1	<b>ORACLE</b> (v1.0): Module to simulate the
2	<b>ORganic Aerosol Composition and Evolution in the atmosphere</b>
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#### 12 Abstract

A computationally efficient module to describe organic aerosol (OA) partitioning and 13 chemical aging has been developed and implemented into the EMAC atmospheric chemistry-14 climate model. The model simulates the formation of secondary organic aerosol (SOA) from 15 semi-volatile (SVOCs), intermediate-volatility (IVOCs) and volatile organic compounds 16 (VOCs). It distinguishes SVOCs from biomass burning and all other combustion sources using 17 two surrogate species for each source category with an effective saturation concentration at 298 18 K of  $C^* = 0.1$  and 10 µg m<sup>-3</sup>. Two additional surrogate species with  $C^* = 10^3$  and  $10^5$  µg m<sup>-3</sup> are 19 used for the IVOCs emitted by the above source categories. Gas-phase photochemical reactions 20 that change the volatility of the organics are taken into account. The oxidation products (SOA-sv, 21 SOA-iv, and SOA-v) of each group of precursors (SVOCs, IVOCs, and VOCs) are simulated 22 separately to keep track of their origin. ORACLE efficiently describes the OA composition and 23 evolution in the atmosphere and can be used to i) estimate the relative contributions of SOA and 24 primary organic aerosol (POA) to total OA, ii) determine how SOA concentrations are affected 25 by biogenic and anthropogenic emissions, and iii) evaluate the effects of photochemical aging 26 and long-range transport on the OA budget. We estimate that the global average near-surface 27 OA concentration is 1.5 µg m<sup>-3</sup> and consists of 7% POA from fuel combustion, 11% POA from 28 biomass burning, 2% SOA-sv from fuel combustion, 3% SOA-sv from biomass burning, 15% 29 SOA-iv from fuel combustion, 28% SOA-iv from biomass burning, 19% biogenic SOA-v, and 30 15% anthropogenic SOA-v. The modeled tropospheric burden of OA components is 0.23 Tg 31 POA, 0.16 Tg SOA-sv, 1.41 Tg SOA-iv, and 1.2 Tg SOA-v. 32

#### 34 **1. Introduction**

Atmospheric aerosols adversely affect air quality and human health, and play an important 35 role in climate change. Depending on the physicochemical properties, aerosols affect the energy 36 budget of the Earth's atmosphere by scattering and absorbing solar radiation (direct effect) and 37 influencing the reflective properties of clouds, their lifetime, and precipitation formation 38 (indirect effects). Organic aerosol (OA) is an important constituent of atmospheric particulates 39 and varies with geographic region, accounting for 20 to 90% of the submicron particulate mass 40 (Zhang et al., 2007). However, the understanding of OA sources, atmospheric processing and 41 removal is limited. Given that anthropogenic carbonaceous emissions from developing and 42 emerging economies are expected to dramatically increase in the future (IPCC, 2013), a better 43 understanding of the chemical evolution of OA is essential to reduce the aerosol-related 44 45 uncertainties in global climate simulations and improve air quality and climate assessments.

OA consists of primary material (POA, directly emitted from sources such as fossil fuel 46 combustion and biomass burning) and secondary particulate matter (SOA, formed within the 47 atmosphere from the oxidation of gas-phase precursors). The relative contribution of POA and 48 49 SOA to the overall OA budget remains controversial. Recent studies show that OA is dominated by SOA not only in tropical regions but also in urban environments (Zhang et al., 2005). The 50 51 formation of SOA is often underestimated in atmospheric chemistry-climate models (CCMs) and chemistry-transport models (CTMs) (Heald et al., 2005). Robinson et al. (2007) attributed this 52 53 "unexplained" SOA to the evaporation of POA, oxidation of the resulting vapors in the gas phase and subsequent recondensation, SOA formation from intermediate-volatility organic compounds 54 55 (IVOCs) and chemical aging of the first generation products of the volatile organic compounds (VOCs) that are not taken into account by models. Laboratory and field studies have confirmed 56 57 that the photo-oxidation of fossil fuel combustion and biomass burning emissions can lead to the formation of substantial SOA mass that cannot be explained by the traditional treatment of SOA 58 in CTMs and CCMs (Kroll and Seinfeld, 2008; Jimenez et al., 2009; Grieshop et al., 2009; 59 Hennigan et al., 2011; Miracolo et al., 2011; May et al., 2012). 60

61 CCMs have traditionally treated POA and SOA as nonvolatile and nonreactive particles that 62 are emitted directly into the atmosphere (Kanakidou et al., 2005). Most models convert 63 "hydrophobic" OA to "hydrophilic" OA using an assumed aging rate constant to account for the 64 chemical conversion of fresh OA to more water soluble compounds and it's mixing with other

soluble aerosol components without any mechanistic detail. In order to describe the OA volatility 65 changes that accompany this chemical conversion, Donahue et al. (2006) developed the volatility 66 basis-set (VBS) framework. This framework describes the OA absorptive partitioning, where OA 67 is assumed to be semi-volatile and photochemically reactive and is distributed in logarithmically-68 spaced volatility bins. With this approach, the intermediate and semi-volatile primary emissions 69 and the SOA formation and its chemical aging can be simulated in a common framework that is 70 well suited for regional and global modeling (Murphy and Pandis, 2009; Tsimpidi et al., 2010; 71 Jathar et al., 2011). 72

Several recent regional scale modeling studies have accounted for the semi-volatile nature and 73 chemical aging of organic compounds demonstrating improvements in reproducing the organic 74 aerosol budget and its chemical resolution (Murphy and Pandis, 2009; Tsimpidi et al., 2010; 75 Hodzic et al., 2010; Tsimpidi et al., 2011; Fountoukis et al., 2011; Shrivastava et al., 2011; 76 Bergstrom et al., 2012; Athanasopoulou et al., 2013; Zhang et al., 2013; Fountoukis et al., 2014). 77 However, such efforts on a global scale are limited. Farina et al. (2010) implemented the VBS in 78 79 the Goddard Institute for Space Studies General Circulation Model II' (GISS II') to describe the 80 formation of SOA from VOCs, while assuming that POA emissions were nonvolatile and nonreactive and not accounting for IVOC emissions. Pye and Seinfeld (2010) estimated the global 81 82 OA production from emissions of semivolatile organic compounds (SVOCs) and IVOCs using the global CTM GEOS-Chem. However, this study simulated only the first generation of 83 reactions neglecting the subsequent chemical aging. In contrast to other studies, their 84 implementation of semivolatile POA led to decreases in modeled total surface-level OA 85 86 concentrations due to the partial evaporation of POA upon emission. Jathar et al. (2011) modified the model of Farina et al. (2010) to account explicitly for the semivolatile and reactive 87 88 POA and predicted a global dominance of SOA, which brought the POA/SOA fractions into better agreement with measurements. Jo et al. (2013) employed the VBS framework in the 89 GEOS-Chem global CTM to study the effect of photochemical aging on global SOA. They 90 found that the model results were in better agreement with all observations relative to the 91 simulations without aging and to those of the traditional two-product approach. All of the above 92 93 studies showed improved representation of POA and/or SOA in the corresponding global CTMs and brought the model predictions closer to field measurements. However, the number of studies 94

that have employed the recent OA findings is rather limited and in some cases lack one or more
potentially important processes (i.e., the aging reactions, the semivolatile character of POA).

97 This study aims to improve the description of organic aerosols in large-scale models, making use of the VBS approach and recent developments based on laboratory and field measurements. 98 A new computationally efficient module for the description of organic aerosol composition and 99 evolution in the atmosphere (ORACLE) has been developed and implemented in the 100 101 ECHAM/MESSy Atmospheric Chemistry (EMAC) model (Jöckel et al., 2006) to simulate POA and SOA formation and growth. ORACLE uses a novel lumping method that allows the use of a 102 small number of species, minimizing the computational cost. At the same time it provides 103 valuable information about the OA sources and physicochemical evolution during its 104 atmospheric lifetime. The ORACLE user has full control of the complexity of the OA scheme 105 and is able to adjust the number of species and reactions, depending on the application and the 106 desired chemical resolution (number of compounds and volatility bins), as well as the 107 physicochemical properties of OA components (aging reaction constants, emission factors, etc.). 108 The application of this advanced OA module can help identify gaps in our understanding of the 109 110 OA formation and composition, improve the predictive capability of OA and help in identifying the major uncertainties that should be addressed in future experimental and modeling studies. 111

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#### 113 **2. Global Model Description**

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#### 115 **2.1 EMAC Model**

The ECHAM5/MESSy Atmospheric Chemistry model (EMAC) (Jöckel et al., 2006; 2010) is 116 used as the host model for ORACLE. This is a numerical chemistry and climate simulation 117 118 system that includes sub-models describing lower and middle atmospheric processes and their interaction with oceans, land and human influences. EMAC combines the Modular Earth 119 Submodel System (MESSy; Jöckel et al., 2005) and the 5th generation of the European Centre 120 Hamburg general circulation model (ECHAM5; Röckner et al., 2006). ECHAM5 simulates the 121 atmospheric flow with the prognostic variables vorticity, divergence, temperature, total moisture 122 123 and the logarithm of the surface pressure, and is integrated in the base model layer of MESSy. The interface structure of MESSy allows the use of different modules for atmospheric chemistry, 124 125 transport and diagnostic tools. EMAC has been extensively described and evaluated against

ground-based and satellite observations, and can be run on a range of spatial resolutions (Jöckel
et al., 2006; Pozzer et al., 2012; de Meij et al., 2012). In this study, the applied spectral resolution
is T42L31, corresponding to a horizontal grid resolution of approximately 2.8°x2.8° and 31
vertical layers extending to 25 km. EMAC is applied for 6 years covering the period 2004-2009
and the first year is used as spin-up.

