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2	Modelling the role of fires in the terrestrial carbon balance by incorporating SPITFIRE into the
3	global vegetation model ORCHIDEE: Part 1. Simulating historical global burned area and fire
4	regimes
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3 Abstract

4 Fire is an important global ecological process that influences the distribution of biomes, with 5 consequences for carbon, water, and energy budgets. Therefore it is impossible to appropriately 6 model the history and future of the terrestrial ecosystems and the climate system without including 7 fire. This study incorporates the process-based prognostic fire module SPITFIRE into the global 8 vegetation model ORCHIDEE, which was then used to simulate burned area over the 20th century. 9 Special attention was paid to the evaluation of other fire regime indicators such as seasonality, fire 10 size and fire length, next to burned area. For 2001-2006, the simulated global spatial extent of fire 11 agrees well with that given by satellite-derived burned area datasets (L3JRC, GLOBCARBON, 12 GFED3.1), and 76–92% of the global burned area is simulated as collocated between the model 13 and observation, depending on which dataset is used for comparison. The simulated global mean annual burned area is 346 Mha yr^{-1} , which falls within the range of 287–384 Mha yr^{-1} as given by 14 the three observation datasets; and is close to the 344 Mha yr⁻¹ by the GFED3.1 data when crop 15 16 fires are excluded. The simulated long-term trend and variation of burned area agree best with the 17 observation data in regions where fire is mainly driven by climate variation, such as boreal Russia 18 (1930-2009), and Canada & US Alaska (1950-2009). At the global scale, the simulated decadal 19 fire variation over the 20th century is only in moderate agreement with the historical 20 reconstruction, possibly because of the uncertainties of past estimates, and because land-use 21 change fires and fire suppression are not explicitly included in the model. Over the globe, the size 22 of large fires (the 95th quantile fire size) is underestimated by the model for the regions of high 23 fire frequency, compared with fire patch data as reconstructed from MODIS 500-m burned area 24 data. Two case studies of fire size distribution in Canada & US Alaska, and southern Africa 25 indicate that both number and size of large fires are underestimated, which could be related with 26 short fire patch length and low daily fire size. Future efforts should be directed towards building 27 consistent spatial observation datasets for key parameters of the model in order to constrain the 28 model error at each key step of the fire modelling.

1 1 Introduction

2 Fire is an important process in the Earth system, that existed long before the large-scale 3 appropriation of natural ecosystems by humans (Bowman et al., 2009; Daniau et al., 2013). Fires 4 have multiple biophysical and ecological consequences, and they are also an important source of 5 atmospheric trace gases and aerosol particles (Langmann et al., 2009; van der Werf et al., 2010). By 6 damaging some plant types and concurrently promoting others, fires play an important role in 7 shaping vegetation structure and function (Bond et al., 2005; Pausas and Keeley, 2009). Fire 8 changes the surface albedo, aerodynamic roughness, and the sensible and latent heat fluxes (Liu et 9 al., 2005; Liu and Randerson, 2008). These fire-induced ecosystem changes could further influence 10 the surface energy budget and boundary-layer climate (Beck et al., 2011; Randerson et al., 2006; 11 Rogers et al., 2013). In addition, gas and aerosol species emitted to the atmosphere from biomass 12 burning modify atmospheric composition and the radiative forcing balance (Tosca et al., 2013; 13 Ward et al., 2012). Fire aerosols also degrade air quality and cause increased health risk (Marlier et 14 al., 2013). Thus fire process and biomass burning emissions need to be included in the Earth 15 system models, which are often used to investigate the role of fire in past, present and future 16 biophysical and biogeochemical processes.

17 The type of fire model embedded in global vegetation models has evolved from simple fire 18 hazard models (Thonicke et al., 2001) to the current state-of-the-art process-based fire models 19 (Andela et al., 2013; Kloster et al., 2010; Lasslop et al., 2014; Li et al., 2013; Pfeiffer et al., 2013; 20 Prentice et al., 2011) and empirical models with optimization by observation (Knorr et al., 2014; Le 21 Page et al., 2014). The majority of fire models explicitly simulate ignitions from natural and 22 human sources, fire propagation, fuel combustion and vegetation mortality, ideally at daily or even 23 finer time steps. Recently, Pfeiffer et al. (2013) have incorporated multi-day burning, 24 "coalescence" of fires and interannual lightning variability in the LPJ-LMfire model. Li et al. 25 (2013) incorporated social and economic factors in the human ignition scheme. Lasslop et al. 26 (2014) investigated the model sensitivity to the climate forcing and a number of spatial parameters. 27 Evaluation of fire models in these studies at the global scale has mainly focused on models' 28 capability in broadly reproducing the large-scale distribution of fire activity during the past decade

as revealed by satellite observations. Less attention was paid to the simulation of long-term
 historical fire trend and variation, and fire regimes, which may include the number, size and
 intensity of fires – essential variables governing fire-climate-vegetation feedbacks (Archibald et al.,
 2013; Barrett et al., 2011; Hoffmann et al., 2012).

5 It is well known that across all fire-prone ecosystems, the magnitude and trend of burned area 6 depend strongly on large fire events that represent only a low fraction in total number of fires 7 (Kasischke and Turetsky, 2006; Keeley et al., 1999; Stocks et al., 2002). These large fires have 8 profound impacts on landscape heterogeneity (Schoennagel et al., 2008; Turner et al., 1994), 9 biological diversity (Burton et al., 2008) and may also induce a higher rate of carbon emissions 10 compared with small fires (Kasischke and Hoy, 2012). Besides, the social and economic 11 consequences of extreme large fires are more severe (Richardson et al., 2012). In some ecosystems, 12 past climate warming was shown to have increased the occurrence of large fires (Kasischke et al., 13 2010; Westerling et al., 2006), and fire regimes were projected to even further deviate from 14 historical range given the future climate change (Pueyo, 2007; Westerling et al., 2011). Given the 15 importance of these large fires, it is essential that we should evaluate the ability of fire models to 16 simulate their occurrence.

17 In this study, we have incorporated the SPITFIRE fire model (Thonicke et al., 2010) into the 18 global land surface model ORCHIDEE (Krinner et al., 2005). This allowed us to simulate global 19 fire activity during the 20th century, and to perform an in-depth model evaluation. In present study, 20 we focus on evaluating the ORCHIDEE-SPITFIRE model performance in simulating fire 21 behaviours and regimes, including ignitions, fire spread rate, fire patch length, fire size 22 distribution, fire season and burned area. Quantification of fire carbon emission as a component of 23 the terrestrial carbon balance will be presented in a companion paper (Part 2). Specifically, the 24 objectives of the present study are: (1) to evaluate simulated burned area using multiple datasets 25 including that from satellite observation, government fire agency, and historical reconstruction 26 over the 20th century (Sect. 3.1, 3.2, 3.3, 3.4, 3.5); (2) to compare simulated fire size distribution 27 with observations; in order to investigate especially the model's ability to simulate large fire 28 occurrence (Sect. 3.6); and (3) to examine potential sources of model error in order to identify

4

1 future research need and potential model improvements (Sect. 4.2).

2 2 Data and methods

3 2.1 Model description

The processes and equations of the fire model SPITFIRE, as described by Thonicke et al. (2010), were implemented in the vegetation model ORCHIDEE (Krinner et al., 2005). The SPITFIRE model operates on a daily time step, consistent with the STOMATE sub-module in ORCHIDEE, which simulates vegetation carbon cycle processes (photosynthates allocation, litterfall, litter and soil carbon decomposition). The major processes within SPITFIRE are briefly described below with applicable minor modifications (see Thonicke et al., 2010 for more detailed description).

11 1. *Daily Potential ignition* includes ignitions from lightning and human activity. Remotely 12 sensed lightning flashes (Cecil et al., 2012) were obtained from the High Resolution Monthly 13 Climatology of lightning flashes by the Lightning Imaging Sensor-Optical Transient Detector 14 (LIS/OTD) (http://gcmd.nasa.gov/records/GCMD lohrmc.html). The LIS/OTD dataset provides 15 annual mean flashes over 1995–2000 on a 0.5° grid at monthly time step. This single annual data 16 was repeated each year throughout the simulation. Following Prentice et al. (2011), the proportion 17 of lightning flashes that reach the ground with sufficient energy to ignite a fire is set as 0.03. This 18 value differs from that in the original SPITFIRE model as implemented in LPJ-DGVM (Thonicke 19 et al., 2010); there a cloud-to-ground (CG) flashing ratio of 0.2 was used, followed by a further 20 ignition efficiency of 0.04.

To estimate potential ignitions by humans, the original equations (3) and (4) in Thonicke et al.
(2010) were modified for the purpose of unit adjustment, as below:

23
$$I_{\rm H} = PD \times 30.0 \times e^{-0.5 \times \sqrt{PD}} \times a(ND)/10000$$
 Eq. (1)

where I_H is the daily ignition number (1 day⁻¹ km⁻²), PD is population density (individuals km⁻²). The parameter a(ND) (ignitions individual⁻¹ day⁻¹) represents the propensity of people to produce ignition events; and is a spatially explicit parameter as in Thonicke et al. (2010) (Supplement Fig. S1). 2. Daily Fire numbers are derived by scaling the potential ignitions (which include human
 and lightning ignitions) with the fire danger index (FDI), which is derived by comparing simulated
 daily fuel moisture to a Plant Functional Type (PFT) dependent moisture of extinction. All fires
 with a fireline intensity less than 50 kW m⁻¹ are assumed unable to propagate and are suppressed
 as stated by Thonicke et al. (2010).

6 3. Daily mean fire size is calculated by assuming an elliptical shape of fire, with the major 7 axis length being the product of fire spread rate and daily fire active burning time (i.e., the time 8 that fires actively burn during that day). Fire spread rate is obtained using the Rothermel equation 9 (Rothermel, 1972; Wilson, 1982). Fire active burning time is modelled to increase as a logistic 10 function of daily fire danger index, with a maximum of 241 minutes (4 hours). Fire frontline 11 intensity is calculated following Byram (1959), as a product of fuel heat content, fuel consumption, 12 and fire spread rate. Note that over a grid cell of 0.5 degree, the model simulates a set of 13 homogeneous fires with all their characters (including fire size) being identical among each other, 14 i.e., the model represents the temporal but not the spatial heterogeneity over a given grid cell. 15 Daily grid cell burned area is calculated as the product of fire number and mean fire size.

16 4. *Fire-induced tree mortality* is determined from the combined fire damage of tree crown 17 and cambium. Crown damage depends on the fraction of tree crown that is affected by crown 18 scorch, which further depends on tree crown length, tree height and fire flame height. Fire flame 19 height is derived from surface fire intensity. Cambial damage depends on fire residence time and a 20 prescribed PFT-dependent critical time. Note that SPITFIRE simulates only crown scorch, but not 21 active crown fires that could propagate through crown fire spread.

