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a global  
nonhydrostatic  
model**

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# Application of a global nonhydrostatic model with a stretched-grid system to regional aerosol simulations around Japan

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## Abstract

An aerosol-coupled global nonhydrostatic model with a stretched-grid system has been developed. Circulations over the global and target domains are simulated with a single model, which includes fine meshes covering the target region to calculate meso-scale circulations. The stretched global model involves relatively low computational costs to simulate atmospheric aerosols with fine horizontal resolutions compared with a global uniform nonhydrostatic model. As opposed to general regional models, neither a nesting technique nor boundary conditions are required. In this study, we developed a new air-quality model for the simulation of areas surrounding Tokyo, Japan, with a maximum horizontal resolution of approximately 10 km. We determined that this model was capable of simulating meteorological fields and anthropogenic primary particles, e.g., elemental carbon, and secondary particles, such as sulfate, with comparable results to those found with in-situ measurements and with other regional models. By combining the meteorological fields obtained from an atmosphere-ocean coupled model, we also applied the new model to a climate scenario experiment of PM<sub>2.5</sub> (aerosol particles with diameters less than 2.5 μm) over Japan with a high horizontal resolution to assess the public health impact at the prefecture scale.

## 1 Introduction

Aerosols can greatly affect regional air quality and contribute to global climate change (Forster et al., 2007). Recently, transboundary aerosol pollution, whereby regions beyond a given country's borders are affected by the aerosols generated in that country, has been of increasing concern (Ramanathan et al., 2008; Yu et al., 2012). The ongoing rapid economic growth in developing countries has the potential to exacerbate this issue (UNEP and WMO, 2011). Air pollution generated by aerosols is a critical public health issue due to the deleterious effects of these particles on human health (Dockery et al., 1993; Pope et al., 2009). Aerosols, which scatter and absorb solar

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radiation and act as cloud condensation nuclei, can directly and indirectly change the Earth's radiation budget. The majority of aerosols are emitted from localized areas, which are referred to as hotspots, such as megacities and biomass-burning regions, and are spread throughout the world via atmospheric transport (e.g., Ramanathan et al., 2008). Therefore, global aerosol-transport models should consider the important regional-scale characteristics of aerosol hotspots to reliably estimate their impacts on air quality and climate change.

Most existing global aerosol-transport models do not address the spatial variability of aerosols in the vicinity of hotspots due to their coarse horizontal resolution of 100–300 km (Kinne et al., 2006; Textor et al., 2006). In addition, global aerosol-transport models with coarse resolutions frequently adopt a spectral transform method with a hydrostatic approximation to effectively calculate atmospheric dynamics. This spectral transform method is less effective than the grid-point method (Stuhne and Peltier, 1996; Taylor et al., 1997; Randall et al., 2000) for high horizontal resolutions (Tomita et al., 2008). Models that employ the grid-point method flexibly define grid points to enable an adaptive focus on study regions. Thus, global models based on the grid-point method seem most appropriate for use in simulating aerosol transport from hotspots to outflow regions.

For this purpose, we utilized the global Nonhydrostatic Icosahedral Atmospheric Model (NICAM) developed by Tomita and Satoh (2004) and Satoh et al. (2008). NICAM has been employed for the global simulation of atmospheric processes with high-resolution grid spacing, whose size is comparable to the typical deep convective cloud scale. Miura et al. (2007) performed a one-week computation with a horizontal resolution of 3.5 km using the Earth Simulator at the Japan Agency for Marine-Earth Science and Technology (JAMSTEC) to successfully simulate a Madden-Julian Oscillation (MJO) event. Suzuki et al. (2008) implemented an aerosol transport model named the Spectral Radiation-Transport Model for Aerosol Species (SPRINTARS; Takemura et al., 2005) in NICAM (we refer to this aerosol-coupled model as NICAM-SPRINTARS) and performed a one-week simulation with a horizontal resolution of 7 km using the Earth



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the simulations of transboundary air pollution, which is expected to increase in Asia (Takemura, 2012), are potentially superior to those obtained by general regional models. Given the heterogeneous distribution of populations in terms of the geography of megacities, Stretch-NICAM-SPRINTARS enables improved estimates of aerosol impacts on human health for future scenarios on a local scale, for example, within prefectures or municipalities of a country. Populations in megacities, particularly those in Asia, are highly susceptible to air pollution (UNEP and WMO, 2011). To predict the extent to which ambient particulates will affect the population in 2030, we performed a scenario experiment involving PM<sub>2.5</sub> (aerosol particles with diameters less than 2.5 μm) around Japan by forcing Stretch-NICAM-SPRINTARS with meteorological fields obtained by an atmosphere-ocean coupled general circulation model (AOGCM), which is referred to as the Model for Interdisciplinary Research on Climate (MIROC) and was developed by the Atmosphere and Ocean Research Institute at the University of Tokyo (AORI/UT), the National Institute for Environmental Studies (NIES) and JAM-STEAC (Watanabe et al., 2010). Based on the results of this experiment, we estimated human mortality impacts attributable to PM<sub>2.5</sub> exposure under the future scenario of Representative Concentration Pathway (RCP) 4.5 in 2030 and accounting for excess mortality, population distributions, and ambient PM<sub>2.5</sub> changes in a given area.

This paper is organized as follows: Stretch-NICAM-SPRINTARS and the experimental design are described in Sect. 2. In Sect. 3, the model results are validated using in-situ measurements in terms of meteorological fields and aerosol species, especially elemental carbon (EC), sulfate and SO<sub>2</sub>. In Sect. 4, we present the validation of total aerosol amounts, i.e., PM<sub>2.5</sub>, and an application of the proposed Stretch-NICAM-SPRINTARS using the results of the future scenario experiment by “MIROC-AOGCM” for the estimation of health impacts. The conclusions are summarized in Sect. 5.

## 2 Model description

### 2.1 Nonhydrostatic Icosahedral Atmospheric Model (NICAM)

NICAM, which employs an icosahedral grid-point method with a nonhydrostatic equation system (Tomita and Satoh, 2004; Satoh et al., 2008), is run with a maximum horizontal resolution of 3.5 km (Tomita et al., 2005; Miura et al., 2007) and can be applied to a transport model of aerosols and gases as a conventional atmospheric general circulation model (Suzuki et al., 2008; Niwa et al., 2011; Dai et al., 2014). NICAM can also be employed for regional-scale simulations by adopting a stretched-grid system (Tomita, 2008a; Satoh et al., 2010). The stretched icosahedral grid was developed from a general grid transformation method, i.e., the Schmidt transformation method, for a horizontal grid system on a sphere. In the Schmidt transformation, the grid interval on a sphere lacks uniformity with a finer horizontal resolution close to the center of the target region. Tomita (2008a) showed that the Schmidt transformation minimizes potential errors involving the isotropy and homogeneity of the target region. The stretched-grid system can solve the main problems associated with commonly used regional models, which occur from artificial perturbations near boundary areas in cases where meteorological and aerosol fields are prescribed. In addition, the computational cost of the stretched-grid system is substantially lower than that of a global calculation under the same horizontal resolution in the target region. For example, when the globally uniform grid with a maximum horizontal resolution of 10 km is applied to the global simulation, the minimum required theoretical computational cost is 256 times higher than the cost of the stretched-grid system in this study. The model framework of the stretched global model is identical to that of the uniformed global model without special modifications. These advantages can facilitate additional developments by testing a new scheme with minimal computational cost. Compared with general regional models, the stretched-grid system is more suitable for the current study, which aimed to extend its use to the global uniform high-resolution NICAM-SPRINTARS.

