

AttentionFire_v1.0: interpretable machine learning fire model for burned-area predictions over tropics

Fa Li^{1,2}, Qing Zhu¹, William J. Riley¹, Lei Zhao³, Li Xu⁴, Kunxiaojia Yuan^{1,2}, Min Chen⁵, Huayi Wu², Zhipeng Gui⁶, Jianya Gong⁶, and James T. Randerson⁴

¹Climate and Ecosystem Sciences Division, Climate Sciences Department, Lawrence Berkeley National Laboratory, Berkeley, CA, USA

²State Key Laboratory of Information Engineering in Surveying, Mapping and Remote Sensing,

Wuhan University, Wuhan, China

³Department of Civil and Environmental Engineering, University of Illinois Urbana-Champaign, Champaign, IL, USA

⁴Department of Earth System Science, University of California Irvine, Irvine, CA, USA

⁵Department of Forest and Wildlife Ecology, University of Wisconsin-Madison, Madison, WI, USA

⁶School of Remote Sensing and Information Engineering, Wuhan University, Wuhan, China

Correspondence: Qing Zhu (qzhu@lbl.gov)

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Abstract. African and South American (ASA) wildfires account for more than 70% of global burned areas and have strong connection to local climate for sub-seasonal to seasonal wildfire dynamics. However, representation of the wildfire-climate relationship remains challenging due to spatiotemporally heterogenous responses of wildfires to climate variability and human influences. Here, we developed an interpretable machine learning (ML) fire model (AttentionFire_v1.0) to resolve the complex controls of climate and human activities on burned areas and to better predict burned areas over ASA regions. Our ML fire model substantially improved predictability of burned areas for both spatial and temporal dynamics compared with five commonly used machine learning models. More importantly, the model revealed strong time-lagged control from climate wetness on the burned areas. The model also predicted that, under a highemission future climate scenario, the recently observed declines in burned area will reverse in South America in the near future due to climate changes. Our study provides a reliable and interpretable fire model and highlights the importance of lagged wildfire-climate relationships in historical and future predictions.

1 Introduction

Wildfires modify land surface characteristics, such as vegetation composition, soil carbon, surface runoff, and albedo, with significant consequences for regional carbon, water, and energy cycles (Benavides-Solorio and MacDonald, 2001; Shvetsov et al., 2019; Randerson et al., 2006). Over African and South American (ASA) regions, where more than 70 % of global burned area occurs, wildfires emit $\sim 1.4 \,\mathrm{PgC}\,\mathrm{yr}^{-1}$ (~65% of global wildfire emissions; van Der Werf et al., 2017) and dust and aerosols that can alter regional climate through radiative processes (Etminan et al., 2016; Ramanathan et al., 2001; van Der Werf et al., 2017). While greenhouse gas emissions contribute to climate change, other toxic species and airborne particulate matter from wildfires lead to substantial health hazards, including elevated premature mortality (Knorr et al., 2017; Lelieveld et al., 2015). In particular, wildfire particulate matter emissions across tropical regions have exceeded current anthropogenic sources and are predicted to dominate future regional emissions (Knorr et al., 2017).

Although total tropical wildfire-burned area has declined over the past few decades due to climate change and human activities (Andela and Van Der Werf, 2014; Andela et al., 2017), e.g., from increases in population density, cropland fraction, and livestock density, wildfire still plays a significant role in mediating surface climate (Xu et al., 2020), biogeochemical cycles, and human health (Andela et al., 2017). Further, 21st century projections of increases in temperature, regional drought (Dai, 2013; Taufik et al., 2017), and precipitation variations may outweigh these direct human impacts and result in unprecedentedly fire-prone environments over a large fraction of Africa (Van Der Werf et al., 2008; Andela and Van Der Werf, 2014; Archibald et al., 2009) and South America (Pechony and Shindell, 2010; Malhi et al., 2008). These factors highlight the need for better understanding, prediction, and management of these critical fire regions to minimize economic losses, human health hazards, and natural ecosystem degradation. Therefore, improved understanding and accurate prediction of wildfire activity is increasingly important for effective fire management and sustainable decision making.

Climate is acknowledged as one of the most dominant controllers of ASA wildfires (Chen et al., 2011; Andela et al., 2017). For example, precipitation variations contribute substantially to burned-area patterns in southern and northern Africa (Andela and Van Der Werf, 2014; Archibald et al., 2009) and are also closely linked to wildfire spatiotemporal dynamics in South America (Chen et al., 2011; Van Der Werf et al., 2008; Malhi et al., 2008). More importantly, the strong controls of climate on wildfires often show time lags, and the time delay can be on the order of multiple months (Van Der Werf et al., 2008; Andela and Van Der Werf, 2014). Meanwhile, ocean dynamics (e.g., El Niño-Southern Oscillation, ENSO) may also exert considerable influences on ASA wildfires through influencing wet- and wet-to-dry-season climate and fuel conditions (Yu et al., 2020; Chen et al., 2016; Andela and Van Der Werf, 2014; Chen et al., 2011, 2017). The time lags between ocean dynamics and wildfires can be even longer than that between climate and wildfires (Chen et al., 2020), which enables wildfire predictions ahead of fire season (Chen et al., 2011, 2016, 2020; Turco et al., 2018). The spatiotemporal responses of wildfires to climate changes are complicated by non-linear interactions among climate, vegetation, and human activities (Van Der Werf et al., 2008; Andela et al., 2017). In more xeric subtropical regions, increasing precipitation during the wet season can be the dominant controller on increasing wildfire during the following dry season (through regulation of fuel availability and fuel spatial structures) (Van Der Werf et al., 2008; Littell et al., 2009; Archibald et al., 2009). In contrast, increasing precipitation in more mesic regions results in excessive fuel moisture, thereby becoming the main limitation of dry-season wildfires (i.e., opposite fire trends are observed with increasing precipitation in northern and southern Africa) (Van Der Werf et al., 2008; Andela and Van Der Werf, 2014). In addition to natural processes, human activities are primary ignition sources and have shaped fire patterns in the ASA regions (Aragao et al., 2008; Archibald et al., 2009; Andela et al., 2017). Fire-use types driven by local socio-economic conditions and fire management policies may also affect the fire– climate relationships (Andela et al., 2017). Therefore, strong climate controls from wet season to dry season need to be considered along with fuel distributions and human activities for continental fire predictions under climate change.