5 *Fire carbon emissions* include emissions from surface fuel and crown combustion. Surface fuels are divided into four classes (1-h, 10-h, 100-h, 1000-h), whose designation in terms of hours describes the order of magnitude of time required to lose (or gain) 63% of its difference with the equilibrium moisture content under defined atmospheric conditions (Thonicke et al., 2010). Fuel combustion completeness is simulated as a function of daily fuel moisture, with a smaller fraction of fuel being consumed at higher fuel moisture. Crown fuel consumption is related to the fraction of crown that is scorched by fire flame. The values of all PFT-dependent parameters follow Table

2 2.2 Further modifications made to the SPITFIRE equations

3 Fires in dry climate regions are limited by the availability of fuel on the ground (Krawchuk 4 and Moritz, 2010; Prentice et al., 2011; van der Werf et al., 2008b). This constraint is implicitly 5 included in SPITFIRE equations because fire occurrence is limited by a fireline intensity of 50 kW m⁻¹. However during model testing, we found that this threshold is not enough to limit fires in 6 7 low-productivity regions (with modelled annual Net Primary Productivity or NPP of 0 - 400 g C m^{-2} yr⁻¹, corresponding to an annual precipitation of 0 – 400 mm); and too much burned area was 8 9 simulated for arid and semi-arid regions (see Supplement Fig. S2). Following Arora and Boer 10 (2005), we therefore introduced a new factor that limits the ignition efficiency, depending on the 11 availability of ground fuel. Ignition efficiency varies linearly between zero when ground fuel is lower than 200 g C m⁻², to unity when ground fuel is above 1000 g C m⁻². Here, ground fuel 12 includes aboveground litter and live biomass for grassland PFTs and aboveground litter only for 13 14 tree PFTs.

15 The equations for surface fuel combustion completeness given by Thonicke et al. (2010) 16 follow Peterson and Ryan (1986), which allow combustion completeness to decrease with 17 increasing fuel wetness and level out when fuel wetness drops below a threshold (Fig. 1). During 18 model testing, we found that because fuel wetness frequently approaches zero, simulated fuel 19 combustion completeness is much higher than field experiment values reported by van Leeuwen et 20 al. (2014). We therefore modified the maximum combustion completeness for fuel classes of 21 100-hr and 1000-hr to be the same as mean combustion completeness by van Leeuwen et al. (2014) 22 depending on different biomes (PFTs). This biome-dependent maximum combustion completeness 23 is 0.48 for tropical broadleaf evergreen and seasonal forests, 0.45 for temperate forest, 0.41 for 24 boreal forest, 0.85 for grassland, and 0.35 for cropland. These values are based on a preliminary 25 version of results by van Leeuwen et al. (2014).

1 2.3 Input dataset and the simulation protocol

2 The six-hourly climate fields used to drive the model were from the CRU-NCEP dataset 3 (http://nacp.ornl.gov/thredds/fileServer/reccapDriver/cru_ncep/analysis/readme.htm). Population 4 density for the 20th century was retrieved from the History Database of the Global Environment 5 (HYDE) compiled by the Netherlands Environmental as Assessment Agency 6 (http://themasites.pbl.nl/tridion/en/themasites/hyde/download/index-2.html). From 1850 till 2005, 7 the HYDE gridded data are available for the beginning of each decade and for 2005. Annual data 8 were linearly interpolated within each decade, and further re-gridded to 0.5° resolution. For 9 2006–2009, population density was set constant at the 2005 value.

A three-step simulation protocol was used. For the first two steps, the atmospheric CO2 was 10 11 fixed at the pre-industrial level (285 ppm) and climate forcing data of 1901–1930 were repeated in 12 loop. The first step was a without-fire spin-up from bare ground lasting for 200 years (including a 13 3000-year run of soil-only processes to speed up the equilibrium of mineral soil carbon). The 14 second step was a fire-disturbed spin-up lasting for 150 years, with fire being switched on to 15 account for fire disturbances in the pre-industrial era. Fire ignitions from human activity were 16 included in the fire-disturbed spin-up, with the human population density being fixed at 1850 level. 17 This procedure assumes that the model reached an equilibrium state under conditions of 18 pre-industrial atmospheric CO₂ and climate and fire disturbance.

19 The third step was a transient simulation from 1850 to 2009 with increasing atmospheric CO_2 , 20 climate change, and varying human population density. The climate data used for the transient 21 simulation of 1850–1900 are a repeat of 1901–1910, for the sake of stability. Before entering the 22 transient simulation, the mineral soil carbon stock was verified to vary within 0.1% (with a slight global carbon sink of 0.13 Pg C yr⁻¹ and a negligible annual trend of 0.003 Pg C yr⁻¹ during the 23 24 last 50 years of the fire-disturbed spin-up, excluding all crops, for which fires are not simulated). 25 For the current simulation, the vegetation dynamics module of ORCHIDEE was turned off, i.e., 26 the simulation used a static current-day vegetation distribution map (converted into the 13-PFT 27 ORCHIDEE IGBP map in based on the 1-km vegetation map, 28 http://webmap.ornl.gov/wcsdown/dataset.jsp?ds_id=930), and no land cover change was included.

8

1 This static land cover could affect the model-observation agreement in terms of long-term trend 2 and variation of burned area. Land cover change has double effects on burned area: fires used for 3 land cover change contribute directly to burned area; and the indirect effect depends on fire 4 frequencies of the land cover types before and after the land cover change.

5 Fires on croplands are not simulated in ORCHIDEE, even though the model has two PFTs 6 that approximately represent C3 and C4 crops (but without realistic species-specific phenology). 7 Magi et al. (2012) show that cropland fire seasons differ significantly from those of natural fires, 8 warranting a special treatment of cropland fires in global fire modelling (Li et al., 2012). Cropland 9 fires make up a rather small proportion of the total global amount in terms of both burned area (4.3% 10 according to the GFED3.1 dataset) and carbon emissions (less than 2% according to GFED3.1), 11 given that the "small" fires (Randerson et al., 2012) are not formally included and recommended in 12 the GFED dataset (http://www.globalfiredata.org/data.html).

13 Further, deforestation fires are not explicitly simulated. Evidence shows that deforestation 14 fires occur during the "time window" when climate is dry enough to allow complete burning of 15 deforested fuels (van der Werf et al., 2008a). We expect the simulated daily fire danger index (i.e., 16 an indicator of suitable climate conditions for burning) is able to partly capture this "fire climate 17 window" that's necessary for deforestation fires to happen. Thus the model is able to implicitly 18 account for some deforestation fire activity in tropical and subtropical forests, but not for all of 19 them, because the use of a static land cover map. Preliminary analysis shows that the model could capture ~67% of the deforested area by fires as given by the GFED3.1 data for closed-canopy 20 forests in the region of 20° S– 20° N for 2000–2005 (2.7 Mh yr⁻¹ vs. 4.0 Mha yr⁻¹) (Fig. S3), with the 21 22 seasonal variation being moderately represented (Fig. S4).

23 2.4 Datasets used to evaluate model performance

Several datasets were prepared and used to compare simulated burned area and fire regimeswith various observations.

1 2.4.1 Spatially gridded burned area data

2 2.4.1.1 Satellite-derived burned area data

The **GFED3.1** dataset provides monthly burned area data at 0.5° resolution for 1997–2009 with global coverage (Giglio et al., 2010). The GFED3.1 burned area was mainly generated using MODIS imagery with additional images from TRMM VIRS and ERS ATSR. The fire carbon emissions were also provided which are model simulation results by applying a modified version of the CASA model (van der Werf et al., 2010).

L3JRC dataset provides daily global burned area data at 1-km resolution for April 2000 to
March 2007; these data were generated from the 1-km SPOT VEGETATION satellite imagery
(Tansey et al., 2008). This dataset was assembled at 0.5° resolution at monthly time step for use in
the present study.

12 **GLOBCARBON** burned area data was produced from a combination of SPOT 13 VEGETATION and ERS2–ATSR2/ENVISAT AATSR data as one of the four land products of the 14 ESA GLOBCARBON initiative (Plummer et al., 2007). Global burned area data were provided at 15 monthly resolution with four different spatial resolutions (1-km/10-km/0.25°/0.5°) covering 16 1998–2007.

17 2.4.1.2 Historical burned area reconstruction for the 20th century

18 To evaluate the simulated burned area for the 20th century, historical burned area data were 19 used. These data, which cover the period 1900–2000, were compiled by Mouillot and Field (2005) at 1° resolution and monthly time step (hereafter referred to as the Mouillot data). The data were 20 21 generated by first synthesizing the burned area information from published data at national or 22 regional scale for the periods of the 1980s or 1990s, further interpolated spatially at the global 23 scale at 1° resolution using the available satellite-derived active fire distribution. Then national fire 24 statistics, historical land-use practices and other fire-relevant quantitative information (such as tree 25 ring reconstruction) were used to build the historical fire temporal trend to interpolate historical 26 burned area.

27 When comparing the burned area given by GFED3.1 and the Mouillot data for their

overlapping period of 1997–2000, the global total burned area given by Mouillot and Field (2005)
is 52% higher than GFED3.1 with significant regional discrepancies. As the satellite-derived data
is considered to be more reliable than the national or regional statistical data at a large spatial scale,
a bias correction was performed on the Mouillot data. We calculated the ratio of burned area by
the Mouillot data to GFED3.1 for 1997–2009 for each region (Fig. 2b) and also the globe. This
ratio was then applied to correct for each decade the burned area data in the Mouillot dataset.

Note that the regional breakdowns of the globe by GFED3.1, and Mouillot and Field (2005)
are different (Fig. 2). For comparison of burned area for the 20th century, the regional breakdown
by Mouillot and Field (2005) was adopted as it is based on maximum temporal stability (error
consistency), a highly important factor when comparing long-term data.

11 2.4.2 Fire patch data

12 Alaskan and Canadian fire management agencies have maintained historical fire monitoring 13 for a relatively long time (dating back to 1950s). The historical fire information for the US Alaska 14 was retrieved from the Alaska Interagency Coordination Center (AICC, 15 http://afsmaps.blm.gov/imf firehistory/imf.jsp?site=firehistory). The fire information for Canada 16 was from the Canadian Wildland Fire Information System (http://cwfis.cfs.nrcan.gc.ca/ha/nfdb). 17 These datasets contain information on fire location, fire size (burned area), fire cause, and for 18 some fires the fire report and out date. Note, in these two datasets fires with all sizes are included.

Archibald et al. (2010) classified the MCD45A1 500-m MODIS fire burned area data into individual fire patches for southern Africa (Africa south to the Equator). This fire patch information includes location and patch size (with minimum fire size of 0.25 km²). The fire patch data for Canada & US Alaska, and southern Africa are used to evaluate simulated fire size distribution.

24 2.4.3 Fire season length and the 95th quantile fire size

The global fire season length and the 95th quantile fire size data are provided by Archibald et al. (2013). The fire season length was quantified as the number of months required to reach 80% of the annual burned area using GFED3.1 data. The MCD45A1 burned area product at 500-m
resolution was used to derive the individual fires by applying a flood-fill algorithm, and the 95th
quantile fire size in each grid cell was extracted to represent the size of large fires.

4 2.5 Methods to compare model simulation with observation

5 2.5.1 Metrics used to evaluate modelled burned area against GFED3.1 data

As the GFED3.1 data is most widely used by the fire modelling community, the model results
are evaluated against GFED3.1 data for 1997–2009. Three aspects were examined: mean annual
burned area, interannual variability (IAV) and seasonality in burned area. The evaluation was done
for each GFED3.1 region (Fig. 2).

10 For the model error in terms of mean annual burned area (BA), we use the relative difference:

11
$$E_{BA} = \frac{BA_{model} - BA_{GFED}}{BA_{GFED}}$$
 Eq. (2)

12 where BA_{model} is the simulated burned area averaged over 1997–2009, and BA_{GFED} is 13 GFED3.1 mean annual burned area for the same period. The similarity in IAV (S_{interannual}) is 14 estimated by the correlation coefficient of the two linearly detrended annual burned area time 15 series by model and GFED3.1 data. Finally, the seasonality similarity (S_{season}) is given by:

16
$$S_{\text{season}} = \sum_{i=1}^{12} \min(\text{frac_model}_i, \text{frac_GFED}_i)$$
 Eq. (3)

where frac_model_i and frac_GFED_i are the fraction of burned area for the ith month relative to 17 18 the annual burned area (i.e., monthly BA normalized by the annual BA). Sseason represents the 19 overlapping area of the two normalized monthly BA series and indicates the fraction of burned 20 area in temporal coincidence. The statistical significance of S_{season} was examined by using a 21 bootstrapping method. First, normalized monthly BA from all 14 regions by the model and 22 GFED3.1 data were pooled together. Second, 100,000 pairs of monthly normalized BA were 23 randomly sampled from the pooled data in order to derive a probability distribution function (PDF) 24 of S_{season}. Third, the single-sided probability (p-value) that the calculated S_{season} is from random 25 distribution is obtained for each region, and a p-value less than 0.05 indicates the model could 26 moderately capture the seasonality of burning (i.e., significantly different from a random

1 distribution).

