

1 **Reply to “Geosci. Model Dev. Discuss., 7, C2669–C673, 2014”**

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3 **S. Multsch, J.-F. Exbrayat, M. Kirby, N. R. Viney, H.-G. Frede and L. Breuer**

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5 **Referee #1:** The paper describes an attempt to study uncertainties in irrigation water  
6 requirements simulated for wheat growing in the Murray-Darling Basin (MDB) in  
7 Australia caused by the structure of the model and by the parameters used in it. In  
8 general the paper is well written and interesting but I have serious doubts that the setup  
9 of the study is useful to address the objectives described in the manuscript. My major  
10 points of criticisms are:

11  
12 **Referee #1:** 1.) The authors use six empirical methods to calculate evapotranspiration in  
13 order to study the structural uncertainty of the model and five crop coefficient sets  
14 (needed to convert the ET of a reference crop to the one of a wheat crop) to study the  
15 parameter uncertainty in the model. Results of the so created ensemble of model runs  
16 (6 ET methods x 5 kc-sets) are weighted then according to their performance in  
17 representing measured data to derive a weighted ensemble mean and the ensemble  
18 range around the weighted mean. Such a setup is useful to compare results of complex  
19 models when the knowledge about the accuracy of the models is limited. However, the  
20 methods used to compute ET in this study have been extensively evaluated in previous  
21 research.

22  
23 **Authors:** We agree that the ET methods we investigate have been widely used and  
24 evaluated in ecosystem model applications, but there is no consensus on which ET  
25 method is the best. This is particular the case, if (a) relevant input data are missing to  
26 drive a more sophisticated method, such as Penman-Monteith or Shuttleworth-Wallace.  
27 Moreover, if (b) the scale of the study is regional to global (see p.7529, lines 19-26), the  
28 ET method becomes a major source of uncertainty which is independent from the model  
29 complexity, because every model in hydrology, climate science and crop modelling  
30 relies on an estimate of ET and differences between methods have been reported by  
31 others (see page 7540, lines 5-16).

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33 Despite this general knowledge, most studies use a single ET method, disregarding that  
34 ET simulations impact model predictions to a large extent. An objective method to  
35 evaluate model performance in combination with ensemble predictions could improve  
36 this limitation. The “Reliability Ensemble Averaging (REA)” technique is such a method,  
37 which has been also applied in hydrology to less complex models. We now use it for the  
38 first time to predict irrigation water requirements in this study. Similar to climate science,  
39 where REA has been developed, the accuracy of our model results is difficult to assess,

1 as direct validation data are not available. We are able to show that REA leads to a  
2 decrease of model uncertainty, which is particular important in data scarce regions  
3 where common physically based methods such as the Penman-Monteith approach are  
4 limited in their application. We therefore think that the selected approach is valuable and  
5 provides interesting insight for modellers from a variety of research disciplines.

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7 **Referee #1:** Based on comparisons with lysimeter measurements performed in different  
8 climatic environments it is well known that the Penman-Monteith (PM) method usually  
9 performs best when high quality measurements of all the required weather variables are  
10 available. The other ET-methods used in the study are less precise because they ignore  
11 some important weather variables or relationships between them determining site  
12 specific evapotranspiration. As such, I don't understand the value of applying additional  
13 inaccurate methods and of creating some "artificial" uncertainty. In other words: what is  
14 the additional value of the weighted ensemble mean as compared to the direct use of  
15 the PM method (maybe with site specific adjustment of the resistance terms)?

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17 **Authors:** We agree that Penman-Monteith performs best. But in many model  
18 applications, e.g. simulations for irrigation management, catchment scale modelling or  
19 estimation of large scale water consumption (see p.7529, lines 19-26), other methods  
20 such as Priestly-Taylor and in particular Hargreaves-Samani are in use. These models  
21 give by far different results. We give an example of the "potential" uncertainty which  
22 arises if relevant information on climate data is missing which is a crucial information.  
23 But we also show that REA is a potential interesting approach to reduce this uncertainty,  
24 if more than one ET method is used in simulation studies.

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27 **Referee #1:** 2.) To "evaluate" the performance of the ET methods the authors used  
28 class-A pan data measured at 34 stations (page 7532, lines 16-23). By doing this the  
29 authors evaluate evapotranspiration calculated for a reference crop (ET) by measured  
30 evaporation from an open water surface (E) which is certainly not the same. The  
31 agreement derived from this comparison is then used as a weighting parameter to  
32 compute the weighted ensemble mean. Again, I doubt that this weighted ensemble  
33 mean will represent an improvement to the direct use of the PM-method. But the authors  
34 can test this by comparing the performance of PM56 to the ensemble mean of the other  
35 methods.

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37 **Authors:** For the application of REA, a comparison of model simulations and  
38 observations is needed to calculate the model performance criterion (*page 7535, lines 2-*  
39 *7 "...capability of each ensemble member to represent real world data by its bias."*). We  
40 could have treated PM56 as being an "observation" in the sense of a benchmark model.  
41 However, we think that a more independent test is more appropriate in the sense of

1 REA and therefore decided to use those observations that are at hand: class-A pan  
2 observations. To account for the difference of class-A pan evaporation and reference  
3 crop ET, we used a commonly applied correction factor (pan-coefficients according to  
4 McMahan et al. (2013)) to derive crop ET from class-A pan measurements. Most often,  
5 ET estimates are not compared to any measurements at all, leaving modelers with no  
6 information on how good their model application is. We therefore think that a  
7 comparison to class-A pan is for sure not perfect, but better than no testing at all. This  
8 will be acknowledged in a revised version of the paper.

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11 **Referee #1:** 3.) Parameter uncertainty is evaluated by using 5 sets of crop coefficients.  
12 These coefficients relate the ET of a wheat crop to the ET of the reference crop surface.  
13 The factors describing the differences between the ET of wheat and the one of the  
14 reference grass surface are described in detail in Allen et al. (1998), for example.  
15 Methods to reduce uncertainties in crop coefficients would be to (i) adjust standard crop  
16 coefficients by considering the local conditions (wheat management, wetting interval,  
17 aridity, growing period length) or (ii) using a process based crop model that directly  
18 accounts for the underlying processes. APSIM, for example, has been developed in  
19 Australia and was frequently applied for the local conditions. I doubt that the set of the  
20 crop coefficients used in this study really provides a representative picture on the  
21 expected parameter uncertainty.

22  
23 **Authors:** We are aware that there are more site specific and regionally adapted Kc  
24 values even if crop coefficients are in first instance meant to adjust ETo to a specific  
25 crop type, reflecting albedo, crop height, surface resistance, soil evaporation (Allen et  
26 al., 1998). But in contrast to the widely existing assumption, that better adapted Kc  
27 values lead to improved crop ET estimations, we show that the uncertainty related to the  
28 choice of Kc is small compared to the uncertainty inherent to the ET model structure (or  
29 method) in itself. We highlight the calculation of uncertainty of irrigation water  
30 requirement in this manuscript and a method how to reduce it. The importance of the  
31 consideration of uncertainty has also been reported elsewhere as stated in the  
32 manuscript (page 7543 lines 21-25): "*Despite the growing importance of IRR for today's*  
33 *agriculture and the effect on surface (Hoekstra et al., 2012) and groundwater (Wada et*  
34 *al., 2010) resources, few studies have dealt with the predictive uncertainty of this*  
35 *requirement (e.g. Wada et al., 2013) and how to reduce it.*"

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38 **Referee #1:** 4.) While the authors focus on potential uncertainties caused by the ET  
39 calculation method and the crop coefficients, there is little explanation why these two  
40 factors were selected and how some other factors may affect the uncertainties in  
41 irrigation water requirement calculated in this study.

1 The model applied here uses some very crude assumptions (e.g. that runoff is fixed to  
2 20% of precipitation, see equation 2). In addition, it does not account for the spatial  
3 heterogeneity in soil or crop conditions. From this perspective it's hard to see what  
4 readers can learn from the results and what can be generalized for other sites, models  
5 and investigated factors.

6  
7 **Authors:** The straight forward single crop coefficient concept has been recently applied  
8 in various studies (page 7528 lines 27-28, page 7529 lines 1-14). The focus is drawn on  
9 crop coefficient parameters and ET methods, as both are crucial features in assessing  
10 irrigation water requirements as is already described on page 7528, 15-26.

11  
12 **Authors:** Maybe we were not clear enough in describing the scope of the study, and we  
13 will certainly address this better in a revised version of the manuscript. We fully agree  
14 with the reviewer that there are crude assumptions in some of the ET methods we  
15 applied. However, these ET methods are used worldwide in many simulation studies,  
16 without any considerations to improve the methods (e.g. the 20% precipitation reduction  
17 by runoff in the CROPWAT model). That is why we implemented them as given. In  
18 almost all studies, researcher use only one ET method with a single, often spatially  
19 independent Kc set. As a result, some scientist ask to at least use better adapted, local  
20 Kc sets. However, we show in our work that for any large scale studies, the uncertainty  
21 introduced by Kc parameterization is small compared to the uncertainty introduced by  
22 the ET method. Of course, this is not the case for any local model application, where  
23 crop conditions and spatial heterogeneity needs to be considered. But this is not done in  
24 large scale model applications.

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26 **Authors:** Both factors, i.e. crop coefficient and evapotranspiration, have been reported  
27 to be important for the performance of models based on the single crop coefficient  
28 concept as reported by others (see 7540 5-19; 7542 1-15) and we address this part of  
29 uncertainty in this study.

30 The fixed fraction of runoff is adapted from the default setting of the Cropwat model.

31  
32 **Referee #1:** Page 7527, lines 9-11: "We find that structural model uncertainty is far more  
33 important than model parametric uncertainty to estimate irrigation water requirement."  
34 Please notice that only one parameter was tested. Therefore this conclusion is to  
35 general.

36  
37 **Authors:** Will be rewritten to "We find that structural model uncertainty among reference  
38 ET is far more important than model parametric uncertainty introduced by crop  
39 coefficients. These crop coefficients are used to estimate irrigation water requirement  
40 following the single crop coefficient approach."

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3 **Referee #1:** Page 7527, lines 16-18: “We conclude that multi-model ensemble  
4 predictions and sophisticated model averaging techniques are helpful in predicting  
5 irrigation demand and provide relevant information for decision making.” To support this  
6 conclusion it is required to show the additional value of the multi-model ensemble  
7 predictions, as compared for example to a single application of the Penman-Monteith  
8 method. I can still not see it here.

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10 **Authors:** We disagree in this point. As explained in our rebuttal to comment 2). Using  
11 REA, we show that we are able to reduce the predictive uncertainty by considering a  
12 number of “uncertain” single models.