Accurate predictive modeling of wildfire with skillful representation of how environmental and anthropogenic factors modulate the burned area is still challenging. State-of-the-art process-based fire models (e.g., the Fire Model Intercomparison Project; Rabin et al., 2017) have reasonably simulated the spatial distribution of burned areas. However, they generally do not accurately capture burned-area seasonal variation and inter-annual trends and variability (Andela et al., 2017). Improving predictability and reducing uncertainties of process-based models require more sophisticated representation of fire processes and parameterization, which remain a long-term challenge (Bowman et al., 2009; Hantson et al., 2016; Teckentrup et al., 2019). In response to this challenge, data-driven statistical or machine learning (ML) approaches have been developed and demonstrated to effectively capture wildfire severity and burned-area dynamics (Archibald et al., 2009; Chen et al., 2020, 2011; Zhou et al., 2020). However, the spatially heterogenous, non-linear, and timelagged controls have been oversimplified, e.g., using linear models or only considering climate variables at specific time lags or seasons (Chen et al., 2011, 2016, 2020; Archibald et al., 2009; Gray et al., 2018) or have been black boxed. For example, the commonly used neural network or deeplearning models (Zhu et al., 2022; Joshi and Sukumar, 2021) themselves are complex and built upon hidden neural layers with non-linear activation functions and thus cannot directly identify the relative importance of different drivers for wildfires (Murdoch et al., 2019; Jain et al., 2020). A few ML models (e.g., decision tree and random forest) provide variable importance; however, such importance scores are constant across the entire dataset rather than spatiotemporally varied (S. S. C. Wang et al., 2021; Yuan et al., 2022b). While post-hoc analyses could interpret ML models (Altmann et al., 2010; Lundberg and Lee, 2017), inconsistent and unstable explanations can be derived with different post-hoc methods or settings (Slack et al., 2021; Molnar et al., 2020). Such limitations impede an interpretable and reliable way to understand the critical spatiotemporal processes from wet season to dry season (Reichstein et al., 2019; Jain et al., 2020).

In this work, we developed a wildfire model (Attention-Fire) leveraging on an interpretable long short-term memory (LSTM) framework to predict wildfire burned areas over northern hemispheric Africa (NHAF), southern hemispheric Africa (SHAF), and southern hemispheric South America (SHSA) (Giglio et al., 2013). We also focused on using the AttentionFire model to explore the dependency of simulated burned area on different drivers from wet season to dry season across different grid cells. We assessed model predictability with observed burned area from the Global Fire Emission Database (GFED) and compared it with five other machine-learning-based fire models.

2 Methods

2.1 AttentionFire model

The AttentionFire model is based on an interpretable attention-augmented LSTM (Liang et al., 2018; Qin et al., 2017; Guo et al., 2019; Li et al., 2020; Vaswani et al., 2017) framework. Like the traditional artificial neural network (ANN) models, the LSTM is also built upon neurons and the non-linear activation functions; specifically, the LSTM uses the gating mechanism (i.e., forget, input, and output gates) (Hochreiter and Schmidhuber, 1997; Wang and Yuan, 2019) to filter out useless information while keeping useful information underlying in the time series as hidden states (Fig. 1). Relative to traditional ANN, the LSTM has shown advantages in capturing short- and long-term dependencies in input time series (Hochreiter and Schmidhuber, 1997), such as the time-lagged controls from wet-to-dryseason climate conditions on wildfires. However, LSTM cannot explicitly and dynamically select important drivers from multiple driving time series to make predictions (Qin et al., 2017; Liang et al., 2018; Guo et al., 2019; Li et al., 2020; Vaswani et al., 2017). Further, LSTM works as a black box, lacking interpretability to identify the relative importance of each driver across different time steps (Guo et al., 2019; Li et al., 2020; Liang et al., 2018). Attention mechanisms overcome these challenges by adaptively assigning larger weights to more important drivers and time steps (Liang et al., 2018; Vaswani et al., 2017). Here, we use attention mechanisms to explicitly capture controlling factors of fire predictions with various time lags (Fig. 1). Below are detailed descriptions of the fire model.

Given four categories of time series, $X = (X^1, X^s, X^f, X^c)^T$, where *T* is the length of time series, we use $X^i = (x_1^i, x_2^i, ..., x_T^i)^T \in \mathbb{R}^T$, where $1 \le i \le n$, to denote the *i*th time series, and we use $X_t = (x_t^1, x_t^2, ..., x_t^n)^T \in \mathbb{R}^n$, where $1 \le t \le T$, to represent the vector at time step *t*. x_t^I , x_t^s , x_t^f , and x_t^c represent the variables of ignition (e.g., population density), suppression (e.g., road network density), fuel availability (e.g., living biomass), and climate (e.g., precipitation) at time step *t*. The AttentionFire model aims to learn a nonlinear function *F* to map the *n* time series to the observed burned area Y_{T+1} at time step T + 1:

$$\hat{Y}_{T+1} = F(X^{\rm l}, X^{\rm s}, X^{\rm f}, X^{\rm c})^{T},$$
(1)

where \hat{Y}_{T+1} is the predicted burned area at time step T + 1.