The peak fire month, which is defined as the month with maximum monthly burned area, is
compared between the model and GFED3.1 data. The difference between simulated and observed
peak month is quantified by the following index, after Prentice et al. (2011):

5
$$D_2 = [1 - (\sum_{j=1,n} A_j \cos \theta_j / \sum_{j=1,n} A_j)]/2$$
 Eq. (4)

6 where θ_j is the angle between vectors representing the simulated and observed peak fire month 7 (with January to December resembling one to twelve on a clock), *n* is the total number of grid 8 cells and A_j is the burned area by GFED3.1 data. According to Eq. (4), the value of D₂ is zero 9 when simulated peak month is perfectly in phase with the observation, 0.5 if the timing is off by 3 10 months in either direction, and one (the maximum) if the timing is off by 6 months.

11 2.5.2 Concatenate simulated consecutive daily fire events into multi-day fire patches

12 The fire patch data for Canada & US Alaska and southern Africa contain fires that span 13 multiple days. In the model, fires are simulated as independent daily events with the active 14 burning time being limited to 4 hours (i.e., all fires have the same size; they start and extinguish 15 within the same day). Pfeiffer et al. (2013) introduced a mechanism to allow fires to span multiple 16 days under suitable weather conditions. Inspired by their study, we developed an approach to 17 concatenate fires during consecutive days into "multi-day fire patches". The size of each 18 "multi-day fire patch" is the cumulative daily fire size during its persistence time. This allows the 19 modelled "multi-day fire patch size" to be compared to observations. For clarity, the period of 20 multiple days that a fire spans is hereafter referred to as "fire patch length". 21 The approach to concatenate independent daily fires into multi-day fire patches is illustrated 22 in Fig. 3, which shows simulated daily fire number and fire size for a 0.5° grid cell in northern 23 Africa for the fire season of October 2001 to April 2002. Fires occur in different consecutive-day 24 periods, i.e., when simulated FDI remains above zero, and when fuel amount and fire intensity 25 exceed the given thresholds. In the example in Fig. 3, the model simulated five such periods, each 26 spanning a different number of days. Within each period, with the increase of FDI (i.e., advancing

into more suitable fire weather, shown in subplot a), the simulated daily number of fires (subplot b)
 and the mean fire size (subplot c) increased as well.

3	Fig. 3 lower panel illustrates in detail how separate daily fires were concatenated into fire
4	patches for the period of Day 740–766 (since 2000-01-01). Rather than viewing fires on a given
5	day as being independent from those of previous days, we now consider part of these fires as
6	"persisting" from previous days and part of these fires as new patches. For example, on Day 741,
7	four fires were originally simulated to start and extinguish within this day (with the exact same fire
8	size). We now consider that two of these four fires persisted from the previous day, i.e., the two fire
9	patches staring on Day 740. We then consider that the other two fires of Day 741 were new fire
10	patches initiated on this day. Similarly, the five fires originally simulated on Day 742 are now
11	considered as:
12	- two extended from the two fire patches of Day 740 (which already persisted into Day 741);
13	- two extended from the two fire patches of Day 741;
14	- one new fire patch started at this day.
15	Following this approach, fires simulated in all following days were identified either as
16	extending from fire patches of previous days, or as new fire patches. As such, the total number of
17	fire patches is equal to the maximum daily fire number during this period. In the example of Fig. 3
18	lower panel, ten fire patches were extracted from Day 740 to 766 (as indicated by the numbers and
19	arrows in red in the subplot b):
20	- two fire patches extended from Day740 to 766;
21	- two extended from Day 741 to 766;
22	- one extended from Day 742 to 766;
23	
24	- the last two fire patches extended from Day 760 to 766.
25	The size for each fire patch was the cumulative daily fire size during its persistence time. The
26	procedure illustrated in Fig. 3 was repeated for all the land grid cells (excluding those with <10

27 days of fire occurrence) over 1997–2009, to generate the "multi-day fire patches" over the globe.

1 3 Results

2 3.1 Comparison of model simulation to satellite observation for the spatial and temporal pattern of
3 burning

4 Fig. 4 shows the spatial distribution of mean annual burned fraction by the model and the 5 three satellite-derived datasets (GFED3.1, GLOBCARBON, L3JRC) for 2001-2006. L3JRC and 6 GLOBCARBON show similar spatial patterns of burning, which is different from the GFED3.1 7 data. Generally, L3JRC and GLOBCARBON have less burned area in the southern hemisphere 8 than GFED3.1 (see also Fig. 5a), with smaller spatial extent of burning in the savanna systems in 9 Africa and Australia. By contrast, in the middle to high latitudes of the northern hemisphere, 10 L3JRC and GLOBCARBON show more burned area than GFED3.1. All three datasets capture the 11 grassland burning in central and eastern Asia.

ORCHIDEE coupled with SPITFIRE is generally able to reproduce the spatial distribution and magnitude of satellite-observed burned fraction. The simulated mean annual global burned area for 2001–2006 is 346 Mha yr⁻¹, which falls within the range of 287–384 Mha yr⁻¹ given by the satellite observation data, and close to the 344 Mha yr⁻¹ by the GFED3.1 dataset when crop fires are excluded.

17 Fires in grassland-dominated systems are well captured by the model, including steppe fires 18 in central and eastern Asia, savanna fires in northern Africa, northern Australia and central to east 19 South America. Two regions could be identified where model simulation is different from all the 20 three observation datasets. One is the woodland savanna (miombo) in southern Africa, where 21 burned area is underestimated by the model (simulated annual burned fraction is ~4%, but 14–24% 22 is observed). The other is western and central continental US (dominated by C3 and C4 grass in 23 the land-cover map used by ORCHIDEE) where fires are overestimated (simulated annual fraction 24 is ~6%, but 1-2% is observed). For the fires in high-latitude (>45°N) boreal forest, sparsely 25 forested area and tundra, the magnitude of burned fraction by ORCHIDEE falls between that 26 found in the GFED3.1 and in the L3JRC/GLOBCARBON datasets (Fig. 4).

27 The simulation pixels are divided into five classes according to their simulation quality, as

1 shown in Fig. 4. Table 1 shows the mean burned area for each category and dataset. The grid cells 2 with burning collocated between model and observation data (labelled as ORC-good, ORC-max, 3 ORC-min in Fig. 4) cover the majority of global burned area (76-92% depending on different 4 datasets), indicating that the model can reproduce the major spatial extent of burning. However, 5 discrepancy still remains, in that 50% of the modelled global burned area is classified as 6 ORC-max (i.e., overestimation of burned fraction by the model), whereas observation datasets 7 have half of the global burned area labelled as ORC-min (i.e., underestimation of burned fraction 8 by the model).

9 Fig. 4 also illustrates the lower burned area found in L3JRC and GLOBCARBON in 10 comparison with GFED3.1 for southern Hemisphere and subtropical Northern Hemisphere, in 11 contrast to the higher burned area in the middle-to-high latitude region in the Northern 12 Hemisphere. The simulated latitudinal distribution of burned area generally falls within the 13 minimum-maximum range of the three observation products. Exceptions are the regions of 14 ~5–15°S and 30–40°N, corresponding to the underestimated burning in southern African savanna 15 and the overestimate in western and central US, discussed above. The annual time series of burned 16 area are shown in Fig. 5b. The correlation coefficient between ORCHIDEE and the GFED3.1 data 17 is highest (0.48; linearly detrended correlation coefficient of 0.59), compared to that of 0.26 18 between GLOBCARBON and the GFED3.1; and -0.59 between L3JRC and the GFED3.1.

19 3.2 Model evaluation against GFED3.1 burned area data

The simulated burned area for 1997–2009 is evaluated against the GFED3.1 data for each region in terms of mean annual burned area, and similarity in interannual variability and seasonality (see metrics in Sect. 2.5). The results are presented in Table 2; and the annual burned areas for different regions are shown in Fig. 6.

The model error for annual burned area (BA) is highest for Middle East (MIDE, by a factor of 41.9, occupying 0.1% vs. 5.6% of global BA by GFED3.1 vs. ORCHIDEE) and lowest for Boreal Asia (BOAS, by a factor of -0.1, occupying 1.6% vs. 1.4% of global BA). The model underestimates burned area in the three biggest fire regions (Northern hemisphere Africa, Southern Hemisphere Africa and Australia, together occupying 86% vs. 46.8% of global BA) by on average
 45.6%. Prominent model overestimation is found in Central Asia (CEAS, by a factor of 3.8,
 occupying 3% vs. 14.9% of global BA), and Southern Hemisphere South America (SHSA, by a
 factor of 1.6, occupying 5.5% vs. 15.7% of global BA).

5 The correlation coefficient for the linearly detrended global annual burned area between 6 ORCHIDEE and GFED3.1 is 0.59, indicating that the model moderately captures the IAV of 7 burned area (although it fails to reproduce the 1998 El Niño peak burning), because errors are 8 compensated among regions (Fig. 5b). On regional scale, the model performs best at regions 9 where the IAV of burned area is known to be mainly driven by climate, such as Boreal North 10 America (BONA), Boreal Asia (BOAS), and the Equatorial Asia (EQAS). However, the model 11 performs rather poor at regions where burned area shows little IAV such as Africa (NHAF and 12 SHAF, see also Fig. 6), or the burned area is unrealistically simulated by the model (such as TENA 13 and MIDE). For most regions the model captures the fire seasonality rather well (Table 2), with 14 S_{season} being significantly different from that of a randomly distributed seasonality, except in 15 NHAF, BOAS and SEAS.

16 3.3 Fire and precipitation relationship

The model captures well the empirical relationship between burned area and precipitation found in tropical and subtropical regions (Fig. 7; see also Prentice et al., 2011; van der Werf et al., 2008b). In low-precipitation regions (<400 mm yr⁻¹), the climate is favourable for fire but burning is limited by the available fuel. In contrast, regions with higher precipitation (>200 mm yr⁻¹) always support sufficient fuel amount but fires are limited by the duration of dry season when fires can occur. Burned area is maximal for regions with intermediate precipitation and productivity (Krawchuk and Moritz, 2010).

Maximum burning occurs around an annual precipitation of 1000 mm according to model simulation, compared to 1200 mm by GFED3.1, and 1400 mm by the GLOBCARBON and L3JRC datasets. GLOBCARBON and L3JRC show the lowest burning in this tropical/subtropical belt, followed by the model simulation, with the burned area by GFED3.1 being the highest. The fire and precipitation relationship was further divided into four sub-regions of America, Africa,
 Asia and Australia following van der Werf et al (2008b) and the results are shown in Supplement
 Fig. S5. The model-observation agreement in fire-precipitation pattern is moderate in Africa and
 Australia, but low-precipitation fires are overestimated in America and Asia.