The EMAC model simulates gas-phase species online through the Module Efficiently 131 132 Calculating the Chemistry of the Atmosphere submodel (MECCA; Sander et al., 2011). MECCA calculates online the concentration of the major oxidants (OH,  $H_2O_2$ , NO<sub>3</sub>, and O<sub>3</sub>) by using a 133 chemical scheme based on the MIM mechanism (Pöschl et al., 2000) including 218 gas phase 134 reactions, 69 photolysis reactions, and 12 heterogeneous reactions of 129 gases in total. 135 Inorganic aerosol microphysics and gas/particle partitioning are calculated by the Global Modal-136 aerosol eXtension aerosol module (GMXe; Pringle et al., 2010). This submodel is 137 computationally efficient and is suitable for medium to long-term simulations with global and 138 regional models. 139

The aerosol microphysics are described using 7 interacting lognormal modes 140 141 (4 hydrophilic and 3 hydrophobic modes). The 4 hydrophilic modes cover the full aerosol size spectrum (nucleation, Aitken, accumulation, and coarse modes). Each size range has fixed size 142 143 boundaries and a variable mean radius. The 3 hydrophobic modes have the same size range as the hydrophilic modes apart from the nucleation mode. The aerosol composition within each 144 mode is uniform with size (internally mixed), though the composition can vary between modes 145 (externally mixed). The removal of gas and aerosol species through dry deposition is calculated 146 147 within the DRYDEP submodel (Kerkweg et al., 2006a) based on the big leaf approach. The sedimentation of aerosols is calculated within the SEDI submodel (Kerkweg et al., 2006a) using 148 149 a first order approach. The formation of sulfate through the aqueous phase oxidation of  $SO_2$  is treated by the SCAV submodel (Tost et al., 2006) that calculates the aqueous phase redox 150 151 reactions based on the prognostically predicted pH of clouds. SCAV also simulates cloud scavenging and rainout of gas and aerosol species. The optical properties of aerosols (optical 152 153 thickness, single scattering albedo, asymmetry factor) are calculated within the AEROPT 154 submodel (Lauer et al., 2007) based on pre-calculated look-up tables from Mie theory. The CLOUD submodel (Jöckel et al., 2006) is used to calculate the cloud cover as well as cloud 155 micro-physics including precipitation. The detailed two-moment liquid and ice-cloud 156

microphysical scheme of Lohmann et al. (2007), which enables a physically based treatment ofaerosol-cloud interactions, is used to compute cloud microphysical processes.

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#### 160 **3. ORACLE description**

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#### 162 **3.1 Module Overview**

The ORACLE module (i) uses logarithmically-spaced saturation concentration bins to 163 describe the OA components based on their volatility (Section 3.3), (ii) simulates the 164 contribution of SVOCs and IVOCs from fuel combustion and biomass burning emissions and 165 anthropogenic and biogenic VOCs to the formation of SOA (Section 3.4), (iii) monitors the gas-166 phase photochemical reactions of SOA precursors (Section 3.5), (iv) assumes bulk equilibrium 167 between the gas and particulate phases (Section 3.6), and (v) distributes the OA in size modes 168 (Section 3.6). A schematic overview of the ORACLE module and how it is implemented in 169 EMAC is provided in Figure 1. 170

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#### 172 **3.2 OA Terminology**

The ORACLE module describes the chemical life cycle of a large number of different species 173 involved in the formation of total atmospheric OA (i.e., 48 species are used in the current 174 configuration) in order to cover and track the distribution of these species in multiple 175 simultaneously occurring phases, their participation in a large number of possible formation 176 177 pathways (i.e., multigenerational oxidation), their volatility distribution and their sources. The complexity of this system demands a systematic classification of the organic compounds that 178 will be in line with ongoing field and laboratory studies and model development. Following the 179 Murphy et al. (2014) naming convention for classifying OA components, the ORACLE 180 181 classification system has the following syntax:

182

#### [Source root name]-[initial volatility]

183 with the "source root name" indicated by an acronym currently used in the field and the "initial 184 volatility" of the species upon emission indicated by an alphabetical lowercase suffix. Table 1 185 lists the root terms and suffixes used for identifying the emitted volatility of the organic 186 components described by ORACLE in this application. The source root name of organic components consist of lowercase alphabetical characters (Modifiers) that describe their source type (i.e., "f" for fuel combustion), followed by three capital letters (Base Term): the first letter declares if the organic component is primary ("P"), coming from primary emissions, or secondary ("S"), formed from the oxidation of gas phase species; the second is always the letter "O" denoting organic; and the third identifies the phase of the species ("A" for aerosol and "G" for gas). The exact names of all the organic compounds simulated by ORACLE are listed in Table 2.

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# 195 **3.3 Volatility Basis Set Theory**

The ORACLE module adopts the volatility basis set approach proposed by Donahue et al. 196 (2006). It subdivides the thousands of organic compounds into groups (surrogate species) with 197 logarithmically-spaced effective saturation concentrations. This framework abandons the 198 traditional distinction between POA and SOA and allows the EMAC model to efficiently treat 199 both semivolatile primary emissions and SOA production, and then simulate the chemical 200 evolution of these species under a unified framework. The exact volatility resolution defined and 201 202 used by ORACLE can be controlled through the interface layer of the module, offering the flexibility to investigate different schemes. For the current application, eight surrogate species 203 are used for POA divided into two groups each with saturation concentrations C<sup>\*</sup> at 298 K equal 204 to  $10^{-1}$ ,  $10^{1}$ ,  $10^{3}$ ,  $10^{5}$ µg m<sup>-3</sup>: i) POA from biofuel combustion, fossil fuel combustion, and other 205 206 urban sources (fPOA) and ii) POA from biomass burning (bbPOA). The least volatile fraction, at  $10^{-1} \mu g m^{-3}$ , describes the low volatility organics in the atmosphere that are mostly in the 207 particulate phase even in remote locations. On the other hand, even under highly polluted 208 conditions the majority of the material in the  $10^5 \mu g m^{-3}$  volatility bin will exist almost 209 210 exclusively in the vapor phase. The user can control the complexity of the ORACLE module and therefore can easily add bins at either end of the range to describe and investigate unique 211 212 conditions. It is worth mentioning that the use of low volatility bins to accurately represent the extremely low volatility organic compounds (e.g., with  $C^*$  lower than  $10^{-3}$ ) would be useful for 213 the study of new particle formation which is outside the scope of the current work and is not 214 included in the current version of ORACLE. A similar approach is followed for SOA formed 215 from VOCs. Following the approach of Lane et al. (2008) it is assumed that the oxidation of the 216 anthropogenic and biogenic VOC species (aVOC and bVOC, respectively) results in four 217

218 products for each precursor distributed in four volatility bins with effective saturation 219 concentrations at 298K equal to 1, 10,  $10^2$  and  $10^3 \ \mu g \ m^{-3}$  at 298K. This range is typically 220 constrained by available smog chamber data (Stanier et al., 2008). A wider range can also be 221 used and tested through the ORACLE interface. The different aerosol types and chemical 222 processes modeled in the proposed framework are illustrated in Figure 2.

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# 224 **3.4 Emission inventory of OA Precursors**

*VOC emissions*: The emissions of isoprene and monoterpenes are calculated online by EMAC 225 with the ONLEM submodel (Kerkweg et al., 2006b) and depend on ecosystem type, temperature, 226 and solar radiation. The AIRSEA submodel estimates the oceanic isoprene emissions from 227 chlorophyll concentrations (Pozzer et al., 2006). In this application, isoprene and monoterpene 228 global emissions are 477 and 62.4 Tg yr<sup>-1</sup>, respectively. The emissions of the aVOCs that are 229 considered SOA precursors are derived from the CMIP5 RCP4.5 emission inventory (Clarke et 230 al., 2007). These compounds were not considered by the original gas phase chemistry submodel 231 MECCA. Therefore, six lumped aVOC species have been added to MECCA to assess the aSOA 232 233 formation in ORACLE. These lumped species are grouped similarly to the Statewide Air Pollution Research Center SAPRC99 chemical mechanism (Carter, 2000) into: two alkane 234 235 species, two olefins, and two aromatics. The aVOCs that are considered SOA precursors and their annual global emissions are listed in Table 3. 236

237 IVOC and SVOC emissions: The EMAC model considers the contribution of fuel combustion and biomass burning to POA emissions while primary marine organic emissions are not included 238 in this application. The AEROCOM database is used for the aerosol emissions from fossil and 239 biofuel combustion based on the year 2000 (Dentener et al., 2006). The biomass burning 240 241 contribution is based on the Global Fire Emissions Database (GFED version 3.1) which is monthly resolved and covers the period 1997-2009 (van der Werf et al., 2010). These emission 242 datasets treat POA as non-reactive and non-volatile. However, the POA concentration upon 243 emission is highly sensitive to ambient conditions, including dilution and temperature 244 (Hildemann et al., 1989; Lipsky and Robinson, 2006). ORACLE accounts for the volatility of 245 POA emissions by distributing the traditional nonvolatile emissions into emissions of SVOCs 246 and IVOCs. SVOCs have saturation concentrations between 0.01 and 100  $\mu$ g m<sup>-3</sup> and exist in 247

both the gas and particulate phases under typical ambient conditions. IVOCs have saturation concentrations between  $10^3$  and  $10^6 \,\mu g \, m^{-3}$  and exist in the gas phase (Pandis et al., 2013).