First, the model iteratively transforms the *i*th driving variable at time step *t* to a hidden state vector \mathbf{h}_{t}^{i} , where $1 \le t \le T$ and $1 \le i \le n$, through LSTM gate mechanisms (please refer to Li et al., 2020, for the details of the gates in Fig. 1). Second, as the importance of each time step varies, temporal at-

tention is applied to h_t^i to calculate its corresponding weight or importance w_t^i . Third, the weighted summation h_{sum}^i of h_t^i is obtained to represent the summarized information for the *i*th driving variable:

$$w_t^i = f_{\text{attn}}(\boldsymbol{h}_t^i)$$
$$\boldsymbol{h}_{\text{sum}}^i = \sum_{t=1}^T w_t^i \boldsymbol{h}_t^i, \tag{2}$$

where $h_t^i \in R^m$ is the hidden state vector of the *i*th driving series at time step *t* that stores the summary of the past input sequence (Hochreiter and Schmidhuber, 1997); w_t^i is the calculated weight for the *i*th driver at time step *t* through attention function f_{attn} :

$$w_{t}^{i'} = \tanh(W_{p}h_{t}^{i})$$

$$w_{t}^{i} = \frac{e^{w_{t}^{i'}}}{\sum_{j=1}^{T} e^{w_{t}^{j'}}},$$
(3)

where $W_p \in \mathbb{R}^{1 \times m}$ is a parameter matrix that needs to be learned.

To further capture the relative importance of the *i*th driving variable compared to other driving variables, variable attention is used for the summarized information \boldsymbol{h}_{sum}^i and \boldsymbol{h}_T^i . Note that \boldsymbol{h}_T^i is also a kind of summarized information derived by the LSTM (Hochreiter and Schmidhuber, 1997; Guo et al., 2019). The weight or importance of the *i*th driving variable w_i is calculated as

$$w_i' = \tanh(W_a[\boldsymbol{h}_{sum}^l, \boldsymbol{h}_T^l])$$

$$w_i = \frac{e^{w_i'}}{\sum_{j=1}^n e^{w_{j'}}}.$$
(4)

Finally, using the weighted sum of all driving variables, the model generates the prediction \hat{Y}_{T+1} :

$$o_i = W_o[\boldsymbol{h}_{sum}^i, \boldsymbol{h}_T^i] + b_o$$
$$\hat{Y}_{T+1} = \sum_{i=1}^n o_i w_i,$$
(5)

where $W_a \in \mathbb{R}^{1 \times 2m}$ is a learnable parameter matrix, and the linear function with weight $W_o \in \mathbb{R}^m$ and bias $b_o \in \mathbb{R}$, along with attention-calculated weight w_i , produce the final prediction result. The parameters of attention-based LSTM are learned via a back-propagation algorithm by minimizing the mean-squared error between predictions and observations (Guo et al., 2019; Leung and Haykin, 1991).

The AttentionFire model is implemented with Python under Python 3 environment. The model is open access at https: //doi.org/10.5281/zenodo.7416437 (Li et al., 2022b) under Creative Commons Attribution 4.0 International license. Detailed code and descriptions are included in the repository, including loading datasets, model initialization, training, predicting, saving parameters, and loading the trained model (see more details in "Code availability" section).



Figure 1. An illustrative workflow for AttentionFire_v1.0 model prediction. Four kinds of drivers are considered: ignition related, suppression related, fuel, and climate. The temporal attention is used to identify important time steps for each kind of driver, while the variable attention is used to identify important drivers for final burned-area prediction.

2.2 Baseline models and model settings

Five other widely used machine learning (ML) models are used as baseline models to compare with AttentionFire model: ANN (Joshi and Sukumar, 2021; Zhu et al., 2022), decision tree (DT) (Amatulli et al., 2006; Coffield et al., 2019), random forest (RF) (Yu et al., 2020; Li et al., 2018; Gray et al., 2018), gradient-boosting decision tree (GBDT) (Coffield et al., 2019; Jain et al., 2020), and naive LSTM (Liang et al., 2019; Natekar et al., 2021; Gui et al., 2021; Mei and Li, 2019). The details of the baseline models selected, including strengths, potential limitations, and their applications in wildfire studies, and references are listed in Table 1. The ANN and LSTM have shown good performance on multiple earth science problems (Yuan et al., 2022a; Reichstein et al., 2019), including wildfires (Joshi and Sukumar, 2021; Liang et al., 2019; Zhu et al., 2022); however, the black-box nature of such models makes them lack interpretability. The DT method provides variable importance and is easily interpretable with its single-tree structure, but it is prone to overfitting compared to RF and GBDT. The RF alleviates the overfitting through feature selection and ensemble learning (Breiman, 2001), while the GBDT avoids overfitting by constructing multiple trees with shallow depth (Ke et al., 2017). DT, RF, and GBDT provide variable importance scores for dominant driver inference; however, such importance scores are constant across the entire dataset and thus impede detailed interpretation of the variable importance, like over space and time. The aforementioned ML models have been commonly used in wildfire science (Jain et al., 2020).

The inputs of climate- and fuel-related variables for the first four models (non-sequence models) are variables of the latest three months available for prediction (Yu et al., 2020), while the corresponding inputs of naive LSTM and AttentionFire models are whole-year historical time sequences which cover dynamics from wet to dry seasons to capture short- and long-term dependencies underlying the input sequence (Qin et al., 2017; Vaswani et al., 2017; Guo et al., 2019; Li et al., 2020). The socioeconomic predictors (i.e., population, road density, livestock) consider only the more recent and available statistics typically reported at a year scale. For each model, we iteratively leave a 1-year dataset (one out of all 19 years' datasets for the period 1997-2015, \sim 5 % of all datasets) out (i.e., a holdout dataset, such as the dataset in 2015, that the model has never seen) for testing, 1 year of data ($\sim 5\%$ of all datasets, such as the dataset in 2014) for validation (the model was stopped for training, and its parameters were saved when it showed the highest performance on the validation dataset to avoid overfitting during training; Yuan et al., 2022b; Jabbar and Khan, 2015) and use the remaining dataset ($\sim 90\%$ of all datasets, such as the dataset during 1997-2013) for model training (i.e., tuning model parameters). Such an evaluation scheme quanti-