5 3.4 Peak fire month and fire season length

6 The spatial distributions of fire peak months by ORCHIDEE and GFED3.1 are compared in 7 Fig. 8. The spatial pattern of simulated peak fire month is in general agreement with GFED3.1 8 data. The model simulation gives $D_2 = 0.3$, indicating that simulated and observed peak fire 9 month differ by on average two months. At regional scale, simulated peak fire months for most 10 regions are within one month of those by GFED3.1 data, except MIDE and SHAF (see the far 11 right-hand column of Table 2). Refer to Supplement Fig. S6 for more detailed information of 12 modelled and observed seasonal pattern of burning for different GFED3.1 regions.

Fig. 9 compares simulated fire season length with the GFED3.1-derived fire season length from Archibald et al. (2013). The spatial pattern of fire season length by model simulation agrees well with that given by Archibald et al. (2013), with fire season length lasting 1–3 months in boreal regions, and 4–7 months in semi-arid grasslands and savannas. The fire season length in eastern Africa, South Africa, Botswana, Namibia, Argentina and Mexico is overestimated by the model by 2–4months; and underestimated in Southeast Brazil by 1–2 months.

19 3.5 Long-term trends of burned area during the 20th century

Over the 20th century, the historical trend of modelled global total burned area generally follows the Mouillot reconstruction data (Mouillot and Field, 2005) as corrected by GFED3.1 data (see Sect 2.4.1), with increasing burned area after the 1930s until the 1990s–2000s, after which global burned area began to decrease (Fig. 10). However, the inter-decadal variability of burned area is underestimated. Regionally, simulated decadal burned area agrees relatively well with the Mouillot data in boreal Russia (beginning from 1930s), although the observation data is subject to great uncertainty, especially before 1950s (Mouillot and Field, 2005). The simulated burned area 1 also agrees well with fire agency data for Canada & US Alaska (Supplement Fig. S7). The 2 correlation coefficient for annual BA between model and fire agency data is 0.44 after 1950, and 3 0.57 between model and Mouillot data, when the observation data are considered to be more 4 reliable. This reflects the model capability to capture fire variability driven by climate variation 5 relatively well.

6 The model fails, however, to capture burned area variation for regions where fires from 7 changed land cover likely played a bigger role in the earlier 20th century according to the Mouillot 8 data (Mouillot and Field, 2005), for example, in Australia and New Zealand, USA and southern 9 South America, mainly because the static land cover was used in the simulation. Strong 10 model-observation disagreement also occurs for regions where the implementation of modern fire 11 prevention has drastically reduced the burned area, as occurred in the 1960s in boreal North 12 America, because the general implicit inclusion of human suppression on ignitions in Eq. (1) on 13 the global scale does not accommodate regional uniqueness.

14 3.6 Compare simulated fire size with observation

15 In this section we compare the simulated fire size distribution against observation over two 16 regions: Canada and US Alaska combined, and southern Africa. The number and size of 17 reconstructed "multi-day fire patches" by the model were used (see Sect. 2.5.2). For both 18 simulated and observed fires, all fires within the test region were pooled together. Fires were 19 binned according to fire patch size in an equal logarithmic distance manner (with the 20 minimum-maximum size range being divided into 100 bins). The mean annual number of fires on 21 an area basis for each bin was calculated. For the fire patches in Canada for which fire start date 22 and fire out date were reported, fire length was calculated and compared with model simulation as 23 well.

Fig. 11 shows the fire size and the corresponding number of fires for each size bin over US Alaska + Canada, and southern Africa. The modelled fire size distribution reaches a maximum at intermediate fire sizes. This is because when the climate is less favourable for fire occurrence, both number of fires and fire size are limited (corresponding to the low fire size end in Fig. 11a &

1 b). While the size and number of fires could be large during the period when climate is dry and 2 large fires are possible (corresponding to the high fire size end in Fig. 11 a & b), the frequency of 3 high-fire period itself might be rare. The similar pattern is shown by the fire agency data of 4 Canada and US Alaska, but not by the satellite-derived fire patch data in southern Africa, which 5 might be due to that the minimum fire size (25ha) is limited by the satellite resolution there. 6 However, for both regions, the frequency and size of extreme large fires were underestimated by 7 the model. Further comparison of fire lengths for Canada (Fig. 11c) reveals that the model 8 underestimated fire length by as much as 60 days for the extreme large fires.

9 We further calculated the cumulative fraction of total burned area by fires below a given 10 quantile of fire size (the minimum size, every tenth quantile from 10th to 90th quantile, and the 11 maximum size) (Fig. 12). According to observation, in boreal Canada & US Alaska, the total 12 burned area is mainly dominated by a few large fires, with the top 10% of fires (90th quantile to 13 the maximum size) accounting for 99.8% of the total burned area. By contrast, the same group of 14 fires (i.e., the highest 10% large fires) account for ~80% of the simulated total burned area, with 15 the remaining being accounted for by many small fires. For southern Africa, the model distribution 16 follows rather relatively well of the observation. The top 30% of fires (70th quantile to the 17 maximum size) make up 90% of the total burned area by satellite data and 85% by the model 18 simulation.

19 Figure 13 compares the simulated 95th quantile fire size with the global observation by satellite. According to observation, fires with biggest fire size (500-10,000 km²) are 20 21 grassland-dominated fires in central and eastern Asia, African savanna and northern Australia, 22 which are followed by fires in Russian and Alaskan boreal forest (and sparsely forested area) or 23 tundra, and savanna-woodland fires in Africa and central South America (50-500 km²). Fires in 24 the rest of the world are relatively small (<50 km²). In terms of spatial fire size distribution, the 25 model could reproduce the biggest fires in grassland-dominated systems in central and eastern Asia (200–1000 km²), northern Australia (50–200 km²), as well as fires in Russian boreal forest 26 27 (and sparsely forested area) or tundra (50-200km²; note that tundra is treated as C3 grassland in 28 the model), but fire size for these regions is generally smaller than observation by up to one magnitude. The fire size in central South America tends to be underestimated, and overestimated
in western to central U.S. Fire size for the rest of the world is comparable between model and the
observation, with general small fire size (<50km²) and low fire frequency.

4 **4 Discussion**

5 4.1 General model performance

6 The SPITFIRE module was first presented by Thonicke et al. (2010). It was later adapted in 7 LPX by Prentice et al. (2011), notably with the removal of ignition sources created by humans, 8 and further by Pfeiffer et al. (2013) and Lasslop et al. (2014). Pfeiffer et al. (2013) developed a 9 scheme to allow fire span of multiple days and fire coalescence, explored the role of lightning 10 interannual variability in the model, and included terrain effects on fire at a broad scale. Lasslop et 11 al. (2014) investigated model sensitivity to climate forcing and to the spatial variability of a 12 number of fire relevant parameters. In the current study, the SPITFIRE module was fully 13 integrated into the global process-based vegetation model ORCHIDEE for the first time. 14 Simulated burned area, fire season and fire patch size distribution were evaluated in a 15 comprehensive way against observation for the recent period (1997-2009), and against 16 reconstructed burned area for the 20th century.

17 The simulated global mean annual burned area for 2001-2006 agrees with the satellite 18 observation data and is most close to the GFED3.1 data. The model could moderately capture the 19 interannual variability in burned area as revealed by GFED3.1 data with the exception of the peak 20 burning of 1998, probably because deforestation and tropical peat fires in the 1997–98 El Niño 21 event (van der Werf et al., 2008a) were not represented. Model-observation discrepancy remains at 22 regional scale, with underestimation mainly in the savanna of Africa and Australia, and 23 overestimation in central Asia, Middle East, the temperate North America, and southern 24 hemisphere South America.

The model can reproduce the maximal burned area for the intermediate range of precipitation for tropical and subtropical regions $(35^{\circ}S-35^{\circ}N)$ (also shown by Prentice et al., 2011; van der Werf et al., 2008b). For boreal regions where climate plays a dominant role in driving the temporal

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variation of burned area, simulated burned area generally agrees well with the historical reconstruction data (boreal Russia for 1930–2009) and government fire agency data (US state of Alaska and Canada for 1950–2009), indicating that the model is capturing the climate as a driver of fire. However, on the global scale because fire trend is determined by multiple factors including climate, land-use practice and fire suppression (Mouillot and Field, 2005), simulated burned area trend only agrees moderately well with the reconstruction data for the 20th century, given that the static land cover and the simple human ignition equation were used in the model.

8 4.2 Potential sources of systematic errors

9 Fire is a complex, regional process that involves diverse dimensions of vegetation, climate 10 and human activity (Bowman et al., 2009). Uncertainties in all these factors will contribute to the 11 overall uncertainty in simulated fire activities. Because SPITFIRE simulates fire occurrence 12 through a complex chain that includes from potential ignition to fire climate, to fire spread rate 13 and fire size and tree mortality, identifying contributions of each modelling step to the ultimate 14 error in simulated fire regime is problematic. A complete error analysis involving all model 15 parameters is beyond the scope of this study, but following sections are intended to serve as 16 preliminary investigation of model errors.

17 4.2.1 Ignition sources

18 On the global scale, due to the limitation of fire by fuel load on the arid and semi-arid regions, 19 modelled annual burned area is more closely related to the total fire numbers rather than the fire 20 danger index (Fig. S8). This might lead to speculation that the potential ignition source is the first 21 identified source of error for simulated burned area. One possible error in ignition is that potential 22 lightning ignitions are not suppressed by human in densely populated areas, which cause 23 lightning-ignited fires being overestimated. We have tested this possibility by applying a 24 population density dependent human suppression of lightning-ignited fires following Li et al. 25 (2012), and the result showed that part of the overestimation of burned fraction in western US and 26 central South America could be reduced (Fig. 4, Fig. S9), but the burned area in Africa and over 1 the whole globe were further underestimated.

2 Published fire models (Kloster et al., 2012; Li et al., 2012; Pechony and Shindell, 2009; 3 Venevsky et al., 2002) generally include ignitions from lightning and human sources, together with 4 explicit or implicit human suppression of fires. However, one common challenge is to properly 5 calibrate ignition parameters. One option is to use active fire counts (as in case of Li et al., 2012; 6 Pechony and Shindell, 2009), but fire counts are not exactly real fire numbers, because a single 7 widespread fire could be seen as many fire counts and the burned area per hotspot vary by an 8 order of magnitude depending on vegetation composition (Hantson et al., 2013). Simulated burned 9 area on the global scale might be comparable with the satellite observation data. However, very 10 little is known on how this agreement was achieved; nor on whether each step of the modelling 11 process was correctly captured, or the ultimate agreement in burned area is mainly thanks to error 12 compensation among different steps. Two recent modelling studies (Lasslop et al., 2014; Yang et 13 al., 2014) used some scaling factor to adjust either directly burned area or the ignitions to match 14 simulated global burned area with observation.

Pfeiffer et al. (2013) argued that interannual lightning variability was critical for their model to capture the trend and interannual variation of burned area, especially for regions where burned area is dominated by lightning fires such as boreal forests in Alaska and Canada. Currently, a single dataset of monthly lightning flashes was repeated each year in our model. However, we have tested the possibility to include the interannual lightning variability by following their approach.

21 The historical lightning variability during the 20th century was reconstructed by using the 22 convective potential available energy (CAPE) output variable from the 20th Century Reanalysis 23 Project (http://portal.nersc.gov/project/20C_Reanalysis/), following Equation 1 on Page 649 of 24 Pfeiffer et al. (2013). A test simulation was run for 1901-2009 over the whole globe with the 25 variable lightning input. We found that shifting from repeated lightning data to CAPE-derived data 26 decreased the Pearson correlation coefficient between simulated decadal burned area and the 27 Mouillot reconstruction for half of the 14 regions and for the globe, but increased the correlation 28 for other regions. Over 1997–2009 with observation data by GFED3.1 more credible than the 20th

century reconstruction, using the CAPE-derived data decreased the Pearson correlation coefficient
 between annual simulated burned area and GFED3.1 for the globe and for most of the regions.
 This surprising result could be due to the fact that model internal errors may compensate for
 possible benefits of using the CAPE-derived data, or because CAPE-derived lightning data does
 not reflect the real lightning variability everywhere. For more detailed results and discussions,
 refer to the Section 2 of the Supplement Material.