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15 **Referee #1:** Page 7527, lines 21-25: “Globally, the proportion of fresh water  
16 consumption by agriculture is large (9087 km<sup>3</sup> yr<sup>-1</sup>) (Hoekstra and Mekonnen, 2012)  
17 and is projected to increase in the future in order to support the increasing world  
18 population. More precisely, most of the change in freshwater consumption will arise from  
19 the increasing irrigation demand by crops (De Fraiture and Wichelns, 2010).” It’s  
20 required to be more precise. The first figure on fresh water use refers to the sum of  
21 irrigation water and natural rainfall while the second statement refers to irrigation only.  
22 That future irrigation water requirements will increase is not sure. Models accounting for  
23 the reduction in transpiration due to increased atmospheric CO<sub>2</sub> concentration show  
24 constant or even declining trends. Therefore this section does not reflect the state of  
25 knowledge.

26  
27 **Authors:** A likely effect of changes of atmospheric CO<sub>2</sub> concentration is not part of this  
28 study. But even by considering the biophysiological effect of reduced transpiration, the  
29 need for additional irrigation water is very likely in the future. This is mainly driven by  
30 demographic development and changes in food diets. Accordingly, we will rewrite the  
31 passage as follows: “*Globally, the proportion of fresh water consumption by agriculture  
32 from rainfall as well as surface and groundwater resources is large (9087 km<sup>3</sup> yr<sup>-1</sup>)  
33 (Hoekstra and Mekonnen, 2012). It is projected that water demand is increasing in the  
34 future, particular by irrigation agriculture, in order to support the increasing world  
35 population with food (Foley et al., 2011; De Fraiture and Wichelns, 2010; Hanjra and  
36 Qureshi, 2010; Wada and Bierkens, 2014).*”

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39 **Referee #1:** Pages 7527-7531 (introduction): The authors describe here what they have  
40 done in the paper but the objectives remain unclear. Is the objective to quantify

1 uncertainties in irrigation water estimates in models of the same type or is it to develop  
2 and present a new method for uncertainty assessment? How does this study compare to  
3 all these crop model comparisons published within the last 2-3 years? Wouldn't it be  
4 better to replace the crop coefficient approach by a real simulation of crop growth  
5 instead of just applying different sets of kc-values with unknown representativeness?  
6

7 **Authors:** We might have not been clear enough with the objectives of our study, which  
8 we certainly will improve in a revised version of our manuscript. The study is more than  
9 a simple model intercomparison, for which a number of studies have been published in  
10 the past years and which are cited in our manuscript. We go beyond a simple  
11 intercomparison: how can we derive better predictions by using an ensemble of well-  
12 known ET methods and which are the likely causes of predictive uncertainty in ET  
13 estimations. We are convinced, that ensemble modeling could overcome some of the  
14 shortcomings of today's global estimations of water resources, given the large  
15 uncertainties in ET estimation.  
16

17 **Authors:** Regarding the concern raised in relation to the Kc approach we argue, that the  
18 Kc approach is widely applied across many model applications in regions worldwide, in  
19 particular for predicting, e.g., irrigation requirements, global water resources,  
20 groundwater depletion, water footprint, virtual water trade. The advantage of this  
21 approach is that it can be used in regions where less data are available where the  
22 application of a comprehensive crop model is not possible. This approach has also been  
23 applied for a number of studies in the Murray-Darling Basin (Barton and Meyer, 2005;  
24 Harris, 2002; Hughes, 1999; Meyer, 1999). Even though we know that the Kc approach  
25 has limitations and that real simulations of crop growth would improve predictions, it  
26 remains unlikely that this will be happening on the macroscale.  
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29 **Referee #1:** Page 7531, lines 15-17: "The applicability of six different ETo methods is  
30 evaluated by using available measured class-A-pan evaporation measurements of 34  
31 stations in the MDB over a 21 years time period" ET is evaluated with E => does not  
32 seem to be very useful  
33

34 **Authors:** In order to make class-A pan measurements comparable with reference ETo  
35 one has to use pan coefficients (Allen et al., 1998). We converted class-A pan  
36 evaporation with pan-coefficients published by McMahon et al. (2013) which are given in  
37 a monthly resolution at 68 sites across Australia. Please see page 12 lines 13-16 in the  
38 manuscript for further details: "*Pan evaporation differs from evaporation from a cropped  
39 surface through a different albedo, heat storage and humidity above the surface. For this  
40 reason, the class-A pan data have been adjusted with monthly pan coefficients  
41 (McMahon et al., 2013) to better compare them with ETo simulations of open surface*

1 *waters. On an annual average, class-A pan evaporation of 1,558 mm yr<sup>-1</sup> were reduced*  
2 *by 9% to 1,422 mm yr<sup>-1</sup> across all stations.”*

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5 **Referee #1:** Page 7532, section 2.1 Study site and data: What about uncertainty in input  
6 data (e.g. land use, weather) and their interaction with model structure? Uncertainties in  
7 humidity and wind speed will likely affect PM but not some other methods like  
8 Hargreaves or Priestley-Taylor

9  
10 **Authors:** We are glad that the reviewer agrees with us that an accounting of the  
11 uncertainty behind ET estimation is complex and includes many sources. A full  
12 accounting of the global uncertainty in a spatial context of ET estimation would be for  
13 sure interesting, but not achievable at the moment; though on the long term it is highly  
14 needed. To our knowledge, our work is one of the few studies that takes a closer look at  
15 a part of this uncertainty in the field of macroscale irrigation requirement studies. We  
16 focus on two important sources of uncertainty, which have been reported to be relevant  
17 for predicting irrigation requirements (Howell et al., 2004; Siebert and Döll, 2010; da  
18 Silva et al., 2013). The other sources of uncertainty, i.e. land use, weather and many  
19 others, are also important but not in the particular scope of this study.

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21  
22 **Referee #1:** Page 7533, equation (2): Which data or findings support the very basic  
23 assumption that 80% of total precipitation becomes effective?

24  
25 **Authors:** The fixed fraction of runoff is adapted from the default setting of the  
26 CROPWAT model according to the FAO56 guidelines. It was not our intention to  
27 improve any of the ET methods, but rather apply them as given.

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30 **Referee #1:** Page 7536, lines 17-18: “The median daily ETo for APET is 3.6 mm d-1,  
31 PM56 3.9 mm d-1, HS 3.8 mm d-1, PPET 5.2 mm d-1, PT 6.4 mm d-1 and TURC 3.4  
32 mm d-1.” Please check the calculation routine and the underlying data for the  
33 calculations with Priestley-Taylor. An overestimate in the here reported range is very  
34 unlikely and not supported by the previous literature!

35  
36 **Authors:** We will again check the amount of the ET predicted by the PT method and  
37 include this in a revised version of the paper.

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### Literature

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1 Wada, Y. and Bierkens, M. F. P.: Sustainability of global water use: past reconstruction and future  
2 projections, *Environ. Res. Lett.*, 9(10), 104003, doi:10.1088/1748-9326/9/10/104003, 2014

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1 **Reply to “Geosci. Model Dev. Discuss., 7, C2866–C2868, 2015”**

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3 **S. Multsch, J.-F. Exbrayat, M. Kirby, N. R. Viney, H.-G. Frede and L. Breuer**

4 **Referee #2:** The authors analyses the uncertainty in estimating irrigation water  
5 requirement by applying six models for ETpot and 5 Kc values (in total 30 simulations).  
6 They found that the uncertainty caused by different model approaches is much larger  
7 that uncertainty caused by Kc values. Furthermore, they state, that multi model  
8 ensemble prediction provide reliable estimates which can be used for management.

9  
10 **Referee #2:** In principle study this is an interesting, well conducted study. Nevertheless,  
11 I do have some concerns with respect to the general approach. Six different ETpot  
12 models were applied and tested against class A pan data although it is well known that  
13 class A pan data may not be the best method to measure ETpot and not for all stations  
14 pan-coefficients were available. Therefore, uncertain class A pan data were used in an  
15 uncertainty study assuming that class A pan data are certain.

16  
17 **Authors:** We are aware of that the utilization of class A pan data in our manuscript  
18 comes along with uncertainties and we did not assume that the data are certain. But  
19 isn't this the case for every kind of measurement? The alternative is not to calibrate and  
20 verify models and apply them in an unobserved fashion (which is most often done when  
21 evapotranspiration is being simulated in hydrology). Class A pan data at least provide  
22 insight into patterns and evaporation behavior.

23 Another reason for using Class A pan data is that there are no other measurements at  
24 hand, which we could use instead. This is a general problem in simulations of  
25 evapotranspiration. Despite that this water balance component significantly contributes  
26 to the total balance, researchers often simulate it without any data for calibration or  
27 validation at all. Knowing that there is not one perfect model, reliability ensemble  
28 averaging (REA) utilizes the information provided by several models of the cohort.

29 Finally, the idea of using the REA method is, that one component of REA, i.e.  $R_B$ ,  
30 weights the different models concerning how good they match observed values. Hence,  
31 an estimate of the natural background variability (i.e. of the bias) of the target variable  
32 (as stated on 7535, lines 4-6) is needed. This is the reason why we have used Class A  
33 pan data in the course of our REA experiment.

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35 **Referee #2:** Furthermore, all other uncertainties related to climate (radiation,  
36 temperature, rainfall, ...) and uncertainty related to regionalization of the punctual  
37 information are ignored.

38

1 **Authors:** We agree that the forcing data itself introduce additional uncertainties.  
2 However, this is not part of this study and it would clearly go beyond the scope of our  
3 work presented here. Nevertheless, on the long term we think that more research needs  
4 to be put in the investigation of the global predictive uncertainty of models, where all  
5 sources of uncertainty are evaluated, i.e. spatial input data uncertainty (e.g. soil and  
6 land use information), model forcing data uncertainty (e.g. climate data), parameter  
7 uncertainty, and model structure uncertainty. This would allow to distinguish between  
8 the different sources and identify those components that contribute most to the  
9 predictive uncertainty of modeling (e.g. Exbrayat et al. 2014).

10 We will extent the discussion and discuss other sources of uncertainty as well, i.e.  
11 regionalization, class-A pan data and forcing data.

12

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14 **Referee #2:** The six ETpot methods differ in data demand and representation of the  
15 underlying processes. Some of them use empirical parameters (like PT). These  
16 parameters were taken as certain although they are also uncertain. One could have  
17 calibrated the empirical parameters of the ETpot equations using the class A pan and  
18 studying the effect on IRR. An interesting question would also how the selection of the  
19 ETpot method (there are much more in literature, see Bormann) do effect the findings.

20

21 **Authors:** Yes, we could have used more ET functions, but we have restricted our  
22 analysis to the most commonly applied methods in the region as described in the  
23 introduction (page 7530 lines 13-25). The consideration of other ET functions would  
24 have extended the picture drawn in this manuscript, but the overall message would have  
25 remained the same.

26 Further, we did not want to calibrate each method. This is almost never done when ET is  
27 simulated on large scale. Moreover, the idea of the REA approach is not to identify one  
28 best model and improve it, but to use the information of several models in a statistical  
29 way. Here we show that this concept is straightforward to use and helps to improve  
30 predictions of water requirement on the large scale.