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fuel sources is composed of externally mixed particles, whereas other carbonaceous particles are emitted and treated as internal mixtures of EC and OC (EC-OC internal mixture). Biogenic secondary organic aerosols (SOAs) from monoterpenes are treated but are greatly simplified. In addition, anthropogenic SOAs from toluene and xylene are disregarded in this study. The particle size distribution of these particles are assumed to be a logarithmic normal size distribution using a 1-modal approach with dry mode radii of 18, 100, 80 and 69.5 nm, for pure EC, EC-OC internal mixture, biogenic SOA and externally mixed sulfate, respectively (Goto et al., 2011a). The hygroscopicities, densities and refractive indices for the aerosols are set to the same values used by Takemura et al. (2002) and Goto et al. (2011a). The combinations of the pre-calculated cross-sections of the extinction and simulated mixing ratios for each aerosol species provide the simulated aerosol extinction coefficient for each timestep of the model (Takemura et al., 2002). The atmospheric removal of aerosols in SPRINTARS includes wet (due to rainout and washout) and dry (due to turbulence and gravity) deposition processes. In this study, the particle mass concentrations for diameters less than  $2.5\ \mu\text{m}$  (defined as  $\text{PM}_{2.5}$ ) are calculated by summing all carbonaceous, sulfate and ammonium aerosols. Because this model cannot directly predict ammonium compounds, we assumed their concentration as the multiplication of the mass concentration of sulfate by 0.27, which is the molar ratio of ammonium ion to ammonium sulfate. The nitrate concentrations in this study, with the target of summer in Japan, can be disregarded.

### 2.3 Design of the standard experiment

The target period comprises one month in August 2007, in which an intensive measurement of aerosol chemical species was conducted under Project FAMIKA (Hasegawa et al., 2008; Fushimi et al., 2011). The six-hour meteorological fields (wind and temperature) were nudged above a height of 2 km using NCEP-FNL reanalysis data (<http://rda.ucar.edu/datasets/ds083.2/>). The one-hour sea surface temperature was also nudged using the NCEP-FNL data. The initial conditions were prescribed by the



NCEP-FNL data for the meteorological fields and the spinup results of the Stretch-NICAM-SPRINTARS model for the aerosol fields, respectively.

The emission inventories of anthropogenic EC, OC and SO<sub>2</sub> in this experiment were prepared by EAGrid2000 with a horizontal resolution of 1 km over Japan (Kannari et al., 2007), REAS version 2 with a horizontal resolution of 0.25° over Asia (Kurokawa et al., 2013) and the AeroCom inventory with a horizontal resolution of 1° over other areas of the world (Diehl et al., 2012). Because EAGrid2000 does not explicitly estimate EC and OC inventories, we estimated the inventories to be consistent with those from previous studies (Morino et al., 2010a, b; Chatani et al., 2011) by modifying the PM<sub>2.5</sub> inventory of EAGrid2000 using scaling factors of EC/PM<sub>2.5</sub> and OC/PM<sub>2.5</sub> based on sources. These inventories of anthropogenic EC and SO<sub>2</sub> in 2007 are described in Figs. 2a and 3a. The emissions of SO<sub>2</sub> from volcanoes in Japan, such as Miyakejima and Sakura-jima, were obtained from statistical reports (<http://www.seisvol.kishou.go.jp/tokyo/volcano.html>) by the Japan Meteorological Agency (JMA). To calculate the sulfur chemistry in SPRINTARS, the distributions of three hourly averaged monthly oxidants (hydroxyl radicals, ozone and hydrogen peroxide) were derived from a global chemical transport model coupled to MIROC, named MIROC-CHASER (Sudo et al., 2002).

In this study, we focused on the aerosol chemical component of EC as the primary particle and sulfate as the secondary particle. To evaluate the model results over the Kanto region, we used observations of the surface mass concentrations of EC and sulfate in four cities under Project FAMIKA: Maebashi/Gunma (139.10° E, 36.40° N), Kisai/Saitama (139.56° E, 36.09° N), Komae/Tokyo (139.58° E, 35.64° N) and Tsukuba/Ibaraki (140.12° E, 36.05° N). The EC particles in PM<sub>2.5</sub> were collected every six hours with quartz fiber filters and analyzed with the thermal/optical method according to the IMPROVE protocol (Chow et al., 2001). The sulfate particles in PM<sub>2.5</sub> were also collected every six hours with Teflon filters and analyzed by ion chromatography. In addition to the limited FAMIKA dataset, we also utilized measurements taken by the EANET (Acid Deposition Monitoring Network in East Asia;

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http://www.eanet.asia/index.html) to assess the monthly mean concentrations of sulfate and SO<sub>2</sub> at ten sites throughout Japan. To validate the concentration of SO<sub>2</sub> for the Kanto region, we accessed monitoring stations operated by Japanese and local governments.

In the validation of the meteorological fields simulated by Stretch-NICAM-SPRINTARS, we obtained measurements for the meteorological parameters (temperature, relative humidity (RH) and wind) at or near the sites of Project FAMIKA and other cities in the Kanto region: Tsuchiura/Ibaraki (140.20° E, 36.07° N), which is the city nearest to Tsukuba; Yokohama/Kanagawa (139.64° E, 35.45° N); Chiba/Chiba (140.12° E, 35.62° N); Adachi/Tokyo (139.82° E, 35.77° N); and Machida/Tokyo (139.43° E, 35.53° N), which is the city nearest to Komae, as shown in Fig. 1b. For precipitation, we used a measurement taken by the Automated Meteorological Data Acquisition System (AMeDAS) at Yokohama; Chiba; Tsukuba; Tokyo, which is near Adachi; Maebashi; Huchu, which is near Machida; and Konosu, which is near Kisai. To evaluate the spatial patterns of the precipitation obtained by Stretch-NICAM-SPRINTARS, we used the quantities of the monthly mean precipitation around Japan that were derived from the Global Satellite Mapping of Precipitation (GSMaP; Okamoto et al., 2005; Kubota et al., 2007; Aonashi et al., 2009; Ushio et al., 2009) and the forecast Grid Point Value (GPV) processed by the JMA.

To evaluate the quantities of the total aerosol amounts, such as PM<sub>2.5</sub>, we compared the simulated PM<sub>2.5</sub> concentrations with the observations at the FAMIKA sites and other monitoring stations operated by the Japanese and local governments of Kawasaki/Kanagawa, which is the city nearest to Yokohama; Machida/Tokyo; Koutou/Tokyo, which is site nearest to the site of Adachi/Tokyo; Osaka/Osaka (135.53° E, 34.68° N); Amagasaki/Hyogo (135.42° E, 34.72° N); and Nonodake/Miyagi (141.17° E, 38.55° N). The PM<sub>2.5</sub> concentrations were continuously observed using tapered element oscillating microbalance (TEOM).

In Tsukuba and Chiba, light detection and ranging (LIDAR) measurements operated by the NIES of Japan were also available (Sugimoto et al., 2003; Shimizu et al., 2004).

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The LIDAR unit measured vertical profiles of the backscattering intensity at 532 and 1064 nm and the depolarization ratio at 532 nm. The backscattering intensity was converted to the extinction coefficient, and the depolarization ratio distinguished the extinction between spherical and non-spherical particles. In this study, we only used vertical profiles of the extinction for spherical particles. A detailed algorithm was provided by Sugimoto et al. (2003) and Shimizu et al. (2004).

## 2.4 Scenario experiment

In previous future-scenario experiments, general climate models such as MIROC-AOGCM (Watanabe et al., 2010) were incapable of estimating the impacts of PM<sub>2.5</sub> on human health within the prefectures of Japan due to their coarse grid sizes. In addition, the model used in the current study, Stretch-NICAM-SPRINTARS, cannot be used for a long-term simulation. Therefore, we combined Stretch-NICAM-SPRINTARS with MIROC-AOGCM by nudging the meteorological results of MIROC-AOGCM under the scenario experiment of Representative Concentration Pathway (RCP) 4.5 in August 2030. In this study, we selected target results of temperature and precipitation fluxes in the representative years within ten years of 2026–2035.