7 4.2.2 Fire number, size and fire patch length

8 The comparison of simulated "multi-day fire patches" against the observation in Canada & 9 US Alaska, and southern Africa shows that the model underestimated the frequency and size of 10 extreme large fires. Given that annual burned area in US Alaska and Canada is slightly overestimated by the model (2.8 Mha yr⁻¹ by model during 1960–2009 against 2.2 Mha yr⁻¹ by fire 11 agency and 1.9 Mha yr⁻¹ by the Mouillot data, also refer to Fig. S7 in Supplement), the total 12 13 number of (the small- and medium-sized) fires must be overestimated (i.e., making compensation 14 for the too small fire size). However, the compensation by fire numbers does not occur for 15 southern Africa, where, given the underestimation of large fire size, the total burned area remains 16 underestimated (Fig. 4, Table 2). Over the globe, despite the fact that the model correctly identifies 17 some regions where large fires occur (mainly with grassland fraction higher than $\sim 70\%$ by the 18 land cover map used in the model), the large fire size remains underestimated – by approximately 19 one magnitude (Fig. 13).

20 The fire size of reconstructed "multi-day fire patches" is defined as the cumulative daily fire 21 size over the corresponding fire patch length, and the underestimation of large fires could be due 22 to underestimation in either fire patch length or daily fire size (i.e., fire patch size grows too 23 slowly over its length). The comparison of simulated fire length with fire agency data in Canada 24 suggests that model generally underestimated fire length. For fires larger than 10,000 ha, the 25 underestimation is as large as 40-60 days. Given similar underestimation of large fire sizes in 26 other ecosystems that are characterized by large-size fires (Fig. 13), the fire length underestimation 27 in Canada is likely to happen elsewhere, though this could not be completely verified mainly because of the lack of fire length observation across the world (the satellite-derived fire size data
 by Archibald et al. 2010, 2013 used in this study does not include the corresponding fire length
 yet).

4 In Pfeiffer et al. (2013), fires were simulated to span multiple days and extinguish when the 5 cumulative precipitation exceeds some threshold, with the daily fire size remaining limited by 241 6 minutes. There is one significant difference between our approach and theirs. In Pfeiffer et al. 7 (2013), fires starting on a given day are always considered as "new fires" to be added on the 8 existing fire count on the previous day. The number of fires over a given grid cell thus 9 accumulates each day as the time advances in a period suitable for fire occurrence. In our model, 10 fires are originally simulated as independent events within individual days (because they are 11 assumed to start and extinguish during the same day). To allow the comparison of simulated fire 12 size against observations, these independent fires were concatenated (regrouped) into fire patches 13 that span different number of days. We made the concatenation by assuming fires at a given day 14 either persist from the previous days or are actual new fire patches. Thus the number of 15 reconstructed fire patches by our approach is the maximum daily fire number during the 16 consecutive days of burning.

17 This difference in accounting for fire patches arose from the different approaches to handle 18 ignitions in the two models, in particular ignitions by lightning activities, because anthropogenic 19 ignitions were excluded in Pfeiffer et al. (2013). Pfeiffer et al. (2013) simulated lightning-ignited 20 fires by finally dropping the physical meaning of the number of lightning flashes in the input data. 21 This information was only used to obtain an all-or-nothing answer, allowing either a single fire 22 over the whole 0.5-degree grid cell or no fire at all. However, the number of cloud-to-ground 23 lightning flashes retained its physical meaning in our approach, and was scaled by the simulated 24 FDI and fuel-limiting ignition efficiency to derive a daily number of fires. As no validation 25 information was provided regarding fire number and fire size in Pfeiffer et al. (2013), it's unclear 26 which approach vields results closer to observations. Overall, it remains challenging to develop a 27 proper approach to represent the heterogeneous fire patches, and the growth of each individual 28 patch in grid-based models (Jones et al., 2009).

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The underestimation of large fire sizes in Canada is partly due to underestimation of fire patch length. A closer comparison of Fig. 11a and c suggests that, while fire length was underestimated by a factor of 2–3 for fires between 10^4 – 10^5 ha, given the same fire number (shown as the vertical axis of Fig. 11a), the fire size was underestimated by roughly an order of magnitude (10 times). This implies the (mean) daily fire size is likely underestimated as well (roughly 3–5 times).

8 Within the model, daily fire size is determined by daily fire active burning time and fire 9 spread rate. Evidence shows that wildfires display a characteristic diurnal cycle, with the most 10 active time being around midday and early afternoon when the humidity is at a minimum and 11 wind speeds are higher (Mu et al., 2011; Pyne et al., 1996; Zhang et al., 2012). Outside this active 12 burning time, fires could persist but propagate at a rather low speed (especially at night) or even 13 turn into smoldering and burn with flame later again when the weather is feasible. For extremely 14 large fires, the active burning period could be longer because these fires often occur when the fuel 15 is extremely dry as a consequence of extended drought weather. Currently, this active burning time 16 is simulated to exponentially increase with the fire danger index (i.e., indicator of daily fire 17 weather) with a maximum time of 241 minutes. This limit might not be feasible for all ecosystems 18 and fire sizes; however, a full exploration of this issue is currently limited by the scarcity of 19 observations.

20 A global map of simulated 95th quantile fire spread rate is provided in Supplement Fig. S10. 21 Li et al. (2012) compiled fire spread information from the literature and reported typical fire spread 22 rates of 12 m min⁻¹ for grasslands, 10.2 m min⁻¹ for shrubs, 9 m min⁻¹ for needle-leaved trees and 23 6.6 m min⁻¹ for other trees. The simulated fire spread in grasslands in central and eastern Asia 24 (20-40 m min⁻¹) is much higher than the observed range. Considering that daily fire sizes are 25 likely underestimated, the limit of 241-minute maximum daily active burning time might be too 26 short to correctly simulated large fire sizes in these grassland ecosystems. By contrast, simulated fire spread rate in savanna vegetation $(1-5 \text{ m min}^{-1})$ is lower than the reported value (10-12 m)27 min⁻¹), and this could help to explain the underestimation of fire size and the broad 28

1 underestimation of burned area in this region.

2 Fire spread in the northern high-latitude boreal forest, sparsely forested area and tundra is modelled to be extremely high (>10 m min⁻¹). This is mainly because herbaceous plants in these 3 4 regions are simulated as C3 grasslands in the model with relatively low bulk density. However the 5 likely high daily burned area due to the high fire spread rate was compensated by simulated short 6 fire patch length in the model (Fig. 11c), so that simulated burned area for the region of 50–75°N 7 agrees well with GFED3.1 data (Fig. 5a). Pfeiffer et al. (2013) proposed to relate the grass fuel 8 bulk density with the annual sum of degree-days over 5° C. We have tested their approach and 9 found that this new approach decreased the simulated burned area for the high-latitude region 10 $(50-70^{\circ}N)$ and for the globe, and thus was not included in the current version of model.

11 To gain more insights into the model's behaviour, the simulated 95th quantile fire patch size 12 was related with other parameters (grassland fraction, fuel bulk density). As shown in Fig. S11, the 13 size of large fires exponentially depends on the fire spread rate. The fire spread rate is very 14 sensitive to the fuel bulk density and grass fraction beyond some threshold (e.g., fire spread rate 15 surges when grass coverage exceeds \sim 70%), with the fuel bulk density being inversely dependent 16 on the grass fraction. Thus the simulated fire size could be sensitive to the land cover map 17 (especially grass fraction) used in the simulation. Besides, as a static land cover map is used in our 18 simulation, the grassland fraction is not allowed to vary as a response to fire disturbance and thus 19 a full fire-climate-vegetation feedback is limited, and this could probably help to explain the 20 underestimation of fire size.

21 4.2.4 Influence of fire-climate-vegetation feedback

In the current simulation the dynamic vegetation module of ORCHIDEE was switched off and a static land cover map was used. Tree mortality was affected by fire-induced tree damage, but tree coverage within a given grid cell was static and not allowed to vary with fire occurrence. A test simulation has been done for Southern Hemisphere Africa following the same simulation protocol as in Sect. 2.3 but with the dynamic vegetation module being switched on, in order to investigate the simulated fire behaviour with dynamic vegetation. Figure S12 compares the simulated annual burned area, grass and tree coverage change and the fire danger index for
 1901–2009 with the model in dynamic and static vegetation modes.

3 In dynamic vegetation mode, the simulated burned area suddenly begins to increase around 4 1965, in response to increased fire danger index. The increase in fire activity further increases the 5 grass coverage and reduces tree coverage, causing a positive feedback to finally induce a peak of 6 burned area around 1975, after which the burned area decreases. In static vegetation mode, the 7 simulated burned area shows a similar peak in response to the peak in the fire danger index, 8 however, with a much smaller peak of burned area than that simulated in dynamic vegetation 9 mode because of the lack of fire-vegetation-climate feedback. Simulated burned area by both 10 simulations is still lower than that given by the GFED3.1 data for the period of 1997–2009, 11 although the peak burned area in the dynamic vegetation mode is comparable with GFED3.1 data. 12 This test indicates that including the fire-climate-vegetation feedback could improve the 13 simulation when the climate is favourable for fire occurrence. At the same time, it also suggests 14 that other factors like climate, and the model mechanisms determining the competitiveness of trees 15 versus grass might also play a role in the error of fire modelling.

16 4.2.5 Potential error sources for regional bias

Given that burned area is simulated in the model as a result of several sequential steps (lightning and human ignitions, fire number & fire size distribution, fire patch length, daily fire size, daily active burning time and fire spread rate), and the scarcity of global coverage observation data, we are not able to give a quantitative estimation of the role of each of these factors in determining the final model error for each GFED region. Rather, here we select several typical regions and briefly discuss the possible reasons for model performance (either good simulation, or over/under-estimation).

The model agrees with the GFED3.1 burned area for 50–75°N relatively well (Fig. 5). Burned area for Canada and US Alaska for 1950–2009 was overestimated by ~27% compared with fire agency data. However for the GFED region of BONA (Fig. 2), the model overestimates BA by 60% (Table 2). This is mainly because spatial extent of BONA includes part of the grassland systems in

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1 northern US, where the burned area is overestimated (Fig. 4), same for TENA. The relatively good 2 simulation of boreal burned area (note in Table 2, E_{BA} for BOAS is -0.1) is mainly because 3 underestimated large fire sizes are compensated by overestimated fire patch numbers. Though 4 simulated fire spread rate for some regions is extremely big, however the daily fire sizes are still 5 likely underestimated, possibly due to too short daily active burning time.

6 The burned area for the northern hemisphere temperate regions is systematically 7 overestimated by the model, including EURO, CEAS, TENA, CEAM, MIDE, and SEAS. Two 8 reasons are suspected to contribute to this overestimation. First, extensive grassland coverage is 9 found in some regions (CEAS, MIDE, TENA, CEAM, part of MIDE), where simulated fire spread 10 rate is much higher than observation, likely creating high daily fire size. Note this is not in 11 contradiction with the underestimated large fire patch sizes (Fig. 13) because fire patch length 12 could be underestimated. Second, for regions where human population density is high and active 13 fire suppression is implemented, such as India in SEAS, China Inner Mongolia in CEAS and 14 Europe, ignitions seem to be excessive, leading to larger burned areas in spite of the small 15 simulated fire sizes. This is partly because the lightning ignitions are not suppressed in the current 16 model version, and because the global human-ignition relationship is not feasible everywhere and 17 the spatial a(ND) dataset used in the model is not able to efficiently handle the spatial 18 heterogeneity.