31

32 **Referee #2:** It seems that the authors assume that nothing is known concerning the  
33 applicability of different ETpot methods to specific regions like the MDB. For me the  
34 argument is not convincing that many models do use these approaches because in this  
35 case one has to train the user to apply only models applicable to specific questions and  
36 regions.

37

38 **Authors:** We argue that in many studies, in particular in macroscale or global studies,  
39 the choice of the ET function is not validated in a regional context, in particular if the  
40 crop coefficient concept is applied as described in the introduction (page 7529, lines 15-

1 25, page 7530, lines 1-12). We recommend to compare different methods in such a  
2 case and suggest to apply a method to reduce the uncertainty, e.g. by reliability  
3 ensemble averaging.

4  
5 **Referee #2:** The data in Tab. 1 already show that the uncertainty related to ETpot is  
6 much larger than the uncertainty related to Kc. Kc, mid for example varies between 1 and  
7 1.15 which is max. 15% compared to the range of 2.4 to 6.4 mm/d in ETpot data (nearly  
8 100%). If this is the story, one could have stopped here.

9  
10 **Authors:** Yes, but in addition to calculate the uncertainty of a single model or a model  
11 parameter, we show an option to reduce the uncertainty by using a method (REA) which  
12 is commonly applied in climate sciences. Again, our objective is not to find one best  
13 model, but rather use the information content of several models. In that sense, we show  
14 that the REA concept is a helpful method in geospatial model applications.

15  
16 **Referee #2:** If the main message is that ensemble averaging improves the prediction of  
17 IRR than I wonder if all ETpot models should be considered although it is clear that  
18 some of them are not reliable. If the argument is that it is not clear for other regions  
19 which ETpot model is reliable (I would not agree with such a statement) then one has to  
20 consider much more approaches as used by Bormann.

21  
22 **Authors:** We agree that Bormann (2011) gives a more complete picture of the structural  
23 differences between ET methods. Instead of using all methods presented by Bormann  
24 (2011), we have considered the most commonly applied methods in Australia. We think  
25 that the consideration of more functions would not have changed the outcome of our  
26 work.

27  
28 **Referee #2:** I recommend repeating the uncertainty analysis but leaving out the two  
29 ETpot methods evaluated as poor. Furthermore, I recommend to “calibrate” the  
30 empirical parameters of the ETpot data using class A pan data and discuss  
31 regionalization as well as other uncertainties.

32  
33 **Authors:** The elimination of one ET method for whatever reason is subjective. The  
34 reliability ensemble averaging method gives an objective criterion to weight the different  
35 models, so that models with a poor performance (weighting factor  $R_B$ ) or models which  
36 differ in a large extent from the simulated ensemble average (weighting factor  $R_D$ ),  
37 receive a lower weight. By doing so, REA automatically punishes poorer performing  
38 models – there is no need to act as suggested by the referee due to the method we  
39 apply.

1 **Referee #2:** The paper is well written. I only wonder why the authors discuss CO2  
2 dependency (pages 7542-7543) because this is a very specific aspect not covered by  
3 the paper. I would delete this part

4

5 **Authors:** Will be deleted.

6

7

8 **Literature**

9 Bormann, H.: Sensitivity analysis of 18 different potential evapotranspiration models to observed climatic  
10 change at German climate stations, Climatic change, 104(3), 729–753, 2011.

11

12

# Reduction of predictive uncertainty in estimating irrigation water requirement through multi-model ensembles and ensemble averaging

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## Abstract

Irrigation agriculture plays an increasingly important role in food supply. Many evapotranspiration models are used today to estimate the water demand for irrigation. They consider different stages of crop growth by empirical crop coefficients to adapt evapotranspiration throughout the vegetation period. We investigate the importance of the model structural versus model parametric uncertainty for irrigation simulations by considering six evapotranspiration models and five crop coefficient sets to estimate irrigation water requirements

1 for growing wheat in the Murray-Darling Basin, Australia. The study is carried out using the  
2 spatial decision support system SPARE:WATER. We find that structural model uncertainty  
3 among reference ET is far more important than model parametric uncertainty introduced by crop  
4 coefficients. These crop coefficients are used to estimate irrigation water requirement following  
5 the single crop coefficient approach. ~~We find that structural model uncertainty is far more~~  
6 ~~important than model parametric uncertainty to estimate irrigation water requirement.~~ Using the  
7 Reliability Ensemble Averaging (REA) technique, we are able to reduce the overall predictive  
8 model uncertainty by more than 10%. The exceedance probability curve of irrigation water  
9 requirements shows that a certain threshold, e.g. an irrigation water limit due to water right of  
10 400mm, would be less frequently exceeded in case of the REA ensemble average (45%) in  
11 comparison to the equally weighted ensemble average (66%). We conclude that multi-model  
12 ensemble predictions and sophisticated model averaging techniques are helpful in predicting  
13 irrigation demand and provide relevant information for decision making.

14

15

## 16 1 Introduction

### 17 1.1 Predicting crop water needs

18 Globally, the proportion of fresh water consumption by agriculture from rainfall as well as  
19 surface and groundwater resources is large ( $9,087 \text{ km}^3 \text{ yr}^{-1}$ ) (Hoekstra and Mekonnen, 2012). It is  
20 projected that water demand is increasing in the future, in particular by irrigation agriculture, in  
21 order to support the increasing world population with food (Foley et al., 2011; De Fraiture and  
22 Wichelns, 2010; Hanjra and Qureshi, 2010; Wada and Bierkens, 2014). ~~Globally, the proportion~~  
23 ~~of fresh water consumption by agriculture is large ( $9,087 \text{ km}^3 \text{ yr}^{-1}$ ) (Hoekstra and Mekonnen,~~  
24 ~~2012) and is projected to increase in the future in order to support the increasing world~~  
25 ~~population. More precisely, most of the change in freshwater consumption will arise from the~~  
26 ~~increasing irrigation demand by crops (De Fraiture and Wichelns, 2010).~~ Therefore, strategies  
27 based on improved irrigation methods and local adaptations of management practices are likely to  
28 be implemented to anticipate this trend. Such strategies are often developed using decision

1 support systems that are informed by mathematical models. For example, irrigation management  
2 has been optimized by modelling and measurements for crops grown in Central Asia (Pereira et  
3 al., 2009) or for irrigated cotton in the High-Plains region of Texas (Howell et al., 2004). Others  
4 have investigated water use efficiency (Wang et al., 2001) or crop water productivity (Liu et al.,  
5 2007) by modelling experiments for irrigated crops grown in China.

6 All these models depend on the calculation of evapotranspiration (ET) which represents the  
7 evaporation from a surface and transpiration from plants. In the case of agricultural crops, ET is  
8 equal to the crop water needed for crop growth and yield production. Globally,  
9 evapotranspiration represents about two thirds of the total rainfall on land, while  
10 evapotranspiration from crops amounts for about 8% (Oki and Kanae, 2006), and is insofar the  
11 most important term of the water balance. The basic concept for deriving crop water needs of  
12 irrigated crops has been initially reported by Jensen (1968) and is proposed by Allen et al. (1998)  
13 as the single crop coefficient concept. The crop specific evapotranspiration ( $ET_c$ ) is derived from  
14 reference evapotranspiration ( $ET_o$ ) and a crop specific coefficient ( $K_c$ ):

$$15 \quad ET_c = ET_o \cdot K_c \quad (1)$$

16 with  $ET_o$  given in [mm] and dimensionless  $K_c$ .  $ET_o$  can be calculated by standardise potential  
17 evapotranspiration (PET) to a short (grass) or tall (alfalfa) reference crop. In the case of the  
18 Penman-Monteith equation (Monteith, 1965; Penman, 1948) standardized fixed values for albedo  
19 (0.23), plant height (0.12 cm) and surface resistance ( $70 \text{ m s}^{-1}$ ) are assumed (Allen et al., 1998;  
20 Jensen et al., 1990).  $K_c$  is commonly calculated on the basis of field experiments (e.g. Ko et al.,  
21 2009; da Silva et al., 2013) and varies with the crop development.

22 Such an approach is part of many irrigation management models, including Cropwat (Smith,  
23 1992), ISAREG (Pereira et al., 2009), ISM (George et al., 2000) or global crop water models  
24 (Siebert and Döll, 2010). Moreover, the single crop coefficient concept is the basis for the  
25 simulation of crop water needs in many studies. For example, Lathuillière et al. (2012) have  
26 derived water use by terrestrial ecosystems and have shown that ET declines over a 10 year  
27 period by about 25% in response to deforestation and replacement by agriculture in Brazil. They  
28 showed that irrigation water requirement (IRR) is relevant for terrestrial water fluxes and a  
29 reliable estimation is crucial for the closure of the water cycle. In another study future climate



1 impacts on groundwater in agriculture areas have been investigated (Toews and Allen, 2009).  
2 They showed that larger return flows to the groundwater can be related to increased IRR under  
3 warmer temperatures and longer vegetation periods. Moreover, the crop coefficient concept is  
4 also the basis for the water footprint (volume of water consumed or polluted to produce one unit  
5 of biomass) assessment of crops (Mekonnen and Hoekstra, 2011) and has been used to determine  
6 water requirements and the water footprint of the agriculture sector in Saudi Arabia (Multsch et  
7 al., 2013). In almost all studies, researcher use only one  $ET_o$  method with a single, often spatially  
8 independent  $K_c$  set. As a result, some scientist ask to ~~at least use locally better adapted, local~~  $K_c$   
9 sets at least (Ko et al., 2009; da Silva et al., 2013). For this reason, the investigation of predictive  
10 uncertainty of IRR is needed, in particular in the frame of large scale assessments.

11

## 12 1.2 Sources of predictive uncertainty

13 Major sources of uncertainties should be considered in the study design, quantified throughout  
14 the modelling process (Refsgaard et al., 2007) and communicated as part of the results to the end  
15 users. Uncertainties related to large scale estimations of the IRR have only rarely been analysed.  
16 For example, Siebert and Döll (2010) have studied the uncertainty in predicting green (rainfall  
17 consumed by crops) and blue (consumed surface and groundwater by crops in terms of irrigation)  
18 water consumption by using different  $ET_o$  equations on a global scale. They observed a  
19 significant difference of blue water consumption, i.e. required irrigation, and only a small change  
20 in green water consumption between model runs while using two classical  $ET_o$  equations. More  
21 recently, Sheffield et al. (2012) pointed out that using a more up-to-date parameterization of PET  
22 to calculate drought indices led to different conclusions on drought occurrence globally.

23 Generally, model predictive uncertainty can be lead back to four sources, input uncertainty,  
24 output uncertainty, structural uncertainty and parametric uncertainty (Renard et al., 2010). The  
25 last two, structural and parametric uncertainty, are addressed in this study with a focus on the  
26 prediction of IRR. As part of the parametric uncertainty, the parameterization of equations to  
27 quantify natural or anthropogenic processes has received considerable interest, particularly in  
28 conceptual rainfall-runoff modelling (Beven, 2006; Vrugt et al., 2009). In case of modelling crop  
29 water needs according to Eq. 1,  $K_c$  is an important model parameter.  $K_c$  values for a large number

1 of crops are provided by the FAO56 irrigation guidelines (Allen et al., 1998) which are  
2 commonly used for irrigation planning. However, it has been highlighted that an adjustment to  
3 the global  $K_c$  is needed if the simulations are used for irrigation planning on a local to regional  
4 scale (Ko et al., 2009; da Silva et al., 2013). Nevertheless, it is still unclear whether a local  
5 adaption of  $K_c$  leads to a better model performance. For this reason, we quantify the parametric  
6 uncertainty of model parameterisation with different  $K_c$  sets.