The emission inventories of anthropogenic EC, OC and SO<sub>2</sub> from anthropogenic sources in the scenario experiment were based on the IPCC inventory, including RCP4.5 with a global horizontal resolution of 0.5° (Lamarque et al., 2010; Moss et al., 2010). However, due to the coarse grid size of the emission inventory around Japan, we multiplied the inventory of EAGrid2000 (Kannari et al., 2007) by scaling factors, which were estimated from the results of IPCC emission inventories for each compound. For the emission inventory, to eliminate an outlier during year 2026–2035 used in the present study, we averaged results from the ten years. Monthly mean oxidant distributions for sulfate formation were also obtained from the MIROC-CHASER model (Sudo et al., 2002) used in the MIROC-AOGCM scenario experiments in the same period. Inventories of the biomass burning over forest fires outside Japan were disregarded in the scenario experiment because a minimal impact on the aerosol distribution around

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Japan during the summer is expected. Figures 2b and 3b also illustrate the emission inventories of EC and SO<sub>2</sub> in 2030 under the RCP4.5 scenario experiment. The quantities of EC and SO<sub>2</sub> emissions in Japan exhibit a decreasing trend from the present to 2030, whereas the quantities of EC and SO<sub>2</sub> emissions in China and Korea exhibit an increasing trend.

To quantify the health impact of exposure to ambient PM<sub>2.5</sub>, we estimated the excess mortality per grid attributable to PM<sub>2.5</sub>, which is denoted by  $D(x)$ , using the following function:

$$D(x) = N(x) \{ \exp(\beta) - 1 \} \{ P(x) - P_0 \}, \quad (1)$$

where  $\beta$  is the linear slope between PM<sub>2.5</sub> and mortality. A similar association was previously employed for the estimation of the excess mortality due to high temperature (Chung et al., 2009). The  $\beta$  value for Japan is set to 0.53 % per 10  $\mu\text{g m}^{-3}$  increase in PM<sub>2.5</sub> (Ueda et al., 2009). The  $P$  value is set to 15  $\mu\text{g m}^{-3}$  based on the environmental quality standards for atmospheric PM<sub>2.5</sub> in Japan because the threshold value cannot be directly derived from the observed dataset.  $P(x)$  is the PM<sub>2.5</sub> mass concentration in units of  $\mu\text{g m}^{-3}$  obtained from Stretch-NICAM-SPRINTARS in one NICAM grid, including the position  $x$ , and  $N(x)$  is the population above 64 yr of age who are at risk due to exposure to air pollutants per grid  $x$  (Ueda et al., 2009). The population distributions were represented by grids with dimensions of 1 km by 1 km. The grid containing the position  $x$  and the excess mortality were estimated in the same horizontal resolution as the population distribution.

### 3 Validation of Stretch-NICAM-SPRINTARS

#### 3.1 Meteorological fields

To evaluate the model performances of the meteorological fields obtained by Stretch-NICAM-SPRINTARS, we used the observed temperature, RH, wind and precipitation

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at each station over the Kanto region shown in Fig. 1b. The results and summary are shown in Figs. 4 to 7 and Table 1. Figure 4 illustrates the temporal variations of temperature at a height of 2 m. The temporal variations in the simulated temperature are generally comparable to those in the observed temperatures with root-mean-square-error (RMSE) values of less than 3 °C, with the exception of the results obtained for Maebashi and Machida. At these two sites, the mean values of the simulated temperatures are lower than those of the observed temperatures by a maximum of 3 °C. The correlation coefficients ( $R$ ) between the simulations and observations range from 0.7–0.8, as shown in Table 1. Figure 5 shows the temporal variations in RH at a height of 2 m. The monthly average temporal variations in the simulated RH are similar to the observations, with the RMSEs in the range of 10–15 %. The  $R$  values of RH between the simulation and observations are approximately 0.6–0.8, as shown in Table 1.

The temporal variation in the wind direction and speed simulated by Stretch-NICAM-SPRINTARS are compared with the observations in Figs. 6 and 7. Near the southern part of the Kanto region (Yokohama, Tsuchiura, Adachi and Machida), with the exception of Chiba, the simulated wind directions are generally comparable to the observations, with a slight overestimation of the simulated wind speed compared with the observations. At these four sites, the  $R$  values and RMSE values range from approximately 0.5–0.6 and from approximately 1.5–2.2  $\text{ms}^{-1}$ , respectively. In Chiba located near the ocean, the  $R$  value of wind speed between the simulation and observations is 0.27, whereas the simulated wind directions generally agree with the observations. Conversely, at Maebashi and Kisai, the daily variations in the simulated wind direction differ significantly from those in the observations, in which the southern winds and northern winds frequently occur during the day and night, respectively, for example, during 5–12 August. The  $R$  values for wind speed between the simulation and observations at these sites are estimated to be approximately 0.3–0.4. The results of the meteorological fields at Maebashi and Kisai, which are surrounded by or are located relatively close to high mountains, indicate that the horizontal resolution of 10 km in this study using Stretch-NICAM-SPRINTARS could not completely resolve the topography.

As a result, it may be inadequate to simulate the wind patterns and diurnal transitions near high mountains.

Figure 8 shows the temporal variations in the amount of precipitation per day at each site. During August 2007 in the Kanto region, the observed precipitation is extremely limited and sometimes localized. The temporal variations in the simulated precipitation are generally similar to those in the observations. However, the precipitation modeled by Stretch-NICAM-SPRINTARS on 19 and 23 August is higher than the observations by more than  $30 \text{ mm day}^{-1}$ . Stretch-NICAM-SPRINTARS does not always capture a sudden shower, as general meteorological models cannot properly simulate this type of precipitation system. This overestimation may have an impact on the monthly mean precipitation shown in Fig. 9, which compares the Stretch-NICAM-SPRINTARS simulated precipitation with the GPV- and GSMaP-derived results. The overestimation of the precipitation obtained by Stretch-NICAM-SPRINTARS compared with the observations is also seen in the Sea of Japan, Kyusyu, and the main island of Japan. All results generally show similar patterns of the occurrence of heavy precipitation in the East China Sea and the Sea of Japan near the Japan coast, especially near Okinawa, the southern part of South Korea and North Korea. Therefore, Stretch-NICAM-SPRINTARS can generally simulate the meteorological fields in the present target regions.

## 3.2 Aerosol fields

### 3.2.1 Evaluation using measurements

Figure 10 illustrates the temporal variations in the surface EC mass concentration obtained by Project FAMIKA at the four stations (Maebashi, Kisai, Komae and Tsukuba). The temporal variation and the average correspond with the observations obtained for Komae, as shown in Fig. 10c. For Tsukuba, which is shown in Fig. 10d, the simulated EC concentrations tend to be underestimated compared with the observed concentrations, especially during the daytime. However, in some instances, these results are comparable with the observations. Conversely, the temporal variation in the simulated

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EC concentrations and the average EC concentrations at Maebashi and Kisai are underestimated compared with the observations by a factor of three to five. At the same sites, simulated sulfur components (sulfate and SO<sub>2</sub>) are compared with the observations in Figs. 11 and 12. The observed SO<sub>2</sub> represents the ensemble results of monitoring stations operated by Japanese and local governments around each FAMIKA site. The averaged differences between the simulated and observed sulfate mass concentrations are within approximately 10 % at Maebashi and Tsukuba, -40 % at Komae and +50 % at Kisai. At all sites, the temporal variations of the simulated sulfate are generally comparable to those of the observations, whereas differences in the sulfate between the simulation and observations are somewhat greater on 7 August in Maebashi and on 6 August in Kisai, Komae and Tsukuba. However, differences between the simulated and observed SO<sub>2</sub> concentrations at all sites are within approximately 30 %. The temporal variations in the simulated SO<sub>2</sub> concur with those in the observations. To assess the performance of Stretch-NICAM-SPRINTARS in simulating the aerosol distribution over Japan, we compared the August averages of the simulated sulfate and SO<sub>2</sub> with the available measurements of EANET (Fig. 13). The results indicate that the simulated sulfate concentrations tend to be underestimated by approximately 40 % compared with the observed sulfate concentrations. However, the correlation between the simulation and observation is adequately acceptable ( $R = 0.79$  or  $R = 0.86$ , with the exception of Hedo). At Hedo located in the southwestern islands of Japan, the overestimation of the simulated precipitation shown in Fig. 9 may cause the underestimation of the simulated sulfate concentrations. The simulated and observed SO<sub>2</sub> concentrations also correlate, with an  $R$  value of 0.95. Figure 14 shows the monthly averaged sulfate and SO<sub>2</sub> in August 2007. The SO<sub>2</sub>, which is a primary product, is localized near the source areas, whereas sulfate, which is as a secondary product, is distributed from the source to the outflow areas. In the Kanto region, for example, sulfate from transboundary and domestic pollution is effectively simulated by Stretch-NICAM-SPRINTARS.