Finally, burned area in the three biggest fire regions of NHAF, SHAF and AUST, which are dominated by savanna and woodland savanna, is underestimated by the model. This underestimation is primarily due to underestimated large fire size, which are not compensated by the ignitions. The simulated fire spread rate in Australia (7.5–15 m min⁻¹) seems comparable with observation (10–12 m min⁻¹), however it's underestimated in Africa. The underestimation of burned area in SHAF is likely also related with its low grassland coverage, given that the simulated fire size is rather sensitive to the grassland fraction (Fig. S11, Fig. S13).

Further, Archibald et al. (2013) showed that two major fire types dominate the burned area of Africa (frequent intense large fires and frequent cool small fires) and their correlation with environmental factors seems to be clearly distinguished by the human impact index. This implies that the a(ND) values should ideally differ as well among these two fire types, which currentlyshare the same value (Fig. S1).

3 The regional pattern of model-observation disagreement in our study is also shared by 4 another SPITFIRE implementation in the JSBACH land surface model by Lasslop et al. (2014), 5 who modified a scalar in the human-ignition equation to match the simulated global burned area 6 with observation. It remains somewhat a common challenge for processed-based fire models to 7 correctly represent the global burned area and its spatial distribution; and in some cases ignitions 8 need to be adjusted or optimized according to the observation data (Lasslop et al., 2014; Yang et al., 9 2014). Li et al. (2013) included the social economic factors in simulating fires for some vegetation 10 types, which could be incorporated in the future development of our model.

11 4.2.6 Uncertainty/error summary

12 The preliminary investigation of modelling error reveals that large fire size is underestimated 13 over regions of high fire frequency; and the ignition error is playing an important role in 14 determining the ultimate simulated burned area. On the regional scale, ignition numbers (fire 15 numbers) are either overestimated to compensate fire size underestimation to cause a moderate or 16 overestimated burned area, or are not enough that the simulated burned area is underestimated as 17 well. The underestimation of large fire patch size is likely due to underestimation in both fire 18 patch length and the daily fire size, which could further be limited by the daily active burning time. 19 Overall, the moderate model-agreement on global burned area could be achieved only when errors 20 among different regions are compensated.

21 4.3 Future model improvement directions and needed datasets

Currently many efforts in global fire modelling are directed at reproducing the temporal and spatial pattern of burned areas (Kloster et al., 2010; Li et al., 2012; Pfeiffer et al., 2013; Prentice et al., 2011; Thonicke et al., 2010). Total burned area is determined by ignition frequency and fire size, which itself is controlled by fire spread rate (fire intensity) and fire duration. More work is needed to investigate if a model can reproduce the mechanisms that drive burned area: i.e. the rate of spread, fire patch length, daily active burning time, fire size, ignition frequency, and fireline intensity. Comparing observed and simulated fire regimes, which combine information on fire timing (fire season), size, numbers and intensity (Gill and Allan, 2008) will help to reveal gaps in this understanding. The present study is a step in this direction, bringing new in-depth model evaluation.

6 In summary, the fire processes in the SPITFIRE model are complex enough to include many 7 aspects of wildland and human-caused fire processes in nature. However, little is known about the 8 parameter sensitivities and their contribution to model error. The simulated intermediate model 9 parameters (e.g., fire numbers, fire patch length, fire size, daily active burning time, fire spread 10 rate, fire intensity) are poorly constrained by the observation data. As a result, error compensation 11 could be prevalent in the model and a wider application of the model is impeded.

12 To advance model development, global measurement datasets of the key fire-relevant 13 parameters, including fire size, fire patch length, fire diurnal variability, fire spread rate, fuel bulk 14 density, wind speed, fire intensity etc., should be established and used to calibrate fire models. On 15 the other hand, the complexity of fire model parameters and the regional nature of fire processes 16 make it unlikely that these parameters could be calibrated in a parameter-by-parameter and 17 site-by-site way, but some more advanced techniques such as data assimilation or model-data 18 fusion could be helpful. Finally, some more mechanistic fire processes should be considered to be 19 included into the model, such as crown fire spread and the mechanistic process of fire extinction.

20 5 Conclusions

We have integrated the SPITFIRE model into a global process-based vegetation model ORCHIDEE. The historical burned area for the 20th century was simulated and the modelled fire regimes were evaluated against observation data. The model was able to capture well the historical climatic drivers of burned area for the 20th century. However, parameter uncertainties such as number of fire ignitions, daily active burning time and fire spread rate result in considerable regional discrepancies. Large fire sizes are generally underestimated, with the error in simulated burned area being partly compensated by overestimated fire numbers. Future model development requires a complete parameter sensitivity analysis for the key processes represented in fire
 modeling. To constrain the model error, consistent spatial observational datasets should be

3 established for validating the key variables in the model at different modelling steps.

4

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11 References

12	Andela, N., Liu, Y. Y., van Dijk, A. I. J. M., de Jeu, R. A. M. and McVicar, T. R.: Global changes in
13	dryland vegetation dynamics (1988-2008) assessed by satellite remote sensing: comparing a
14	new passive microwave vegetation density record with reflective greenness data,
15	Biogeosciences, 10(10), 6657-6676, doi:10.5194/bg-10-6657-2013, 2013.
16	Archibald, S., Lehmann, C. E. R., Gómez-Dans, J. L. and Bradstock, R. A.: Defining pyromes and
17	global syndromes of fire regimes, PNAS, 110(16), 6442–6447, doi:10.1073/pnas.1211466110,
18	2013.
19	Arora, V. K. and Boer, G. J.: Fire as an interactive component of dynamic vegetation models,
20	Journal of Geophysical Research: Biogeosciences, 110(G2), G02008,
21	doi:10.1029/2005JG000042, 2005.
22	Barrett, K., McGuire, A. D., Hoy, E. E. and Kasischke, E. S.: Potential shifts in dominant forest
23	cover in interior Alaska driven by variations in fire severity, Ecological Applications, 21(7),
24	2380–2396, doi:10.1890/10-0896.1, 2011.
25	Beck, P. S. A., Goetz, S. J., Mack, M. C., Alexander, H. D., Jin, Y., Randerson, J. T. and Loranty, M.
26	M.: The impacts and implications of an intensifying fire regime on Alaskan boreal forest
27	composition and albedo, Global Change Biology, no-no,
28	doi:10.1111/j.1365-2486.2011.02412.x, 2011.
29	Bond, W. J., Woodward, F. I. and Midgley, G. F.: The global distribution of ecosystems in a world
30	without fire, New Phytol., 165(2), 525–537, doi:10.1111/j.1469-8137.2004.01252.x, 2005.
31	Bowman, D. M. J. S., Balch, J. K., Artaxo, P., Bond, W. J., Carlson, J. M., Cochrane, M. A.,
32	D'Antonio, C. M., DeFries, R. S., Doyle, J. C., Harrison, S. P., Johnston, F. H., Keeley, J. E.,
33	Krawchuk, M. A., Kull, C. A., Marston, J. B., Moritz, M. A., Prentice, I. C., Roos, C. I., Scott, A.
34	C., Swetnam, T. W., van der Werf, G. R. and Pyne, S. J.: Fire in the Earth System, Science,

35 324(5926), 481 –484, doi:10.1126/science.1163886, 2009.

Burton, P. J., Parisien, M., Hicke, J. A., Hall, R. J. and Freeburn, J. T.: Large fires as agents of ecological diversity in the North American boreal forest, Int. J. Wildland Fire, 17(6), 754–767, 2008.

1	Cecil, D. J., Buechler, D. E. and Blakeslee, R. J.: Gridded lightning climatology from TRMM-LIS
2	and OTD: Dataset description, Atmospheric Research, doi:10.1016/j.atmosres.2012.06.028,
3	2012.
4	Daniau, AL., Goñi, M. F. S., Martinez, P., Urrego, D. H., Bout-Roumazeilles, V., Desprat, S. and
5	Marlon, J. R.: Orbital-scale climate forcing of grassland burning in southern Africa, PNAS,
6	doi:10.1073/pnas.1214292110, 2013.
7	Giglio, L., Randerson, J. T., van der Werf, G. R., Kasibhatla, P. S., Collatz, G. J., Morton, D. C. and
8	Defries, R. S.: Assessing variability and long-term trends in burned area by merging multiple
9	satellite fire products, [online] Available from:
10	http://adsabs.harvard.edu/abs/2010BGeo7.1171G (Accessed 9 January 2011), 2010.
11	Gill, A. M. and Allan, G.: Large fires, fire effects and the fire-regime concept, Int. J. Wildland Fire,
12	17(6), 688–695, 2008.
13	Hantson, S., Padilla, M., Corti, D. and Chuvieco, E.: Strengths and weaknesses of MODIS hotspots
14	to characterize global fire occurrence, Remote Sensing of Environment, 131, 152-159,
15	doi:10.1016/j.rse.2012.12.004, 2013.
16	Hoffmann, W. A., Geiger, E. L., Gotsch, S. G., Rossatto, D. R., Silva, L. C. R., Lau, O. L.,
17	Haridasan, M. and Franco, A. C.: Ecological thresholds at the savanna-forest boundary: how
18	plant traits, resources and fire govern the distribution of tropical biomes, Ecology Letters, 15(7),
19	759–768, doi:10.1111/j.1461-0248.2012.01789.x, 2012.
20	Jones, B. M., Kolden, C. A., Jandt, R., Abatzoglou, J. T., Urban, F. and Arp, C. D.: Fire Behavior,
21	Weather, and Burn Severity of the 2007 Anaktuvuk River Tundra Fire, North Slope, Alaska,
22	Arctic, Antarctic, and Alpine Research, 41(3), 309–316, doi:10.1657/1938-4246-41.3.309,
23	2009.
24	Kasischke, E. S. and Hoy, E. E.: Controls on carbon consumption during Alaskan wildland fires,
25	Global Change Biology, 18(2), 685–699, doi:10.1111/j.1365-2486.2011.02573.x, 2012.
26	Kasischke, E. S. and Turetsky, M. R.: Recent changes in the fire regime across the North American
27	boreal region—Spatial and temporal patterns of burning across Canada and Alaska,
28	Geophysical Research Letters, 33(9), L09703, doi:10.1029/2006GL025677, 2006.
29	Kasischke, E. S., Verbyla, D. L., Rupp, T. S., McGuire, A. D., Murphy, K. A., Jandt, R., Barnes, J.
30	L., Hoy, E. E., Duffy, P. A., Calef, M. and Turetsky, M. R.: Alaska's changing fire regime —
31	implications for the vulnerability of its boreal forestsThis article is one of a selection of papers
32	from The Dynamics of Change in Alaska's Boreal Forests: Resilience and Vulnerability in
33	Response to Climate Warming., Can. J. For. Res., 40(7), 1313–1324, doi:10.1139/X10-098,
34	2010.
35	Keeley, J. E., Fotheringham, C. J. and Morais, M.: Reexamining Fire Suppression Impacts on
36	Brushland Fire Regimes, Science, 284(5421), 1829–1832, doi:10.1126/science.284.5421.1829,
37	1999.
38	Kloster, S., Mahowald, N. M., Randerson, J. T. and Lawrence, P. J.: The impacts of climate, land
39	use, and demography on fires during the 21st century simulated by CLM-CN, Biogeosciences, 9,
40	509–525, doi:10.5194/bg-9-509-2012, 2012.
41	Kloster, S., Mahowald, N. M., Randerson, J. T., Thornton, P. E., Hoffman, F. M., Levis, S.,
42	Lawrence, P. J., Feddema, J. J., Oleson, K. W. and Lawrence, D. M.: Fire dynamics during the