Model (acronym)	Strengths	Potential limitations	Applications
Random forest (RF) (Breiman, 2001)	Provides variable importance; alleviates overfitting through feature se- lection and ensemble learning	Constant variable importance rather than var- ied; time-consuming when building large trees; may not perform well on time series with lags	Gray et al. (2018); Yu et al. (2020)
Decision tree (DT) (Safavian and Landgrebe, 1991)	Provides variable importance; easily in- terpretable with its single-tree structure	Prone to overfitting; constant variable impor- tance rather than varied; time-consuming when building a large tree; may not perform well on time series with lags	Amatulli et al. (2006); Coffield et al. (2019)
Gradient-boosting decision tree (GBDT) (Ke et al., 2017)	Alleviates overfitting by building multi- ple shallow trees; generally fast because of the shallowness of each tree built	Constant variable importance rather than var- ied; may not perform well on time series with lags	Coffield et al. (2019); Jain et al. (2020)
Artificial neural network (ANN) (Ke et al., 2017)	Shows good performance on com- plex and non-linear problems; allevi- ates overfitting through techniques like dropout and regularization	Lack of interpretability; hard to know the op- timal neural network structures for different problems	Joshi and Sukumar (2021); Zhu et al. (2022)
Long short-term mem- ory (LSTM) (Hochreiter and Schmidhuber, 1997)	Shows good performance on time se- ries predictions; alleviates overfitting through techniques like dropout and regularization	Lack of interpretability; may not be suitable for non-time-series problems; vanishing gradi- ent problem when deployed to long time series (Li et al., 2020; Liang et al., 2018)	Liang et al. (2019); Natekar et al. (2021)

Table 1. Strengths, potential limitations, and applications of selected baseline models in wildfire studies.

fied model performance on deducing the temporal dynamics of fires at the annual scale, which is critical for future projections, while leveraging as much data as possible for model training. Details of the settings for used models in the experiments are listed in Table S1.

2.3 Datasets and experiments

The satellite-based global burned-area dataset (Global Fire Emissions Database; Giglio et al., 2013) is used as prediction target, and datasets of various socio-environmental drivers are used as model inputs. Population density, livestock density, road-network density, and land use are considered as anthropogenic factors on fire ignition and spread. Fuel variables include fuel moisture and live- and dead-vegetation biomass. Seven meteorology variables from National Centers for Environmental Prediction-Department of Energy (NCEP-DOE) Reanalysis are considered, including air temperature, precipitation, surface pressure, wind speed, specific humidity, downward shortwave radiation, and vapor pressure deficit. Details of each dataset and corresponding references are listed in Table 2. The raw datasets were unified to the same spatial resolution (T62 resolution: ~ 210 km at the Equator) at the monthly scale, with a covering period from 1997 to 2015.

In addition to the local socio-environmental drivers, we also explored the impacts of ocean indices on burned-area predictions. Chen et al. (2011) found that wildfires in South America were closely linked to the Oceanic Niño Index (ONI) and the Atlantic Multidecadal Oscillation (AMO) index. The ONI and AMO reflected the sea surface temperature (SST) anomalies in the tropical Pacific and north Atlantic. The SST anomalies directly affected ocean–atmosphere interactions and thus the wet-, wet-to-dry-, and onset-of-dryseason climate in South America (Chen et al., 2011). The two indices were significantly correlated with peak fire month wildfires 3 to 7 months later and could predict fire season wildfires in many regions of South America with lead times of 3 to 5 months (Chen et al., 2011). The controls of SST anomalies in the tropical Pacific on climate and thus on wildfires were also found in northern and southern Africa (Andela and Van Der Werf, 2014). In addition, SST anomalies in the tropical northern and southern Atlantic could also affect wildfires in South America (Chen et al., 2016) and Africa (Yu et al., 2020; Chen et al., 2020). Therefore, we included ocean indices (Table 2) and investigated their impacts on wildfire predictions with the AttentionFire model (see Sect. 3.4).

For future projection (2016-2055) of burned area with the AttentionFire model, land use changes (Hurtt et al., 2020), population growth, projected climate, and fuel from five fully coupled Earth system model (ESM) simulations of CMIP6 (O'Neill et al., 2016) under low- (SSP126) and highemission (SSP585) scenarios were used as the ML model input, respectively. The reason to select 2016-2055 as the projected period was that, during 2016-2055, the 99th percentiles of precipitation, temperature, and vapor pressure deficit were within the range of corresponding historical observations, which means that the trained model has covered the range of most projected drivers in the near future and can alleviate extrapolation uncertainty caused by climate change. We also made a longer projection till the end of 21st century and analyzed its longer-term trend (see Sect. 3.4). All available ESMs with outputs of historical and future (SSP126 and SSP585) fuel availability (i.e., biomass of coarse wood debris, vegetation, and litter) and climate variables (Table 2)

Rothman-Ostrow et al. (2020)

LUH2 (Hurtt et al., 2020)

NOAA climate indices

(Enfield et al., 1999, 2021)