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### 3.2.2 Uncertainty in the simulation

Sensitivity tests were conducted to examine potential uncertainties derived from prescribed datasets related to EC and sulfate for the Stretch-NICAM-SPRINTARS simulations. For the EC sensitivity tests, the emission quantities were set to half and twice of those used in the standard run in this study. The results for the FAMIKA sites are shown in Fig. 15a in which the bars indicate the simulated EC concentrations for both sensitivity tests. For the majority of the sites, with the exception of Komae, the results obtained by the sensitivity experiments remain underestimated compared with the measurements. The underestimation of the EC mass concentrations at Maebashi and Kisai was also shown by the previous studies of Morino et al. (2010a, b) and Shimadera et al. (2013), who calculated EC concentrations using the Community Multiscale Air Quality (CMAQ) driven by the Weather Research and Forecasting (WRF) model named WRF-CMAQ with a horizontal resolution of 5 km. WRF-CMAQ employs an emission inventory that is similar to that in the present study. The difference in the EC concentrations at Maebashi between the present study and the previous studies using WRF-CMAQ is partly caused by the difference in the horizontal resolution, which is most likely critical for properly simulating the air pollution delivered by the meteorological wind fields from the center of the Kanto region. However, Fushimi et al. (2011) and Chatani et al. (2014) suggested that the difference in the EC concentrations between WRF-CMAQ and the measurements is largely attributed to an underestimation of the EC emission inventory, especially open biomass burning from domestic sources. Therefore, the same factor may be applicable to the present results using Stretch-NICAM-SPRINTARS.

Sensitivity experiments of the SO<sub>2</sub> emissions and the prescribed hydroxyl radical used in sulfur chemistry were executed under half and twice the amounts used in the standard experiment. Figure 15b shows that the sensitivity of the hydroxyl radical concentrations to the simulated sulfate concentration is substantially smaller than that to the SO<sub>2</sub> emissions. Compared with the SO<sub>2</sub> emissions used in the standard

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experiment, the doubled amount of SO<sub>2</sub> emissions can overcome the slight underestimation of the simulated sulfate compared with the observations. We also determined that the sensitivities of the other oxidants to the simulated sulfate concentrations were minimal (not shown). These results from the sensitivity experiments indicate that the offline prescribed oxidant used in this study is not as critical to the proper prediction of the sulfate concentrations over the Kanto region as the uncertainty in the quantity of SO<sub>2</sub> emissions. Therefore, we conclude that the simulations of Stretch-NICAM-SPRINTARS are generally successful in simulating the air pollution over Japan and are adequate as a new regional model for simulations over the Kanto region.

## 4 Application of Stretch-NICAM-SPRINTARS

### 4.1 PM<sub>2.5</sub>

Figure 16 shows the temporal variation in the surface PM<sub>2.5</sub> mass concentration at the 11 sites over the Kanto region and in western and northern Japan. At all sites, the temporal variations in the simulated PM<sub>2.5</sub> are generally similar to those in the observed values; however, the simulated PM<sub>2.5</sub> concentrations are underestimated compared with the observations by a factor of two or three at the majority of sites and by approximately a factor of four at Maebashi. In addition to the issue of the poor model performance of the meteorological fields at Maebashi, the underestimation of secondary OC may be a critical issue, as suggested by previous studies (Matsui et al., 2009; Morino et al., 2010c). According to previous studies that employed regional aerosol-transport models (Morino et al., 2010b; Chatani et al., 2011), the underestimation of PM<sub>2.5</sub> is common because the measured concentrations of PM<sub>2.5</sub> include undefined chemical species with mean fractions ranging from approximately 30–50 % in the total PM<sub>2.5</sub> in the summer (datasets from the Tokyo Environment Agency and the Kawasaki Municipal Research Institute for Environmental Protection). Therefore, the undefined

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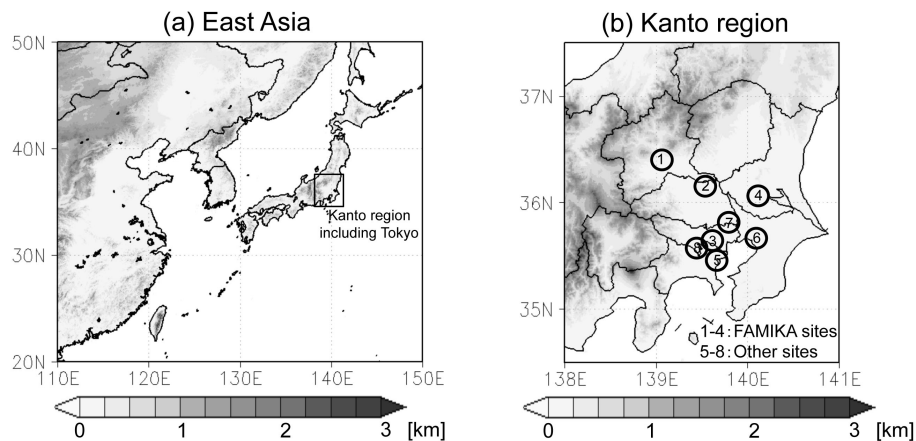
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**Table 1.** Statistical values (mean of the observation and simulation, absolute bias Ba, correlation coefficient  $R$  and root-mean-square-error RMSE) for meteorological fields using the Stretch-NICAM-SPRINTARS simulation and observations at seven sites during the same period, as shown in Figs. 4 to 7.

	Yokohama	Chiba	Tsuchiura	Adachi	Maebashi	Machida	Kisai
Temperature [°C]							
Observation	27.9	30.1	28.1	29.7	29.1	29.1	27.9
Simulation	26.9	28.3	28.3	27.3	25.5	25.9	25.4
Ba	−1.1	−1.8	0.2	−2.3	−3.6	−3.2	−2.6
R	0.71	0.83	0.83	0.79	0.78	0.73	0.79
RMSE	2.2	2.5	2.2	3.2	4.6	4.0	3.2
RH [%]							
Observation	73.5	79.	73.3	75.2	73.7	75.7	71.4
Simulation	83.6	77.5	76.4	77.9	82.7	82.5	82.2
Ba	10.0	−1.5	3.0	2.8	9.0	6.8	10.7
R	0.57	0.65	0.67	0.70	0.70	0.70	0.78
RMSE	13.9	10.6	12.7	10.5	16.1	12.3	14.6
Wind Speed [ $\text{ms}^{-1}$ ]							
Observation	2.9	2.6	1.6	2.6	1.2	2.7	1.9
Simulation	4.2	3.8	3.1	3.4	3.1	3.0	2.8
Ba	1.3	1.1	1.4	0.9	1.9	0.3	0.9
R	0.63	0.34	0.57	0.44	0.11	0.54	0.15
RMSE	2.1	2.3	1.9	1.9	2.5	1.6	2.0

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**Fig. 1.** Topographical maps of **(a)** Japan and **(b)** the Kanto region, including observation sites for the validation of the model. The circles represent eight sites (1. Maebashi/Gunma, 2. Kisai/Saitama, 3. Komae/Tokyo, 4. Tsukuba/Ibaraki, 5. Yokohama/Kanagawa, 6. Chiba/Chiba, 7. Adachi/Tokyo and 8. Machida/Tokyo).