1	20th century simulated by the Community Land Model, Biogeosciences, 7(6), 1877–1902,
2	2010.
3	Knorr, W., Kaminski, T., Arneth, A. and Weber, U.: Impact of human population density on fire
4	frequency at the global scale, Biogeosciences, 11(4), 1085–1102,
5	doi:10.5194/bg-11-1085-2014, 2014.
6	Krawchuk, M. A. and Moritz, M. A.: Constraints on global fire activity vary across a resource
7	gradient, Ecology, 92(1), 121-132, doi:10.1890/09-1843.1, 2010.
8	Krinner, G., Viovy, N., Noblet-Ducoudré, N. de, Ogée, J., Polcher, J., Friedlingstein, P., Ciais, P.,
9	Sitch, S. and Prentice, I. C.: A dynamic global vegetation model for studies of the coupled
10	atmosphere-biosphere system, Global Biogeochem. Cycles, 19, 33 PP.,
11	doi:200510.1029/2003GB002199, 2005.
12	Langmann, B., Duncan, B., Textor, C., Trentmann, J. and van der Werf, G. R.: Vegetation fire
13	emissions and their impact on air pollution and climate, Atmospheric Environment, 43(1),
14	107-116, doi:10.1016/j.atmosenv.2008.09.047, 2009.
15	Lasslop, G., Thonicke, K. and Kloster, S.: SPITFIRE within the MPI Earth system model: Model
16	development and evaluation, J. Adv. Model. Earth Syst., (06), n/a-n/a,
17	doi:10.1002/2013MS000284, 2014.
18	Van Leeuwen, T. T., van der Werf, G. R., Hoffmann, A. A., Detmers, R. G., Rücker, G., French, N.
19	H. F., Archibald, S., Carvalho Jr., J. A., Cook, G. D., de Groot, W. J., Hély, C., Kasischke, E. S.,
20	Kloster, S., McCarty, J. L., Pettinari, M. L., Savadogo, P., Alvarado, E. C., Boschetti, L.,
21	Manuri, S., Meyer, C. P., Siegert, F., Trollope, L. A. and Trollope, W. S. W.: Biomass burning
22	fuel consumption rates: a field measurement database, Biogeosciences Discuss., 11(6),
23	8115–8180, doi:10.5194/bgd-11-8115-2014, 2014.
24	Li, F., Levis, S. and Ward, D. S.: Quantifying the role of fire in the Earth system – Part 1: Improved
25	global fire modeling in the Community Earth System Model (CESM1), Biogeosciences, 10(4),
26	2293–2314, doi:10.5194/bg-10-2293-2013, 2013.
27	Li, F., Zeng, X. D. and Levis, S.: A process-based fire parameterization of intermediate complexity
28	in a Dynamic Global Vegetation Model, Biogeosciences, 9(7), 2761–2780,
29	doi:10.5194/bg-9-2761-2012, 2012.
30	Liu, H. and Randerson, J. T.: Interannual variability of surface energy exchange depends on stand
31	age in a boreal forest fire chronosequence, J. Geophys. Res., 113(G1), G01006,
32	doi:10.1029/2007JG000483, 2008.
33	Liu, H., Randerson, J. T., Lindfors, J. and Chapin, F. S.: Changes in the surface energy budget after
34	fire in boreal ecosystems of interior Alaska: An annual perspective, J. Geophys. Res., 110(D13),
35	D13101, doi:10.1029/2004JD005158., 2005.
36	Magi, B. I., Rabin, S., Shevliakova, E. and Pacala, S.: Separating agricultural and non-agricultural
37	fire seasonality at regional scales, Biogeosciences, 9(8), 3003-3012,
38	doi:10.5194/bg-9-3003-2012, 2012.
39	Marlier, M. E., DeFries, R. S., Voulgarakis, A., Kinney, P. L., Randerson, J. T., Shindell, D. T.,
40	Chen, Y. and Faluvegi, G.: El Nino and health risks from landscape fire emissions in southeast
41	Asia, Nature Clim. Change, 3(2), 131–136, doi:10.1038/nclimate1658, 2013.
42	Mouillot, F. and Field, C. B.: Fire history and the global carbon budget: a 1 degrees x 1 degrees fire

- 2 doi:10.1111/j.1365-2486.2005.00920.x, 2005.
- Mu, M., Randerson, J. T., van der Werf, G. R., Giglio, L., Kasibhatla, P., Morton, D., Collatz, G. J.,
 DeFries, R. S., Hyer, E. J., Prins, E. M., Griffith, D. W. T., Wunch, D., Toon, G. C., Sherlock, V.
 and Wennberg, P. O.: Daily and 3-hourly variability in global fire emissions and consequences
 for atmospheric model predictions of carbon monoxide, J. Geophys. Res., 116(D24), D24303,
 doi:10.1029/2011JD016245, 2011.
- 8 Le Page, Y., Morton, D., Bond-Lamberty, B., Pereira, J. M. C. and Hurtt, G.: HESFIRE: an explicit
 9 fire model for projections in the coupled Human–Earth System, Biogeosciences Discuss., 11(7),
 10 10779–10826, doi:10.5194/bgd-11-10779-2014, 2014.
- Pausas, J. G. and Keeley, J. E.: A Burning Story: The Role of Fire in the History of Life, BioScience,
 59(7), 593–601, doi:10.1525/bio.2009.59.7.10, 2009.
- Pechony, O. and Shindell, D. T.: Fire parameterization on a global scale, Journal of Geophysical
 Research: Atmospheres, 114(D16), n/a–n/a, doi:10.1029/2009JD011927, 2009.
- Peterson, D. L. and Ryan, K. C.: Modeling postfire conifer mortality for long-range planning,
 Environmental Management, 10(6), 797–808, doi:10.1007/BF01867732, 1986.
- 17 Pfeiffer, M., Spessa, A. and Kaplan, J. O.: A model for global biomass burning in preindustrial time:
- 18 LPJ-LMfire (v1.0), Geosci. Model Dev., 6(3), 643–685, doi:10.5194/gmd-6-643-2013, 2013.
- Plummer, S., Arino, O., Ranera, F., Tansey, K., Chen, J., Dedieu, G., Eva, H., Piccolini, I., Borstlap,
 G., Beusen, B., Fierens, F., Heyns, W., Benedetti, R., Lacaze, R., Garrigues, S., Quaife, T., De
- 21 Kauwe, M., Quegan, S., Raupach, M., Briggs, P., Poulter, B., Bondeau, A., Rayner, P., Schultz,
- M., Gobron, N. and McCallum, I.: An update on the Globcarbon initiative: multi-sensor
 estimation of global biophysical products for global terrestrial carbon studies., [online]
- Available from: http://eprints.ucl.ac.uk/179082/ (Accessed 30 June 2011), 2007.
- Prentice, I. C., Kelley, D. I., Foster, P. N., Friedlingstein, P., Harrison, S. P. and Bartlein, P. J.:
 Modeling fire and the terrestrial carbon balance, Global Biogeochem. Cycles, 25, GB3005,
 doi:201110.1029/2010GB003906, 2011.
- Pueyo, S.: Self-Organised Criticality and the Response of Wildland Fires to Climate Change,
 Climatic Change, 82(1-2), 131–161, doi:10.1007/s10584-006-9134-2, 2007.
- Randerson, J. T., Chen, Y., van der Werf, G. R., Rogers, B. M. and Morton, D. C.: Global burned
 area and biomass burning emissions from small fires, Journal of Geophysical Research:
 Biogeosciences, 117(G4), G04012, doi:10.1029/2012JG002128, 2012.
- Randerson, J. T., Liu, H., Flanner, M. G., Chambers, S. D., Jin, Y., Hess, P. G., Pfister, G., Mack, M.
 C., Treseder, K. K., Welp, L. R., Chapin, F. S., Harden, J. W., Goulden, M. L., Lyons, E., Neff,
 J. C., Schuur, E. a. G. and Zender, C. S.: The Impact of Boreal Forest Fire on Climate Warming,
- 36 Science, 314(5802), 1130–1132, doi:10.1126/science.1132075, 2006.
- Richardson, L. A., Champ, P. A. and Loomis, J. B.: The hidden cost of wildfires: Economic
 valuation of health effects of wildfire smoke exposure in Southern California, Journal of Forest
- 39 Economics, 18(1), 14–35, doi:10.1016/j.jfe.2011.05.002, 2012.
- Rogers, B. M., Randerson, J. T. and Bonan, G. B.: High-latitude cooling associated with landscape
 changes from North American boreal forest fires, Biogeosciences, 10(2), 699–718,
- 42 doi:10.5194/bg-10-699-2013, 2013.

¹ history reconstruction for the 20th century, Global Change Biology, 11, 398–420,