250 For the current application, the model distinguishes SVOCs from biomass burning and other combustion sources (biofuel and fossil fuel combustion, and other urban sources) using two 251 252 surrogate species for each emission category with effective saturation concentration at 298 K of  $C^* = 0.1$  and 10 µg m<sup>-3</sup> to cover the volatility range of SVOCs (0.01 to 100 µg m<sup>-3</sup>). For the 253 IVOCs two additional surrogate species with  $C^* = 10^3$  and  $10^5 \mu g m^{-3}$  are used for each of the 254 above two source categories to cover the corresponding volatility range ( $10^3$  to  $10^6 \,\mu g m^{-3}$ ). The 255 emission factors used for the distribution of traditional POA emissions into SVOCs and IVOCs 256 are based on the work of Tsimpidi et al. (2010). These emission factors also account for the 257 258 additional IVOC emissions that were not included in the original emission inventory used by EMAC. Traditional emission inventories account only for a small fraction of the IVOCs since 259 they are based on samples using quartz and/or Teflon filters collected at aerosol concentrations 260 up to  $10^4$  µg m<sup>-3</sup> (Shrivastava et al., 2008; Robinson et al., 2010). The amount of IVOC 261 emissions missing from traditional inventories is estimated to be between 0.25 and 2.8 times the 262 263 traditional POA emissions (Schauer et al., 1999; 2001; 2002). In this study, we assume that the missing IVOC emissions are 1.5 times the traditional emission inventory (Shrivastava et al., 264 2008; Tsimpidi et al., 2010) and are assigned to the fourth volatility bin with  $C^* = 10^5 \,\mu g \, m^{-3}$ . 265 Table 4 lists the emission factors for each SVOC and IVOC and their total annual global 266 267 emissions from fuel combustion use and biomass burning. SVOCs are assumed to be emitted as POA while IVOCs are emitted as POG. Then, they are allowed to partition between the gas and 268 particle phase, which can result in some POG from the evaporation of POA. 269

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# 271 **3.5 Photochemical Reactions**

*Photo-oxidation of VOCs*: The photochemical oxidation of VOCs that are considered as SOA precursors has been implemented into MECCA. The updated mechanism considers the oxidation of alkenes, aromatics and isoprene by OH, and the oxidation of olefins and monoterpenes by O<sub>3</sub>, O, OH, and NO<sub>3</sub>. The oxidation products from anthropogenic (alkenes, aromatics, and olefins) and biogenic (monoterpenes and isoprene) VOCs are lumped into two groups: 1) secondary organic aerosols from the oxidation of anthropogenic VOCs (aSOA-v), and 2) secondary organic aerosols from the oxidation of biogenic VOCs (bSOA-v). These groups are further distributed

into volatility bins with logarithmically-spaced effective saturation concentrations (Figure 2).
The photo-oxidation of aVOCs and bVOCs is described by the following reactions:

281 
$$aVOC + Oxidants \rightarrow \sum_{i=1}^{n} a_i aSOG-v_i$$
 (R1)

282

284

$$aSOG-v_i \leftrightarrow aSOA-v_i$$
 (R2)

283 
$$bVOC + Oxidants \rightarrow \sum_{i=1}^{n} a_i bSOG - v_i$$
 (R3)

$$bSOG-v_i \leftrightarrow bSOA-v_i$$
 (R4)

where  $a_i$  is the aerosol yield, *i* is the corresponding volatility bin, and *n* is the total number of volatility bins. In this application, four volatility bins are considered and the corresponding aerosol yields are those used by Tsimpidi et al. (2010) based on laboratory results from smogchamber experiments under high-NO<sub>x</sub> conditions for aVOCs and low-NO<sub>x</sub> conditions for bVOCs (Table 5). This is based on the assumption that urban areas, where most of aVOCs are emitted (~90%), are characterized by high-NO<sub>x</sub> conditions (Tsimpidi et al., 2008; Karl et al., 2009) and forested regions by low-NO<sub>x</sub> conditions (Pugh et al., 2010; Browne et al., 2013).

292 Chemical Aging Reactions: ORACLE treats all OA compounds as chemically reactive. However, only homogeneous gas-phase aging is considered since it is rapid compared to 293 294 heterogeneous reactions with OH (Donahue et al., 2013). Smog-chamber results indicate a net average decrease in volatility (and increase in SOA production) of aSOA-v after 295 296 multigenerational aging (Hildebrandt et al., 2009). In this application, the volatilities of aSOA-v reacting with OH are reduced by a factor of 10 (Figure 2) with a rate constant of  $1 \times 10^{-11}$  cm<sup>3</sup> 297 molecule<sup>-1</sup> s<sup>-1</sup> (Tsimpidi et al., 2010) and a 7.5% increase in mass to account for one added 298 oxygen, assuming a  $C_{15}$  precursor. aSOA-v can participate in up to three generations of oxidation 299 300 reaching a final organic matter/organic carbon ratio (OM/OC) up to 2.2 (assuming an initial OM/OC of 1.8), which is within the limits (OM/OC:1.8-2.4) of the observed OM/OC of the 301 oxygenated organic aerosols (Aiken et al., 2008). Existing evidence suggests that the aging of 302 303 bSOA-v does not result to an important change in its mass concentration (Ng et al., 2006; Donahue et al., 2012). Murphy et al. (2012) attributed this to a balancing of fragmentation and 304 305 functionalization effects during the photochemical aging of bSOA-v. Therefore, it is assumed here that the chemical aging of bSOA-v does not result in a net increase of the corresponding 306 SOA concentration. The chemical aging of aSOA-v is described by the following reaction: 307

308  $aSOG-v_i + OH \rightarrow 1.075 aSOG-v_{i-1}$  (R5) 309  $aSOG-v_{i-1} \leftrightarrow aSOA-v_{i-1}$  (R6)

volatilities of SVOCs and IVOCs are reduced by a factor of 100 310 The (Figure 2) as a result of the OH reaction with a rate constant of  $2 \times 10^{-11}$  cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup> (Pve 311 and Seinfeld, 2010) and a 15% increase in mass to account for two added oxygens, assuming a 312  $C_{15}$  precursor. This represents a more aggressive addition of oxygen and reduction in volatility 313 compared to aSOA-v. In the present application, SVOCs and IVOCs can participate in up to three 314 315 generations of oxidation reaching a final OM/OC up to 1.8 (assuming an initial OM/OC of 1.2), which is consistent with the observed OM/OC (OM/OC:1.8-2.4) of the oxygenated organic 316 aerosols (Aiken et al., 2008). The oxidation products of SVOCs and IVOCs are called SOA from 317 the oxidation of fuel combustion and biomass burning SVOCs (fSOA-sv and bbSOA-sv, 318 respectively) and SOA from the oxidation of fuel combustion and biomass burning IVOCs 319 320 (fSOA-iv and bbSOA-iv, respectively). The photo-oxidation of SVOCs and IVOCs is described by the following reactions: 321

322	$fSVOC_i + OH \rightarrow 1.15 fSOG-sv_{i-1}$	(R7)
323	$fSOG\text{-}sv_{i\text{-}1} \leftrightarrow fSOA\text{-}sv_{i\text{-}1}$	(R8)
324	$fIVOC_i + OH \rightarrow 1.15 fSOG-iv_{i-1}$	(R9)
325	$\mathrm{fSOG}\text{-}\mathrm{iv}_{\mathrm{i}\text{-}\mathrm{l}} \leftrightarrow \mathrm{fSOA}\text{-}\mathrm{iv}_{\mathrm{i}\text{-}\mathrm{l}}$	(R10)
326	$bbSVOC_i + OH \rightarrow 1.15 \ bbSOG\text{-}sv_{i-1}$	(R11)
327	$bbSOG\text{-}sv_{i\text{-}1} \leftrightarrow bbSOA\text{-}sv_{i\text{-}1}$	(R12)
328	bbIVOC <sub>i</sub> + OH $\rightarrow$ 1.15 bbSOG-iv <sub>i-1</sub>	(R13)
329	bbSOG-iv <sub>i-1</sub> ↔ bbSOA-iv <sub>i-1</sub>	(R14)

The products of reactions R7-R14 can be further oxidized by OH forming species with lower saturation concentration until they reach the lowest volatility bin (i.e., the oxidation of fSOA-sv<sub>i</sub> will produce the fSOA-sv<sub>i-1</sub>). Overall, all OA surrogate species are assumed to have a constant hygroscopicity parameter kappa of 0.14 that remains constant during their atmospheric aging.