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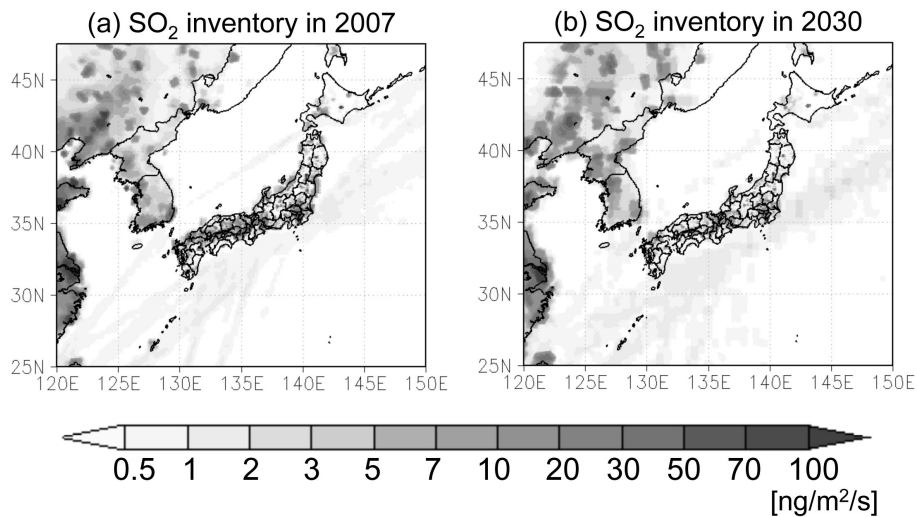
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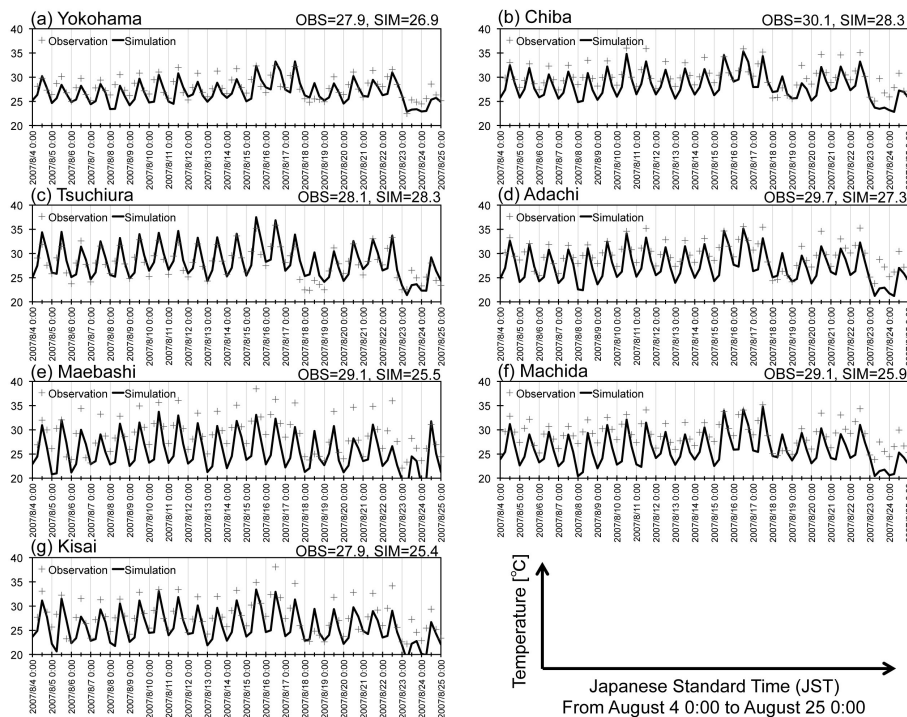
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**Fig. 3.** Same as Fig. 2 but for SO<sub>2</sub> emission inventories.

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**Fig. 4.** Temporal variation in the Stretch-NICAM-simulated and observed air temperature for a height of 2 m at **(a)** Yokohama, **(b)** Chiba, **(c)** Tsuchiura, **(d)** Adachi, **(e)** Maebashi, **(f)** Machida and **(g)** Kisai from 4–24 August 2007. The numbers located in the upper right corner of each panel show the simulated and observed mean values.

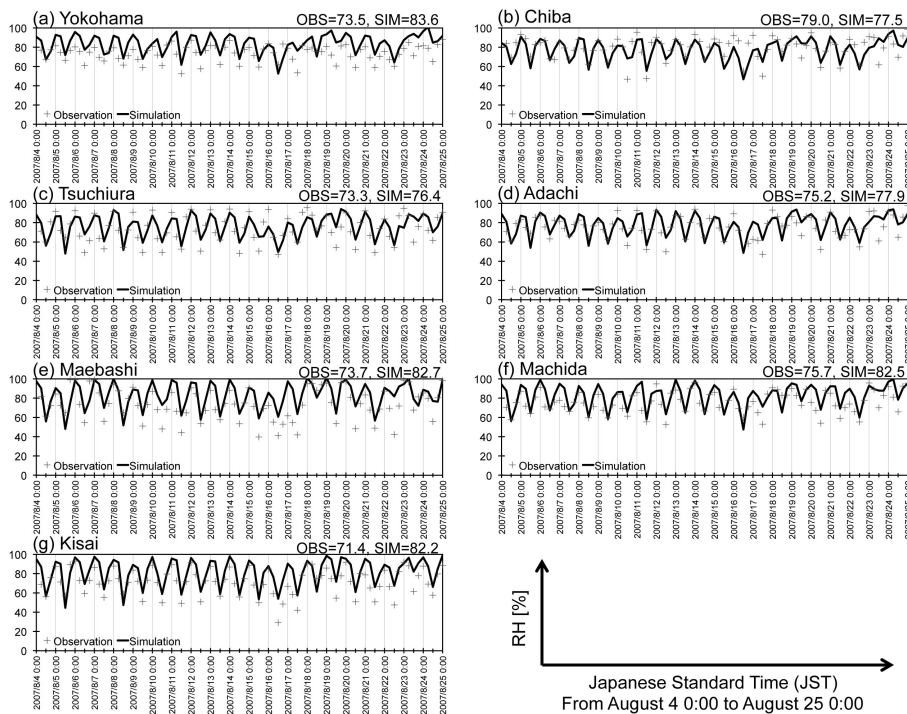


Fig. 5. Same as Fig. 4 but for relative humidity (RH).

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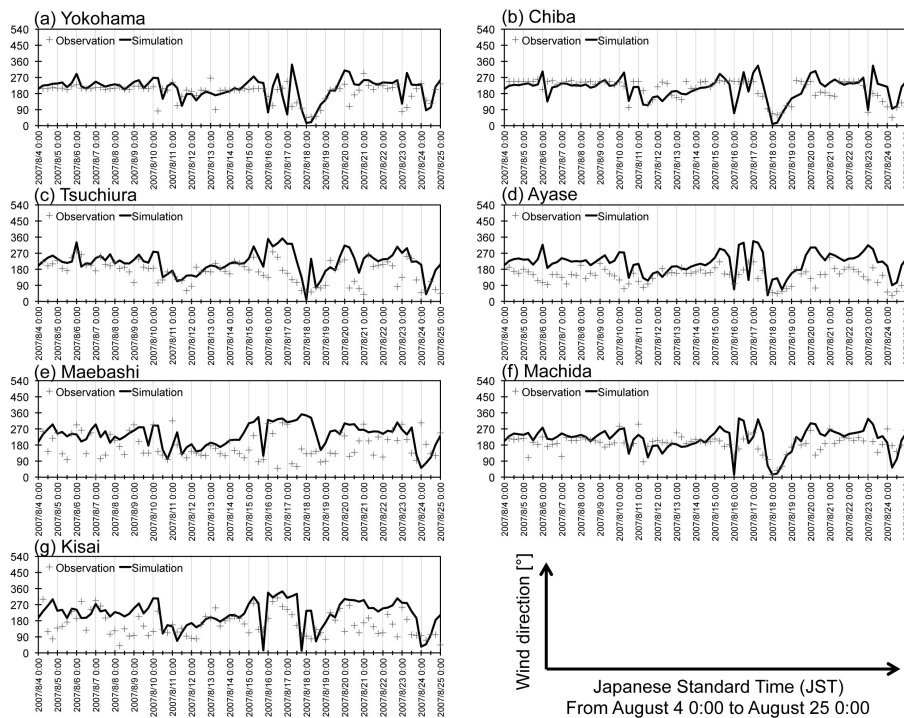


Fig. 6. Same as Fig. 4 but for wind direction.

