of each ORC-HL off-line runs. OLD follows the same simulation protocol as NEW. EXP1–3 and STAT is similar to NEW except for one different setting for each run.

Name	Model	Module	Spin-up		Simulation (1850–2010)	
			Climate forcing	CO ₂ level	Climate forcing	CO ₂ level
NEW	ORC-HL-NVD	Activate ORC-VD, soil freezing and fire schemes	CRU-NCEP 1901–1920 cycle	285 ppm	CRU-NCEP 1901–2010 (for 1850–1900: randomly select from 1901–1920)	rising
OLD	ORC-HL-OVD	–	–	–	–	–
EXP1	ORC-HLNVD	Deactivate soil freezing	–	–	–	fixed at 285 ppm
EXP2	ORC-HL-NVD	–	–	–	–	–
EXP3	ORC-HL-NVD	–	CRU-NCEP 1901–1920 average climatology	–	–	–
STAT	ORC-HL-NVD	Deactivate ORC-VD	–	–	–	–

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Table 3. Model skills at simulating vegetation distribution (S_V), GPP (S_G) and forest biomass (S_B), averaged over different countries/regions. STAT1 and STAT2 are static runs prescribing different PFT maps, ESA and SYNMAP.

		Asian Russia	European Russia	Canada	USA	Europe	China	Northern Hemisphere (20–90° N)
Vegetation distribution	OLD	0.69	0.63	0.53	0.66	0.62	0.57	0.61
	NEW	0.89	0.89	0.70	0.69	0.65	0.61	0.72
GPP	OLD	0.53	0.70	0.59	0.63	0.60	0.57	0.63
	NEW	0.58	0.68	0.50	0.65	0.60	0.56	0.62
	STAT1	0.52	0.63	0.63	0.54	0.55	0.53	0.60
	STAT2	0.50	0.68	0.65	0.65	0.53	0.50	0.63
Forest biomass	OLD	0.52	0.54	0.37	0.49	0.49	0.56	0.46
	NEW	0.62	0.73	0.62	0.57	0.55	0.56	0.59
	STAT1	0.58	0.55	0.50	0.53	0.54	0.52	0.56
	STAT2	0.57	0.47	0.46	0.47	0.54	0.53	0.54

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Table 4. Forest areas (Mkm²) for different countries/regions simulated by models (OLD and NEW) and estimated from land cover products (ESA, GLC, VCF), in comparison with that from Pan et al. (2011). The relative differences compared to Pan et al. (2011) are given in parentheses.

	Asian Russia	European Russia	Canada	USA	Europe	China
Pan et al. (2011)	6.77	1.69	2.30	2.57	2.05	1.56
OLD	10.0 (48%)	1.96 (16%)	6.00 (160%)	3.33 (30%)	2.14 (5%)	2.80 (80%)
NEW	5.00 (−26%)	1.80 (7%)	3.44 (50%)	2.61 (2%)	1.56 (−24%)	1.23 (−21%)
ESA	6.54 (−3%)	1.58 (−6%)	3.64 (58%)	3.00 (17%)	1.81 (−12%)	2.19 (41%)
GLC	8.42 (25%)	2.02 (20%)	4.50 (96%)	4.73 (84%)	2.40 (17%)	2.23 (43%)
VCF	3.43 (−49%)	1.18 (−30%)	2.54 (10%)	2.00 (−22%)	1.19 (−42%)	1.10 (−30%)

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Table 5. Forest biomass (PgC) for different countries/regions simulated by models (OLD, NEW and two static runs) and estimated from Thurner et al. (2013), in comparison with that from Pan et al. (2011). STAT1 and STAT2 prescribe different PFT maps, ESA and SYNMAP. The relative differences compared to Pan et al. (2011) are given in parentheses.