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### 335 **3.6 Gas/Aerosol Partitioning**

ORACLE calculates the partitioning of organic compounds between the gas and particle phases by assuming bulk equilibrium and that all organic compounds form a pseudo-ideal solution. The gas/aerosol partitioning is performed in two steps as follows:

Bulk equilibrium: ORACLE calculates the bulk equilibrium gas and aerosol concentrations following the approach of the SOAM II model of Strader et al. (1999). Considering partitioning of n organic compounds and assuming pseudo-ideal solution, a set of *n* nonlinear equations is obtained:

343 
$$c_{a,i} = c_{t,i} - x_i c_i^* \text{ for } i = 1, m$$

344

345 
$$x_{i} = \frac{c_{a,i}/M_{i}}{\sum_{j=1}^{n} c_{a,j}/M_{j}},$$

where  $c_{i,i}$  and  $c_{a,i}$  are the total and aerosol-phase concentrations of product *i* in µg m<sup>-3</sup>, respectively,  $c_i^*$  is the effective saturation concentration of product *i*,  $x_i$  is the mole fraction of product *i* in the absorbing organic phase, and  $M_i$  is the molecular weight of product *i*. The molecular weights of all POA, SOA-sv, and SOA-iv components are assumed to be 250 g mol<sup>-1</sup> while the molecular weights of bSOA-v and aSOA-v are 180 and 150 g mol<sup>-1</sup>, respectively (Tsimpidi et al., 2010). The temperature dependence of saturation concentrations is described by the Clausius–Clapeyron equation:

353 
$$c_i^* = c_{i,0}^* \frac{T_0}{T} \exp\left[\frac{\Delta H}{R} \left(\frac{1}{T_0} - \frac{1}{T}\right)\right],$$

where  $c_i^*$  and  $c_{i,0}^*$  are the saturation concentrations at temperature T and T<sub>0</sub>, respectively, R is the 354 gas constant, and  $\Delta H$  is the enthalpy of vaporization. In this application an effective  $\Delta H$  of 30 kJ 355 mol<sup>-1</sup> is used for all aSOA-v and bSOA-v species based on data for  $\alpha$ -pinene (Pathak et al., 356 2007), and a  $\Delta H$  of 106, 94, 82, and 70 kJ mol<sup>-1</sup> is used for the 10<sup>-1</sup>, 10<sup>1</sup>, 10<sup>3</sup>, 10<sup>5</sup> µg m<sup>-3</sup> 357 volatility bins, respectively, for all organic compounds from fuel combustion and biomass 358 burning sources based on data for large saturated species commonly found in primary emissions 359 (Donahue et al., 2006). ORACLE solves this equation set, which yields the bulk aerosol 360 composition at equilibrium. 361

Aerosol size distribution: The aerosol size distribution is determined by distributing the change in aerosol mass after the bulk equilibrium into each size mode using a weighting factor (Pandis et al., 1993). Assuming pseudo-ideal solution, the fraction,  $f_{i,k}$ , of total flux of species *i* between gas and aerosol phases that condenses onto or evaporates from an aerosol mode *k* is given by:

367 
$$f_{i,k} = \frac{N_k d_k \left(c_i - x_{i,k} c_i^*\right) / (\beta_k + 1)}{\sum_{l=1}^m N_l d_l \left(c_i - x_{i,l} c_i^*\right) / (\beta_l + 1)},$$

where  $N_k$  and  $d_k$  are the number and mean diameter of particles in the mode k, respectively, m is the total number of aerosol modes,  $\beta_k = 2\lambda/\alpha d_k$ ,  $\alpha$  is the aerosol accommodation coefficient, and  $\lambda$  is the mean free path of air molecules (Pandis et al., 1993). The above equation is solved iteratively at each time step and determines the OA composition of each mode. Overall, the user is allowed to use up to 3 hydrophilic modes (Aitken, accumulation, coarse) for the size distribution of all OA surrogate species. In the current application, only the accumulation mode is used to limit the computational cost of the module.

375

#### 376 **4. Model results**

377

#### 378 **4.1 Total OA concentrations**

The model calculated global average surface OA concentration is 1.5  $\mu$ g m<sup>-3</sup> (Figure 3). 379 High OA concentrations are modeled over regions affected by biomass burning and biogenic 380 VOC emissions: the tropical forests and savannas of South America (Amazon Basin), Africa 381 (Congo Basin), and Southeast Asia. These areas are characterized by a pronounced annual 382 cycle due to the influence of two distinct seasons, the wet and dry periods. The dry season is 383 characterized by intense wildfires and OA consists mainly of biomass burning OA. The wet 384 season is characterized by low biomass burning emissions; therefore, OA consists mainly of 385 386 biogenic SOA. Considerable OA concentrations are also calculated over the industrialized regions of the Northern Hemisphere (i.e., China, Europe, and Eastern US), where strong fossil 387 and biofuel combustion related sources are located. The model simulates a continental 388 background OA concentration of 1-2  $\mu$ g m<sup>-3</sup>, which is mainly formed by the condensation of 389

oxidized low volatility organic gases. These gases are formed from the photochemical aging
of VOCs, IVOCs and SVOCs, which have been emitted in the gas phase and have been
efficiently transported remote from their sources. The effects of trade wind transport of bbOA
off the subtropical west coasts of Africa and America are discernible in Figure 3. In these
regions, the OA concentration is increased by long-range transport of continental air masses
to adjacent ocean areas.

396

### **397 4.2 Model performance evaluation**

Figure 4 shows the comparison of model calculated OA concentrations with measurements 398 from 399 the European Monitoring and Evaluation Programme (EMEP; http://nilu.no/projects/ccc/onlinedata/pm/) and the Interagency Monitoring of protected 400 http://vista.cira.colostate.edu/improve Visual Environments 401 (IMPROVE; /Data/IMPROVE/improve data.htm), as well as short-term measurement data collected over 402 East Asia as summarized by Jo et al. (2013). Each point of the scatter plots represents a monthly 403 average value at a particular monitoring station. The mean bias (MB), mean absolute gross error 404 405 (MAGE), normalized mean bias (NMB), normalized mean error (NME), and the root mean square error (RMSE) were also calculated (Table 6) to assess the model performance: 406

407 
$$MAGE = \frac{1}{N} \sum_{i=1}^{N} |P_i - O_i| \qquad MB = \frac{1}{N} \sum_{i=1}^{N} (P_i - O_i)$$

408

409 
$$NME = \frac{\sum_{i=1}^{N} |P_i - O_i|}{\sum_{i=1}^{N} O_i} \qquad NMB = \frac{\sum_{i=1}^{N} (P_i - O_i)}{\sum_{i=1}^{N} O_i}$$

410

411 
$$RMSE = \left[\frac{1}{N}\sum_{i=1}^{N} (P_i - O_i)^2\right]^{\frac{1}{2}},$$

where  $P_i$  is the modeled OA mass,  $O_i$  is the observed value of OA at the same monthly averaged time, and *N* is the total number of data points used for the comparison. NME (in %) and MAGE (in µg m<sup>-3</sup>) provide an assessment of the overall discrepancy between model predictions and observations, while NMB (in %) and MB (in µg m<sup>-3</sup>) are indicative of systematic errors. RMSE 416 (in  $\mu$ g m<sup>-3</sup>) incorporates both the variance of the prediction and its bias. Both NME and MAGE 417 inherently include the corresponding bias, which is the reason why their magnitude is equal or 418 larger than NMB and MB, respectively. For an unbiased calculation, NME and MAGE express 419 the variance. When NME and NMB or MAGE and MB, respectively, are close to each other in 420 magnitude, the discrepancy is explained as a systematic bias rather than scatter. When the 421 magnitude of NME/MAGE is larger than NMB/MB, part of the discrepancy between predictions 422 and observations is explained as scatter.

The model, despite its coarse resolution, captures relative well the monthly average 423 concentrations of OA over the USA, Europe, and Asia (Table 6). This is rather encouraging 424 given the expected uncertainties in the emission inventory and in a number of parameters used 425 by the model such as the emission fractions for POA from fuel combustion and biomass burning, 426 the reaction rates and aerosol yields that accompany the formation of SOA from SVOCs, IVOCs 427 and VOCs, etc. In addition, the fact that the formation of SOA from aqueous-phase reactions and 428 heterogeneous reactions, including processes like oligomerization, have not been accounted for, 429 adds to the model bias. However, despite these limitations, the modeled total OA concentrations 430 431 are generally in reasonable agreement with the measurements. More precisely, over the USA, the model overestimates the OA with a NMB of 57%. The high NME (89%) indicates that part of 432 433 the discrepancy between model results and observations is explained as scatter, which is visible in Figure 4a. Over Europe, the model overestimates OA with an NMB of 22%. The 434 measurements of OA in the station of Ispra, Italy, are systematically high (up to 22  $\mu$ g m<sup>-3</sup>) and 435 have been excluded from the statistical analysis. The model performs worst over East Asia 436 (RMSE = 5.5) and underestimates OA concentrations with an NMB of -36% since it cannot 437 capture the high values measured over Beijing and Shijiazhuang (up to 32  $\mu$ g m<sup>-3</sup>) possibly due 438 439 to its limited spatial resolution.