7 The model structure also introduces uncertainties, as any model remains a simplification of the  
8 real world. In the context of modelling water resources, all hydrological and crop growth models  
9 rely on the estimation of ET. According to equation 1,  $ET_o$  is required to estimate crop specific  
10 evapotranspiration.  $ET_o$  equations are often divided into categories according to the input data  
11 (Bormann, 2011; Tabari et al., 2013): temperature based equations such as Hargreaves-Samani  
12 (HS) equation (Hargreaves and Samani, 1985), radiation based equations such as Priestley-Taylor  
13 (PT) (Priestley and Taylor, 1972) or combined equations such as the FAO56 Penman-Monteith  
14 (PM56) equation (Allen et al., 1998), that further takes wind speed into account. Nevertheless, in  
15 many cases it was shown that the variability among PET methods is large (Fisher et al., 2011;  
16 Kite and Droogers, 2000). Because most water resources models rely on some calculation of  $ET_o$ ,  
17 we see it as a crucial source of structural uncertainty that is rarely considered.

### 18 **1.3 Reduction of predictive uncertainty by ensemble modelling**

19 Ensembles of model predictions can be developed by different sets of model parameterization  
20 (single-model ensemble) and model structures (multi-model ensemble). The weighting of model  
21 ensembles according to their fit to observational data has become of interest to reduce the  
22 uncertainty and to derive a more robust predictions and projections. Giorgi and Mearns (2002)  
23 have introduced the reliability ensemble averaging technique (REA) in climate research.  
24 Basically, different models are weighted according to their performance in representing measured  
25 data and according to the distance of individual models to the ensemble average prediction to  
26 quantify the convergence of different models. This approach has been applied more recently for  
27 predicting catchment nitrogen fluxes (Exbrayat et al., 2013) and calculating water balances and  
28 land use interaction (Huisman et al., 2009).

1 In a first step, we analyse the relative contributions of the structural and parametric model  
2 uncertainty in hind casts of IRR of wheat across the Murray-Darling-Basin (MDB), Australia.  
3 Simulations are calculated using the spatial decision support system SPARE:WATER (Multsch  
4 et al., 2013). In a second step, we apply the REA methodology to reduce the predictive  
5 uncertainty of IRR. The general procedure is as follows:

- 6 • The applicability of six different  $ET_0$  methods is evaluated by using available measured  
7 class-A-pan evaporation measurements of 34 stations in the MDB over a 21years time  
8 period;
- 9 • 30 different model realisations are setup in a multi-model ensemble by combining various  
10  $ET_0$  equations (n=6) and crop coefficient data sets (n=5);
- 11 • IRR is calculated by forcing the multi-model ensemble with climate time series of 21  
12 years (monthly data) for 3,969 sites (each 1 km<sup>2</sup> x 1km<sup>2</sup>) in the MDB where irrigated  
13 wheat has been grown according to the land use allocation in 2000;
- 14 • The 30 model realisations are weighted according to their performance in representing  
15 measured data and their distance to the ensemble average.

16 By doing so, we quantify structural ( $ET_0$  method) and parametric ( $K_c$  set) uncertainty and apply  
17 REA to provide a robust estimate of IRR and the confidence interval around it. **The underlying**  
18 **research question is how can we derive better predictions by using an ensemble of well-known**  
19  **$ET_0$  methods as well as  $K_c$  sets and which are the likely causes of predictive uncertainty in IRR**  
20 **estimations. Finally, we show a procedure to reduce predictive uncertainty of IRR.**

21

## 22 **2 Methods and data**

### 23 **2.1 Study site and data**

24 The MDB covers about 1 million km<sup>2</sup> of south-east Australia (Fig. 1). Irrigation agriculture in the  
25 MDB sums up to 17,600 km<sup>2</sup>, which is equal to 65% of the total irrigation agriculture in  
26 Australia. Total water withdrawal for irrigation in 2006 amounted to 7.36 km<sup>3</sup> yr<sup>-1</sup> (ABS, 2006).  
27 Wheat is the second most important crop grown in MDB after grazing pastures, covering 3,969

1 km<sup>2</sup> in 2006 and was therefore selected for this case study for which IRR and its underlying  
2 uncertainty was calculated. The cropping areas have been taken from a land use map from 2006  
3 (ABARES, 2010) with a spatial resolution of 0.01° x 0.01° (~1 km x 1 km). We assume a fixed  
4 land use distribution over time in our model study to clearly target the uncertainty in ET<sub>o</sub> method  
5 and crop coefficients. Climate data for 1986-2006 were taken from the SILO Data Drill of the  
6 Queensland Department of Natural resources and Water (<https://longpaddock.qld.gov.au/silo/>  
7 (Jeffrey et al., 2001)) with a spatial resolution of 0.05° x 0.05° (~5 km x 5 km). We used the  
8 same weather dataset over all 3,969 1 x 1 km land grid cells overlapped by a 5 x 5 km grid cell in  
9 the weather data. The model was forced with monthly data. For validation, we compared  
10 simulated ET<sub>o</sub> to measured class-A pan data from 34 stations throughout the MDB. The class-A  
11 pan data were obtained from Patched Point Dataset of the Queensland Department of Science,  
12 Information Technology, Innovation and the Arts,  
13 (<http://www.longpaddock.qld.gov.au/silo/ppd/>). Measured data have been adjusted with monthly  
14 pan-coefficients according to McMahon et al. (2013) to represent evaporation from open surface  
15 water. For stations where no pan-coefficient was available we used the one from the nearest  
16 station.

17

## 18 **2.2 Simulation of irrigation requirement with SPARE:WATER**

19 SPARE:WATER (Multsch et al., 2013) is a spatial decision support system for the calculation of  
20 crop specific water requirements and water footprints from local to regional scale. Input  
21 parameter for the simulation are climate data, irrigation management (irrigation water quality,  
22 irrigation efficiency, irrigation method), a digital elevation model and crop characteristics such as  
23 maximum crop height and length of growing season as well as sowing and planting date. In a first  
24 step, the water requirement of growing a crop is simulated for each grid cell according to the  
25 spatial resolution of the input data. In a second step, the water footprint for spatial entities such as  
26 administrative boundaries or catchments is calculated considering statistical data on crop yield  
27 and harvest area. Water footprints for geographic entities are given as volume of water consumed  
28 per year (e.g. km<sup>3</sup> yr<sup>-1</sup>) and water footprints for specific crops as volumes of water consumed per  
29 biomass (m<sup>3</sup> t<sup>-1</sup>).

1 In this study the calculation of the IRR is calculated as the difference between  $ET_c$  and effective  
2 rainfall ( $P_{eff}$ ). The latter one is estimated from the difference of surface run-off (RO) and  
3 precipitation (P). RO is derived as a fixed fraction of 20% of total P. **The fixed fraction of runoff**  
4 **is adapted from the default setting of the FAO CROPWAT model (Smith, 1992).** On this basis,  
5 IRR is calculated according to Eq. 2:

$$6 \quad IRR = \max(ET_c - P_{eff}, 0) \quad (2)$$

7 with IRR,  $ET_c$  and  $P_{eff}$  given in [mm].  $ET_c$  is calculated based on the single crop coefficient  
8 approach initially proposed by Jensen (1968) and recommended by Allen et al. (1998) according  
9 to Eq. 1. The input parameters for this method are the length of four individual stages (initial  
10 season, growth season, mid-season and late season) during the growing season and three related  
11 crop coefficients ( $K_c$ ). These define the ratio between  $ET_o$  and  $ET_c$  for each part of the growing  
12 season. We have considered five different  $K_c$  data sets (Table 1). The most common dataset has  
13 been proposed from the FAO56 Irrigation and Drainage Guidelines (Allen et al 1998). This  
14 approach has been applied for calculating crop water footprints (Mekonnen and Hoekstra, 2011)  
15 and is part of the widely used Cropwat model (Smith 1992). It has been discussed that locally  
16 adapted  $K_c$  sets are superior in simulating site-specific crop water requirement than global ones  
17 (Ko et al., 2009; da Silva et al., 2013). Thus, further data sets have been collected from various  
18 sources which represent site-specific relationships between  $ET_o$  and  $ET_c$  for areas in the MDB.

19  $ET_o$  has been calculated with six different methods (Table 2). Two of them are classified as  
20 combined methods (PM56, PPET), three are radiation-based methods (PT, TURC, APET) and  
21 one is a temperature based method (HS). All of them are commonly applied function, e.g. PM56  
22 and HS are included in Cropwat (Smith, 1992) and Aquacrop (Steduto et al., 2009), two models  
23 to quantify crop water and IRR, widely used and promoted by the FAO. The cropping system  
24 model EPIC (Williams, 1989) additionally allows the use of the PT equation, while the global  
25 vegetation model LPJmL (Fader et al., 2010) and the global water model WaterGap (Döll et al.,  
26 2003) are restricted to PT. APET and PPET have been particularly tested for the utilisation under  
27 Australian weather conditions in several (Chiew et al., 2002; Chiew and Leahy, 2003; Donohue  
28 et al., 2010).

29

### 1 **2.3 Reliability Ensemble averaging**

2 We used two types of ensemble averaging techniques, which differ in the weighing technique.  
3 We calculated an equally weighted average of all 30 model realisations (6 ET<sub>o</sub> methods x 5 K<sub>c</sub>  
4 datasets) for every grid cell which sum up to 3,969 cells (1 x 1 km) in the MDB where irrigated  
5 wheat is grown according to the land use allocation in 2006. However, this method does not  
6 consider the capability of its ensemble members to predict a target value nor does it value the  
7 agreement of model predictions amongst each other. Therefore, we apply the REA technique that  
8 was initially proposed by Giorgi and Mearns (2002) to reduce uncertainties in climate change  
9 projections (see appendix C for details). Moreover, it was used in impact studies targeting land  
10 use change impacts on hydrology (Huisman et al., 2009) and water quality scenario projections  
11 (Exbrayat et al., 2013).