Variable category	Variables (abbreviation, units)	Spatial (temporal) resolution	Dataset and reference
Wildfire	Burned area (BA, ha per month)	0.25° (monthly)	Global Fire Emissions Database 4 (Giglio et al., 2013)
Climate	Precipitation (RAIN, $mm s^{-1}$), temperature (TA, K), surface air pressure (PA, Pa), specific humidity (SH, kg kg ⁻¹), downward shortwave radiation (SW, W m ⁻²), wind speed (WIND, $m s^{-1}$), vapor pressure deficit (VPD, hPa) (VPD calculated according to Bolton; 1980)	$\sim 1.9^{\circ}$ (monthly)	NCEP–DOE Reanalysis 2 (Kanamitsu et al., 2002)
Fuel conditions	Fuel moisture (FUELM, %), coarse wood debris (CWDC, gC m ^{-2} s ^{-1}), vegetation biomass (VegC, gC m ^{-2} s ^{-1}), litter biomass (LitterC, gC m ^{-2} s ^{-1})	$\sim 1.9^{\circ}$ (monthly)	ELM prognostic simulations (Zhu et al., 2019)
Human activities	Population density (Popu, persons per grid)	~ 1 km (yearly)	Dobson et al. (2000)
	Road density (Road, kmkm^{-2})	0.5° (yearly)	Meijer et al. (2018)

Table 2. Input and output variables and datasets of the AttentionFire model

Livestock density (LS, number of livestock per

Bare soil (Bare, %), Forest (Forest, %), and

Ocean Niño Index (ONI), Atlantic Multidecadal

Oscillation (AMO) index, tropical Northern Atlantic (TNA) index, and tropical Southern At-

were selected, including ACCESS-ESM1-5 (Ziehn et al., 2020), CESM2 (Danabasoglu et al., 2020), NorESM2-LM (Seland et al., 2020), NorESM2-MM (Seland et al., 2020), and TaiESM1 (Y. C. Wang et al., 2021). For each ESM, the variable bias was corrected with the mostly used linear scaling method (Maraun, 2016; Dangol et al., 2022; Shrestha et al., 2017), which adjusted the bias in model simulations based on the ratio of modeled- and observed-variable mean value. Then the bias-corrected variables of each ESM were used to drive the AttentionFire model for future burnedarea projection. Finally, given the uncertainty of each ESM, the multi-model ensemble (MME) mean of projected burned area was calculated (Li et al., 2022a) and analyzed. Details of the bias correction method can be found in Maraun (2016). For future projections, temporally constant road and livestock density were used due to the lack of future data in the two scenarios (i.e., SSP585 and SSP126), and the Attention-Fire model was not coupled in the ESMs. Such limitations and uncertainties were discussed in Sect. 3.5.

grid)

Grass (Grass, %)

lantic (TSA) index

Land cover

Oceanic indices

3 Results and discussions

 0.5° (yearly)

 0.25° (yearly)

monthly

3.1 Model predictability on burned-area spatial-temporal dynamics

The AttentionFire model accurately captured the spatial distribution and temporal variations (Figs. 2 and S1) of wildfireburned areas over NHAF, SHAF, and SHSA regions. The AttentionFire model had the lowest mean absolute errors (MAEs) between model-predicted and observed (GFED) gridded monthly burned areas among the six ML approaches. The gridded MAEs of burned area for AttentionFire were 110, 142, and 39 Kha yr⁻¹ in NHAF, SHAF, and SHSA regions, which were respectively 6 %-66 %, 13 %-65 %, and 11 %-42% lower than the other five ML approaches in the three regions. These results highlight the capability of the AttentionFire model to capture critical driving factors of burned area across time and space.

The fact that the AttentionFire model outperformed the other five models (Fig. 2g-i) indicates the benefit of skillfully integrating time-lagged and spatially heterogenous controls from critical drivers on wildfires. Compared to non-sequence models (i.e., RF, MLP, DT, and GBDT), the AttentionFire model adaptively captured historical dependencies of wild-



Figure 2. The AttentionFire model accurately captured burned-area spatial dynamics. Spatial distribution of observed and AttentionFirepredicted fire season mean burned area (BA) with one-month lead time in northern hemispheric African (NHAF) (**a–b**), southern hemispheric African (SHAF) (**c–d**), and southern hemispheric South American (SHSA) (**e–f**) regions. (**g–i**) Performance (in terms of mean absolute error between predicted and observed burned area) of AttentionFire and other five baseline models, including long short-term memory (LSTM), random forest (RF), artificial neural network (ANN), decision tree (DT), and gradient-boosting decision tree (GBDT).

fires on climate conditions from wet to dry seasons (Van Der Werf et al., 2008; Archibald et al., 2009; Andela and Van Der Werf, 2014; Chen et al., 2011). A more detailed analysis is provided in next section. Compared to the naive LSTM models, the variable and temporal attention mechanisms integrated in AttentionFire has proven to be beneficial to model performance.

The spatial heterogeneity and temporal variation of wildfire responses to complex environmental and human factors have made wildfire predictions challenging, especially at large spatial scales (Chen et al., 2016; Littell et al., 2016; Andela and Van Der Werf, 2014; Chen et al., 2011; Zhou et al., 2020). The capability of the AttentionFire model to reasonably predict spatial and temporal distributions of burned area ahead of fire season allows more time to explore and implement management options, such as allocation of firefighting resources, fuel clearing, or targeted burning restrictions (Chen et al., 2011).

3.2 Dominant drivers of tropical burned-area dynamics

The AttentionFire model dynamically weights variable importance and highlights critical temporal windows (Qin et al., 2017; Vaswani et al., 2017; Liang et al., 2018; Guo et

al., 2019; Li et al., 2020) that maximize model predictability. Therefore, the variable weights could inform dominant physical processes, while the temporal weights reflect the temporal dependency structure, making it interpretable for spatial-temporal analysis. For the AttentionFire model predictions, the variable weights showed that climate wetness exerted strong and spatially heterogenous controls on burned areas. Specifically, precipitation (for SHAF and SHSA regions) and vapor pressure deficit (VPD; for NHAF region) played the most important roles (Fig. 3) in burned-area prediction during fire seasons (defined as the four months with the largest burned areas, Fig. S2), and the control strengths from those climate wetness variables on fires were significantly (one-tailed t test, p value < 0.05) stronger in regions with larger burned areas (grid cells with top 10 % burned areas) than those with smaller burned areas (grid cells with last 90 % burned areas; Fig. 4a-f).