1	Rothermel, R. C.: A mathematical model for predicting fire spread in wildland fuels, Res. Pap.
2	INT-115. Ogden, UT: U.S. Department of Agriculture, Intermountain Forest and Range
3	Experiment Station. 40 p. [online] Available from: internal-pdf://A mathematical model for
4	predicting fire spread in wildland fuels-0846678785/A mathematical model for predicting fire
5	spread in wildland fuels.pdf, 1972.
6	Schoennagel, T., Smithwick, E. A. H. and Turner, M. G.: Landscape heterogeneity following large
7	fires: insights from Yellowstone National Park, USA, Int. J. Wildland Fire, 17(6), 742-753,
8	2008.
9	Stocks, B. J., Mason, J. A., Todd, J. B., Bosch, E. M., Wotton, B. M., Amiro, B. D., Flannigan, M.
10	D., Hirsch, K. G., Logan, K. A., Martell, D. L. and Skinner, W. R.: Large forest fires in Canada,
11	1959–1997, Journal of Geophysical Research: Atmospheres, 107(D1), 8149,
12	doi:10.1029/2001JD000484, 2002.
13	Tansey, K., Grégoire, JM., Defourny, P., Leigh, R., Pekel, JF., van Bogaert, E. and Bartholomé,
14	E.: A new, global, multi-annual (2000–2007) burnt area product at 1 km resolution, Geophys.
15	Res. Lett., 35(1), doi:10.1029/2007GL031567, 2008.
16	Thonicke, K., Spessa, A., Prentice, I. C., Harrison, S. P., Dong, L. and Carmona-Moreno, C.: The
17	influence of vegetation, fire spread and fire behaviour on biomass burning and trace gas
18	emissions: results from a process-based model, Biogeosciences, 7(6), 1991–2011, 2010.
19	Thonicke, K., Venevsky, S., Sitch, S. and Cramer, W.: The role of fire disturbance for global
20	vegetation dynamics: Coupling fire into a Dynamic Global Vegetation Model, Global Ecology
21	and Biogeography, 10, 661–677, 2001.
22	Tosca, M. G., Randerson, J. T. and Zender, C. S.: Global impact of smoke aerosols from landscape
23	fires on climate and the Hadley circulation, Atmos. Chem. Phys., 13(10), 5227–5241,
24	doi:10.5194/acp-13-5227-2013, 2013.
25	Turner, M. G., Hargrove, W. W., Gardner, R. H. and Romme, W. H.: Effects of fire on landscape
26	heterogeneity in Yellowstone National Park, Wyoming, Journal of Vegetation Science, 5(5),
27	731–742, doi:10.2307/3235886, 1994.
28	Venevsky, S., Thonicke, K., Sitch, S. and Cramer, W.: Simulating fire regimes in human-dominated
29 20	ecosystems: Iberian Peninsula case study, Global Change Biology, 8(10), 984–998,
30 21	doi:10.1046/j.1365-2486.2002.00528.x, 2002.
31	Ward, D. S., Kloster, S., Mahowald, N. M., Rogers, B. M., Randerson, J. T. and Hess, P. G.: The
32 33	changing radiative forcing of fires: global model estimates for past, present and future,
33 34	Atmospheric Chemistry and Physics Discussions, 12(4), 10535–10621,
34 35	doi:10.5194/acpd-12-10535-2012, 2012.
36	Van der Werf, G. R., Dempewolf, J., Trigg, S. N., Randerson, J. T., Kasibhatla, P. S., Giglio, L., Murdiyarso, D., Peters, W., Morton, D. C., Collatz, G. J., Dolman, A. J. and DeFries, R. S.:
37	Climate regulation of fire emissions and deforestation in equatorial Asia, Proc. Natl. Acad. Sci.
38	U.S.A., 105(51), 20350–20355, doi:10.1073/pnas.0803375105, 2008a.
39	Van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M., Kasibhatla, P. S., Morton,
40	D. C., DeFries, R. S., Jin, Y. and van Leeuwen, T. T.: Global fire emissions and the contribution
41	of deforestation, savanna, forest, agricultural, and peat fires (1997–2009), Atmos. Chem. Phys.,
42	10(23), 11707–11735, doi:10.5194/acp-10-11707-2010, 2010.
. –	10(20), 11, 07 11, 00, 00, 10, 017 halp 10 11707 2010, 2010.
1	Van der Werf, G. R., Randerson, J. T., Giglio, L., Gobron, N. and Dolman, A. J.: Climate controls
----	---
2	on the variability of fires in the tropics and subtropics, Global Biogeochemical Cycles, 22(3),
3	GB3028, doi:10.1029/2007GB003122, 2008b.
4	Westerling, A. L., Hidalgo, H. G., Cayan, D. R. and Swetnam, T. W.: Warming and Earlier Spring
5	Increase Western U.S. Forest Wildfire Activity, Science, 313, 940-943, 2006.
6	Westerling, A. L., Turner, M. G., Smithwick, E. A. H., Romme, W. H. and Ryan, M. G.: Continued
7	warming could transform Greater Yellowstone fire regimes by mid-21st century, PNAS,
8	201110199, doi:10.1073/pnas.1110199108, 2011.
9	Wilson, R. A. J.: A reexamination of fire spread in free-burning porous fuel beds [Wildland fuels,
10	forest fire management, model], USDA Forest Service Research Paper INT (USA) [online]
11	Available from:
12	http://agris.fao.org/agris-search/search.do?f=1983/US/US83048.xml;US8236661 (Accessed
13	15 February 2014), 1982.
14	Yang, J., Tian, H., Tao, B., Ren, W., Kush, J., Liu, Y. and Wang, Y.: Spatial and temporal patterns
15	of global burned area in response to anthropogenic and environmental factors: Reconstructing
16	global fire history for the 20th and early 21st centuries, Journal of Geophysical Research:
17	Biogeosciences, n/a-n/a, doi:10.1002/2013JG002532, 2014.
18	Zhang, X., Kondragunta, S., Ram, J., Schmidt, C. and Huang, HC.: Near-real-time global biomass
19	burning emissions product from geostationary satellite constellation, Journal of Geophysical
20	Research: Atmospheres, 117(D14), D14201, doi:10.1029/2012JD017459, 2012.

21 Figures and Tables

Fig. 1 Surface-fuel combustion completeness as a function of fuel wetness in (a) original SPITFIRE model by Thonicke et al., (2010) and (b) as modified in the present study, , taking the tropical forests as an example. The combustion of live grass biomass is assumed to follow that of 1-h dead fuel.

26

27 Fig. 2 The regional breakup of the globe according to (a) the GFED3.1 dataset: BONA, Boreal 28 North America; TENA, Temperate North America; CEAM, Central America; NHSA, Northern 29 Hemisphere South America; SHSA, Southern Hemisphere South America; EURO, Europe; MIDE, 30 Middle East; NHAF, Northern Hemisphere Africa; SHAF, Southern Hemisphere Africa; BOAS, 31 Boreal Asia; CEAS, Central Asia; SEAS, Southeast Asia; EQAS, Equatorial Asia; AUST, 32 Australia and New Zealand. And (b) Mouillot and Field (2005): (1) Australia and New Zealand; (2) 33 Boreal North America; (3) Boreal Russia; (4) India; (5) South East Asia; (6) Central Asia; (7) USA 34 (West Mississippi); (8) USA (East Mississippi); (9) East Asia; (10) Middle East and Northern

1 Africa; (11) Africa (sub-Saharian); (12) Central South America; (13) Southern South America; (14)

2 Europe.

3

Fig. 3 Schematic diagram showing the concatenation of fires within consecutive days into multi-day fire patches. An example is given for a 0.5° grid cell in northern Africa for the fire season from October 2001 to April 2002. Upper panel: fires in different consecutive-day periods as simulated by the model. Lower panel: zooming for the period of Day 740–766 since 2000-01-01. Ten different fire patches were extracted from fires in this period. The number of fire patches and their persistence time were indicated by the numbers and arrows in red in the subplot b, respectively. Refer to Sect. 2.5.2 for detailed explanations.

11

12 Fig. 4 Mean annual burned fraction (in percentage) over 2001-2006 (a) as simulated by 13 ORCHIDEE, and by the satellite-derived burned area datasets: (b) GFED3.1, (c) L3JRC and (d) GLOBCARBON. The subplot (e) shows for each grid cell the quality flag of 14 15 ORCHIDEE-simulated burned fraction in comparison with observation datasets. ORC-err-burn, 16 where ORCHIDEE shows burning but the other three observation datasets do not; 17 ORC-err-noburn, where at least two of the three observation datasets do show burning, but 18 ORCHIDEE does not; ORC-min, where ORCHIDEE simulates lower burned fraction than the 19 other three datasets; ORC-max, where ORCHIDEE simulates higher burned fraction; ORC-good, 20 where ORCHIDEE-simulated burned fraction falls within the range given by the three observation 21 datasets. When calculating the minimum and maximum burned fraction of the observation datasets, 22 an arbitrary tolerance margin of 25% was applied around the min/max value to take into account 23 the observation uncertainty.

24

Fig. 5 (a) Latitudinal distribution of burned area (Mha yr⁻¹) according to GFED3.1 (blue), ORCHIDEE (thick black), GLOBCARBON (orange) and L3JRC (green). Data are shown for the mean annual value for 2001–2006. (b) Annual burned area time series for different datasets.

28

Fig. 6 Annual burned area by ORCHIDEE (grey) and GFED3.1 data (black) for 1997–2009 for the
14 GFED regions.

1	
2	Fig. 7 Burned fraction distribution as a function of annual precipitation according to: the model
3	simulation (black), GFED3.1 (blue), GLOBCARBON (orange) and L3JRC (green) for the tropical
4	and subtropical regions (S35–N35). The annual precipitation data are from CRU data and binned
5	in 200-mm intervals.
6	
7	Fig. 8 Spatial pattern of the peak fire month by (a) ORCHIDEE and (b) GFED3.1 data over
8	1997–2009. Only grid cells with fire collocated in both datasets are shown.
9	
10	Fig. 9 Fire season length (months) by (a) Archibald et al. (2013) derived from GFED3.1 data, and
11	(b) ORCHIDEE simulation for 1997–2009.
12	
13	Fig. 10 The annual burned area for 1901–2009 as simulated by ORCHIDEE (grey bar), reported
14	by the Mouillot data (Mouillot and Field, 2005, black bar), and by GFED3.1 data (dashed white
15	bar). Data are shown for the mean values over each decade for 1901-2000, and for 2001-2005
16	(2000sA) and 2006–2009 (2000sB). Refer to Sect. 2.4.1 for the correction of the Mouillot data by
17	using GFED3.1 data.
18	
19	Fig. 11 Fire size distribution as simulated by the model and derived from (a) fire agency data for
20	US Alaska and Canada, and (b) MODIS 500-m burned area data by Archibald et al. (2010) for
21	southern Africa. The horizontal axis indicates fire size (ha) and the vertical axis indicates the
22	corresponding number of fires (in units of ha ⁻¹ yr ⁻¹) for the given fire size. (c) The fire patch size
23	and corresponding mean fire patch length (in unit of days) by the model simulation and Canadian
24	fire agency data (using only the fire patches for which fire report and out date are available).
25	
26	Fig. 12 Fire size and the corresponding cumulative fraction of the total burned area by fires below
27	a given fire size for (a) Canada & US Alaska, and (b) southern Africa. Data are shown for a series
28	of equally distanced 10th quantile fire sizes. Numbers in the curves show the location of every
29	10th quantile fire size from 0th quantile (the minimum fire size) to 100th quantile (the maximum
30	fire size).



8 Fig. 1



- 9
- 10 Fig. 2



12 Fig. 3







3 Fig. 5



5 Fig. 6









3 Fig. 9











3

- 4
- 5

6 Table 1 Mean annual burned area (Mha yr^{-1}) for 2001–2006 for different ORCHIDEE simulation

7 quality flags as shown in Fig. 4.

	ORCHIDEE	GFED3.1	GLOBCARBON	L3JRC
ORC-err-burn	29	-	-	-
ORC-err-noburn	-	27	57	92
ORC-good	93	135	73	96
ORC-max	194	32	23	30
ORC-min	30	150	167	167
Global (Total)	346	344	287	384

9 Table 2 Model error characterization in comparison with the GFED3.1 data for 1997–2009. E_{BA} ,

10 the model error of mean annual burned area in relative to the GFED3.1 data; $S_{interannual}$, the

11 correlation coefficient of linearly detrended annual simulated and GFED3.1 burned area series;

⁸

Region	$E_{BA} \\$	$S_{\text{interannual}}$	$\mathbf{S}_{\text{season}}^{*}$	Burned	Percentage of	Percentage of	Peak fire
(in				area by	the global	the global	month
Fig.				GFED3.1	total burned	total burned	(GFED3.1,
2a)				(ha yr ⁻¹)	area	area	model)
					(GFED3.1)	(ORCHIDEE)	
BONA	0.6	0.53	0.7	2.1	0.6	1.0	(7,7)
TENA	18.1	0.52	0.75	1.3	0.4	7.4	(8,7)
CEAM	4.4	0.55	0.63	1.2	0.3	1.8	(5,4)
NHSA	3.2	-0.12	0.96	2.1	0.6	2.6	(2,2)
SHSA	1.8	0.41	0.71	19.2	5.5	15.7	(8,8)
EURO	4.9	0.01	0.86	0.4	0.1	0.7	(8,8)
MIDE	41.9	0.22	0.73	0.4	0.1	5.6	(8,6)
NHAF	-0.3	0.01	<u>0.59</u>	125.1	35.8	24.9	(12,12)
SHAF	-0.6	0	0.68	123.2	35.2	14.4	(8,6)
BOAS	-0.1	0.43	<u>0.55</u>	5.6	1.6	1.4	(7,7)
CEAS	3.8	0.08	0.75	10.5	3.0	14.9	(8,7)
SEAS	0.5	-0.12	<u>0.52</u>	4.7	1.3	2.0	(3,4)
EQAS	-0.8	0.97	0.73	1.7	0.5	0.1	(9,9)
AUST	-0.5	0.37	0.75	52.4	15.0	7.5	(10,11)

1 S_{season}, the seasonal similarity of burned area by the model and the GFED3.1 data (see Sect. 2.5

2 for the definition of each metrics).

3 * A bootstrapping method was used to derive a probability distribution function of S_{season} by randomly sampling from the normalized monthly burned area of GFED3.1 and the model for 100,000 times. The underlined number indicates that the S_{season} is not significantly different from a randomized monthly distribution of burned area at a significant level of 0.05.