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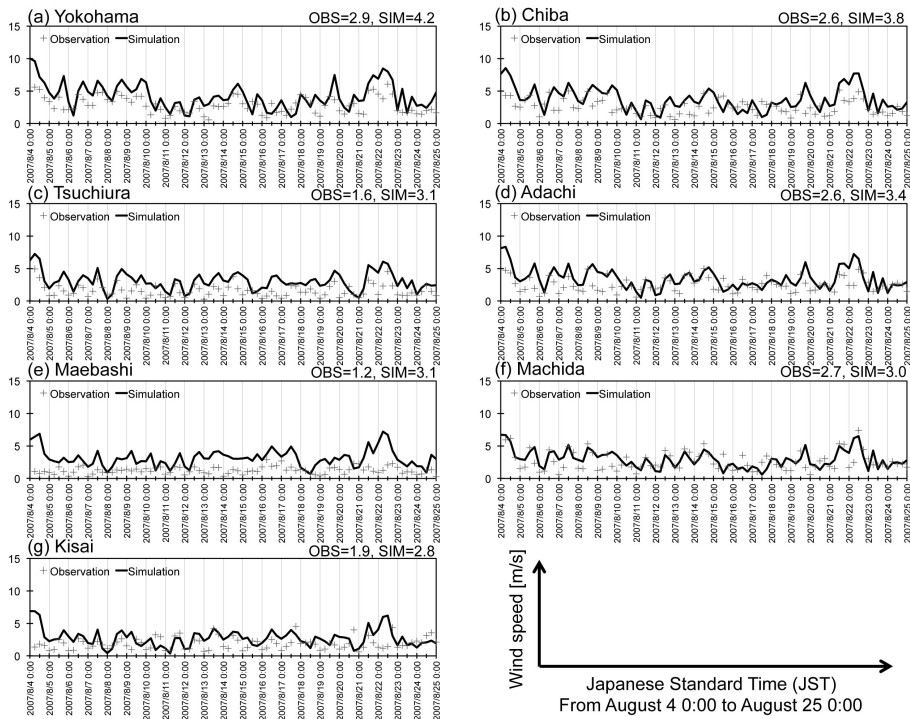


Fig. 7. Same as Fig. 4 but for wind speed.

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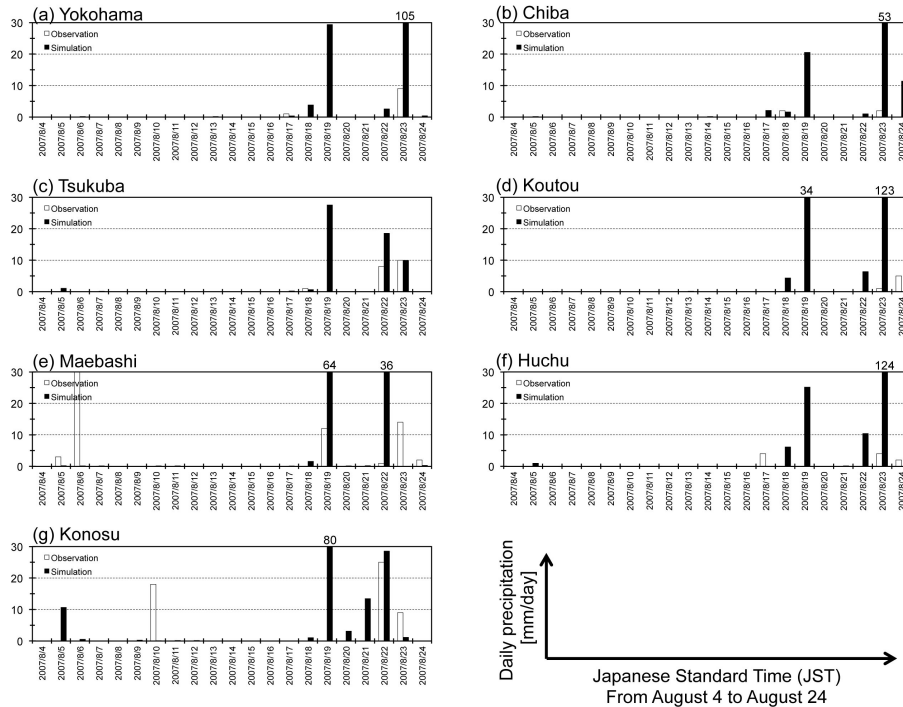
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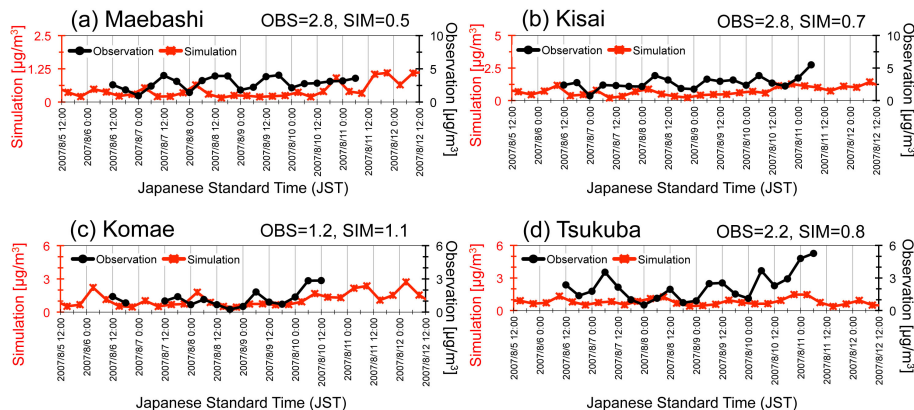


**Fig. 8.** Temporal variation in the simulated Stretch-NICAM-SPRINTARS and observed precipitation amounts at (a) Yokohama, (b) Chiba, (c) Tsukuba, (d) Koutou, (e) Maebashi, (f) Huchu and (g) Konosu from 4–24 August 2007.



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**Fig. 10.** Temporal variation in Stretch-NICAM-SPRINTARS-simulated and observed EC mass concentrations in units of  $\mu\text{g}/\text{m}^3$  near the surface at (a) Maebashi, (b) Kisai, (c) Komae and (d) Tsukuba from 12 p.m. on 5 August 2007, to 12 p.m. on 12 August 2007. The left axis in red represents the simulated values, and the right axis in black represents the observed values.

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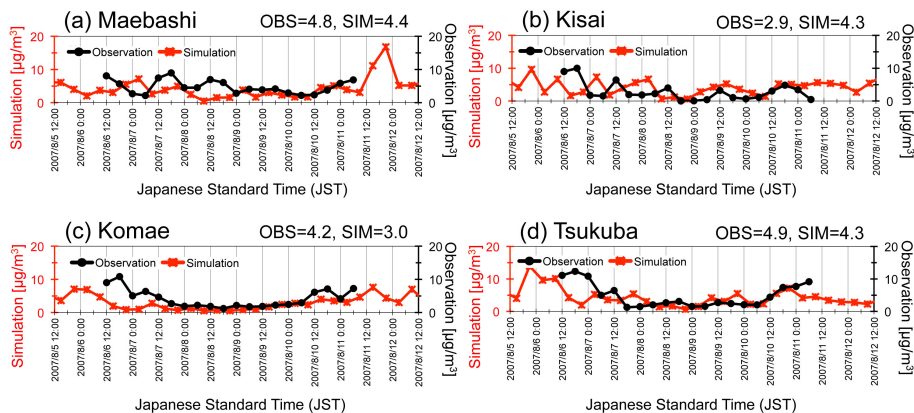


Fig. 11. Same as Fig. 10 but for sulfate.