In AttentionFire model predictions, the precipitation and VPD explained $\sim 66\%$ to $\sim 80\%$ (Fig. S3) of the annual mean fire season wildfire-burned areas. Variations of VPD and precipitation not only affect fire season ignition likelihood and fire spread (Sedano and Randerson, 2014; Holden et al., 2018) through fuel moisture but also regulate vegetation growth, fuel structure (e.g., fuel composition and spatial



Figure 3. Ranked top-five important variables for fire season burned area in northern hemispheric Africa (NHAF) (**a**), southern hemispheric Africa (SHAF) (**b**), and southern hemispheric South America (SHSA) (**c**). For each grid cell within each study region, there is a mean variable weight, representing the importance of the variable for fire prediction in the grid cell. For each region, the variable weights are summed, weighted by their corresponding mean burned areas, and normalized.



Figure 4. Spatial-temporal importance of climate wetness variables for burned-area dynamics. (**a**–**c**) Spatial importance of climate wetness variables for fire season burned areas. (**d**–**f**) Statistical comparison of the climate wetness variable importance over regions with large and small burned areas. (**g**–**i**) Fire season burned-area dependency on the history of the climate wetness driver over northern hemispheric African (NHAF), southern hemispheric African (SHAF), and southern hemispheric South American (SHSA) regions.

connectivity; Gale et al., 2021), and fuel availability (Mueller et al., 2020; Littell et al., 2009, 2016; Van Der Werf et al., 2008). The importance of these climate wetness variables confirms the dominant roles of local water balances and air dryness for wildfire prediction from sub-seasonal to seasonal scales (Littell et al., 2016; Archibald et al., 2009; Chen et al., 2011), especially in regions with large burned areas.

Furthermore, we found that the emergent functional relationships between climate wetness and wildfire-burned area were parabolic (Fig. S3): i.e., enhancement of historical precipitation or decline of historical VPD (indicating wetter conditions) first increased burned area in more xeric conditions, then suppressed burned area under more mesic conditions, consistent with previous findings in subtropical regions (Andela and Van Der Werf, 2014; Van Der Werf et al., 2008). The transition points of these emergent functional relationships (thresholds at which the relationships reverse) were region specific, and these relationships may be useful for developing, tuning, and benchmarking wildfire models (Zhu et al., 2022).

For the time lags between those dominant climate wetness variables and fire season burned areas, our results demonstrated that burned area over NHAF was more modulated by relatively short-term wetness (VPD during wet-to-dry and onset of dry season, from September to December), while SHAF and SHSA burned areas depended more on long-term wetness (precipitation during wet and wet-to-dry season, December to March in SHAF and November to April in SHSA) (Fig. 4g–i). The short-term variations of climate wetness can directly affect near-surface temperature and moisture availability, which affect fuel flammability (Littell et al., 2016; Holden et al., 2018), while the long-term wetness (e.g., during rainy season) can affect fuel availability, composition, and spatial connectivity, which can result in even stronger long time-lagged controls on dry-season burned areas (Abatzoglou and Kolden, 2013; Littell et al., 2016; Chen et al., 2011; Van Der Werf et al., 2008; Archibald et al., 2009; Andela and Van Der Werf, 2014).

Previous work has shown that when and where fires occurred during the dry season can be affected by precipitationinduced fuel availability patterns during the wet season and during wet-to-dry transition seasons in savannah ecosystems (Van Der Werf et al., 2008; Archibald et al., 2009; Andela and Van Der Werf, 2014). Also, precipitation variations during the wet season and wet-to-dry transition seasons in the tropical forest ecosystem can affect soil recharge during the wet season and further affect plant transpiration, local surface humidity, and precipitation during the following dry season (Chen et al., 2011; Ramos da Silva et al., 2008; Malhi et al., 2008). The exact responses of fires to short- and longterm climate variations depend on both local wetness and fuel conditions (e.g., fires in wetter ecosystems with enough fuel availability can be mainly limited by the length of dry season, while fires in drier ecosystems can be limited by fuel availability during wet season; Van Der Werf et al., 2008; Andela and Van Der Werf, 2014). Therefore, an effective way of integrating the climate wetness history (i.e., AttentionFire model) can lead to more accurate predictions of burned-area spatial-temporal dynamics.

3.3 Possible usage of oceanic index for long-leading-time predictions

In ASA regions, large-scale variations of oceanic dynamics can directly influence local climate (e.g., precipitation variations during wet seasons; Chen et al., 2011; Andela and Van Der Werf, 2014) through time-lagged controls of teleconnections and indirectly influence fires during the following dry seasons (Chen et al., 2016, 2011; Andela et al., 2017). Therefore, we hypothesized that ocean dynamics might benefit AttentionFire model predictions, especially for long-leadingtime fire predictions, through providing additional information that has not been reflected in local climate and land surface conditions (Chen et al., 2016, 2011; Andela et al., 2017; Chen et al., 2020).

We compared model performance for short-term (1–4 months ahead) and long-term (5–8 months ahead) fire predictions with and without considering the four oceanic indices (OIs). Relative to the MAE of short-term predictions, the mean MAE of long-term predictions without and with teleconnections increased by $\sim 34\%$ and $\sim 14\%$ in NHAF, $\sim 34\%$ and $\sim 15\%$ in SHAF, and $\sim 17\%$ and $\sim 7\%$ in SHSA, respectively, indicating the decline of system predictability with longer leading time (Fig. 5). However, for long-term predictions, including OIs could decrease the mean MAE by $\sim 20\%$, $\sim 19\%$, and $\sim 11\%$ in NHAF, SHAF, and SHSA regions, respectively, compared with the case without oceanic indices. While the mean variable importance of OIs was consistently lower than that of local cli-

mate (Fig. S4) across the three regions, the OIs did provide additional information for long-term predictions with lower biases (Fig. 5). The results demonstrated the potential usage of teleconnections for long-leading-time burned-area predictions (Chen et al., 2020, 2016, 2011).