12 The strength of the REA method is that it considers both the quality of a model prediction  
13 (performance) and its position within an ensemble of prediction (convergence). The aim is to  
14 provide a best estimate of predictions and a robust assessment of the confidence interval around  
15 it. The REA weighting scheme estimates two factors, model performance ( $R_B$ ) and model  
16 convergence ( $R_D$ ).  $R_B$  represents the capability of each ensemble member to represent real world  
17 data by its bias  $B$ .  $R_D$  is a measure of the distance  $D$  of a single model to the equally weighted  
18 ensemble average. Both are limited by the natural background variability ( $\varepsilon$ ). The combined  
19 effect known as reliability factor ( $R$ ) is derived as:

$$20 \quad R = \left[ \frac{\varepsilon}{abs(B)} \right]^{R_B} * \left[ \frac{\varepsilon}{abs(D)} \right]^{R_D} \quad (3)$$

21 In this study,  $\varepsilon$  is calculated from measured class-A pan evaporation for 34 climate stations in the  
22 study region for the time period from 1986 to 2006. The class-A pan data has been adjusted with  
23 monthly pan coefficients for climate stations in Australia (McMahon et al., 2013). We calculated  
24 the annual mean evaporation [mm] for each year and each station and used the 50% confidence  
25 interval (difference between the 25% and 75% percentile) of 224 mm to define  $\varepsilon$ . The  
26 consideration of the difference between upper and lower percentiles has been recommended by

1 Giorgi and Mearns (2002). Model performance is measured by the RMSE between measured  
2 (class-A pan) and predicted  $ET_o$  for each model ( $i$ ).

3 The convergence criterion  $R_D$  is calculated in an iterative procedure. The difference between the  
4 average IRR of each ensemble member  $i$  and the ensemble average is calculated. Under the  
5 consideration of the natural background variability  $\varepsilon$  a first guess of  $R_D$  (for each ensemble  
6 member) is predicted as well as a first guess of the REA average. This procedure is repeated by  
7 considering the newly derived REA average until the ensemble convergence, so that the  
8 difference between ensemble members and the REA average cannot be reduced by additional  
9 iterations (see Giorgi and Mearns (2002) for a complete methodological description). The error of  
10 the equally weighted ensemble average is described by the RMSE between  $IRR_i$  predicted by  
11 model  $i$  (with  $n=30$  models) and the equally weighted ensemble average irrigation water  
12 requirement ( $\overline{IRR}$ ). The error of the reliability ensemble average ( $RMSE_{REA}$ ) is derived from the  
13 reliability factor of each model ( $R_i$ ), the irrigation water requirement predicted by model  $i$  ( $IRR_i$ )  
14 and the REA weighted ensemble average ( $\overline{IRR}_{REA}$ ). The RMSE represents an approximate 60-  
15 70% confidence interval under the assumption that the amount of irrigation is distributed  
16 somewhere between normal and uniform.

17

## 18 **3 Results**

### 19 **3.1 Validation of $ET_o$ methods**

20 We applied six  $ET_o$  equations to 34 sites in the MDB for which measured class-A pan  
21 evaporation data were available from 1986 to 2006 (Fig. 2). Class-A pan data represent the  
22 evaporation from an open water surface and integrate all climate factors driving evaporation such  
23 as radiation, wind speed, humidity and temperature. Pan evaporation differs from evaporation  
24 from a cropped surface through a different albedo, heat storage and humidity above the surface.  
25 For this reason, the class-A pan data have been adjusted with monthly pan coefficients  
26 (McMahon et al., 2013) to better compare them with  $ET_o$  simulations of open surface waters. On  
27 an annual average, class-A pan evaporation of  $1,558 \text{ mm yr}^{-1}$  were reduced by 9% to  $1,422 \text{ mm}$   
28  $\text{yr}^{-1}$  across all stations.

1 The median daily  $ET_o$  for APET is  $3.6 \text{ mm d}^{-1}$ , PM56  $3.9 \text{ mm d}^{-1}$ , HS  $3.8 \text{ mm d}^{-1}$ ,  
2 PPET  $5.2 \text{ mm d}^{-1}$ , PT  $6.4 \text{ mm d}^{-1}$  and TURC  $3.4 \text{ mm d}^{-1}$ . According to the root-mean-squared-  
3 error (RMSE) PM56 gave the most reliable results. The median of  $ET_o$  for APET, PM56 and HS  
4 are close to the median of the measured evaporation rate of  $3.7 \text{ mm d}^{-1}$ . Apart from PT and PPET,  
5 the other methods underestimate  $ET_o$ , especially where class-A pan data are larger than  $6 \text{ mm d}^{-1}$ .  
6 The relationship between measured and simulated  $ET_o$  is linear as shown by the coefficients of  
7 determination  $r^2$  ranging from 77% (PT) to 88% (PPET).

8 The simulated  $ET_o$  is normally distributed if a single station and one year is tested (Shapiro test  
9 for normality:  $\alpha > 0.1$  for each year and station). The difference between the 34 stations is up  
10 to two times larger than the inter-annual difference in the 21 years period. Thus, spatial  
11 variability is larger than temporal variability in the MDB. The intra-annual variability shows a  
12 different picture. The median  $ET_o$  in the summer months is up to four times larger than the  $ET_o$   
13 during winter months for all  $ET_o$  methods, except PPET and PT with a six times larger  $ET_o$  in  
14 summer than in winter months.

15 Four of the six methods simulate the measured data with a high  $r^2$  and a low RMSE. The  
16 difference between the methods itself is large, in particular through the high  $ET_o$  estimates by PT  
17 and PPET. Thus, the structural uncertainty through the  $ET_o$  method is substantial and needs to be  
18 considered for the prediction of IRR which is addressed in the next chapters.

19

## 20 **3.2 Irrigation water requirement and its variability**

21 The IRR of wheat has been simulated using an ensemble of thirty model realisations for each of  
22 the 3,969  $1 \text{ km} \times 1 \text{ km}$  irrigated cells in the MDB for 21 years. Average values of IRR for all  
23 model realisations are shown in Table 3. In most cases, the largest estimates are given by the  
24 combinations of the Kc set Hughes with the  $ET_o$  method PT. These are almost 2.5 times higher  
25 than the lowest average IRR calculated by the combination of TURC with the Kc set Harris. It is  
26 obvious that changing  $ET_o$  method results in a larger variation of calculated IRR than using a  
27 different Kc set. Hence, the average IRR give a first idea about variability due to model structures  
28 and parameters.



1 Over a large watershed such as the MDB local differences in IRR may be large while catchment  
2 wide water management plans define thresholds for water withdrawal, for example due to water  
3 rights or water resources protection measures. A given threshold may require heterogeneous local  
4 adaptations of irrigation management and a change in cropping patterns. Figure 3 shows the  
5 probability that a certain amount of IRR is exceeded in the MDB on average over the 21 year  
6 period. It illustrates the range of IRR predicted by the ensemble of all 30 model realisations for  
7 each grid cell. Two groups can be identified that are separated by  $ET_0$  methods. The first group is  
8 composed of PPET and PT calculations. In this case, IRR is up to twice as high as compared to  
9 predictions by other models. The second group is formed by APET, HS, PM56 and TURC with  
10 substantially lower calculations of less than 500 mm in most cases. We note that the parametric  
11 uncertainty is almost negligible compared to the uncertainty introduced by the various  $ET_0$   
12 methods.

### 13 **3.3 Ensemble averaging, uncertainty and weighting**

14 Ensemble predictions have become an important tool to account for different model structures  
15 and parameters (Exbrayat et al., 2013; Huisman et al., 2009; Wada et al., 2013). The  
16 consideration of ensembles is especially helpful to increase our confidence in simulations when  
17 no validation data are at hand, such as projections of Earth's future climate under specified  
18 emission scenarios. Here we apply the concept of ensemble prediction to simulations of IRR.  
19 Two different ensemble averages, expressed as the exceedance probability of the IRR of wheat  
20 are shown in Fig. 4. The first one represents the equally weighted average of irrigation ( $\overline{IRR}$ ,  
21 black line). The second one represents a weighted average using the reliability ensemble  
22 averaging ( $\overline{IRR}_{REA}$ , red line, see methods description) that weights predictions based on their  
23 performance and agreement with other ensemble members. This prevents dismissing some model  
24 structure, a process that can be rather subjective. Also, even an overall poorly performing model  
25 can contribute to the optimal information extracted from the ensembles (Viney et al., 2009), or  
26 may outperform better performing models once boundary conditions are changed (Exbrayat et al.,  
27 2013).

28 We use the inverse of the cumulative daily RMSE (Fig. 2) of the  $ET_0$  methods during the  
29 growing season to calculate the criterion  $R_B$  (RMSE 154 mm for APET, 123 mm for PM56,

1 142 mm HS, 232 mm PPET, 373 mm PT, 166 mm TURC). The convergence criterion  $R_D$  was  
2 calculated based on the difference of the predicted irrigation given by a single ensemble member  
3 and the equally weighted ensemble average (see Methods description). Overall, the PT model  
4 combinations have the lowest reliability factors of between 0.51 and 0.6 followed by PPET with  
5 0.96, a result driven by the poorer performance of these methods to simulate pan-evaporation  
6 (Fig. 2), and the outlying positions of simulations using PT and PPET (Fig. 3). All other models  
7 are weighted similarly, a result in accordance with the similar performance and simulated values  
8 exhibited by these methods (see Table 4 for details).

9 The application of the reliability factor leads to a decrease of the calculated total IRR in each grid  
10 cell as well as to a decrease of its overall uncertainty (Fig. 4). The uncertainty range is given by  
11 the ensemble average plus/minus the RMSE in each grid cell, assuming that modelling errors are  
12 normally distributed.

13 Exceedance probability curves might support defining thresholds in irrigation planning with  
14 consequences for decision makers through, for example, the adaptation of improved irrigation  
15 practice (e.g. from full to deficit irrigation, installation of advanced irrigation techniques) or the  
16 purchase of additional water rights. For example, a limit of available irrigation water of 400 mm  
17 per growing season will be exceeded less frequently in the MDB if the REA average IRR is  
18 considered (45%) in comparison to the equally weighted average (66%).

19 The spatial distribution of the equally weighted and the REA weighted ensemble averages are  
20 shown in Fig. 5a and b. The equally weighted average of IRR ranges between 124 and 691 mm  
21 with an average across the MDB of 424 mm (Fig. 5a). Thus, spatial variability is large and  
22 western and northern areas require five to six times more irrigation than in the south-east. The  
23 REA derived average IRR ranges between 104 mm and 663 mm across the river basin (Fig. 5b)  
24 with an average of 405 mm. Depending on the location this value is up to 18% lower as  
25 compared to simulations based on the equally weighted average (Fig. 5c). Also, the uncertainty  
26 range decreases as consequence of the REA method by about 10 % across the MDB with  
27 maximum values of around 26% when comparing equally and REA weighted RMSE (Fig. 5d-f).  
28 The largest change in uncertainty can be found in the south-east of the MDB and also in areas  
29 towards the east (Fig. 5f). Thus, REA not only leads to a decrease of predicted IRR but also to a

1 reduction of its uncertainty. The uncertainty is reduced because the REA is drawn toward the  
2 group of the better  $ET_o$  methods that also agree well between themselves.