In contrast to highly populated areas, long-term observations of OA over forests are unavailable, therefore we have collected OA data measured during the short-term field campaigns of DABEX, DODO, and AMMA over the subtropical West Africa (Capes et al., 2008; Capes et al., 2009), AMAZE-08 and SAMBBA over the Amazonian rainforest (Chen et al., 2009; Brito et al., 2014), and measurement data collected over the Canadian boreal forest (Schwartz et al., 2010; Takahama et al., 2011). Capes et al. (2008) performed aircraft measurements during the DABEX and DODO field experiments (January and February of 2006) 447 over subtropical West Africa at altitudes up to 4000 m. This period was characterized by intense agricultural fires in the sub-Sahelian part of West Africa resulting in high OA concentrations 448  $(4-16 \ \mu g \ m^{-3})$ . EMAC agrees reasonably well with the observations since the calculated OA 449 concentrations during January and February of 2006 range between 2 and 12.5 µg m<sup>-3</sup>, mainly 450 451 due to high biomass burning emissions over the area. During the wet season of the same year (July and August 2006), Capes et al. (2009) carried out aircraft measurements over subtropical 452 453 West Africa in the frame of the AMMA project. This period is characterized by low biomass burning emissions; therefore, OA consists according to EMAC mainly of biogenic SOA. The 454 observed median concentration is 1.1  $\mu$ g m<sup>-3</sup> while EMAC calculates an average value of 2.7  $\mu$ g 455  $m^{-3}$ . In the Northwestern Amazon basin, Chen et al. (2009) reported OA concentrations of 0.6  $\mu$ g 456 m<sup>-3</sup> and 0.9 µg m<sup>-3</sup> on average, during the wet season of 2008 (February and March) as part of 457 the AMAZE-08 experiment. The corresponding average OA concentration by EMAC is 2.7 µg 458 m<sup>-3</sup>. During the dry season (September 2012), Brito et al. (2014) reported an average OA 459 concentration of 13.7 µg m<sup>-3</sup> in the Southwestern Amazon basin during the SAMBBA field 460 experiment. Since this value is mostly affected by biomass burning emissions (which vary 461 462 significantly interannually) and the EMAC model applied emissions during the years 2005-2009, a direct comparison between observations and measurements is not possible. However, EMAC 463 captures the increased concentrations during the dry seasons since OA varies between 9.8  $\mu$ g m<sup>-3</sup> 464 and 30  $\mu$ g m<sup>-3</sup> over the same area during the months of September. Finally, Schwartz et al. 465 466 (2010) and Takahama et al. (2011) performed measurements of OA over the boreal forest of British Columbia, Canada during May-June 2008 and March-September 2009, respectively. The 467 reported values were 1.3 µg m<sup>-3</sup> (May-June 2008), 0.6 µg m<sup>-3</sup> (March-April 2009), and 4.1 µg m<sup>-3</sup> 468 <sup>3</sup> (May-September 2009). The OA concentrations calculated by EMAC fit well with this range of 469 observations with values of 1.4  $\mu$ g m<sup>-3</sup>, 0.9  $\mu$ g m<sup>-3</sup>, and 4.5  $\mu$ g m<sup>-3</sup>, respectively. 470

471

#### 472 **4.3 Primary Organic Aerosol**

473 POA is the fraction of the organic emissions that is emitted and contained in the aerosol phase 474 and has not undergone chemical reactions. POA in the present application is divided into fPOA 475 and bbPOA depending on its source as described in section 3.3. fPOA and bbPOA have high 476 concentrations close to their sources; further downwind the concentrations rapidly decrease due 477 to dilution and evaporation (Figure 5).

*fPOA*: The global and land average surface concentrations of fPOA are 0.11  $\mu$ g m<sup>-3</sup> and 0.22 478 μg m<sup>-3</sup>, respectively (Table 7; Figure 5a). Fossil fuels and biofuels are used in the industrial, 479 480 residential, and transport sectors, therefore the fPOA concentration peaks over densely populated and highly industrialized areas. In fact, the highest fPOA concentration is calculated over Eastern 481 China (up to 8  $\mu$ g m<sup>-3</sup> over Beijing), India and Bangladesh (1-4  $\mu$ g m<sup>-3</sup>). In Europe the highest 482 concentration is found over Eastern European countries (1-2  $\mu$ g m<sup>-3</sup> over Poland and Romania), 483 Central Europe (0.5-1 µg m<sup>-3</sup> over the greater Paris and Rhine-Ruhr areas), and in the vicinity of 484 Moscow (0.5-1 µg m<sup>-3</sup>). Over Africa fPOA concentrations peak over the western part of the 485 continent (1-3  $\mu$ g m<sup>-3</sup> over Nigeria), while it is also high over the greater areas of Addis Ababa, 486 Kampala and Johannesburg (0.5-1.5 µg m<sup>-3</sup>). Over North America, the highest fPOA levels are 487 modeled around Los Angeles (0.5-1  $\mu$ g m<sup>-3</sup>), the northeastern USA (0.5-1.5  $\mu$ g m<sup>-3</sup>) and the 488 Mexico City metropolitan area (0.5-1  $\mu$ g m<sup>-3</sup>), while over South America fPOA is high over the 489 greater area of Rio de Janeiro (0.5-1  $\mu$ g m<sup>-3</sup>). 490

*bbPOA*: The global and land average surface concentrations of bbPOA are 0.12  $\mu$ g m<sup>-3</sup> and 491  $0.39 \ \mu g \ m^{-3}$ , respectively (Table 7; Figure 5b). The bbPOA levels are affected by emissions from 492 forest, woodland, peatland, and savannah fires as well as agricultural waste burning. Therefore, 493 high concentrations are modeled over the tropical rainforest and savannas in the Amazon and 494 Congo Basins (5-10 µg m<sup>-3</sup>), Southeast Asia (3-6 µg m<sup>-3</sup>), and the boreal forests of Alaska, 495 Canada, and Russia. During the dry season, bbPOA concentrations increase significantly (e.g., 496 10-50 µg m<sup>-3</sup> during January over the Congo Basin) while during the wet season they are much 497 lower (e.g.,  $0.5-3 \mu g m^{-3}$  during July over the Congo Basin). 498

499

### 500 4.4 Secondary Organic Aerosol from SVOCs

501 SVOCs are emitted in both the aerosol and gas phase and can be transferred from phase to phase due to cooling and dilution as the air masses travel from their sources. SVOCs in the gas 502 503 phase can be oxidized and possibly re-condense to the aerosol phase forming secondary organic aerosols (SOA-sv). In ORACLE SOA-sv is subdivided into fSOA-sv and bbSOA-sv depending 504 505 on its source as described in section 3.5. fSOA-sv and bbSOA-sv concentrations are high according to our simulation downwind of polluted urban areas and megacities and the major 506 rainforests (Figure 6). However, as most of the SVOCs are in the aerosol phase after emissions 507 (as POA) and remain there without undergoing chemical reactions during their atmospheric 508

509 lifetime, fSOA-sv and bbSOA-sv concentrations are significantly lower compared to the 510 corresponding fPOA and bbPOA concentrations (Figures 5, 6).

511 fSOA-sv: The global and land average surface concentrations of fSOA-sv are 0.03 µg m<sup>-3</sup> and 512 0.04 µg m<sup>-3</sup>, respectively (Table 7; Figure 6a). fSOA-sv disperses over the continents but remains 513 at low levels in most cases (lower than 0.2 µg m<sup>-3</sup>). The highest concentrations of fSOA-sv are 514 modeled over India and Bangladesh (0.3-0.7 µg m<sup>-3</sup>), and in the greater Beijing area and Nigeria 515 (0.1-0.5 µg m<sup>-3</sup>).

bbSOA-sv: The global and land average surface concentrations of bbSOA-sv are 0.05 µg m<sup>-3</sup> 516 and 0.07 µg m<sup>-3</sup>, respectively (Table 7; Figure 6b). bbSOA-sv has higher concentrations in the 517 Southern Hemisphere where the main rainforests of the world are located (Amazon, Congo 518 519 Basin, part of SE Asia) and the meteorological conditions favor the partition of SVOC emissions into the gas phase where they are subject to oxidation. The highest bbSOA-sv concentrations (up 520 to 1 µg m<sup>-3</sup>) are calculated over the west coast of Central Africa, affected by the biomass burning 521 emissions from the Congo Basin rainforest. These emissions are the source of the relatively high 522 bbSOA-sv concentrations (0.3-0.7  $\mu$ g m<sup>-3</sup>) that are calculated over the Southern Atlantic Ocean. 523 The concentration of bbSOA-sv over the Amazon basin is up to 0.7  $\mu$ g m<sup>-3</sup>. The atmosphere over 524 the Pacific and Indian Oceans is also affected by biomass burning emissions from the Amazon 525 rainforest and the rainforests of southeastern Asia, respectively (bbSOA-sv concentration is 0.1-526  $0.3 \text{ }\mu\text{g m}^{-3}$ ). Over the boreal forests, bbSOA-sv reaches low levels (less than 0.1  $\mu\text{g m}^{-3}$ ) as most 527 528 of the biomass burning SVOCs are emitted directly into the aerosol phase forming bbPOA.