3.4 Future trends of burned area over Africa and South America

Due to climate change and human activities (Andela et al., 2017), strong but opposing trends of burned areas have been observed in northern (decreasing) and southern (increasing) hemispheric Africa (Andela and Van Der Werf, 2014) and within different regions of southern hemispheric America (Andela et al., 2017) during the recent two decades, resulting in an overall declining burned-area trend in Africa and South America. However, whether this decline will persist is under debate. On the one hand, the projected increases in population, expansion of agriculture, mechanized (fire-free) management, and fire suppression policies will likely continue to decrease burned areas (Andela and Van Der Werf, 2014), e.g., human activities were regarded as one of the main drivers for fire decline in NHAF region. On the other hand, future climate change (Dai, 2013; Taufik et al., 2017) could outweigh human impacts and result in unprecedented fire-prone environments in the tropics (Pechony and Shindell, 2010; Malhi et al., 2008), e.g., fires showed strong dependency on climate wetness in NHAF, SHAF (Andela and Van Der Werf, 2014; Archibald et al., 2009) and SHSA (Chen et al., 2011) regions.

Considering land use changes, population growth, and projected climate and fuel conditions under the SSP585 highemission scenario, our model predicted that burned areas in the NHAF region will continue to decline; the currently increasing trend will be dampened in the SHAF region, and the currently decreasing trend will be reversed in the SHSA region (Fig. 6). The increasing trend in SHSA, decreasing trend in NHAF, and dampened trend in SHAF under SSP585 were robust when projecting burned area till the end of the 21st century (Fig. S5). Over NHAF and SHSA, burned-area trends at the grid cell level were mostly robust (Fig. 6a, c; p < 0.05) and of the same sign, thus resulting in a robust trend at the regional scale. Under the low-emission scenario (i.e., SSP126), the decreasing trend in NHAF disappeared (Fig. S5a), and the increasing trend in SHSA was reduced by $\sim 69\%$ (Fig. S5c), implying the big influences of climate changes and socioeconomic development pathways on future burn-area changes in the two regions.

To investigate what drives future burned-area changes under SSP585, we iteratively surrogated each driver with its climatology while keeping the other factors the same. Burned-area changing trends in NHAF and SHSA were mostly affected by VPD changes because removing VPD inter-annual changes resulted in non-significant burned-area trends throughout the whole of the NHAF and SHSA regions (Fig. 6a, c). VPD was projected to continuously in-



Figure 5. Performance of AttentionFire burned-area predictions with 1–4 months leading time (short term) and with 5–8 months leading (long term). MAE is mean absolute error. Four ocean indices which have been widely used for fire prediction over South American and African regions were considered for long-term forecasting, including Oceanic Niño Index, Atlantic Multidecadal Oscillation index, tropical Northern Atlantic index, and tropical Southern Atlantic index.



Figure 6. Future burned-area trends under the SSP585 high-emission scenario. (**a**–**c**) Spatial distribution of fire season burned-area trends using drivers with interannual variations; dots in (**a**)–(**c**) indicate grid cells with statistically significant changes in the trend. (**d**–**f**) Regionally aggregated burned-area changes with historical mean subtracted. Blue and red lines respectively represent burned-area anomaly in history and the future; the black line represents the future burned-area trend while removing the interannual variations of the dominant variable. Solid lines represent significant BA trends (*p* value <0.05), while dashed lines represented non-significant BA trends.

crease due to warming but had different implications over NHAF and SHSA. Over the relatively fuel-abundant SHSA region, increased VPD will likely increase burned area (Pearson r = 0.64, p value <0.05, Fig. S6) through increasing fuel dryness and combustibility (Kelley et al., 2019; Chen et al., 2011; Malhi et al., 2008; Van Der Werf et al., 2008). In contrast, over the semi-arid savannah-dominated NHAF region (less fuel, compared with SHSA), higher VPD could decrease burned area (Pearson r = -0.71, p value <0.05, Fig. S6) through limiting plant growth and fuel availability (Van Der Werf et al., 2008; Andela and Van Der Werf, 2014; Andela et al., 2017). For the SHAF, population growth and climate changes showed stronger influences on burned-area changes (Andela and Van Der Werf, 2014), while the heterogeneity of wildfire responses finally led to a non-significant trend at the regional scale (Fig. 6). Our findings highlight the importance of climate changes for understanding future burned-area dynamics and motivate for better representation of climate wetness effects on wildfire dynamics in processbased and machine-learning-based wildfire prediction models.

3.5 Directions for future research

The time-lagged controls of climate on ASA wildfires are critical for sub-seasonal to seasonal wildfire prediction (Chen et al., 2020; Andela and Van Der Werf, 2014; Chen et al., 2011) but remain less well represented due to the complex interactions among fire, climate, fuel, and human activities. Here, we deployed the interpretable AttentionFire model to understand and predict fire dynamics in the ASA region. We revealed the dominant, spatially heterogenous, and timelagged controls of climate wetness on ASA wildfires. Such climate wetness importance on ASA wildfires was consistent with previous findings (Andela and Van Der Werf, 2014; Chen et al., 2011) and was also confirmed by the other three tree-based ML models (i.e., DT, RF, and GBDT) with variable importance (e.g., precipitation and VPD were regarded as the top-five most important variables in Fig. S7). However, differences existed across the model-identified most important drivers (Fig. 3 versus Fig. S7). The variable importance of the AttentionFire model was spatiotemporally varied (Fig. 4), while tree-based-model-provided variable importance was constant over the entire dataset. We showed that the climate wetness was more (less) important in areas with large (small) burned areas, and its importance also varied over time (Fig. 4), but the other MLs did not explicitly distinguish such differences. Albeit, the higher accuracy and generally acceptable computation speed of AttentionFire (Table S2), its memory consumption, and its model training time could be up to 141 % and 22 times higher than the other ML models. The implementation of LSTM in the Attention-Fire model is a serial model instead of a parallel model; therefore, future work could improve the model efficiency by deploying some easy-for-parallel-computing time series prediction frameworks, e.g., temporal convolutional attention (Lin et al., 2021) and self attention (Mohammadi Farsani and Pazouki, 2020; Vaswani et al., 2017).