3

#### 4 **4 Discussion and conclusions**

5 The simulation of IRR strongly varies amongst  $ET_o$  methods. Bormann (2011) recommended that  
6 the selection of the  $ET_o$  method should be based on the validation of  $ET_o$  with real world  
7 observations rather than only on the availability of climate input data. This is due to the general  
8 large variability among  $ET_o$  methods, which was also revealed in a study where PT was set as a  
9 benchmark model and the RMSE between  $ET_o$  methods was analysed (McMahon et al., 2013).  
10 Likewise, the influence of a single  $ET_o$  method on the prediction of crop yields was also reported  
11 for an agriculture site in Europe (Balkovič et al., 2013) where  $ET_o$  estimates by PT were 40%  
12 higher and those by Penman-Monteith 10% lower in comparison to HS. We also found a large  
13 variability among  $ET_o$  methods in our study. However, similar ranges across Australia for  $ET_o$   
14 have been reported by others (Chiew et al., 2002) for APET, PPET and PM56 as well as lower  
15 values for PT. Lascano *et al* (2010) as well as Lascano and Van Bavel (2007) have shown that  
16 methods to calculate ET based on combination methods, i.e., Penman-Monteith, tend to  
17 underestimate ET by as much as 25%, especially in dry climates.

18 Bormann (2011) further recommended that the reliability of  $ET_o$  equations should be tested in a  
19 spatial context, especially if applied on large scale. For various regions across Australia, a large  
20 range of mean annual  $ET_o$  between 1,700 mm (PT) and 3,670 mm (PPET) was reported  
21 (Donohue et al., 2010). To investigate the spatial heterogeneity within the MDB we analysed  
22 results of the 34 class-A pan stations. Overall, the performance of four of the  $ET_o$  methods was  
23 good with RMSEs around 1 mm day<sup>-1</sup>, except for three stations in the north. PPET performed less  
24 well with RSME increasing to 2 mm day<sup>-1</sup> while the PT value ranged up to 4 mm day<sup>-1</sup>. However,  
25 we found no consistent spatial pattern. **We are aware of that the utilization of class-A pan data  
26 comes along with uncertainties. We ~~and we~~ did not assume that the data are error-free certain, but  
27 for the application of REA, a comparison of model simulations and observations is needed to  
28 calculate the model performance criterion. We could have treated PM56 as being an  
29 “observation” in ~~termsthe sense~~ of a benchmark model. However, we think that a more**

1 independent test is more appropriate in the sense of REA and therefore decided to use those  
2 observations that are at hand: class-A pan observations. To account for the difference of class-A  
3 pan evaporation and reference crop ET, we used a commonly applied correction factor (pan-  
4 coefficients according to McMahon et al. (2013)) to derive crop ET from class-A pan  
5 measurements. Most often, ET estimates are not compared to any measurements at all, leaving  
6 modelers with no information on how good their model application is. We therefore think that a  
7 comparison to class-A pan is for sure not perfect, but better than not testing at all.

8  $ET_o$  estimates using the PM56 method revealed the best performance criteria in our study. PM56  
9 considers the most meteorological input parameters thereby possibly best representing the  
10 altering dry and wet conditions across the MDB over the year. The better performance of  
11 physically based equations in comparison to more empirical approaches for the simulation of  $ET_o$   
12 has also been reported by others (Donohue et al., 2010). PT performed least well in our study and  
13 resulted in up to two times larger estimates than other  $ET_o$  methods. This is somewhat contrasting  
14 with other studies (Chiew et al., 2002; Donohue et al., 2010) where PT gave lower  $ET_o$  values in  
15 comparison to methods such as APET and PPET.

16 One reason is that Donohue et al. (2010) have considered the actual albedo from remotely sensed  
17 vegetation cover (Donohue et al., 2008) for the estimation of the net incoming solar radiation. In  
18 our calculations, an albedo of a reference crop 0.23 (short crop, i.e. grass) has been considered  
19 according to the guidelines for  $ET_o$  from Allen et al. (1998). Another likely reason for this  
20 observation is that the PT equation is based on the Penman-Monteith equation in which the  
21 aerodynamic term is replaced by a constant (alpha) which is commonly set to 1.26 under  
22 Australian climatic conditions (Chiew and Leahy, 2003) and which we also applied. The  
23 consideration of region-specific alpha for the MDB could have increased the performance of PT  
24 in our study. The HS equation is commonly applied in situations where meteorological data are  
25 scarce, because the equation depends on more readily available temperature and extra-terrestrial  
26 radiation derived from latitude and day of the year. A reason for its good performance in our  
27 study could be that the semi-arid climate in most of the MDB is favourable for the HS equation,  
28 which is supported by Tabari (2010) who conclude that HS is a good candidate model for warm  
29 humid and semi-arid sites, but fails under cold humid climates. However, the poor response of

1 HS to changing climatic boundary conditions has also been criticized in a study on global  
2 drought simulations (Sheffield et al., 2012).

3 We combined the six  $ET_o$  methods with five  $K_c$  sets to address stochastic parametric uncertainty  
4 for irrigated wheat in the MDB. We show that the  $ET_o$  method uncertainty range exceeded the  
5 uncertainty range of  $K_c$  sets. Thus, the  $K_c$  sets have a minor influence on predicted IRR. At first  
6 sight, this seems to be contrasting to others who have stated that adapted, regional  $K_c$  sets are  
7 required to estimate reliable IRR rates. For instance, da Silva et al. (2013) reported that  $K_c$  sets  
8 from FAO56 lead to errors in plot scale irrigation planning under tropical conditions. Similar  
9 observations were reported for semi-arid conditions in the Texas High Plains region (Ko et al.,  
10 2009), highlighting the importance of regionally based  $K_c$  sets. While regional adaptation of  $K_c$   
11 might be important at smaller scales, e.g. on the farm level, we conclude that large scale  
12 applications do not necessarily need to focus on this potential contribution of uncertainty. Rather,  
13 effort should be put into finding appropriate  $ET_o$  methods, or even better, utilize ensemble  
14 predictions to cover a more realistic range of predictions. Our study confirms this latter  
15 recommendation, as we could not identify a single best  $ET_o$  method for the MDB. Especially in  
16 cases where no data for a direct evaluation of model results are available the application of model  
17 ensembles gives insight to the predictive uncertainty, e.g., being helpful in the development of  
18 best management practices (Exbrayat et al., 2013), study of land use (Huisman et al., 2009) or  
19 climate change (Exbrayat et al., 2014).

20 Besides the uncertainty introduced by local to global  $K_c$  values the utilisation of the single crop  
21 coefficient concept itself comes along with errors, which are not addressed in this study. For  
22 example, Lascano (2000) shows how  $K_c$  varies as a function of time (50 days) and how it  
23 changes when using a daily, 3 and 8-day moving average. Moreover, the temporal resolution of  
24  $ET_o$  calculation, i.e., hourly vs. daily is an important component and errors associated with the  
25 method of irrigation (surface, drip, sprinkler) cannot be neglected, but are beyond the uncertainty  
26 calculation of this study. We acknowledge that we do not consider uncertainties in boundary  
27 conditions (e.g. ~~relevance of CO<sub>2</sub> concentration,~~ land-use management options, climatic  
28 variability) although these may be non-negligible. For example, ~~atmospheric CO<sub>2</sub> has been~~  
29 ~~reported as a driving factor of ET in North America, South America and Asia regions besides~~  
30 ~~climate forcing (Shi et al., 2013). Others-~~ Bocchiola et al. (2013) reported that changes in future

1 precipitation regimes will have the greatest impact on the calculated water footprint (reflecting  
2 high ET rates) of maize in Italy and that changes in CO<sub>2</sub> and warming were less important  
3 (~~Bocchiola et al., 2013~~). Conversely, water use was more driven by agricultural management than  
4 by regional climatic variation in a water footprint analysed for an irrigation district in China (Sun  
5 et al., 2013). Statistical correction of model forcing data (such as bias correction of precipitation)  
6 has also been reported to alter ET estimates as shown by Ye et al. (2012) for the Upper Yellow  
7 River in China with changes of up to 29% of ET. ~~Beyond that, the forcing data themselves~~  
8 ~~introduce additional uncertainties. However, this is not part of this study and it would clearly go~~  
9 ~~beyond the scope of our work presented here. Nevertheless, on the long term more research needs~~  
10 ~~to be put in the investigation of the global predictive uncertainty of models, where all sources of~~  
11 ~~uncertainty are evaluated, i.e. spatial input data uncertainty (e.g. soil and land use information),~~  
12 ~~model forcing data uncertainty (e.g. climate data), parameter uncertainty, and model structure~~  
13 ~~uncertainty.~~ Thus, an even more complete picture of global model uncertainty can only be shown  
14 by considering all sorts of predictive uncertainty, including model input data, validation data, and  
15 spatial input data in addition to the impact of model structural and parametric uncertainty as  
16 presented here.

17 However, we argue that future management practices or the impact of climate change cannot be  
18 reliably evaluated due to the large uncertainty that exists in the ET<sub>o</sub> method, the basis of water  
19 resources modelling. We partially cope with this problem by applying the REA technique to  
20 extract the most relevant information from our simulations. The advantage of REA in decision  
21 making has already been shown for other fields of research, such as the development of N  
22 reduction scenarios to improve surface water quality (Exbrayat et al., 2013) or estimation of the  
23 effect of land use change on water budgets and hydrological fluxes (Huisman et al., 2009).  
24 Despite the growing importance of IRR for today's agriculture (Siebert and Döll, 2010) and the  
25 effect on surface (Hoekstra et al., 2012) and groundwater (Wada et al., 2010) resources, few  
26 studies have dealt with the predictive uncertainty of this requirement (e.g. Wada et al. (2013))  
27 and how to reduce it.

28

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2 Table 1. The five crop parameter sets for  $K_c$ .

Name (Reference)	Spatial reference	$K_{c_{ini}}$	$K_{c_{mid}}$	$K_{c_{end}}$
FAO56 (Allen et al., 1998)	Global	0.7	1.15	0.25
Harris (Harris, 2002)	Queensland	0.3	1.15	0.25
Kirby (Kirby et al., 2012)	Murray-Darling Basin	0.4	1.15	0.4
Meyer (Meyer, 1999)	Griffith, MDB	0.4	1.05	0.5
Hughes (Hughes, 1999)	Murray and Murrumbidgee valleys	0.3	1.0	0.6

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2 Table 2. The six equations applied for the calculation of reference evapotranspiration.