529

# 530 4.5 Secondary Organic Aerosol from IVOCs

IVOCs are emitted in the gas phase where they react with OH, becoming less volatile and 531 532 more chemically processed, and condense to the aerosol phase to produce secondary organic aerosol (SOA-iv). SOA-iv is subdivided into fSOA-iv and bbSOA-iv depending on the source as 533 534 described in section 3.5. In the current application, IVOC emissions are assumed to be 4 times higher than of SVOC (Table 4). Therefore, IVOCs are a significant source of SOA and the SOA-535 536 iv concentration by far exceeds that of SOA-sv and POA (Figures 5-7). Furthermore, the effective long-range transport of IVOCs results in the formation of significant SOA-iv amounts 537 remote from the sources (Figure 7). This is consistent with several field campaigns over 538 megacities which have shown that the concentration of primary combustion organic particles 539

decreases with distance from the urban source areas due to evaporation and deposition,
remaining at low levels in surrounding areas, while secondary oxygenated and photo-chemically
processed organics remain in high concentrations in suburban and rural sites (Aiken et al., 2009;
Hildebrandt et al., 2010; Morgan et al., 2010).

fSOA-iv: The global and land average surface concentrations of fSOA-iv are 0.22 µg m<sup>-3</sup> and 544 0.35 µg m<sup>-3</sup>, respectively (Table 7; Figure 7a). This is two times higher than the corresponding 545 average surface concentration of fPOA, even though the peak concentrations of fSOA-iv and 546 fPOA are similar (7  $\mu$ g m<sup>-3</sup> over Bangladesh and 8  $\mu$ g m<sup>-3</sup> over the greater Beijing area, 547 respectively). This difference indicates that fSOA-iv is more regionally distributed compared to 548 fPOA, due to the effects of long-range transport of IVOC emissions on fSOA-iv formation. The 549 model calculates a continental background of around 0.5 µg m<sup>-3</sup> for fSOA-iv. The highest fSOA-550 iv concentrations are modeled over India and Bangladesh (3-7 µg m<sup>-3</sup>), eastern China (1-4 µg m<sup>-1</sup> 551 <sup>3</sup>) and western Africa (1-4  $\mu$ g m<sup>-3</sup>). The Balkan Peninsula and eastern Mediterranean are strongly 552 influenced by long-range transport of IVOC emissions from Eastern Europe (the fSOA-iv 553 concentration is 0.5-1 µg m<sup>-3</sup>). fSOA-iv concentrations are also significant over the Arabian 554 Peninsula (around 1 µg m<sup>-3</sup>), eastern and South Africa (0.5-1 µg m<sup>-3</sup>). Over North America 555 fSOA-iv concentrations are relatively high over a large area covering the eastern USA, the 556 Californian Peninsula and North Mexico (0.5-1 µg m<sup>-3</sup>), while over South America high fSOA-iv 557 concentrations occur mostly over Rio de Janeiro (0.5-1 µg m<sup>-3</sup>). Long-range transport is also 558 559 important for the modeled fSOA-iv concentrations over the oceans. This is mostly evident over the Arabian Sea where the fSOA-iv concentration is 1-2  $\mu$ g m<sup>-3</sup>, due mainly to long-range 560 transport of fSOA-iv from India. The Atlantic Ocean atmosphere is influenced by IVOC 561 emissions from western Africa and the fSOA-iv concentration over this region is estimated at 562 0.5-1 µg m<sup>-3</sup>. Over the Pacific Ocean fSOA-iv is sensitive to long-range transport from the Los 563 Angeles and Mexico City areas and from Beijing and other big cities in eastern China, resulting 564 in significant concentrations off the west coast of California (up to 0.5-1 µg m<sup>-3</sup>) and the Pacific 565 Rim (up to  $0.5-1.5 \text{ µg m}^{-3}$ ). 566

bbSOA-iv: The modeled global and land average surface concentrations of bbSOA-iv are 0.42
μg m<sup>-3</sup> and 0.74 μg m<sup>-3</sup>, respectively (Table 7; Figure 7b). Similar to fSOA-iv, bbSOA-iv is two
times higher than bbPOA and disperses over a wide area covering most of South America,
Central and South Africa, Eastern Russia, Southeastern Asia and Indonesia due to emissions

571 from the major rainforests and savannas in the Southern Hemisphere, the tropics and the boreal forests in the Northern Hemisphere. The highest bbSOA-iv concentrations are modeled over the 572 Amazon and Congo Basin rainforests (5-10  $\mu$ g m<sup>-3</sup>). bbSOA-iv concentrations are also high over 573 southeastern Asia (up to 1-5 µg m<sup>-3</sup>) and the boreal forests of Alaska, Canada, and Russia (0.5-574 1.5  $\mu$ g m<sup>-3</sup>). The subtropical west coasts of Africa, South America and Indonesia are strongly 575 influenced by long-range transport of bbSOA-iv. The strongest transport effect is expected over 576 577 the Atlantic Ocean, where biomass burning IVOC emissions can travel thousands of kilometers from the sources (e.g., the Congo Basin rainforest), resulting in significant bbSOA-iv 578 concentrations (2-6  $\mu$ g m<sup>-3</sup>). 579

580

# 581 **4.6 Secondary Organic Aerosol from VOCs**

SOA-v is formed from the oxidation of biogenic and anthropogenic VOCs. Global VOC
emissions are dominated by biogenic compounds from vegetated areas (Kanakidou et al., 2005).
However, anthropogenic VOCs, emitted in urban areas, can also contribute significantly to SOA
formation, especially considering the aging reactions of aSOA-v.

*bSOA-v*: The global and land average surface concentrations of bSOA-v are 0.28  $\mu$ g m<sup>-3</sup> (Figure 8a). EMAC calculates highest bSOA-v concentrations over the Amazon Basin rainforest (5-10  $\mu$ g m<sup>-3</sup>) mostly due to the oxidation of isoprene. Over the Congo Basin rainforest, where isoprene emissions are similar to monoterpene emissions, the bSOA-v average surface concentration is 3-6  $\mu$ g m<sup>-3</sup>. bSOA-v concentrations are also relatively high over Southeast Asia (2-4  $\mu$ g m<sup>-3</sup>), Southeast USA and Australia (2-5  $\mu$ g m<sup>-3</sup>), and Europe (0.5-1.5  $\mu$ g m<sup>-3</sup>).

aSOA-v: The global and land average surface concentrations of aSOA-v are 0.23  $\mu$ g m<sup>-3</sup> 592 (Figure 8b). Photochemical aging adds significantly to aSOA-v since only 10% of the modeled 593 594 aSOA-v is formed from the first photooxidation step of aSOA-v. The remainder 90% is formed through the continued oxidation of aSOA-v. The aggressive aging of aSOA-v results in a wider 595 596 distribution of aSOA-v than bSOA-v and in a significant contribution of aSOA-v to the total SOA-v (45%). The aSOA-v is modeled to have a continental background of around 0.5  $\mu$ g m<sup>-3</sup> 597 598 and relatively high concentrations in the vicinity of urban areas. The highest concentrations occur over India, Bangladesh and the Persian Gulf (2-3 µg m<sup>-3</sup>). Over India, aromatics dominate 599 600 the aVOC emissions (57%), followed by alkanes (29%) and olefines (14%). Over the Persian Gulf region alkenes are the dominant aVOC (65%), followed by aromatics (33%) and olefines 601

- 602 (2%). However, these numbers do not proportionally reflect the VOC contributions to the aSOA-603 v formation since different types of aVOC have different aerosol yields. Over Europe, the 604 highest aSOA-v concentrations occur over the Mediterranean Basin (0.5-1  $\mu$ g m<sup>-3</sup>) due to the 605 long-range transport of
- aSOA-v from Central European emissions. Over the USA, aSOA-v concentrations are relatively
- high over California and the eastern USA (around  $1 \mu \text{g m}^{-3}$ ).
- 608

# 609 4.7 Chemical composition of OA

Table 7 summarizes the modeled global average chemical composition of OA at the surface. 610 Most of OA is SOA (82%) formed from the oxidation of organic compounds in the gas phase. 611 SOA consists of 52% SOA-iv, 42% SOA-v, and 6% SOA-sv. The important contribution of 612 SOA-iv to the overall SOA mass is attributed to the fact that strong fuel combustion related and 613 biomass burning IVOC emissions react with atmospheric oxidants to form low volatility 614 products that condense into the particle phase as SOA-iv. This result further emphasizes the 615 importance of the oxidation of IVOCs as an additional source of OA and implies that global and 616 regional models that do not account for IVOCs could underestimate OA formation by as much as 617 40%. In addition, the model results highlight the importance of anthropogenic emissions to 618 global OA formation. In particular it calculates that one third of OA originates from 619 anthropogenic sources. Even if this fraction may be overestimated due to the aggressive aging of 620 IVOCs and aVOCs applied in this study, it corroborates the findings of recent studies that 621 highlight the potentially large impact of anthropogenic OA, especially SOA, on the global 622 aerosol load (Spracklen et al., 2011; Carslaw et al., 2013; Lee et al., 2013). 623