	Asian Russia	European Russia	Canada	USA	Europe	China
Pan et al. (2011)	27.9	9.6	14.0	19.4	13.0	6.5
Thurner et al. (2013)	25.2 (−10%)	9.0 (−6%)	15.9 (14%)	–	10.6 (−18%)	–
OLD	24.3 (−13%)	14.7 (53%)	33.4 (138%)	17.7 (−9%)	16.2 (27%)	8.1 (25%)
NEW	11.0 (−61%)	7.6 (−21%)	13.2 (−6%)	12.0 (−38%)	8.3 (−36%)	3.5 (−47%)
STAT1	6.9 (−75%)	10.8 (12%)	21.2 (52%)	8.7 (−55%)	8.0 (−39%)	3.6 (−44%)
STAT2	13.7 (−51%)	15.5 (62%)	36.1 (158%)	17.9 (−8%)	13.8 (6%)	4.5 (−31%)

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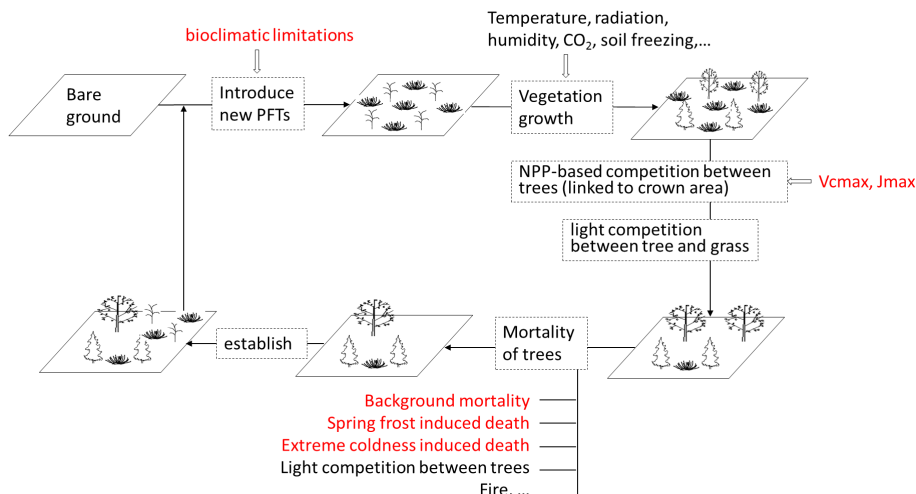


Figure 1. Schematic of ORCHIDEE vegetation dynamics module (ORC-VD). The modifications in this study are marked red.

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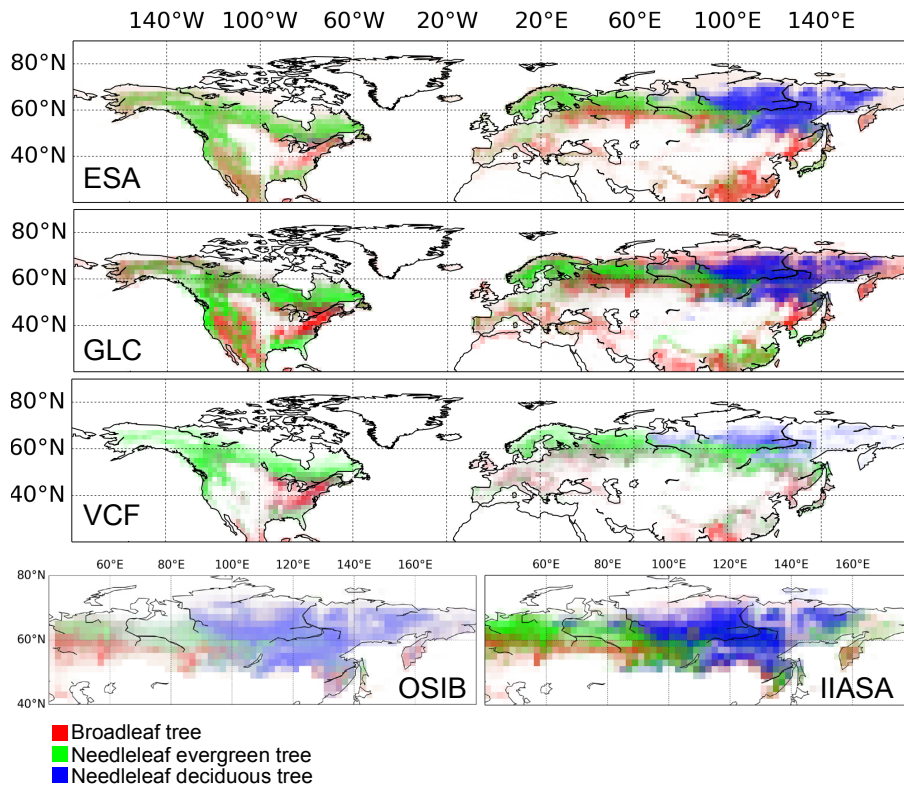


Figure 2. Composite-color map of the fractional vegetation cover in PFT maps converted from the five land-cover products based on Poulter et al. (2011). Color indicates the relative fraction of three PFT groups: broadleaf (including evergreen and deciduous, red), needleleaf evergreen (green), and needleleaf deciduous (blue) trees. Brighter colors represent higher fractional covers.