Method	Abbreviation	Equation
FAO-56 Penman-Monteith (Allen et al., 1998)	PM56	$PET_{PM56} = \frac{0.408 \cdot \Delta \cdot (R_n - G) + \gamma \cdot \frac{900}{T_{mean} + 273} \cdot u_2 \cdot (e_s - e_a)}{\Delta + \gamma \cdot (1 + 0.34 \cdot u_2)}$
Priestley-Taylor (Priestley and Taylor, 1972)	PT	$PET_{PT} = \alpha \cdot \left[ \frac{\Delta}{\Delta + \gamma} \right] \cdot \frac{(R_n - G)}{\lambda}$
Hargreaves-Samani (Hargreaves and Samani, 1985)	HS	$PET_{HS} = 0.0023 \cdot (T_{mean} + 17.8) \cdot (T_{max} - T_{min})^{0.5} \cdot R_a \cdot 0.408$
Turc (Allen, 2003; Turc, 1961)	TURC	$PET_{TURC} = \alpha_T \cdot \frac{T_{mean}}{T_{mean} + 15} \cdot \frac{23.8856 \cdot R_s + 50}{\lambda}$
Areal – PET (Morton, 1983)	APET	$PET_{APET} = b_1 + b_2 \left( \frac{1 + \gamma \cdot p}{\Delta} \right)^{-1} \cdot R_{TP}$
Point – PET (Morton, 1983)	PPET	$PET_{PPET_{Energy-balance}} = R_n - \lambda_p \cdot f_T \cdot (T_p - T_{mean})$ $PET_{PPET_{vapor-transfer}} = f_T \cdot (e_s - e_a)$

3 With  $PET_{PM56}$ ,  $PET_{PT}$ ,  $PET_{HS}$ ,  $PET_{TURC}$ ,  $PET_{APET}$ ,  $PPET_{Energy-Balance}$  and  $PPET_{Vapor-Transfer}$  in [mm], extra-terrestrial  
4 radiation  $R_a$ , solar radiation  $R_s$ , net radiation  $R_n$ , soil heat flux density  $G$  and net radiation at equilibrium temperature  
5  $R_{TP}$  in [ $\text{MJ m}^{-2}$ ], equilibrium temperature  $T_p$ , mean  $T_{mean}$ , minimum  $T_{min}$  and maximum  $T_{max}$  air temperature in [ $^{\circ}\text{C}$ ],  
6 wind speed  $u_2$  at 2 m height [ $\text{m s}^{-1}$ ], atmospheric pressure  $p$ , saturated  $e_s$  and actual  $e_a$  vapour pressure in [kPa], slope  
7 of vapour pressure curve  $\Delta$  and the psychrometric constant  $\gamma$  in [ $\text{kPa } ^{\circ}\text{C}^{-1}$ ], latent heat of vaporization  $\lambda$  in [ $\text{MJ kg}^{-1}$ ],  
8 and the dimensionless empirical constants  $b_1$  and  $b_2$  [-], the heat transfer coefficient  $\lambda_p$  [-], the vapour transfer  
9 coefficient  $f_T$  [-] and the humidity based value  $\alpha_T$ .

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2 Table 3. Average equally weighted irrigation water requirement ( $\overline{IRR}$ ) [mm] during the growing  
3 season of wheat in all cells [n=3,969] of the MDB grouped by  $ET_o$  methods and  $K_c$  sets over the  
4 period 1986-2006 (APET: Areal potential evapotranspiration; PM56: FAO56 Penman Monteith;  
5 HS: Hargreaves-Samani; PPET: Point potential evapotranspiration; PT: Priestly-Taylor; TURC:  
6 Turc).

		Kc					$\overline{IRR}$
		Kirby	Hughes	Meyer	FAO56	Harris	
$ET_o$ method	HS	381	381	372	349	336	364
	PT	661	671	654	618	580	637
	PPET	577	577	565	534	514	551
	PM56	365	362	355	344	324	350
	APET	357	354	347	329	315	340
	TURC	315	316	308	289	279	301
	$\overline{IRR}$	443	443	433	410	391	424

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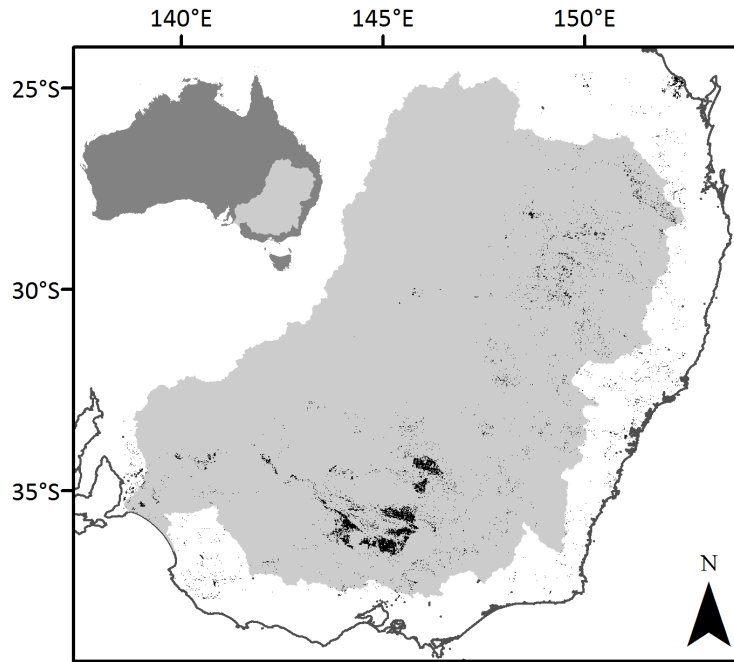
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2 Table 4. Performance ( $R_B$ ) and convergence ( $R_D$ ) and reliability ( $R$ ) coefficient of the ensemble  
3 members.

		FAO56	Harris	Hughes	Kirby	Meyer
APET	$R_B$	1	1	1	1	1
	$R_D$	1	1	1	1	1
	$R$	1	1	1	1	1
PPET	$R_B$	0.96	0.96	0.96	0.96	0.96
	$R_D$	0.99	1.00	0.99	0.99	0.99
	$R$	0.96	0.96	0.96	0.95	0.96
HS	$R_B$	1	1	1	1	1
	$R_D$	1	1	1	1	1
	$R$	1	1	1	1	1
PM56	$R_B$	1	1	1	1	1
	$R_D$	1	1	1	1	1
	$R$	1	1	1	1	1
T	$R_B$	1	1	1	1	1
	$R_D$	1	1	1	1	1
	$R$	1	1	1	1	1
PT	$R_B$	0.60	0.60	0.60	0.60	0.60
	$R_D$	0.98	1.00	0.85	0.88	0.90
	$R$	0.59	0.60	0.51	0.53	0.54

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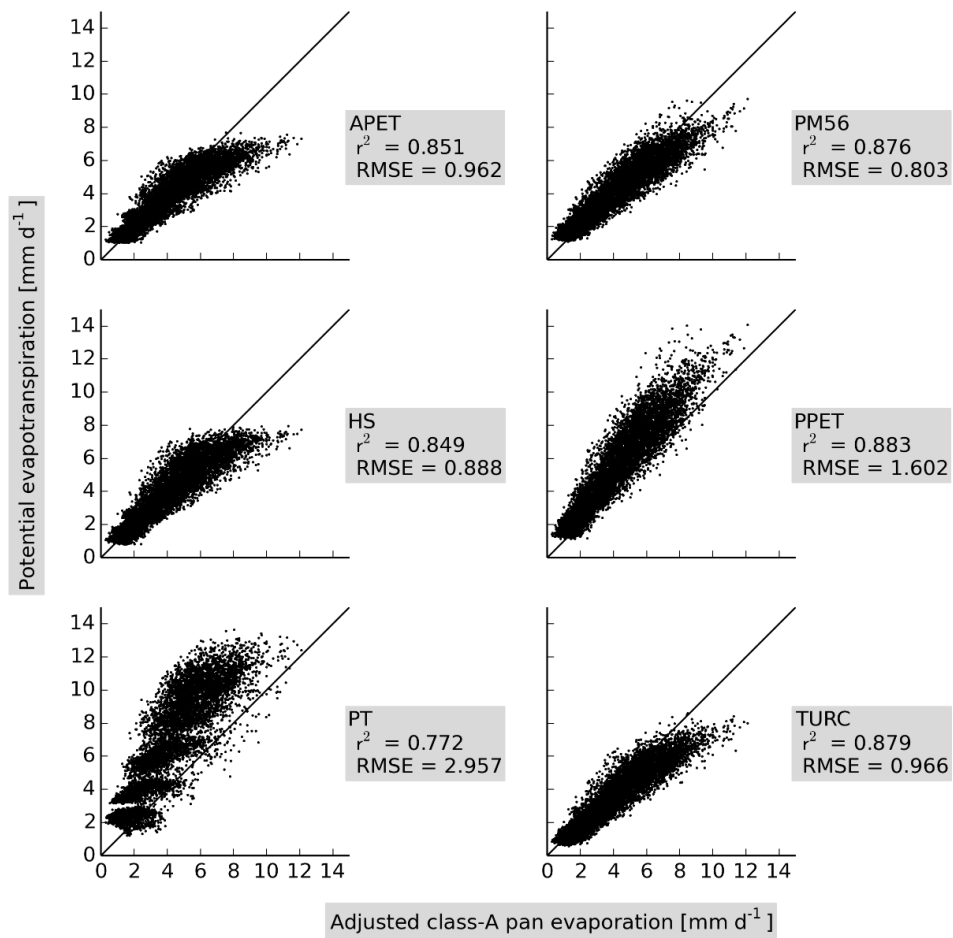


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3 Figure 1. The Murray-Darling basin (MDB) is located in south-east Australia. Irrigated wheat  
4 areas (2005/06) across the MDB are indicated as black dots, n=3,969; cell size=1 x 1 km.