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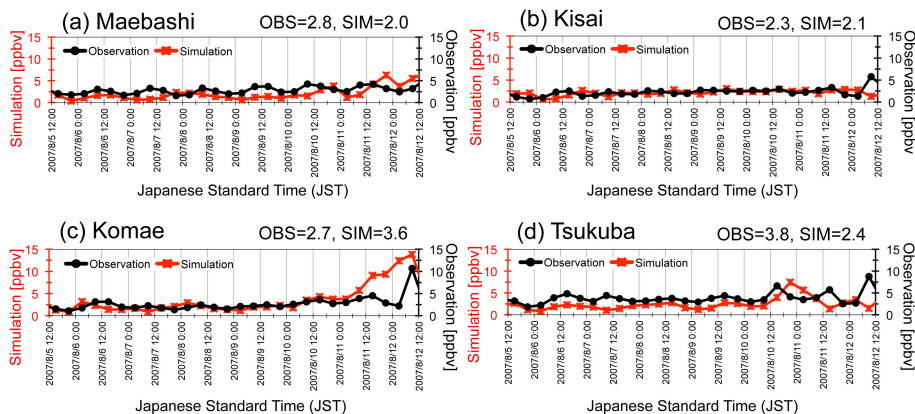


Fig. 12. Same as Fig. 10 but for  $\text{SO}_2$ , in units of ppbv.

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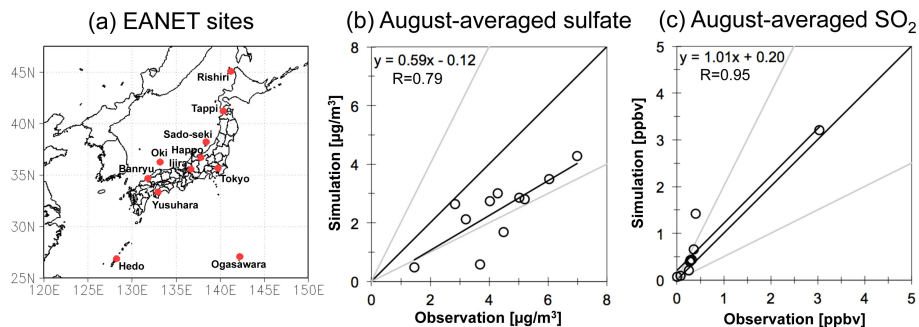
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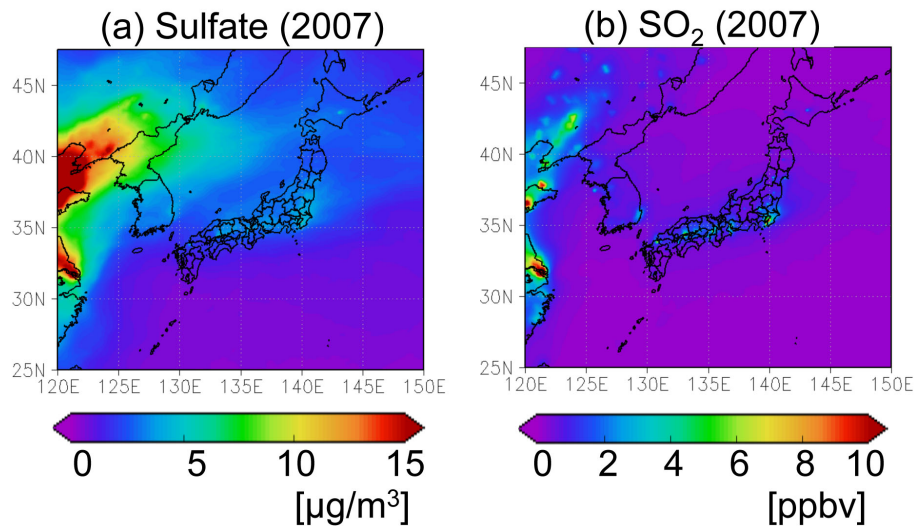
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**Fig. 13.** Comparison of average concentrations in August for (b) sulfate and (c)  $\text{SO}_2$  between the Stretch-NICAM-SPRINTARS simulations and the observations at EANET sites shown in panel (a).

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**Fig. 14.** Average simulated Stretch-NICAM-SPRINTARS concentrations in August for **(a)** sulfate and **(b)**  $\text{SO}_2$  near the surface.

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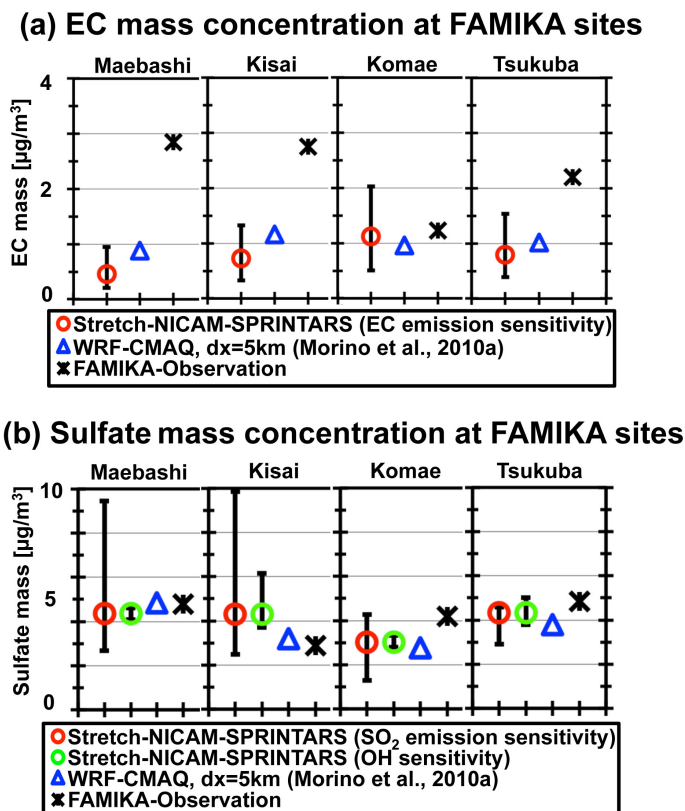
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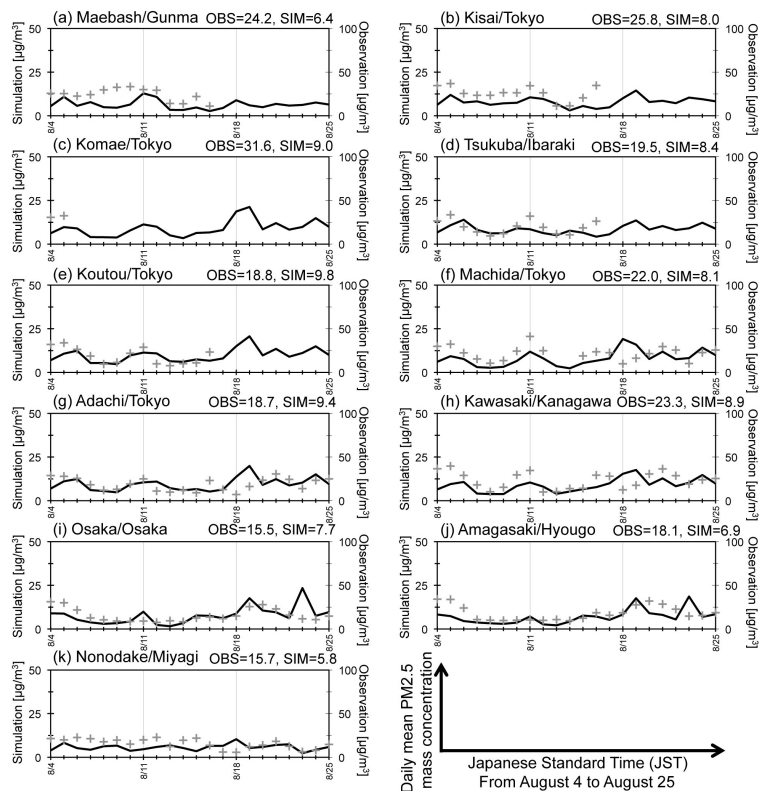
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**Fig. 15. (a)** EC and **(b)** sulfate mass concentrations at the surface of FAMIKA sites using Stretch-NICAM-SPRINTARS results of the sensitivity experiments, WRF-CMAQ results obtained by Morino et al. (2010a) with a horizontal resolution of 5 km and the FAMIKA observation. The value and bar represent mean values from the period shown in Figs. 10 and 11.