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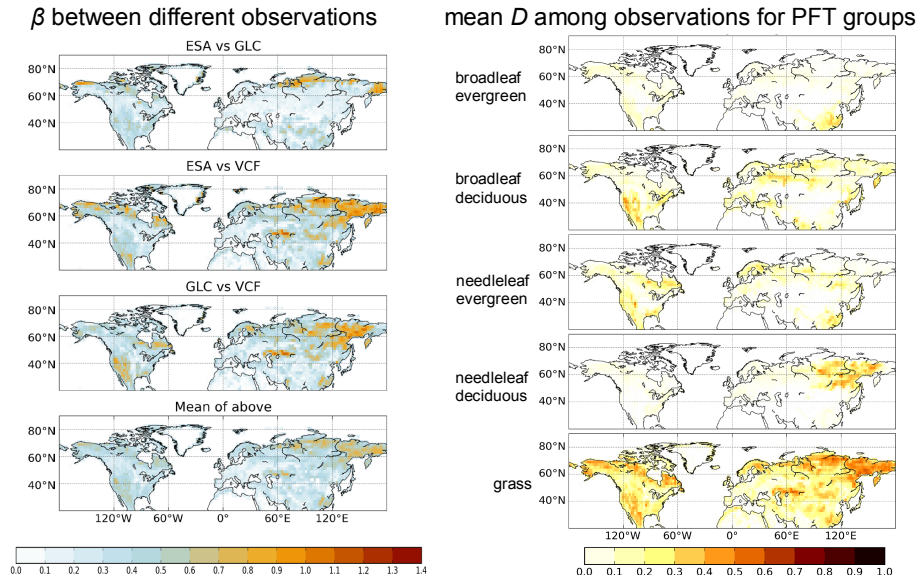


Figure 3. Beta diversity (β) between the three observational datasets (ESA, GLC and VCF) (left panel), and mean dissimilarity index (D) among them for different PFT groups (right panel). β ranges from 0 to $\sqrt{2}$, and D ranges from 0 to 1, both with higher values representing larger disagreement.

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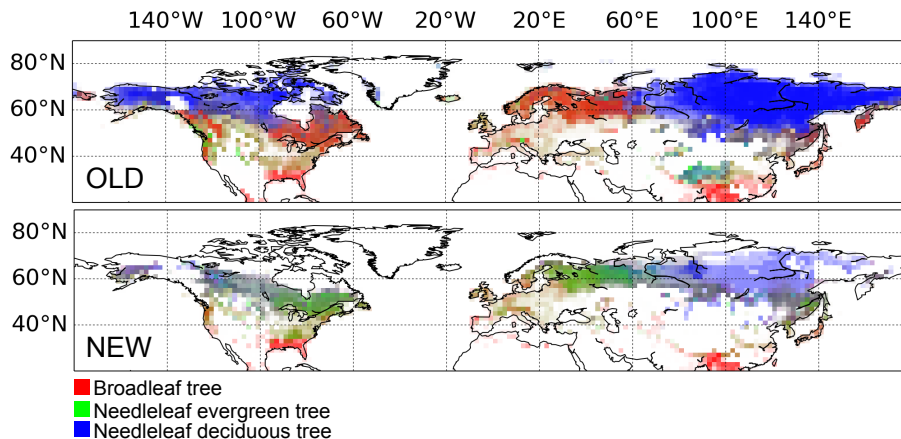


Figure 4. Composite-color map of the fractional vegetation cover in OLD and NEW. Color indicates the relative fraction of three PFT groups: broadleaf (including evergreen and deciduous, red), needleleaf evergreen (green), and needleleaf deciduous (blue) trees.