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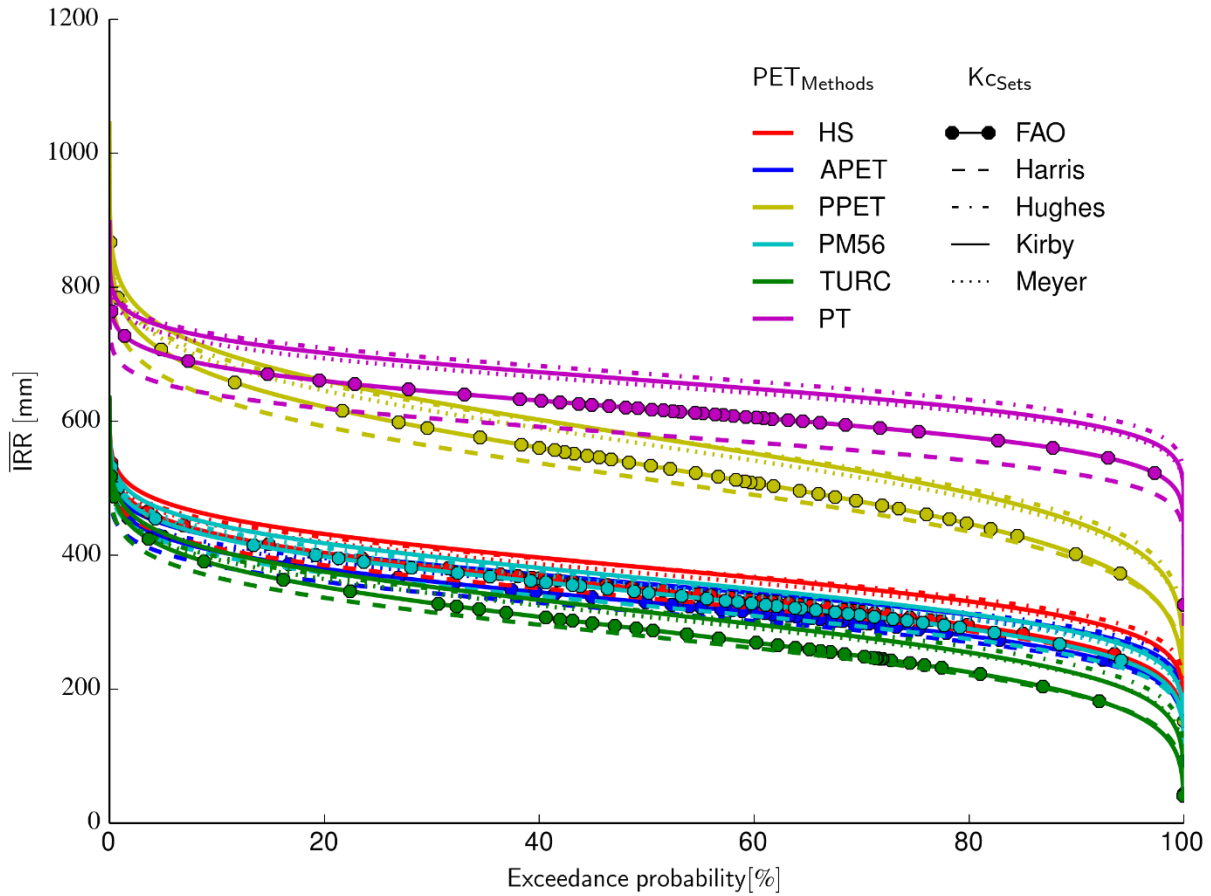
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3 Figure 2. Comparison of daily measured class-A pan evaporation with simulated potential  
4 evapotranspiration at 34 sites in the MDB during the time period from 1986 to 2006. The class-A  
5 pan measurements have been adjusted with site-specific pan coefficients. The coefficient of  
6 determination (r<sup>2</sup>) and the root mean square error (RMSE) are depicted for each ET<sub>0</sub> method  
7 (APET: Areal potential evapotranspiration; PM56: FAO56 Penman-Monteith; HS: Hargreaves-  
8 Samani; PPET: Point potential evapotranspiration; PT: Priestly-Taylor; TURC: Turc).

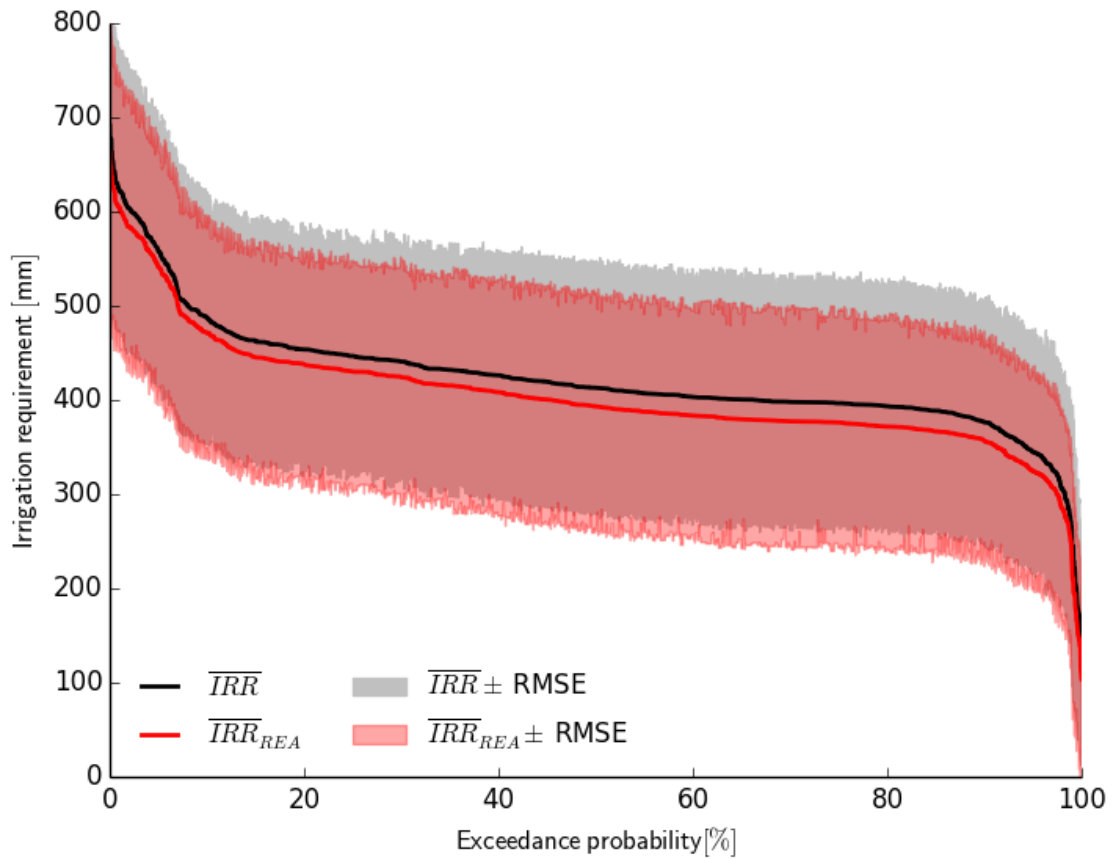
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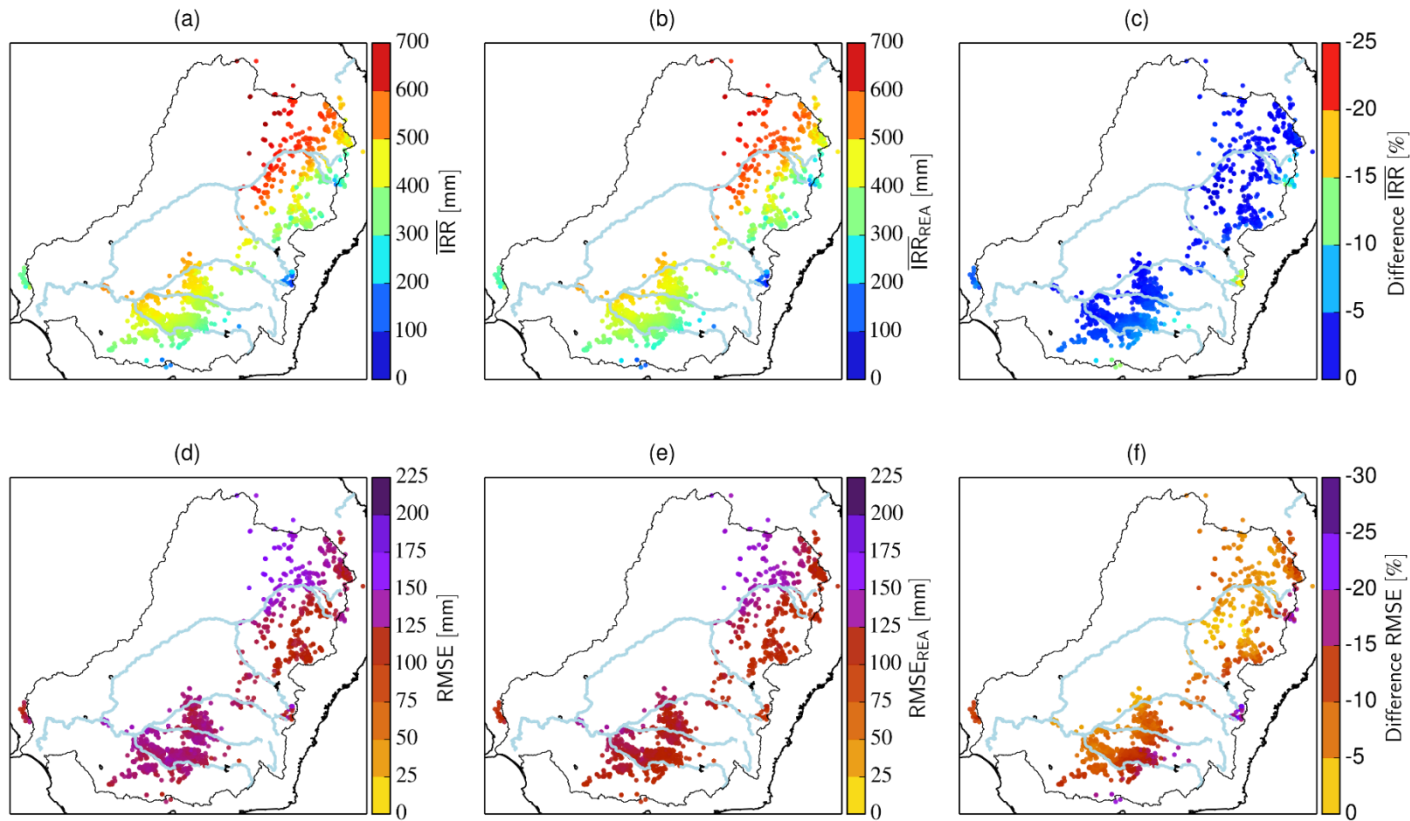
3 Figure 3 Exceedance probability of equally weighted average irrigation water requirement ( $\overline{IRR}$ )  
4 for wheat during the growing season. Averages have been calculated for each cropping area [n =  
5 3969 = 100%] for the period 1986-2006. Colours indicate different ET<sub>o</sub> methods (APET: Areal  
6 potential evapotranspiration; PM56: FAO56 Penman Monteith; HS: Hargreaves-Samani; PPET:  
7 Point potential evapotranspiration; PT: Priestly-Taylor; TURC: Turc) and symbols differentiate  
8 K<sub>c</sub> sets.

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3 Figure 4 Cumulative density function of equally weighted ( $\overline{IRR}$ ) and REA weighted ( $\overline{IRR}_{REA}$ )  
4 average irrigation water requirement for wheat during the growing season. Averages have been  
5 calculated for each cropping area [ $n = 3969 = 100\%$ ] for the period 1986-2006. Colours indicate  
6 the predicted root mean square difference (RMSE) of the ensemble of  $ET_0$  methods and Kc sets  
7 (APET: Areal potential evapotranspiration; PM56: FAO56 Penman Monteith; HS: Hargreaves-  
8 Samani; PPET: Point potential evapotranspiration; PT: Priestly-Taylor; TURC: Turc).



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2 Figure 5. Average equally weighted (a) and REA weighted (b) irrigation water requirement during the growing season of wheat (1986-  
 3 2006). Dots indicate irrigated cropping areas [n=3,969; cell size=1 x 1 km] (note: a buffer has been used to increase the visibility of the  
 4 single grid cells). (c) illustrates the difference between both IRR calculations (b-a). (d) and (e) show the root mean square error between  
 5 the 30 realizations and the equally weighted (d) and REA weighted (e) averages as well as the difference (f) between both calculations,  
 6 respectively.