624 At higher altitudes the production of SOA is enhanced since organic gases can be efficiently transported vertically and oxidized forming lower volatility SOA (Figure 9). In addition, the 625 produced SOA at higher altitudes has a relatively long lifetime since it is less subjected to wet 626 and dry deposition. This results in a higher fraction of SOA in total OA in the free troposphere 627 628 than at the surface (92% compared to 82% at the surface). Furthermore, in contrast to the surface OA composition the fraction of aSOA-v in the free troposphere is calculated to be higher than 629 630 bSOA-v since the latter is not allowed to participate in additional photochemical reactions (Table 7). The modeled tropospheric burden of POA is 0.23 Tg, the sum of SOA-sv and SOA-iv is 1.57 631

Tg, and SOA-v is 1.2 Tg. Pye and Seinfeld (2010) estimated 0.03 Tg of POA, 0.90 Tg of SOAsv and SOA-iv, and 0.71 Tg of SOA-v, while Jathar et al. (2011) found 0.09 Tg of POA, 1.25 Tg of SOA-sv and SOA-iv, and 1.02 Tg of SOA-v. Farina et al. (2010) and Jo et al. (2013) did not account for SOA-sv and SOA-iv and calculated 1.1 Tg and 0.8 Tg of POA, respectively, and 0.98 Tg and 1.16 Tg of SOA-v, respectively.

637

# 638 5. Conclusions

A new module describing the organic composition of aerosols and the evolution in the 639 atmosphere has been developed and implemented into the global EMAC model. ORACLE treats 640 fossil fuel, biofuel and biomass burning related aerosol emissions as semi-volatile and 641 chemically reactive, and also accounts for the emissions and oxidation of IVOCs as an 642 additionally important source of OA. Furthermore, it considers the formation of SOA from the 643 oxidation of anthropogenic and biogenic VOC precursors. The model employs the volatility 644 basis set theory to simulate the emissions, chemical reactions and phase partitioning of all OA 645 components. 646

647 The comparison of the model results with measured ambient OA mass concentrations illustrates that by treating fuel combustion and biomass burning aerosol emissions as semi-648 649 volatile and reactive, and accounting for emissions and oxidation of IVOCs and traditional VOCs, the model realistically represents the sources, chemistry and properties of OA, and 650 651 reproduces the measured concentrations of OA over urban and forested areas. In contrast to previous CCMs that treat POA as non-volatile and non-reactive, the ORACLE module calculates 652 653 that most of the OA is SOA (82%). Furthermore, approximately one third of OA is estimated to originate from anthropogenic sources. Such high anthropogenic OA fractions challenge the 654 655 results of the traditional CCMs and add to the most recent findings that suggest a strong contribution of anthropogenic sources to global OA concentrations. 656

657 On model calculated OA at the surface consists of 18% POA. average, 5% SOA-sv, 43% SOA-iv, and 34% SOA-v. The tropospheric burdens of OA components are 658 modeled to be 0.23 Tg POA, 0.16 Tg SOA-sv, 1.41 Tg SOA-iv and 1.2 Tg SOA-v. POA 659 660 concentrations are particularly high close to biomass burning regions (forest and savannah fires) and fuel combustion dominated sources (urban and industrial regions), though decrease rapidly 661 662 during atmospheric transport. SOA-sy concentrations are four to five times lower than POA

663 concentrations since most of the SVOCs are emitted directly in the aerosol phase. IVOCs, on the 664 other hand, are expected to be important SOA precursors, resulting in high SOA-iv 665 concentrations that are widely dispersed over the continents and the tropical oceans, being 666 subject to the long-range transport of IVOC emissions. SOA-v accounts for 34% of the total 667 SOA concentration and originates from biogenic sources (55%) and anthropogenic sources 668 (45%). In the current application, only anthropogenic SOA-v has been allowed to participate in 669 aging reactions, which considerably increases the modeled aSOA-v concentration.

The ORACLE module subdivides OA into several compounds allowing the quantification of 670 primary versus secondary as well as biogenic versus anthropogenic contributions to OA 671 concentrations. Such fundamental information can shed light on long-term changes in OA 672 abundance, and hence project the effects of OA on future air quality and climate. ORACLE is 673 well suited for studying the climatic impact of OA as it captures the dynamic aging of OA and its 674 atmospheric evolution by becoming increasingly oxidized, less volatile, and more hygroscopic. 675 Future applications will also include the effects on cloud condensation nuclei and indirect 676 aerosol effects on climate. Furthermore, the detailed composition of OA, computed by 677 678 ORACLE, can be used to directly compare model predictions with the latest aerosol mass spectrometer (AMS) factor analysis data, which is planned as a next application. 679

Overall, ORACLE is a flexible module that efficiently describes the organic aerosol composition and chemical evolution in the atmosphere. Sensitivity studies of the influence of several parameters (such as reaction constants, aerosol yields, SVOC and IVOC emission fractions) on the model calculations will be conducted in a planned future study and used, along with AMS factor analysis data taken around the world, to optimize the model configuration to further improve the estimates of OA concentrations and investigate their role in air quality and climate.

687

### 688 **Code Availability**

The complete ORACLE code can be obtained upon request by emailing the first author. To use ORACLE as part of EMAC, please first apply for an ECHAM5 and a MESSy license. The GCM ECHAM5 has been developed at the Max Planck Institute for Meteorology in Hamburg (see: <u>http://www.mpimet.mpg.de/en/wissenschaft/modelle/echam/echam5.html</u>). The Modular Earth Submodel System (MESSy) is continuously further developed and applied by a consortium 24 of institutions, initiated by the Max Planck Institute for Chemistry. The usage of MESSy and access to the source code is licensed to all affiliates of institutions that are members of the MESSy Consortium. Institutions can be a member of the MESSy Consortium by signing the MESSy Memorandum of Understanding. More information can be found on the MESSy Consortium Website (http://www.messy-interface.org).

699

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1020	Table 1: Organic component name classifications in the ORACLE module
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Term	Description
Source root name	
Modifiers	
А	Mass from anthropogenic sources (i.e. aSOA)
В	Mass from biogenic sources (i.e. bSOA)
F	Mass from fossil and biofuel combustion (i.e. fPOA)
Bb	Mass from biomass burning processes (i.e. bbPOA)
Base Terms	
POA	Primary organic aerosol. This is emitted in the particle phase and has not
POG	Primary organic gas that has not undergone chemical reaction
SOA	Secondary organic aerosol formed from the oxidation of gas-phase organic
SOG	species Secondary organic gas. The gas-phase mass produced by at least one chemical
	reaction in the atmosphere
Initial Volatility	
Suffix	
-SV	Product of the oxidation of SVOCs
-1	Product of the oxidation of IVOCs
-v	Product of the oxidation of vOCs
-V	Product of the oxidation of VOCs

# **Table 2:** Description of ORACLE organic compounds in gas and aerosol phases.

1048	Gas-phase compounds	Particle phase compounds	Description
1049	fPOG	fPOA	Primary organic compounds from fuel combustion
1050	bbPOG	bbPOA	Primary organic compounds from biomass burning
1050	fSOG-sv	fSOA-sv	Secondary organic compounds from the oxidation
1051			of fuel combustion SVOCs
	bbSOG-sv	bbSOA-sv	Secondary organic compounds from the oxidation
1052			of biomass burning SVOCs
1053	fSOG-iv	fSOA-iv	Secondary organic compounds from the oxidation
1000			of fuel combustion IVOCs
	bbSOG-iv	bbSOA-iv	Secondary organic compounds from the oxidation
			of biomass burning IVOCs
	aSOG-v	aSOA-v	Secondary organic compounds from the oxidation
			of anthropogenic VOCs
	bSOG-v	bSOA-v	Secondary organic compounds from the oxidation
			of biogenic VOCs

**Table 3.** Global annual emissions of the lumped anthropogenic VOC species introduced in theORACLE module.

VOC lumped species	RCP4.5 species	Emission Rates (Tg yr <sup>-1</sup> )
ARO1	Benzene, Toluene	18,3
ARO2	Trimetyl Benzenes, Xylene, other Aromatics	13.9
ALK4	Pentanes	15.1
ALK5	Hexanes, other higher Alkenes	21.2
OLE1	Propene	7.4
OLE2	Other Alkenes	8.2

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**Table 4.** Emission factors of the primary organic compounds from fuel combustion and biomass
burning sources for each volatility bin used in ORACLE. The total global annual emission rates
are also shown. The SVOC/IVOC emission rates are estimated by multiplying the emission
factors of each bin with the non-volatile emissions rates.