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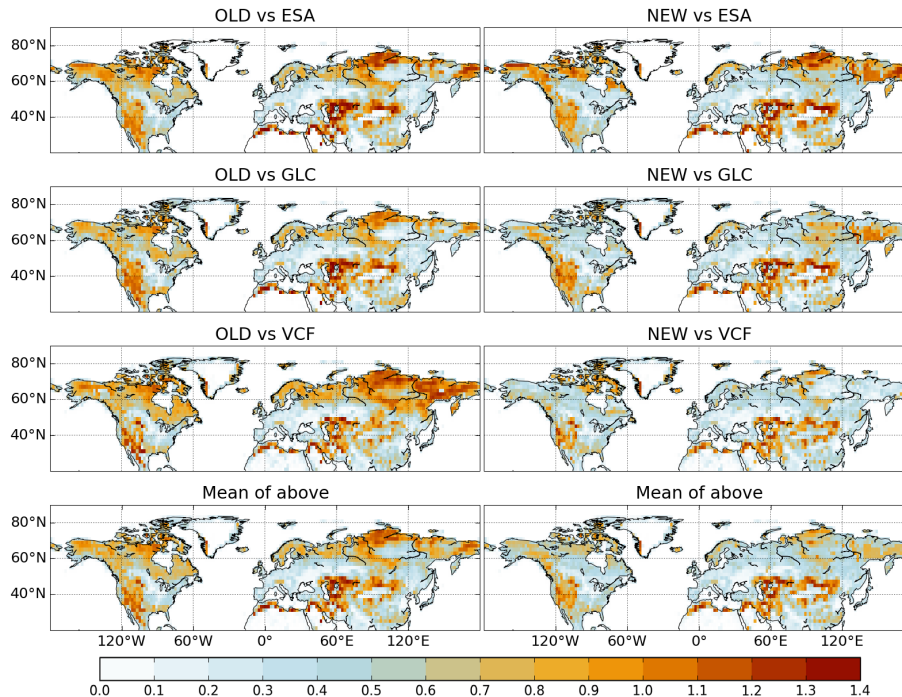


Figure 5. Beta diversity (β) to quantify the disagreement in vegetation distribution between model and observational datasets. β ranges from 0 to $\sqrt{2}$, with higher values representing larger disagreement.

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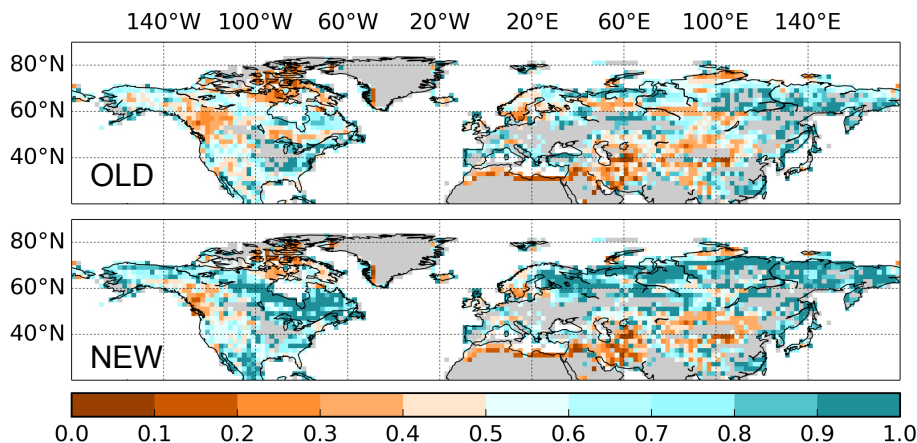


Figure 6. Model skill at simulating vegetation distribution (S_V , Eq. 9) for OLD and NEW. S_V ranges from 0 to 1, with higher values representing better model performances, integrating observational uncertainty. Three kinds of grid cells are masked out (in gray): (1) the grid cells where $S_V > 1$ for both models, indicating that the observational data have too large uncertainties to be qualified for model evaluation (13 % of the total land points for the studied region), (2) the grid cells where all models and datasets have 100 % bare ground in Sahara Desert and Greenland (10 %), and (3) the grid cells where crop fraction is higher than 0.5 (8 %).