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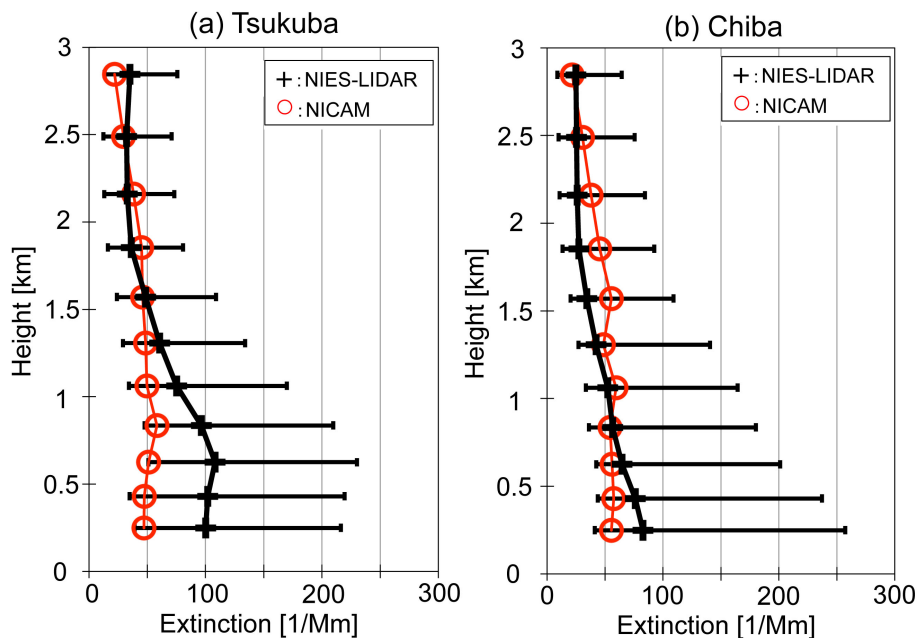


**Fig. 16.** Temporal variation in simulated Stretch-NICAM-SPRINTARS and observed  $\text{PM}_{2.5}$  near the surface at **(a)** Maebashi, **(b)** Kisai, **(c)** Komae, **(d)** Tsukuba, **(e)** Koutou, **(f)** Machida, **(g)** Adachi, **(h)** Kawasaki, **(i)** Osaka, **(j)** Amagasaki and **(k)** Nonodake from 4–25 August 2007. The left axis in red represents the simulated values, and the right axis in black represents the observed values. The numbers located in the upper right corner of each panel show the simulated and observed mean values.

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**Fig. 17.** Average extinction coefficients in August for the spherical particles simulated by Stretch-NICAM-SPRINTARS (shown in red) and the spherical particles observed by the NIES-LIDAR network (shown in black) at **(a)** Tsukuba and **(b)** Chiba, in units of  $\text{Mm}^{-1}$ . The bars represent the 25th and 75th percentiles of the LIDAR observations.

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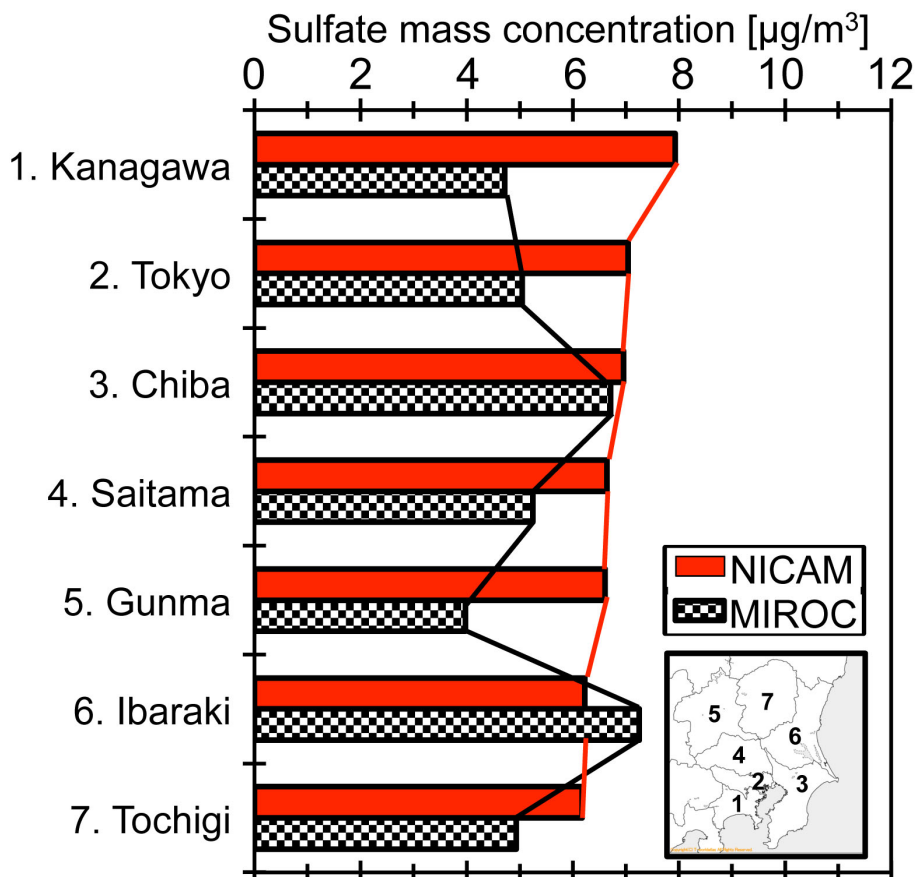
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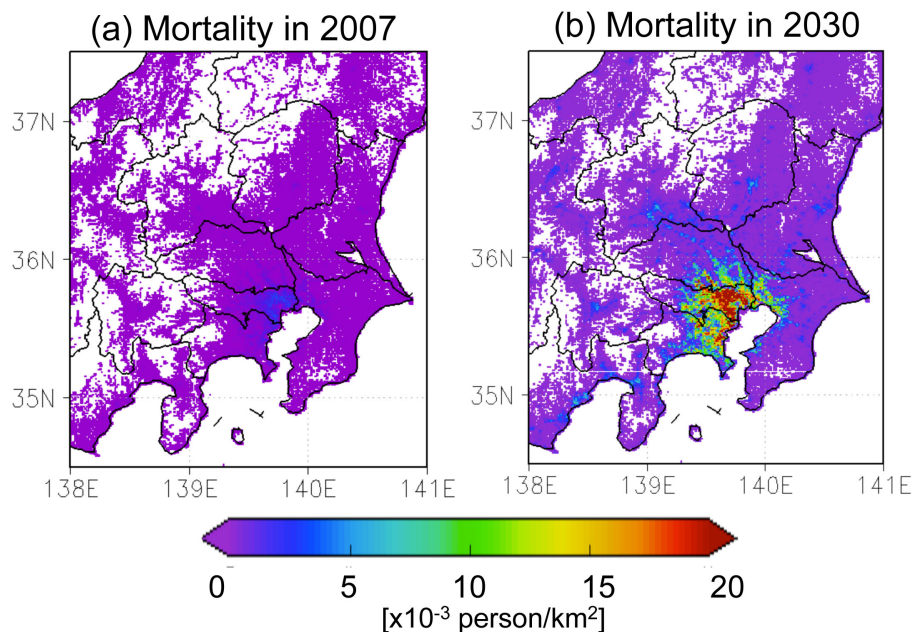


**Fig. 19.** Stretch-NICAM-SPRINTARS- and MIROC-AOGCM-simulated sulfate mass concentrations in each prefecture for August 2030 in the RCP4.5 scenario experiment.



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**Fig. 20.** Examples of the number of deaths caused by  $\text{PM}_{2.5}$  in each grid of 1 km by 1 km for **(a)** the standard experiment in 2007 and **(b)** the scenario experiment of RCP4.5 in 2030 using the Stretch-NICAM-SPRINTARS simulation in the present study.

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