	Volatility D	istribution	Non Volatile POA emission rates	SVOC emission rates	IVOC emission rates		
					$Tg y^{-1}$	$Tg y^{-1}$	$Tg y^{-1}$
C <sup>*</sup> (µg m <sup>-3</sup> ) at 298 K	10-1	10 <sup>1</sup>	$10^{3}$	10 <sup>5</sup>			
Fuel combustion OA	fPOA1	fPOA2	fPOG3	fPOG4	12.3	6.2	24.6
Biomass burning OA	bPOA1	bbPOA2	bbPOG3	bbPOG4	24.8	12.4	49.6
Emission factors	0.18	0.32	0.5	1.5			
.069							
.070							
.071							

**Table 5.** Secondary organic aerosol mass yield<sup>1</sup> parameters

	C* in µg m <sup>-3</sup> at 298K					
VOC lumped Species	1	10	10 <sup>2</sup>	10 <sup>3</sup>		
ARO1	0.003	0.165	0.300	0.435		
ARO2	0.002	0.195	0.300	0.435		
ALK4	0.000	0.038	0.000	0.000		
ALK5	0.000	0.150	0.000	0.000		
OLE1	0.001	0.005	0.038	0.150		
OLE2	0.003	0.026	0.083	0.270		
ISOP	0.009	0.030	0.015	0.000		
TERP	0.107	0.092	0.359	0.600		

1074 <sup>1</sup> The SOA yields are based on an assumed particle density of  $1.5 \text{ g cm}^{-3}$ .

**Table 6.** Statistical evaluation of monthly averaged modeled OA against IMPROVE, EMEP and

- 1082 East Asia observations during 2005-2009

Network	Number of stations	Mean Observed	Mean Modeled	MAGE	MB	NME	NMB	RMSE
		$(\mu g m^{-3})$	(µg m <sup>-3</sup> )	(µg m <sup>-3</sup> )	(µg m <sup>-3</sup> )	(%)	(%)	(µg m <sup>-3</sup> )
IMPROVE	193 <sup>a</sup>	1.93	3.02	1.71	1.09	89	57	2.6
EMEP	10 <sup>b</sup>	1.81	2.20	1.21	0.39	67	22	2.09
East Asia	$18^{\rm c}$	11.02	7.09	4.76	-3.93	43	-36	5.53

<sup>a</sup> 10202 measurements

<sup>b</sup> 301 measurements

1086 <sup>c</sup>74 measurements

**Table 7.** Global and land average surface concentration and tropospheric burden of organic
aerosol components. The fractional contribution of each component to total OA is listed in
parenthesis.

OA component	fPOA	bbPOA	fSOA-sv	bbSOA-sv	fSOA-iv	bbSOA-iv	aSOA-v	bSOA-v
Global Average (µg m <sup>-3</sup> )	0.11 (7%)	0.16 (11%)	0.03 (2%)	0.05 (3%)	0.22 (15%)	0.42 (28%)	0.23 (15%)	0.28 (19%)
Land Average (µg m <sup>-3</sup> )	0.22 (8%)	0.39 (14%)	0.04 (1%)	0.07 (2%)	0.35 (12%)	0.74 (26%)	0.39 (13%)	0.70 (24%)
Tropospheric Burden (Tg)	0.11 (4%)	0.12 (4%)	0.06 (2%)	0.1 (3%)	0.5 (17%)	0.91 (30%)	0.65 (22%)	0.55 (18%)

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- 1102

1103	Figure 1: Diagram	m showing ORA	CLE as part of t	the EMAC chemistr	y-climate model.
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1104

Figure 2: Schematic of the VBS resolution and the formation procedure of SOA from SVOC, 1105 IVOC and VOC emissions. Red indicates that the compound is in the vapor phase and blue in the 1106 1107 particulate phase. The circles correspond to primary organic material that can be emitted either in the gas or in the aerosol phase. The triangles indicate the formation of SOA from SVOCs by fuel 1108 combustion and biomass burning sources, while the squares show SOA from IVOCs by fuel 1109 combustion and biomass burning sources, and the diamonds the formation of SOA from 1110 1111 anthropogenic and biogenic VOC sources. The partitioning processes, the aging reactions and the names of the species used to track all compounds are also shown. 1112

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\*For this application it is assumed that SOA formed by biogenic VOCs does not participate in aging reactions (Lane
et al., 2008; Tsimpidi et al., 2010).

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**Figure 3:** Modeled average surface concentration of total OA ( $\mu$ g m<sup>-3</sup>) during the years 2005-2009.

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**Figure 4:** Scatter plot comparing model predictions of total OA concentration (in  $\mu$ g m<sup>-3</sup>) with observations from the (**a**) IMPROVE network in the United States, (**b**) EMEP network in Europe and East Asian sites from 2005 to 2009. Each point represents a monthly average value. Also shown the 1:1, 2:1 and 1:2 lines.

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**Figure 5:** Modeled average surface concentrations (in  $\mu$ g m<sup>-3</sup>) of (**a**) POA from fuel combustion (fPOA) and (**b**) POA from biomass burning emissions (bbPOA) during the years 2005-2009.

- **Figure 6:** Modeled average surface concentrations (in  $\mu g m^{-3}$ ) of (a) SOA-sv from the oxidation
- of SVOCs from fuel combustion (fSOA-sv) and (b) SOA-sv from the oxidation SVOCs from
- biomass burning emissions (bbSOA-sv) during the years 2005-2009.
- 1131
- **Figure 7:** Modeled average surface concentrations (in  $\mu g m^{-3}$ ) of (a) SOA-iv from the oxidation
- 1133 of IVOCs from fuel combustion (fSOA-iv) and (b) SOA-iv from the oxidation of IVOCs from
- biomass burning emissions (bbSOA-iv) during the years 2005-2009.
- 1135
- **Figure 8:** Modeled average surface concentrations (in  $\mu$ g m<sup>-3</sup>) of (a) SOA-v from the oxidation
- 1137 of biogenic VOC emissions (bSOA-v) and (b) SOA-v from the oxidation of anthropogenic VOC
- emissions (aSOA-v) during the years 2005-2009.
- 1139
- 1140 Figure 9: Modeled average zonal concentrations (in  $\mu g m^{-3}$ ) of (a) POA (sum of fPOA and
- bbPOA) and (b) SOA (sum of fSOA-sv, bbSOA-sv, fSOA-iv, bbSOA-iv, aSOA-v, and bSOA-v)
  during the years 2005-2009.
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Figure 2: Schematic of the VBS resolution and the formation procedure of SOA from SVOC, 1180 IVOC and VOC emissions implemented in ORACLE. Red indicates that the organic compound 1181 is in the vapor phase and blue in the particulate phase. The circles correspond to primary organic 1182 material that can be emitted either in the gas or in the aerosol phase. The triangles indicate the 1183 formation of SOA from SVOCs by fuel combustion and biomass burning sources, while the 1184 squares show SOA from IVOCs by fuel combustion and biomass burning sources, and the 1185 1186 diamonds the formation of SOA from anthropogenic and biogenic VOC sources. The 1187 partitioning processes, the aging reactions of the organic compounds, and the names of the species used to track all compounds are also shown. 1188

\*For this application it is assumed that SOA formed by biogenic VOCs does not participate in aging reactions (Lane et al., 2008; Tsimpidi et al., 2010).



**Figure 3:** Modeled average surface concentration of total OA ( $\mu$ g m<sup>-3</sup>) during the years 2005-2009.



**Figure 4:** Scatter plot comparing model predictions of total OA concentration (in  $\mu$ g m<sup>-3</sup>) with observations from the (**a**) IMPROVE network in the United States, (**b**) EMEP network in Europe and East Asian sites from 2005 to 2009. Each point represents a monthly average value. Also shown the 1:1, 2:1 and 1:2 lines.



**Figure 5:** Modeled average surface concentrations (in  $\mu g m^{-3}$ ) of (a) POA from fuel combustion

1230 (fPOA) and (b) POA from biomass burning emissions (bbPOA) during the years 2005-2009.



**Figure 6:** Modeled average surface concentrations (in μg m<sup>-3</sup>) of (**a**) SOA-sv from the oxidation

1242 of SVOCs from fuel combustion (fSOA-sv) and (b) SOA-sv from the oxidation SVOCs from 1243 biomass burning emissions (bbSOA-sv) during the years 2005-2009.



Figure 7: Modeled average surface concentrations (in  $\mu g m^{-3}$ ) of (a) SOA-iv from the oxidation 

of IVOCs from fuel combustion (fSOA-iv) and (b) SOA-iv from the oxidation of IVOCs from biomass burning emissions (bbSOA-iv) during the years 2005-2009.



**Figure 8:** Modeled average surface concentrations (in  $\mu$ g m<sup>-3</sup>) of (a) SOA-v from the oxidation

of biogenic VOC emissions (bSOA-v) and (b) SOA-v from the oxidation of anthropogenic VOC
emissions (aSOA-v) during the years 2005-2009.





**Figure 9:** Modeled average zonal concentrations (in  $\mu g m^{-3}$ ) of (a) POA (sum of fPOA and bbPOA) and (b) SOA (sum of fSOA-sv, bbSOA-sv, fSOA-iv, bbSOA-iv, asOA-v, and bSOA-v)

1283 during the years 2005-2009.

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