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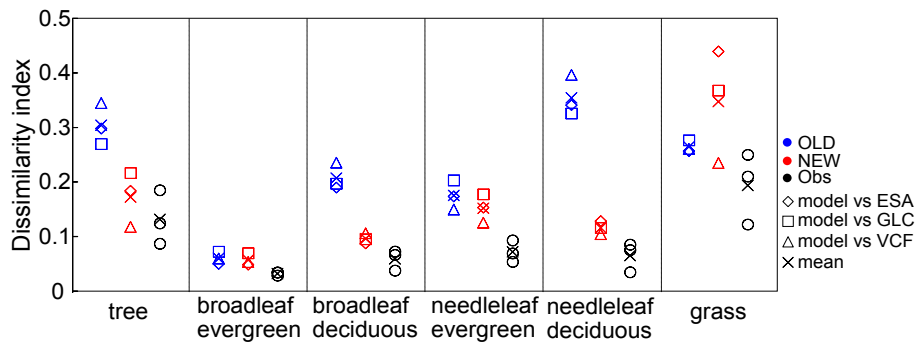


Figure 7. Dissimilarity index (D , Eq. 10) for fractional cover of PFT groups including total tree, grass, and four tree subtypes between model (OLD, blue, and NEW, red) and observations (black), averaged over Northern Hemisphere (20–90° N). D ranges from 0 to 1, with higher values representing larger disagreement.

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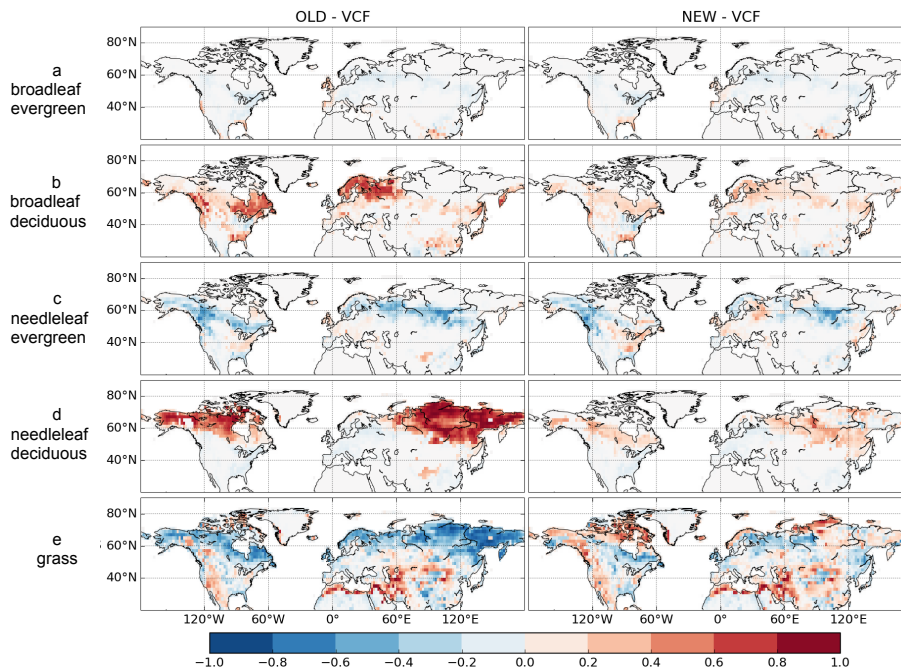


Figure 8. Difference in fractional cover of PFT groups between model (OLD and NEW) and observation-derived PFT map (VCF). Similar map for the difference between model and ESA/GLC is shown in Fig. S3.

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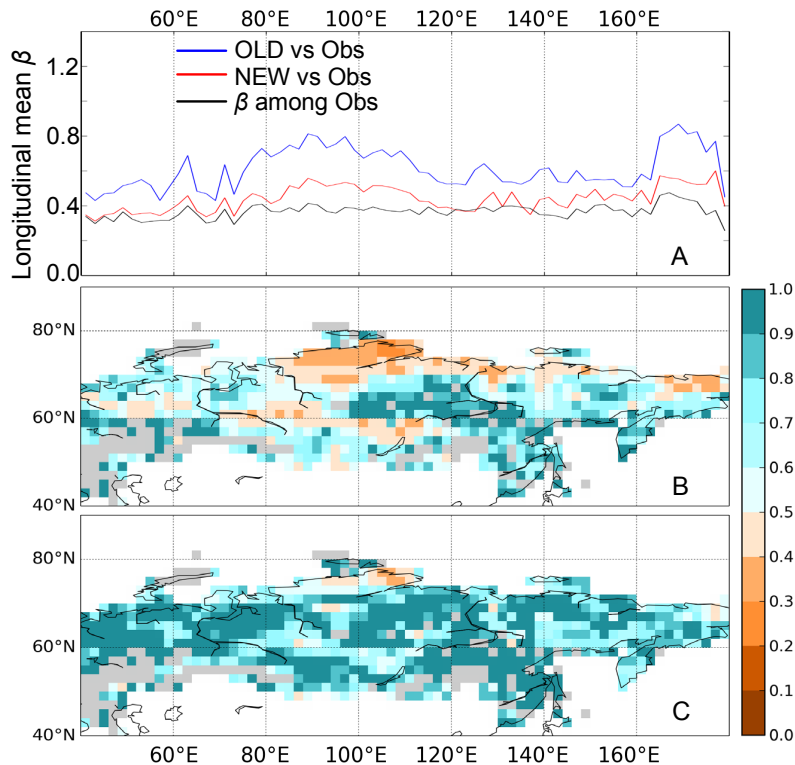