This study focused on wildfire prediction in the ASA region, and we showed the performance improvement of the AttentionFire model by representing the time-lagged controls of climate on wildfires. Whether the AttentionFire model can also outperform other ML models in other regions may depend on the dependency strength and time lags between wildfires and climate variables. For example, in North American boreal forests, lightning was identified as the major driver of the interannual variability in burned areas by influencing the number of ignitions in the dry season (Veraverbeke et al., 2017). In such regions, the AttentionFire model might not outperform other ML models due to the lesser dominance of time-lagged controls. In regions like the western US and India, where wildfires showed time-lagged dependencies with local climate (Littell et al., 2009; Kale et al., 2022) and where some extreme wildfires were caused by persistent drought from wet to dry seasons with multi-month lags (Taufik et al., 2017; Littell et al., 2016), the Attention-Fire model could potentially be useful.

With the fully coupled ESMs of CMIP6, we analyzed future burned-area changes under high- (SSP585) and lowemission (SSP126) scenarios in the ASA region. While the MME mean was considered, substantial uncertainty has been found across different ESMs in history (Yuan et al., 2022a, 2021; Wu et al., 2020) and the future (Li et al., 2022a; Lauer et al., 2020). Therefore, further work is needed to narrow the projection uncertainty of ESMs, e.g., with constraints of causality (Nowack et al., 2020; Li et al., 2022a) and observations (Tokarska et al., 2020; Lauer et al., 2020). Meanwhile, for future projections, although land use and land cover changes, population growth, and climate and fuel changes were considered, constant livestock and road density were adopted due to lack of data. The impacts of livestock and road density therefore need further exploration with available data under different future scenarios. In addition, the Attention-Fire model is currently not coupled with the ESM; therefore, the feedback among fires, climate, and biomass was ignored. To analyze such feedback, the AttentionFire model needs to surrogate the original fire module and be coupled with the ESM (Zhu et al., 2022).

4 Conclusions

This study developed an interpretable machine learning model (AttentionFire_v1.0) for burned-area predictions over African and South American regions. Compared with observations and another five widely used machine learning baseline models, we demonstrated the effectiveness of the AttentionFire model to capture the magnitude, spatial distribution, and temporal variation of burned areas. Attention mechanisms enabled the interpretation of complex but critical spatial-temporal patterns (Li et al., 2020; Guo et al., 2019; Liang et al., 2018; Vaswani et al., 2017; Qin et al., 2017), thus uncovering the black-box relationships in machine learning models for burned-area predictions. We demonstrated the spatiotemporally heterogenous and strong time-lagged controls from local climate wetness on burned areas. Furthermore, under the SSP585 high-emission scenario, our results suggested that the increasing trend in burned area over southern Africa will be dampened, and the declining trend in burned area over fuel-abundant southern America will reverse. This study highlights the importance of the skillful representation of spatiotemporally heterogenous and strong time-lagged climate wetness effects on understanding wildfire dynamics and developing advanced early fire warning models.

Code availability. The source code of AttentionFire_v1.0 and all baseline machine learning models is archived at Zenodo repository: https://doi.org/10.5281/zenodo.7416437 (Li et al., 2022b) under Creative Commons Attribution 4.0 International license, with four zip files: data, data_preparation, model, and example. The "data" file contains the links to all raw datasets used to drive the model (e.g., burned areas, climate forcing). The "data_preparation" file contains the code to preprocess the raw datasets and make them ready for training and testing of the AttentionFire model. The "model" file contains the Python code of the AttentionFire model. The "example" file gives a detailed example of how to use the AttentionFire model for burned-area predictions.

There is also a tutorial file "Data_Model_Tutorial" that contains descriptions on (1) how to load the raw datasets, (2) how to prepare the input and output datasets for the ML model, (3) how to initialize the ML model and run the model, (4) how to train the ML model and use the trained ML model for predictions, and (5) how to save and load the model parameters and save the predicted results.

Data availability. We used NCEP-DOE Reanalysis Climate forcings: https://psl.noaa.gov/data/gridded/data.ncep.reanalysis2.html (Kanamitsu et al., 2002) and NOAA oceanic index data: https://psl.noaa.gov/data/climateindices/list/ (NOAA, 2022). Population density data are from https://landscan.ornl.gov/ (Dobson et al., 2000). Road density data are from https://www.globio.info/download-grip-dataset (Meijer et al., 2018). Livestock density data could be found at https://www.fao.org/dad-is/en/ (Robinson et al., 2014). We used LUH2 land cover change data: https://luh.umd.edu/data.shtml (Hurtt et al., 2020). The grid-cell-level burned area data are from the Global Fire Emissions Database https: //daac.ornl.gov/VEGETATION/guides/fire_emissions_v4.html (https://doi.org/10.3334/ORNLDAAC/1293, Randerson et al., 2018).

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Author contributions. QZ and FL designed the study. QZ, FL, and MC designed the model experiments. FL wrote the code and ran the experiments. LZ, WJR, JTR, LX, KY, HW, ZG, and JG all contributed to the interpretation of the results and writing of the paper.

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