Figure 9. (a) Longitudinal average beta diversity (β) between model (OLD, blue and NEW, red) and observational datasets (including ESA, GLC, VCF, OSIB and IIASA) and between different observations (black) in Siberia. β ranges from 0 to $\sqrt{2}$, with higher values representing larger disagreement. (b) and (c): model skill at simulating vegetation distribution (S_V) for OLD and NEW in Siberia. S_V ranges from 0 to 1, with higher values representing better model performances, integrating observational uncertainty. The pixels where $S_V > 1$ for both models, indicating that the observational data have too large uncertainties to be qualified for model evaluation (12% of the total land points in Siberia), were masked out (in gray).

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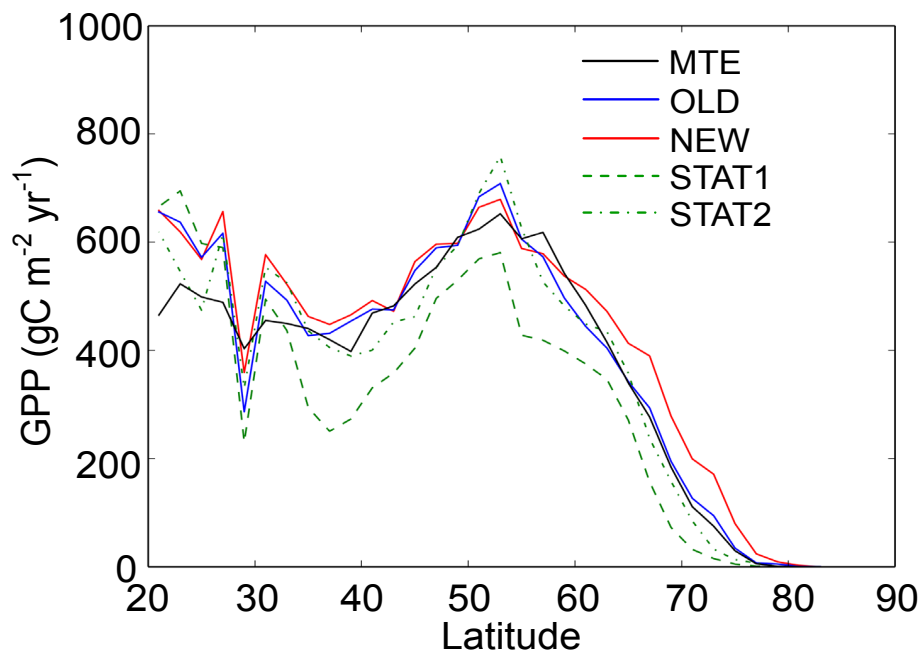


Figure 10. Latitudinal mean annual GPP (2° bands) during 1999–2008 from OLD (blue) and NEW (red), compared with that from STAT (static run in which ORC-VD is deactivated, green dashed lines) and MTE (Jung et al., 2011, black). In STAT1 and STAT2, PFT map is prescribed from ESA and SYNMAP respectively.

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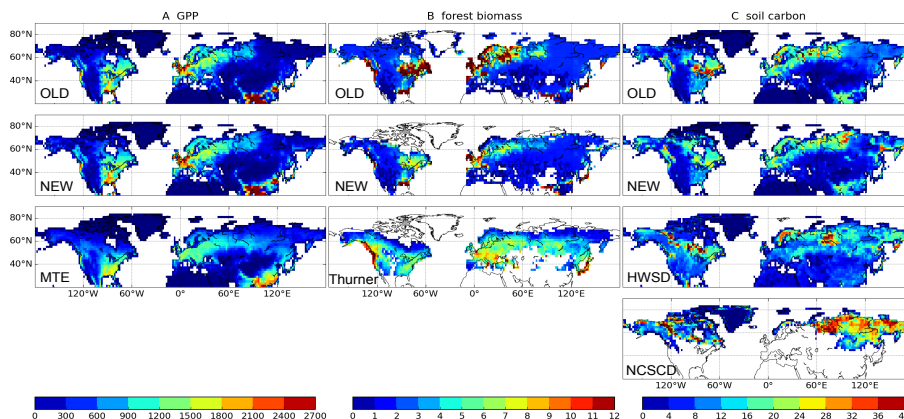


Figure 11. Spatial pattern of **(a)** mean annual GPP ($\text{gC m}^{-2} \text{yr}^{-1}$) during 1999–2008 from OLD, NEW, and MTE (Jung et al., 2011); **(b)** forest biomass density (per forest area, kgC m^{-2} forest) from OLD, NEW and Thurner et al. (2013); and **(c)** total soil carbon density (kgC m^{-2}) simulated by OLD and NEW (0–2 m depth), and from HWSD (0–1 m depth) and NCSCD (0–1 m depth).

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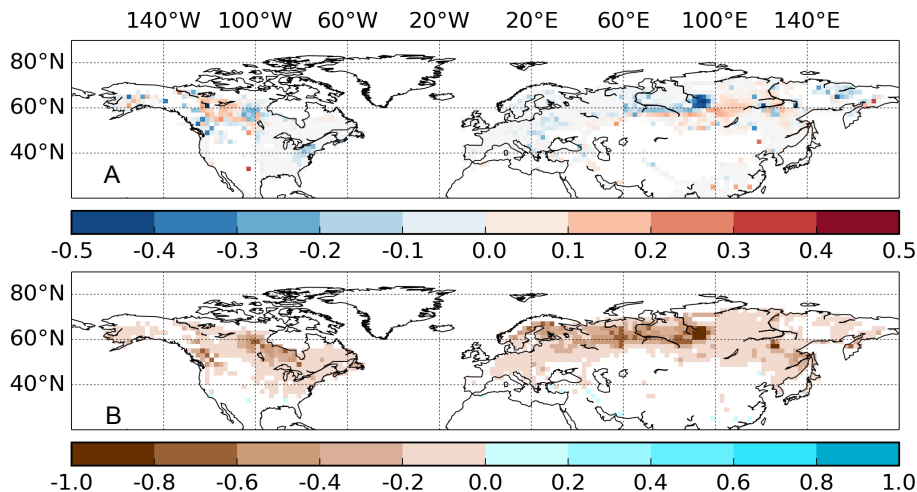


Figure 12. Difference of tree fractional cover (**a**) and water availability (WA, **b**) between with and without soil freezing (NEW-EXP1). WA is averaged over the growing season (May–September) and over tree PFTs (PFT 2–9) weighted by their fractions.

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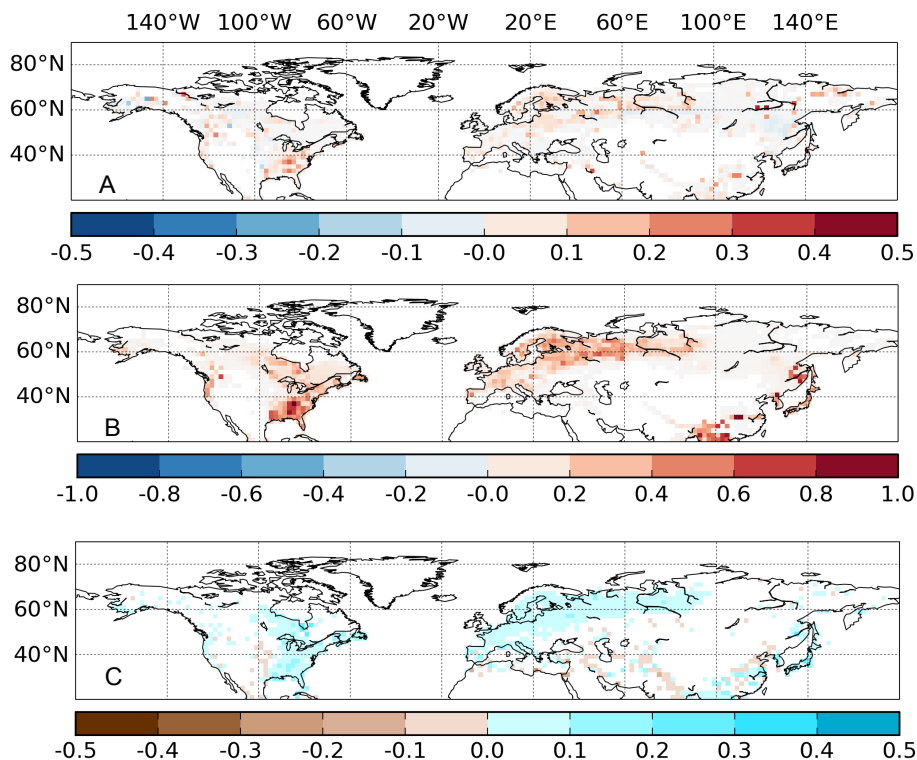


Figure 13. Difference of tree fractional cover (**a**), tree NPP ($\text{gCm}^{-2}\text{yr}^{-1}$, **b**) and water availability (WA, **c**) between with and without CO₂ rising (NEW-EXP2). NPP is averaged over tree PFTs (PFT 2–9) weighted by their fractions. WA is averaged over the growing season (May–September) and over tree PFTs (PFT 2–9) weighted by their fractions.

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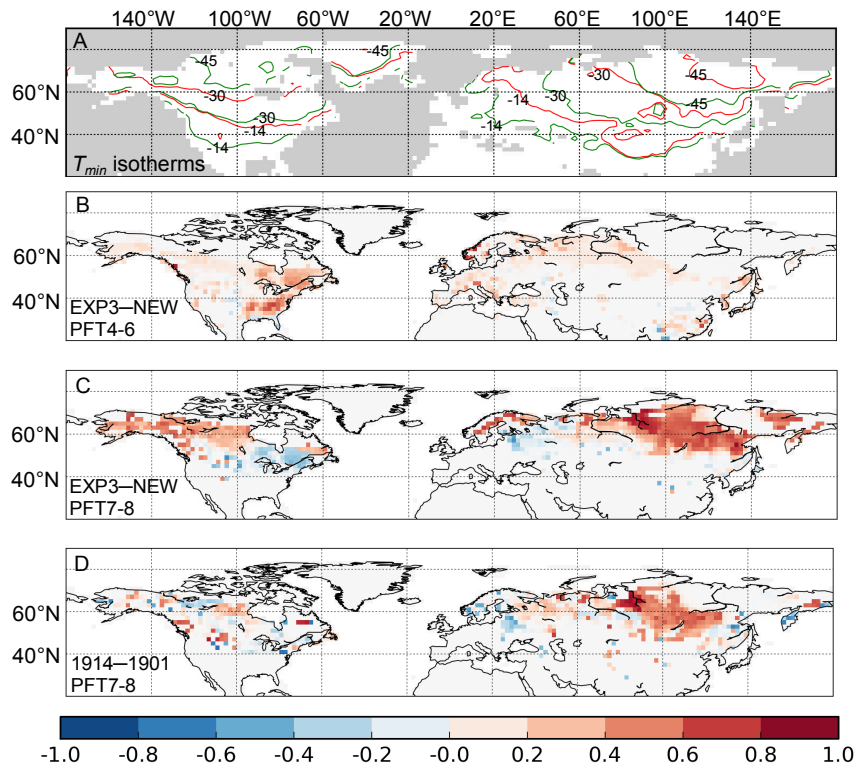


Figure 14. (a) Minimum temperature (T_{min}) isotherms calculated from the 20 year average climatology (red lines) and the mean of the twenty T_{min} for each year (green lines). The T_{min} values are labeled on the lines, corresponding to the PFT-dependent $T_{min,crit}$ for temperate and boreal trees (see Table 1). (b, c) Difference of the vegetation fractional cover between EXP3 (using 20 year climatology as forcing file in spinup) and NEW for temperate trees (PFT4–6, b) and boreal broadleaf deciduous/needleleaf evergreen trees (PFT7–8, c). (d) Difference in fraction of PFT7–8 between spinup results forced by climatic data of two different single years (1914 and 1901).